

South32 - Illawarra Metallurgical Coal

DENDROBIUM MINE

End of Panel Surface Water and Shallow Groundwater
Assessment: Longwall 18 (Area 3B)



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Cover photo: Cordeaux River tributary location CR29_S1, looking upstream on 27/10/2016

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TABLE OF CONTENTS

| | | |
|-----|---|----|
| 1. | Introduction..... | 11 |
| 1.1 | Reporting Objectives | 11 |
| 1.2 | Longwall 18..... | 11 |
| 1.3 | Feedback from agencies on previous assessment | 12 |
| 2. | Surface water and groundwater management | 13 |
| 2.1 | Surface Water Monitoring..... | 13 |
| 2.2 | Improvements to monitoring network in reporting period | 14 |
| 2.3 | Surface water flow data update | 15 |
| 2.4 | Shallow Groundwater Monitoring | 16 |
| 2.5 | Soil moisture monitoring | 18 |
| 2.6 | Weather conditions during the assessment period | 22 |
| 3. | Longwall subsidence effects..... | 24 |
| 3.1 | Measured subsidence | 24 |
| 3.2 | Observed surface impacts..... | 25 |
| 3.3 | Specialist advice in relation to observed impacts..... | 26 |
| 3.4 | Valley closure at Waterfall 54 | 26 |
| 4. | Assessment of surface water quality effects | 28 |
| 4.1 | Overview of surface water quality | 28 |
| 4.2 | Quantitative assessment of water quality trends..... | 30 |
| 4.3 | Catchments within Longwall 18 area of influence | 32 |
| 4.4 | TARP triggers at tributary LA4..... | 32 |
| 4.5 | Iron-staining in Wongawilli Creek | 33 |
| 4.6 | Sandy Creek..... | 35 |
| 5. | Assessment of surface water flow effects | 36 |
| 5.1 | Surface Water Flow TARPs..... | 36 |
| 5.2 | Performance Measures | 37 |
| 5.3 | Assessment for Longwall 18..... | 37 |
| 5.4 | Assessment against surface water flow TARPs..... | 38 |
| 5.5 | Assessment D: flow reduction Wongawilli Creek..... | 68 |
| 5.6 | Assessment against surface water flow Performance Measures..... | 69 |
| 5.7 | Watercourse pool water levels | 70 |
| 5.8 | Pool outflow status | 72 |
| 6. | Assessment of shallow groundwater (swamps) | 76 |
| 6.1 | Shallow groundwater levels..... | 76 |
| 6.2 | Soil moisture..... | 87 |
| 7. | Conclusions..... | 89 |
| 7.1 | Effects on surface water quality..... | 89 |
| 7.2 | Effects on surface water flow..... | 89 |
| 7.3 | Effects on swamps | 92 |
| 8. | References..... | 93 |
| | Appendix A1: Water quality hydrographs..... | 94 |
| | Appendix A2: Water quality trend analysis..... | 95 |
| | Appendix B: Rainfall data | 96 |

| | |
|---|-----|
| Appendix C: Flow gauge data..... | 101 |
| Appendix D: Shallow groundwater hydrographs | 115 |
| Appendix E: Soil moisture hydrographs..... | 116 |
| Appendix F: Stream pool level hydrographs | 117 |
| Appendix G: Watercourse flow observations | 118 |
| Appendix H: Rainfall-runoff modelling..... | 119 |

LIST OF TABLES

| | |
|--|----|
| Table 1. Summary of Surface Water flow TARPs – Longwall 18 | 8 |
| Table 2. Summary of surface water flow Performance Measures– Longwall 18 | 9 |
| Table 3. Comments on previous EOP assessment..... | 12 |
| Table 4. Surface Water Flow Monitoring Sites in Area 3A and 3B | 13 |
| Table 5. Stream gauges that have been re-rated..... | 15 |
| Table 6. Summary of Swamp Monitoring..... | 17 |
| Table 7. Surface water features within area of Longwall 18 influence..... | 25 |
| Table 8. Reported subsidence impacts to stream beds during Longwall 18..... | 26 |
| Table 9. Summary of Water Quality TARPs for the monitoring period | 28 |
| Table 10. Summary of surface water quality observations and trends | 29 |
| Table 11. Summary of flow-corrected water quality trends | 31 |
| Table 12. Recommendations of the IEPMC (2018) (revised as IEPMC, 2019a) | 36 |
| Table 13. Area 3B Surface flow Performance Measures | 37 |
| Table 14. Flow assessments A, B and C for the sub-catchment to DC13S1 | 40 |
| Table 15. Flow assessments A, B and C for the sub-catchment to DCS2..... | 42 |
| Table 16. Flow assessments A, B and C for the sub-catchment to DCU | 44 |
| Table 17. Flow assessments A, B and C for the sub-catchment to WC12S1 | 46 |
| Table 18. Flow assessments A, B and C for the sub-catchment to WC15S1 | 48 |
| Table 19. Flow assessments A, B and C for the sub-catchment to WC21S1 | 50 |
| Table 20. Flow assessments A, B and C for the sub-catchment to WWL | 52 |
| Table 21. Flow assessments A, B and C for the sub-catchment to LA4S1..... | 54 |
| Table 22. Flow assessments A, B and C for the sub-catchment to LA3S1..... | 56 |
| Table 23. Flow assessments A, B and C for the sub-catchment to LA2S1..... | 58 |
| Table 24. Flow assessments A, B and C for the sub-catchment to ND1S1 | 60 |
| Table 25. Flow assessments A, B and C for the sub-catchment to SC10C..... | 62 |
| Table 26. Flow assessments A, B and C for the sub-catchment SC10 | 64 |
| Table 27. Flow assessments A, B and C for the sub-catchment to SCL2/2122205..... | 66 |
| Table 28. Assessment D for Wongawilli Creek: Longwall 18 | 68 |
| Table 29. Performance criteria related to shallow groundwater levels at swamp monitoring sites | 76 |

| | |
|--|----|
| Table 30. Summary of cumulative shallow groundwater effects and TARP status at <i>Impact Sites</i> | 84 |
| Table 31. Summary of shallow groundwater level trends at <i>Reference Sites</i> | 85 |
| Table 32. TARP trigger conditions related to soil moisture at swamp monitoring sites..... | 87 |
| Table 33. Cumulative assessment of soil moisture hydrographs in Areas 3A and 3B..... | 88 |
| Table 34. Area 3B watercourse flow assessment summary..... | 91 |

LIST OF FIGURES

| | |
|--|----|
| Figure 1. Monitoring sites – Field monitoring and sampling sites..... | 19 |
| Figure 2. Monitoring Sites – Hydrographic gauging stations..... | 20 |
| Figure 3. Monitoring sites – Swamp shallow groundwater piezometers..... | 21 |
| Figure 4. Rainfall and potential evapotranspiration (EVT) at Area 3 for the reporting period..... | 22 |
| Figure 5. Calculated soil moisture from BOM’s AWRA Landscape Model..... | 23 |
| Figure 6. Long-term rainfall trends at Dendrobium (1960-2021)..... | 23 |
| Figure 7. Predicted Subsidence above Area 3B (from MSEC, 2020)..... | 24 |
| Figure 8. Observed surface impacts..... | 27 |
| Figure 9. Flow-corrected time-series plots of EC and pH at DCC_FR6..... | 31 |
| Figure 10. Flow-corrected time-series plots of Al and Zn at DCC_FR6..... | 32 |
| Figure 11. Dissolved iron concentration in Wongawilli Creek..... | 34 |
| Figure 12. Groundwater hydrograph at monitoring bore S2220 (total piezometric head)..... | 34 |
| Figure 13. DC13S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 39 |
| Figure 14. DCS2 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 41 |
| Figure 15. Comparison of DCU against Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 43 |
| Figure 16. WC12S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 45 |
| Figure 17. Comparison of WC15S1 against Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 47 |
| Figure 18. WC21S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 49 |
| Figure 19. WWL vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 51 |
| Figure 20. Comparison of LA4S1 against Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 53 |
| Figure 21. LA3S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 55 |
| Figure 22. LA2S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 57 |
| Figure 23. ND1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 59 |
| Figure 24. SC10C vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 61 |
| Figure 25. SC10 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 63 |
| Figure 26. SCL2/2122205 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]..... | 65 |

Figure 27. Time series plot of water level observations in Pool 50.....71

Figure 28. Groundwater hydrographs for lower HBSS adjacent to Wongawilli Creek 71

Figure 29. Flow status of pools on the LA2 watercourse.....73

Figure 30. Flow status of pools on the ND1 watercourse.....73

Figure 31. Relationship between water level (stage) at WWU and WF5474

Figure 32. Comparison of WF54 stage with range in “expected” stage.....75

Figure 33. Overview of swamp saturation levels by month, Area 3B (south).....77

Figure 34. Overview of swamp saturation levels by month, Area 3B (north)78

Figure 35. Shallow groundwater hydrographs for Swamp 14.....80

Figure 36. Shallow groundwater hydrograph for Swamp 23, piezometer 0280

Figure 37. Shallow groundwater hydrograph for Swamp 35a, piezometer 0181

Figure 38. Shallow groundwater hydrograph for Swamp 35b, piezometer 0182

Figure 39. Shallow groundwater hydrograph for swamp piezometers 150_01 and 151_01 ..83

Summary

This report summarises the observed, measured and estimated effects on hydrological features resulting from the extraction of Dendrobium Longwall 18. Longwall 18 is the tenth and final panel to be extracted from Dendrobium Area 3B. Extraction of Longwall 18 commenced on 2/12/2021 and was completed on 17/5/2022. Rainfall during Longwall 18 extraction was well above average, totalling 2281 mm in the calendar year to the end of the longwall (18/5/2021 – 17/5/2022). Extremely heavy rainfall was experienced in March 2022 when 1010 mm was recorded in a single month. This follows similarly high rainfall in 2020 (1436 mm) and 2021 (1448 mm). As a result, there has been a full recovery in stream flow, shallow groundwater levels and soil moisture across all catchments since the severe drought of 2017-2019.

The Illawarra Metallurgical Coal Environmental Field Team (IMCEFT) conducts monitoring and inspections on landscape features including watercourses and swamps within Dendrobium Area 3B. This monitoring is conducted in accordance with the Dendrobium Area 3B Subsidence Management Plan (SMP) and monitoring and contingency plans contained therein. Trigger Action Response Plans (TARPs) contained in the SMP form the basis of the impact assessments in this report. A total of 25 new (or updated) surface impacts attributed to the extraction of Longwall 18 were recorded. No new fracturing or diversion of surface water flows was identified along the streams. However, iron staining was identified in one location along LA3 to the west of LW16.

Surface water quality

At many stream monitoring sites including reference sites, water electrical conductivity (EC) has decreased over the last three years due to higher-than-average rainfall and associated increase in runoff. The decreasing trend follows slightly more saline conditions at most locations during the 2017-2019 drought which resulted in low flows and evaporative concentration of salts. Similarly, Dissolved Oxygen (DO) has trended higher over the last two-year period due to higher flows.

Anomalous water quality effects are noted in streams that have been directly mined under by previous longwalls (e.g. WC21, SC10C, LA4, Donalds Castle Creek). Those effects include transient or persistent increases in EC, increases (or decreases) in pH and increases in dissolved metal concentrations such as Fe, Mn, Al and Zn. Water quality TARPs were triggered at Lake Avon tributary site LA4_S1 for EC, pH and DO. Analysis of flow-corrected trends in water quality indicate concentrations of Fe, Mn, Zn and SO₄ are slightly above baseline in Sandy Creek (Rockbar 5). IMC has initiated a longitudinal study to assess the cause of the increase. No adverse changes in water quality are noted in Lake Avon and Lake Cordeaux.

Iron staining in creek beds is commonly associated with watercourses that have been directly mined beneath or are within the mining area of influence. Over the last two years, new or recurrent iron staining has been noted on Wongawilli Creek, WC21, LA5 and SC10C. The observations of iron staining are likely related to recovery of groundwater levels which has resulted in reactivation of iron-rich springs near creek channels. Iron staining has also been observed in natural catchments on the Woronora Plateau that are located outside mining influence.

Stream flow

Assessment of stream flow gauging records has identified mining-related effects on the flow regime in tributaries to Donalds Castle Creek (DCS2 and DC13S1 – Level 3), tributaries to Wongawilli Creek (WC21S1 and WC15S1 – Level 1 to 3) as well as in Lake Avon tributary sites - LA4S1 and LA3S1 (Level 3) and LA2S1 (Level 0, 2 and 3). As in previous assessments, no changes in flow

characteristics were detected at WC12S1 which is close to Longwalls 16 and 17, noting that rainfall runoff modelling shows an apparent mild change (Level 1).

The upper reaches of sub-catchment to ND1 (a tributary of Native Dog Creek) was mined under by Longwall 18. Comparison against Reference Sites did not indicate an impact, however comparison against rainfall-runoff modelling suggested that a significant decline in flow occurring in the latter half of Longwall 18. This finding will be reviewed in future.

Table 1. Summary of Surface Water flow TARPs – Longwall 18

| Site | Watercourse | Catchment Mined | Position of sub-catchment relative to mining | A) Low flow Q%ile outside Reference Site Q%ile | B) Change in cease-to-flow frequency (beyond natural) | C) Change in median flow, Q50 (beyond natural) | Comment |
|--------|----------------------|-----------------|--|--|---|--|---|
| DC13S1 | DC13 | Yes | Above LWs | ●●●● Level 3 | ●●●● Level 2 | ●●●● Level 3 | Similar to LW14-17. |
| DCS2 | Donalds Castle Creek | Yes | Above LWs | ●●●● Level 3 | ●●●● Level 3 | ●●●● Level 3 | Similar to LW14-17. |
| DCU | Donalds Castle Creek | Yes | Downstream | ●●●● Not triggered | ●●●● Level 1 | ●●●● Not triggered | Similar to LW14-17. Rainfall-runoff modelling supports this finding. |
| WC21S1 | WC21 | Yes | Above LWs | ●●●● Level 3 | ●●●● Level 1 | ●●●● Level 3 | Similar to LW14-17, slight improvement in Assessment B.. |
| WC15S1 | WC15 | Yes | Above LWs | ●●●● Level 3 | ●●●● Level 2 | ●●●● Level 3 | Similar to LW15-17. * Flow monitoring method means that Method B assessment assess low flows, not true 'cease-to-flow'. |
| WC12S1 | WC12 | Yes | Above LWs | ●●●● Not triggered | ●●●● Not triggered | ●●●● Not triggered | Second panel under catchment. No discernible effect. Rainfall-runoff modelling suggests Level 1 impact. |
| WWL | Wongawilli Creek | Yes | Downstream | ●●●● Not triggered | ●●●● Not triggered | ●●●● Not triggered | Similar to LW14-17. Rainfall-runoff modelling supports this finding. |
| LA4S1 | LA4 | Yes | Above LWs | ●●●● Level 3 | ●●●● Level 3 | ●●●● Level 3 | Similar to LW14-17, with improved data availability. * Flow monitoring method means that Method B assessment assess low flows, not true 'cease-to-flow'. |
| LA3S1 | LA3 | Yes | Above LWs | ●●●● Level 3 | ●●●● Level 3 | ●●●● Level 3 | Similar to LW16-17. |
| LA2S1 | LA2 | Yes | Above LWs | ●●●● Level 2 | ●●●● Not triggered | ●●●● Level 3 | Pattern of flow changed during LW18, but broadly similar to LW17. |
| NDS1 | ND1 | Yes | Headwater | ●●●● Not triggered | ●●●● Not triggered | ●●●● Not triggered | LW18 mines under part of ND1 tributaries. No discernible effects. However, rainfall-runoff modelling suggests Level 3 impact. |

●●●● = result of previous longwalls (LW14-17)

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| Site Watercourse | Position of sub-catchment relative to mining | D) Surface flow observations | | Comment |
|------------------|--|------------------------------|------------------|--|
| Wongawilli Creek | Between A3A and A3B | 4 months / 7 | Not triggered | Refer to Performance Measures |
| | | 3 months / 7 | Catchment closed | Wet conditions leading to catchment closure mean this assessment unlikely to be triggered. |

No change to catchment flow characteristics was identified at the Wongawilli Creek gauge downstream of Area 3B (WWL). The TARP assessment methods indicate a continuation of modified

low-flow characteristics at the downstream gauge of Donalds Castle Creek (DCU), which remains at TARP Level 1. Reductions in median flow (Q50) at sites upstream of DCU are obvious, and total approximately 40-60% of median flow at DCU, and so should be able to be detected at DCU, but no reduction in Q50 is apparent.

While noting 'no change' was detected, it is acknowledged that the scale of impacts in headwater streams overlying longwalls (e.g. WC21, DC13) may be impossible to detect further downstream given natural variability, larger contributing (and un-mined) catchments downstream at WWL, as well as the inherent uncertainties of the assessment methods. However, the assessments of WWL and DCU appear to indicate that there is clear potential for returned or re-emergent flow that has been identified as lost from upstream headwater catchments.

Table 2. Summary of surface water flow Performance Measures– Longwall 18

| | |
|---|----------------------------------|
| Wongawilli Creek – minor environmental consequences | This Performance Measure is met. |
| Donalds Castle Creek – minor environmental consequences | This Performance Measure is met. |
| Lake Avon – negligible reduction in the quantity of surface water inflows to Lake Avon | This Performance Measure is met. |
| Cordeaux River – negligible reduction in the quantity of surface water inflow to the Cordeaux River at its confluence with Wongawilli Creek | This Performance Measure is met. |

Pool levels

The water level in Pool 50 (previously Pool 43A) on Wongawilli Creek declined since 2012, and a TARP Level 3 impact was recorded when water levels were recorded below baseline on 20/11/2017, following the identification of a stream bed fracture in 2013. The decline in pool levels started prior to the formation of the fracture and was coincident with a decline of groundwater pressures in lower Hawkesbury Sandstone associated with mining in Area 3A and Area 3B. Piezometric levels in the sandstone substrate adjacent to Wongawilli Creek have recovered as mining in Area 3B moves south and away from Pool 50 and since 2021 are above elevation of the creek bed. Pool water levels trended higher since 2020 period.

LA2 is a second order tributary to Lake Avon. Longwall 17 passed beneath 730 m of the main second order watercourse between 13/12/2020 and 2/3/2021 and Longwall 18 passed beneath the southernmost 17% of the LA2 catchment between 5/12/2021 and 6/3/2022. A review of field observations of pool level and outflow status indicates an apparent mining effect at Pools 24 and 25 and likely at Pool 5 following Longwall 17. No further effects are apparent due to Longwall 18.

ND1 is a second-order watercourse that flows west to join Native Dog Creek below the FSL of Lake Avon. Longwall 18 commenced at a distance of 265 m from ND1_Pool2 and mined directly beneath the upper reaches of ND1C, including ND1C_Pool2 in early April 2022. A review of field observations indicates that ND1C_Pool2 became dry following Longwall 18. No mining affect is apparent in pools along the main second-order watercourse.

Hydrology at Waterfall WF54

Initial analysis suggests that the relationship between WF54 pool levels and those at WWU has changed since mid-December 2021. The effect is observed at moderate to high water levels following rainfall, but not during lowest pool levels, and the flow status of the pool has not changed. The manner of the effect and the distance from active longwalls means that further investigation into the

cause and magnitude of any change is required. This is ongoing and will be reported to agencies once completed.

Swamps

At reference swamp sites, shallow groundwater levels recovered between 2020 and 2022 following the 2017-2019 drought period.

Longwall 18 passed beneath, or within 400 m of Swamps 149 and 35a, 35b, 150 and 151. No shallow groundwater TARPs were triggered in monitored swamps within the zone of influence of Longwall 18; however, given the high rainfall during the review period, these should be reassessed when more data are available. TARPs were previously triggered at Swamps 01a, 01b, 03, 05, 10, 11, 13, 14 and 23. The observed effects at swamps are in line with impacts anticipated in the SMP.

Soil moisture levels have recovered significantly at reference and impact swamp sites since 2020. No soil moisture effects are apparent at monitoring sites within the Longwall 18 zone of influence. Given the high rainfall over the review period, soil moisture at swamps 149, 35a and 35b should be reassessed when more data are available (next EoP review).

I. Introduction

Illawarra Metallurgical Coal (IMC) operates the Dendrobium underground coal mine, located approximately 12 km west of Wollongong (NSW) in the Southern Coalfield. IMC is required under the conditions of mining approval to submit regular reviews of the local hydrological data, including water quantity and quality, for watercourses and water bodies above and adjacent to Dendrobium Mine.

Surface water monitoring has been undertaken by IMC since 2003. Field parameter measurements and sampling for more detailed laboratory chemical analyses were collected by the IMC Environmental Field Team (IMCEFT). Field observation sites include hydrographic gauging stations, shallow groundwater piezometers, soil moisture sensors and surface water sampling sites.

This End of Panel (EoP) assessment reviews hydrographic and water quality data for the Wongawilli Creek catchment, upper Donalds Castle Creek catchment, Lake Avon sub-catchments and the Sandy Creek catchment up to one month after the completion of Longwall 18. Data are assessed against baseline and impact criteria defined in the Trigger Action Response Plan (TARP) which forms part of the Subsidence Management Plan for Area 3B (BHP Billiton, 2015) and the Swamp and Watercourse management plans contained therein.

1.1 Reporting Objectives

This EoP surface water assessment report has been prepared to form part of IMC's EoP Review which satisfies Condition 3-9 of the Approval for Dendrobium Mine (DA 60-03-2001). The EoP Review:

- reports all subsidence effects (both individual and cumulative) for the longwall panel and compares subsidence effects with predictions;
- describes in detail all subsidence impacts (both individual and cumulative) for the panel;
- discusses the environmental consequences for watercourses, swamps, water yield, water quality, aquatic ecology, terrestrial ecology, groundwater, cliffs and steep slopes; and
- compares subsidence impacts and environmental consequences with predictions.

This report provides the following assessment for the EoP Review:

- Impacts to water flow, water levels and water quality in watercourses, including: Wongawilli Creek, Donalds Castle Creek, Sandy Creek, Native Dog Creek and relevant tributaries to Avon Reservoir.
- Impact to flows at Waterfall 54.
- Impacts to shallow groundwater levels and soil moisture levels in mapped Coastal Upland Swamps within the mining area of influence, compared with reference swamps.

1.2 Longwall 18

Longwall mining at Dendrobium has been carried out in three designated areas: Area 1 (east of Lake Cordeaux), Area 2 (west of Lake Cordeaux), and Areas 3A and 3B (between Lake Cordeaux and Lake Avon). Coal is extracted from the Wongawilli Seam in Areas 1 through 3B. Previous workings in the Wongawilli Seam are located to the south at Elouera and Nebo, and to the east at Kemira. The overlying Bulli Seam was mined previously at Mt Kembla to the east of and partially overlapping Area 1.

Extraction of Longwall 18 commenced on 2/12/2021 and was completed on 17/5/2022. Longwall 18 is the tenth panel to be extracted in Area 3B, with an extracted length of 1018 m, a void width of 305 m (including first workings) and a cutting height of up to 3.9 m.

1.3 Feedback from agencies on previous assessment

WaterNSW provided feedback to DPIE in relation to Surface Water components of the Longwall 17 End of Panel Reporting in a letter dated 1/5/2022. WaterNSW noted that previous comments relating to Longwall 16 reporting had been adequately addressed in the previous End of Panel report. WaterNSW recommendations in relation to Surface Water are listed in Table 3.

Table 3. Comments on previous EOP assessment

| WaterNSW comment | Response |
|--|---|
| Investigate a potential link between anomalous water quality results observed in the tributary of Banksia Ck (SC10C_Pool1) and increasing 3-year trends in sulphate and dissolved metal (manganese, zinc) concentrations at Sandy Creek (SCK_Rockbar5) | IMC has engaged HGE0 to investigate a potential link between SC10C_POOL1 and SCK_ROCKBAR5. The investigation will include a monthly longitudinal survey from SC10C_POOL3 to SCK_ROCKBAR5 for a period of 6 months (pending catchment access). Water quality, water chemistry (major ions and metals) and visual observations will be collected at various pools along the stream. It is proposed that HGE0 will complete a specialist report once results have been collated. |
| Continue reviewing and improving stream flow monitoring and when uncertainty results are reported, include a brief discussion of results. | <p>Flow site upgrades are progressing as approved by Activity Approval D2021/29912 and D2021/130997. Site installs have commenced. On 30 August and 1 September 2022 work was completed at site WC20S2 and LC6S1, respectively. Work proposed for 2021 and 2022 were delayed due to COVID and extensive catchment closures.</p> <p>Flow sites experienced damage due to the flood events in 2022. IMC has commenced investment into repairs/ upgrades to these sites.</p> <p>A measurement uncertainty analysis is included in this report.</p> |

In a more recent email to IMC, dated 22/8/2022, WaterNSW expressed concern with the number of new impacts reported in Area 3B associated with longwalls 16 to 18, including in Lake Avon tributary LA2 and Wongawilli Creek Pool 14, Pool 42, Waterfall 54 and water quality impacts. WaterNSW requested that IMC prepare a specialist report on the cumulative impacts on Wongawilli Creek due to extracted, approved, and proposed longwalls in Dendrobium Area 3. It was requested that the report be included as an appendix to the Longwall 18 End of Panel Report.

2. Surface water and groundwater management

This section outlines the network of monitoring infrastructure and sites operated by IMC at and around the Dendrobium Mine. Further details of monitoring sites and procedures are outlined in the Dendrobium Area 3B Watercourse Impact Monitoring Management and Contingency Plan (South32, 2020a).

2.1 Surface Water Monitoring

Monitoring includes a selection of sites downstream and within the mining area, as well as sites located away from the mining area to provide control sites and act as a comparison to impact sites. Pools within streams are monitored monthly before and following mining and weekly (when site access available) during active subsidence, and in response to any observed impacts. Surface water monitoring sites fall into four categories:

- **Flow gauge sites** at which stream flow is monitored at a calibrated gauge or weir.
- **Water chemistry sites** at which samples are collected for laboratory analysis (DOC, Na, K, Ca, Mg, Filt. SO₄, Cl, T. Alk., Total Fe, Mn, Al, Filt. Cu, Ni, Zn, Si), in addition to water observations, field parameters.
- **Water field parameter sites** at which water quality field parameters are measured (pH, Electrical Conductivity (EC), temperature, Dissolved Oxygen (DO), Oxygen Reduction Potential (ORP), in addition to water observations.
- **Water observation sites** at which pool water levels and flow status are noted and photographs taken upstream and downstream.

At a subset of sites, data loggers are installed in pools to allow monitoring of pool water level and temperature at hourly intervals. The monitoring of water quality parameters provides a means of detecting and assessing the effects of streambed fracturing or induction of ferruginous springs.

Figure 1 shows the location of surface water monitoring and sampling sites in relation to the extracted and planned longwall panels. Figure 2 shows the locations of hydrographic gauging stations which extend beyond the mining lease.

A summary of water flow monitoring sites in Areas 3A and 3B is presented in Table 4 and a full list of all installations is included in Appendix B. Several more sites have been installed or are planned in Area 3C in advance of operations commencing there.

Table 4. Surface Water Flow Monitoring Sites in Area 3A and 3B

| Area | Site | Installation | Catchment | Easting (MGA z56) | Northing | Catchment area (km ²) |
|------|--------|--|------------------|-------------------|----------|-----------------------------------|
| A3B | WWU | Natural control; Stainless Steel housing; Diver logger | Wongawilli Creek | 290808 | 6189716 | 3.211 |
| A3B | WWL | Natural control; Stainless Steel housing; Diver logger | Wongawilli Creek | 290975 | 6197526 | 20.079 |
| A3B | WWL_A | Installed August 2019. Weir and half pipe; PVC housing; Orpheus logger | Wongawilli Creek | 290962 | 6197370 | 19.602 |
| A3B | WC21S1 | Natural control; Stainless Steel housing; Diver logger | Wongawilli Creek | 290529 | 6194255 | 2.434 |
| A3B | WC15S1 | Natural control; PVC housing; Diver logger | Wongawilli Creek | 290754 | 6192239 | 1.192 |

| | | | | | | |
|-------|--------------------------|---|----------------------------|--------|---------|-------|
| A3B | WC12S1 | Weir and half pipe flume; Polypipe housing; Orpheus logger | Wongawilli Creek | 290964 | 6191459 | 0.38 |
| A3B | LA2S1 | Weir and half pipe flume; Polypipe housing; Orpheus logger | Lake Avon | 288364 | 6191364 | 0.824 |
| A3B | LA3S1 | Weir and half pipe flume; Polypipe housing; Orpheus logger | Lake Avon | 288385 | 6191548 | 0.375 |
| A3B | LA4S1 | Modified control; Stainless Steel housing; Diver logger | Lake Avon | 288134 | 6192565 | 0.817 |
| A3B | NDT1S1 | Weir and half pipe flume; Polypipe housing; Orpheus logger | Lake Avon/Native Dog Creek | 288607 | 6190491 | 1.13 |
| A3B | NDCS1 | Weir and half pipe flume; Polypipe housing; Orpheus logger | Native Dog Creek | 288473 | 6190484 | 3.75 |
| A3B | DC13S1 | Natural control; PVC housing; Diver logger | Donalds Castle Creek | 289401 | 6194605 | 1.638 |
| A3B | DCS2 | Natural control; PVC housing; Diver logger | Donalds Castle Creek | 289502 | 6194572 | 1.084 |
| A3B | DCU | Natural control; Stainless Steel housing; Diver logger | Donalds Castle Creek | 289407 | 6195577 | 6.219 |
| A3A | SC10S1 | Natural control; Stainless Steel housing; Diver logger | Sandy Creek | 293608 | 6192516 | 2.771 |
| A3A | SC10CS1 | Natural control; Stainless Steel housing; Diver logger | Sandy Creek | 293358 | 6192433 | 0.817 |
| A3A | SCL2 | Modified control; Stainless Steel housing; Diver logger | Sandy Creek | 293819 | 6192648 | 7.029 |
| A3A,C | LC5S1 | Reference site until Area 3C is mined. Weir and half pipe flume; Polypipe housing; Orpheus logger | Lake Cordeaux | 293043 | 6195327 | 1.861 |
| Ref | CR36S1 | Weir and half pipe flume; Polypipe housing; Orpheus logger | Cordeaux River | 291482 | 6197652 | 1.75 |
| Ref | O'Hares Ck at Wedderburn | | | 300411 | 6217387 | |

2.2 Improvements to monitoring network in reporting period

| Type of change / improvement | Description of recent change | Reference / comment |
|---------------------------------|--|--|
| New surface water gauging sites | None in Area 3B during recent EOP period. New sites approved, and some installed. | None installed. Figure B1 (Appendix B) shows network. |
| Upgrade of existing sites | None during recent EOP period | |
| Gauge rating curves | More gaugings taken at most sites. Rating curves updated at most sites. | Details from ALS (consultants) can be requested via IMC. Methods to estimate uncertainty in surface water flow estimation has been developed by Enviromon (consultant), and is being rolled out to all sites. See Appendix C5 for sites assessed thus far. |
| Pool monitoring sites | Installation of additional water level data loggers in key pools. | Additional water level loggers installed in pools in Wongawilli Ck (more relevant to Area 3C). |
| Revision of assessment methods | Surface flow TARPs (Assessments A-D) not change since agreement in early 2020. IAPUM requested that old method (comparison of rainfall-runoff modelling) be re-instated. | Section 5.4, WIMMCP (IMC, 2020a) and Watershed HydroGeo, 2019a. Section 2.3.1. Peer-review of methods planned for early 2022. |
| WWL vs WWL_A correlation | No change. Enviromon analysed the common period of WWL and WWL_A records in order to allow cessation of monitoring at WWL. Due to the shorter record at WWL_A and uncertainties at WWL it is recommended to continue | See separate document (Enviromon, 2021). |

| | | |
|--|--|--|
| | to rely on data from WWL until the end of Area 3A (Longwall 19), and use WWL_A thereafter. | |
|--|--|--|

2.3 Surface water flow data update

IMC's contract hydrographers, ALS, provided the most recent flow data for assessment for sites in and around Area 3B (details in Table C1 of Appendix C). This has been augmented by flow data from sites managed by WaterNSW, specifically one of the primary reference flow gauges (O'Hares Creek at Wedderburn) and for WaterNSW's Sandy Creek gauging station (GS 2122205). The WaterNSW Sandy Creek gauging station is co-located with IMC's SCL2 gauging site, but has a longer record and, based on comments from ALS, relies on higher accuracy monitoring equipment.

This data was then assessed based on the quality of records provided before some further processing was conducted. A discussion of this assessment is provided below. As is standard, data is available to agencies on request.

2.3.1 Re-rating of flow records

ALS updates the rating curves of flow monitoring sites as new manual gaugings are taken and added to the dataset that correlates 'stage' (water level at a monitoring site) and flow at the site. In recent times, WaterNSW has granted limited access to the Special Area during wetter periods in order to improve the moderate/high flow sections of the rating curves. This has meant that historical records of estimated flow can change when a rating curve is updated. ALS has confirmed that no sites were re-rated since November 2020 (prior to the Longwall 16 EOP).

Hydrographers ALS took over the contract for flow monitoring at Dendrobium on 11/05/2016. ALS provide the record of daily flow for each IMC site based on the latest rating curve and the historical record of stage (level) at each site. ALS do not provide re-rated data from before their contract date, i.e. before 11/05/2016.

Table 5. Stream gauges that have been re-rated

| DATE | A3B GAUGES RE-RATED | OTHER GAUGES RE-RATED |
|---------------|--|--|
| October 2020 | WWLA, DCU, WWU, DC13S1, DCS2, WC21S1, WC15S1 | SCL2, SC10S1, SC10CS1 |
| November 2020 | LA2S1, LA3S1, WC12S1, NDT51 | AR31S1, AR32S1, LA13AS1, LA13S1, AR19S1, DC8S1, CR29S1, CR31S1, LC5S1, LA8S1, CR36S1 |

It is apparent from review of recent data obtained from WaterNSW for O'Hares Creek (WaterNSW site 213200) that a similar re-rating process occurs in WaterNSW data but has not been confirmed by WaterNSW.

There are two implications of the re-rating process:

1. Estimates of flow included in previous EOP reports may be different to that reported in the current (or future) EOP report. For example, median flow for sub-catchment WWU for the period May-2016 to June-2020 was 0.068 ML/d in the EOP for Longwall 15 but was revised to 0.202 ML/d EOP for Longwall 16 due to changes to the rating curve.

2. For gauging sites that commenced operation before the contract date of ALS (11/05/2016), time-series data prior to that date need to be adjusted to account for re-rating. This pre-processing step was accomplished by comparing the 'old' (pre-ALS) flow data and the new rating curve in order to derive a flow record that is based on a consistent rating curve across the entire record.

2.3.2 Data quality assessment

An analysis of the data received from ALS was performed to assess the reliability and continuity of data collected at each flow gauge. The data quality code recorded by ALS for flow measurements was used for this purpose. A summary of these data quality codes has been provided in Table C2 of Appendix C, alongside the data quality assessment of each flow gauge.

Each daily flow recorded is the average flow determined from multiple sub-daily (typically 15-minute interval) stage measurements. The Hydstra database maintained by ALS will assign the 'worst' data quality code from any of the sub-daily records to the aggregated or averaged daily record. It is for this reason that Hydstra will sometimes assign quality code 140 ("Level below cease-to-flow") to days where there is a small, non-zero average flow.

For each flow gauge the percentage of available daily flow measurements was calculated. This value indicates the number of measurements that exist between the first date of data collection and the last available date. From this the percentage of 'suspect' data was calculated. Based on the ALS quality codes, suspect data refers to any flow data with a code that falls between 104 and 255. A summary of the data quality assessment for each flow gauge is included in Table C3 and C4 of Appendix C.

Data processing was then undertaken for flow data where entries were blank or entered as text and these could be confidently infilled. These entries were associated with the following quality codes:

- 151 ("data not yet available"): associated with comments of 'rating exceeded', commonly following high regional rainfall events;
- 161 ("poor quality data from debris affecting sensor"): occurred only at flow gauge WWU for the period 23/01/2019 to 27/02/2019;
- 205 ("data lost"): associated with comments such as 'logger dead', 'data lost';
- 255 ("no data exists"): associated with comments of 'rating exceeded', 'logger dead'.

For these entries an infilling procedure was used to estimate the flow value, if the record could be confidently estimated (e.g. flows were consistent through time and compared to other gauging stations, especially at higher flows when the "rating exceeded" flag was assigned. Flow estimates were calculated using either the average flow from the preceding and following days, or the flow recorded at another gauged sub-catchment for the same day, scaled by catchment size. The percentage of infilled data is recorded for the relevant gauges in Appendix C. The results of processing, with comparison against 'raw' data are illustrated on charts in Appendix C.

2.4 Shallow Groundwater Monitoring

Figure 2 shows Longwall 18 in relation to the locations of shallow groundwater monitoring sites in Areas 3B and 3A. Typically, these sites are piezometers approximately 1 - 3 m deep that monitor groundwater levels within the swamp deposits located around the Dendrobium area.

Figure 2 also shows swamp areas: broadly mapped by NSW Office of Environment and Heritage (OEH) and refined through site-scale mapping for IMC carried out by Biosis and Niche Environment and Heritage. Note that the TARP assessment relates only to those piezometers that are located within swamp sub-communities mapped as Banksia Thicket, Sedgeland-heath complex and Tea Tree

Thicket; being listed as Coastal Upland Swamp Endangered Ecological Community (EEC). Piezometers located within other areas, such as fringing Eucalypt Woodland, are excluded from the TARP assessment as per the advice from OEH (17/01/2014).

A summary of the shallow groundwater monitoring sites is presented in Table 6. Swamps that overlap Longwall 18 are indicated by pink shading.

Table 6. Summary of Swamp Monitoring

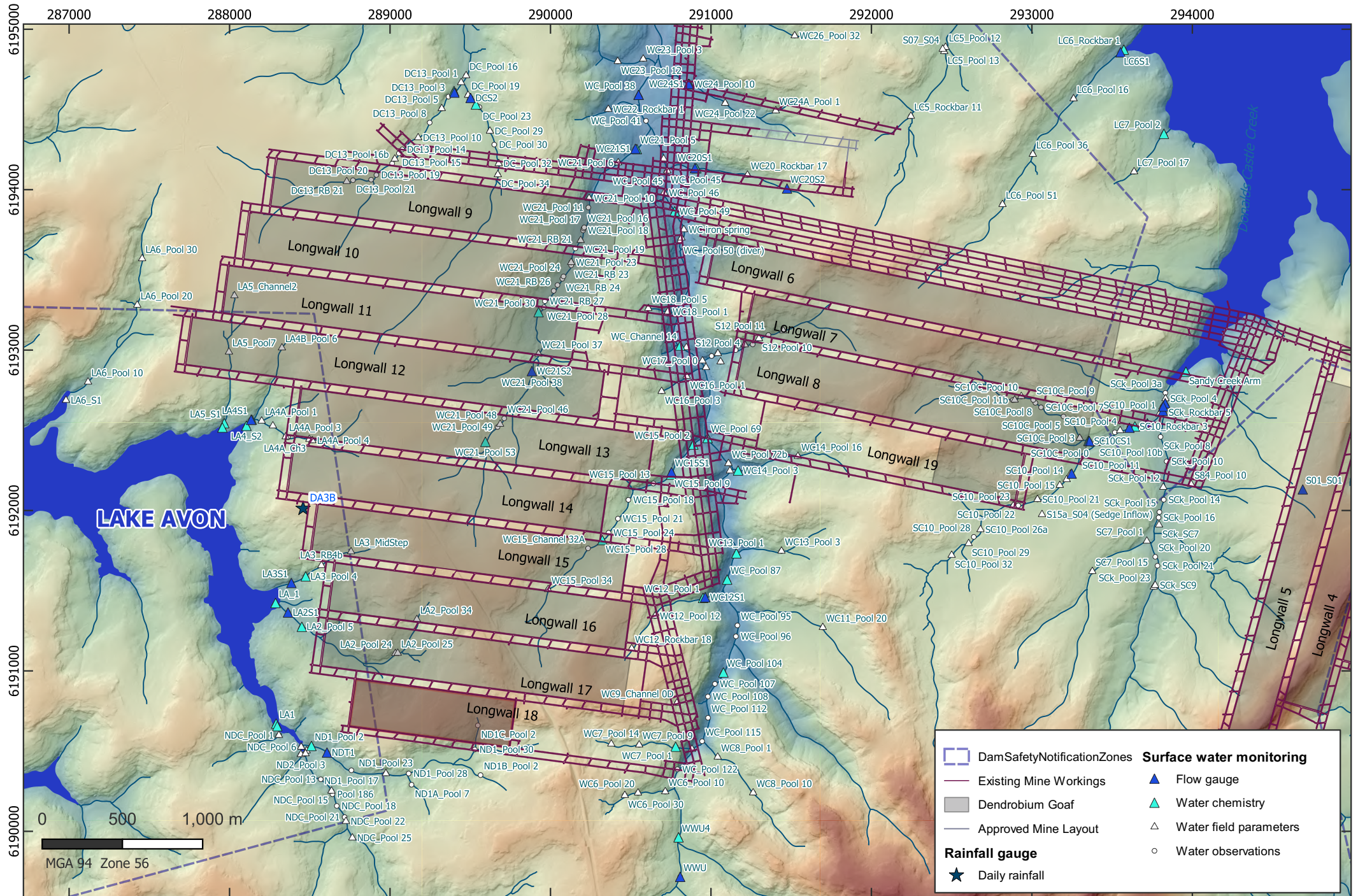
| Swamp | Site type | Number of piezometers | | Within mining area of influence (<400 m) |
|-------|--|-----------------------|--|--|
| | | Total | TARP (Within Coastal Upland Swamp EEC) | |
| 01a | Impact | 9 | 6 | Longwalls 9,10 |
| 01b | Impact | 7 | 6 | Longwall 9 |
| 02 | Reference | 1 | n/a | No |
| 03 | Impact | 1 | 1 | Longwalls 11,12 |
| 05 | Impact | 9 | 6 | Longwalls 9-12 |
| 07 | Reference | 2 | n/a | No |
| 08 | Impact | 6 | 0 | Longwalls 9-11 |
| 10 | Impact | 1 | 1 | Longwall 12 |
| 11 | Impact | 3 | 3 | Longwall 13,14 |
| 13 | Impact | 1 | 1 | Longwalls 13-16 |
| 14 | Impact | 2 | 2 | Longwalls 15-17 |
| 15a | Impact (lower section within 400m mining area) | 16 | 7 | Longwall 8,19 |
| 15b | Impact | 23 | 10 | Longwall 7, 8 |
| 22 | Reference | 2 | n/a | No; Elouera Colliery |
| 23 | Impact | 2 | 2 | Longwalls 15-17 |
| 25 | Reference | 1 | n/a | No |
| 33 | Reference | 2 | n/a | No |
| 35a | Impact | 1 | 1 | Longwalls 17,18 |
| 35b | Impact | 1 | 1 | Longwall 18 |
| 84 | Reference | 1 | n/a | No |
| 85 | Reference | 2 | n/a | No |
| 86 | Reference | 2 | n/a | No |
| 87 | Reference | 2 | n/a | No; Avon Colliery |
| 88 | Reference | 2 | n/a | No; Huntley Colliery |
| 149 | Impact | - | n/a | Longwalls 16-18 |
| 150 | Impact | 1 | 1 | Longwall 18 |
| 151 | Impact | 1 | 1 | Longwall 18 |

Pink shading: swamps within mining area of influence of Longwall 18

2.5 Soil moisture monitoring

Soil moisture profiles are monitored at most swamps, with sensor arrays typically positioned near shallow piezometers (where possible). Where possible the monitoring arrays are numbered according to the corresponding piezometer (if present) with the addition of an 'S' prefix. At most locations, sensors are installed up to a maximum depth of 1.2 m.

Soil moisture is measured using Sentek sensors which monitor changes in the dielectric constant within a cylinder of soil extending to a radial distance of 10 cm from the access tube. Soil moisture is reported as mm water per 100 mm soil depth (or volumetric % water) at each monitored depth (Sentek, 2017). The most recent installations are equipped with automated data loggers set to record moisture levels every hour. The remaining installations are recorded manually during scheduled site visits.



Dendrobium DND Shaft 5 Groundwater Contamination Assessment
Groundwater contours and monitoring sites

Figure 1

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280000

285000

290000

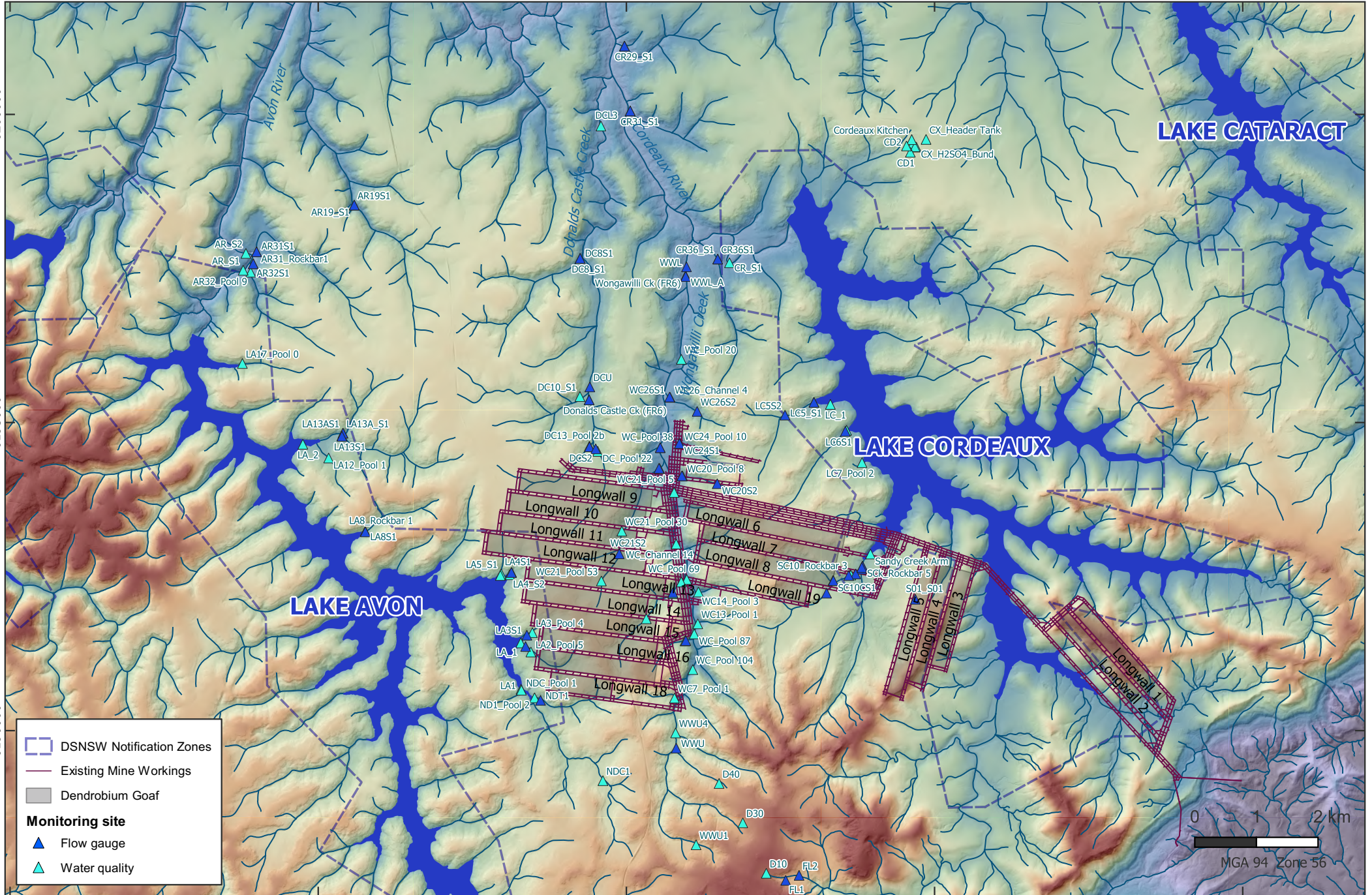
295000

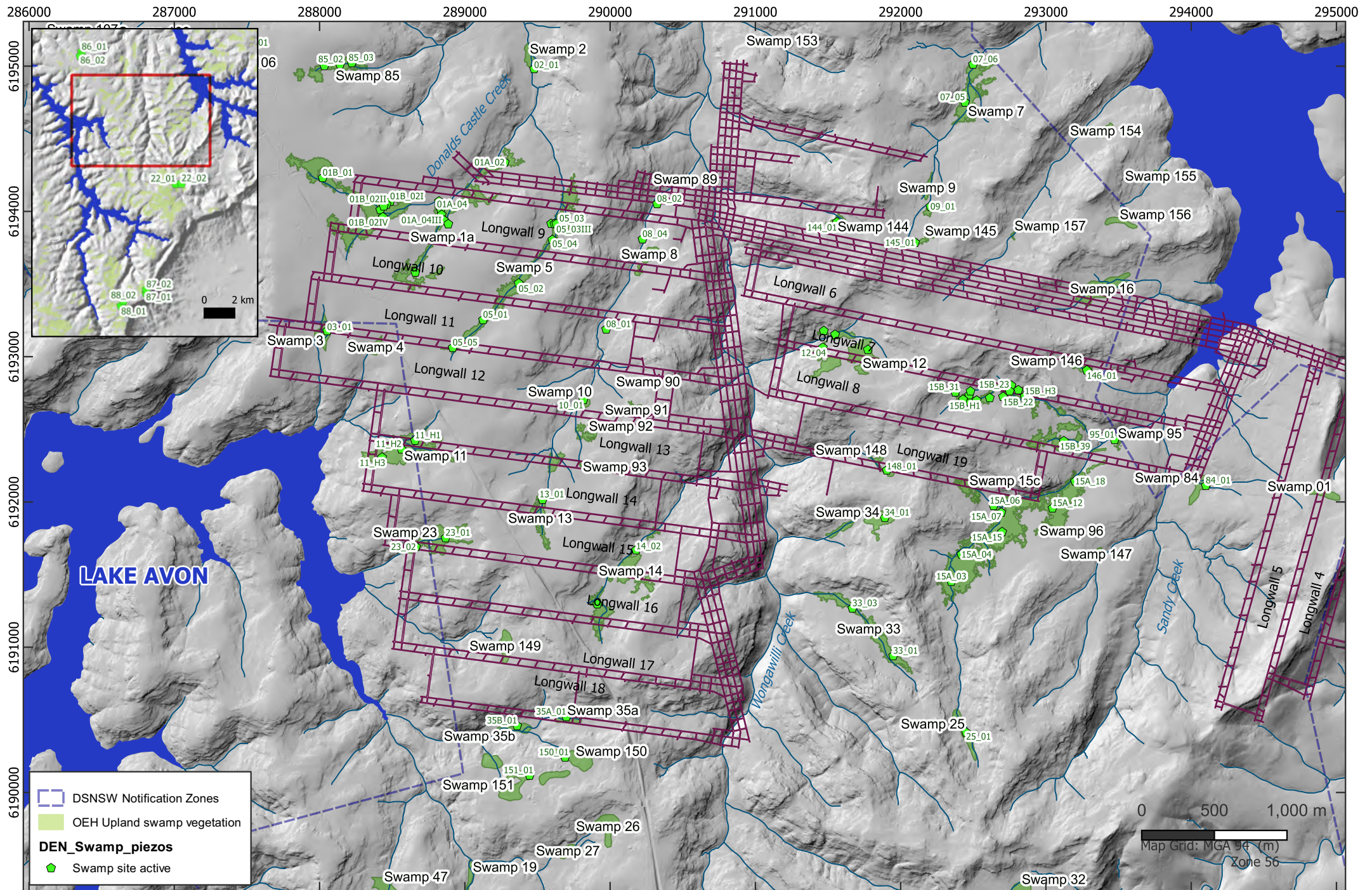
300000

620000

619500

619000





Dendrobium Mine End of Panel Surface Water Assessment
Swamp monitoring sites

Figure 3

file: Dendrobium5.qgz

2.6 Weather conditions during the assessment period

Rainfall data are collected from several gauging stations across the mining lease. Weather observations at Area 3B since the start of Longwall 9 are summarized in Figure 4. Potential evapotranspiration (EVT) is calculated from SILO data (DSITI, 2011) derived for Dendrobium Area 3B, using the FAO Penman-Monteith formula (Allen et al., 1998). The average annual rainfall for Dendrobium is 1050 mm (2002 – 2021; Dendrobium site data). Rainfall events occur year-round but tend to be more frequent in the summer and early autumn months. It is common for a substantial proportion of the annual rainfall to be delivered in a small number of large rainfall events, during which significant surface water runoff and groundwater recharge is generated.

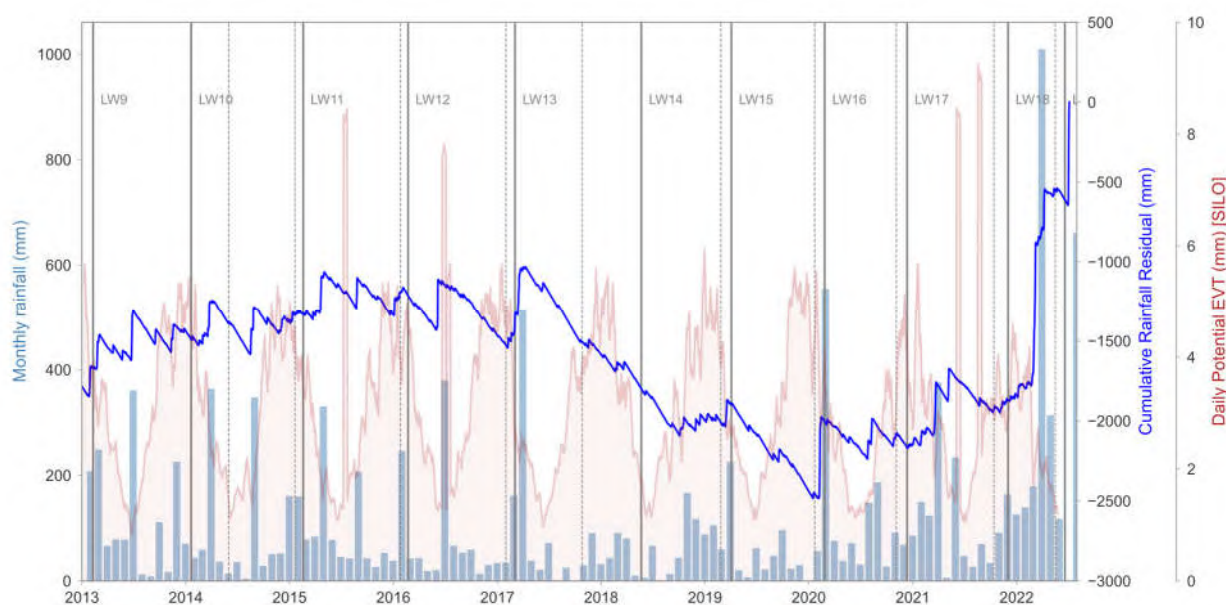


Figure 4. Rainfall and potential evapotranspiration (EVT) at Area 3 for the reporting period

Maximum daily temperature varies seasonally from approximately 20 °C in the winter months (June – August) to 40 °C or higher in the summer (December – February). Evapotranspiration varies seasonally in line with temperature and solar radiation, peaking during the summer months.

Rainfall during Longwall 18 extraction was well above average, totalling 2281 mm in the calendar year to the end of the longwall (18/5/2021 – 17/5/2022). Extremely heavy rainfall was experienced in March 2022 when 1010 mm was recorded in a single month. This follows similarly high rainfall in 2020 (1436 mm) and 2021 (1448 mm) due to sustained La Nina conditions over that period. As a result, there has been a full recovery in stream flow, shallow groundwater levels and soil moisture across all catchments since the severe drought of 2017-2019.

Soil moisture levels derived from the Australian Water Resources Assessment Landscape model (AWRA-L) are published by the Bureau of Meteorology (BOM). A timeseries of estimated soil moisture storage for the Woronora Plateau in the vicinity of Dendrobium Mine is shown in Figure 5. Soil moisture storage declined to record low levels during the 2017-2019 drought. Due to the higher-than-average rainfall since 2020, soil moisture levels are at their highest levels than at any other time during mining in Area 3B.

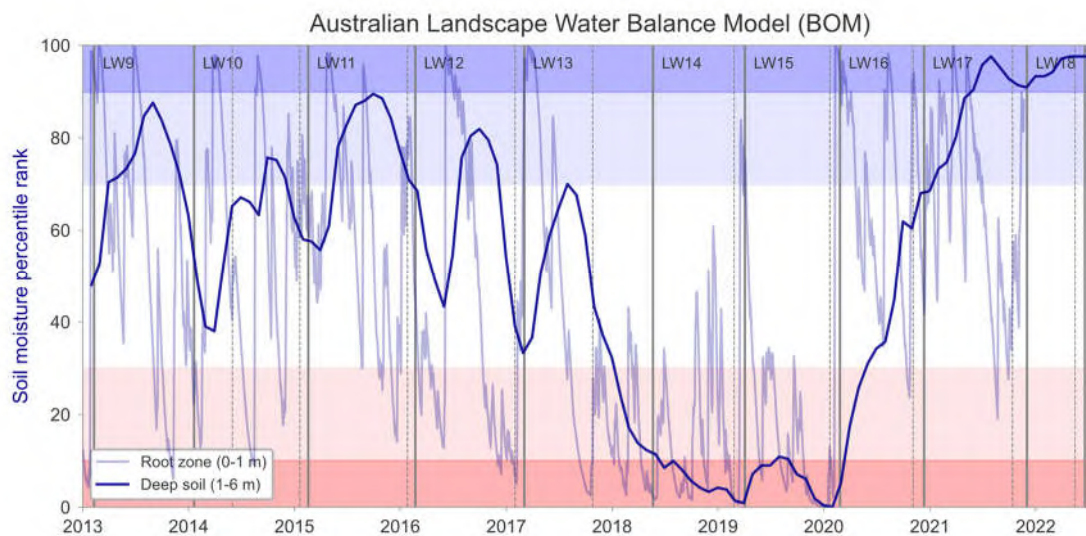


Figure 5. Calculated soil moisture from BOM’s AWRA Landscape Model

Longer term rainfall trends for the plateau are shown in Figure 6 which is based on data from SILO (slightly different from site measurements). Since 1960 there have been four periods of drought on the plateau (yellow shading), the most severe being the Millennium Drought which ended in 2004. Severe drought conditions returned between 2017-2019. The high rainfall since 2020 has been similar to rainfall conditions during much of the 1980s and early 1990s, prior to the Millennium Drought.

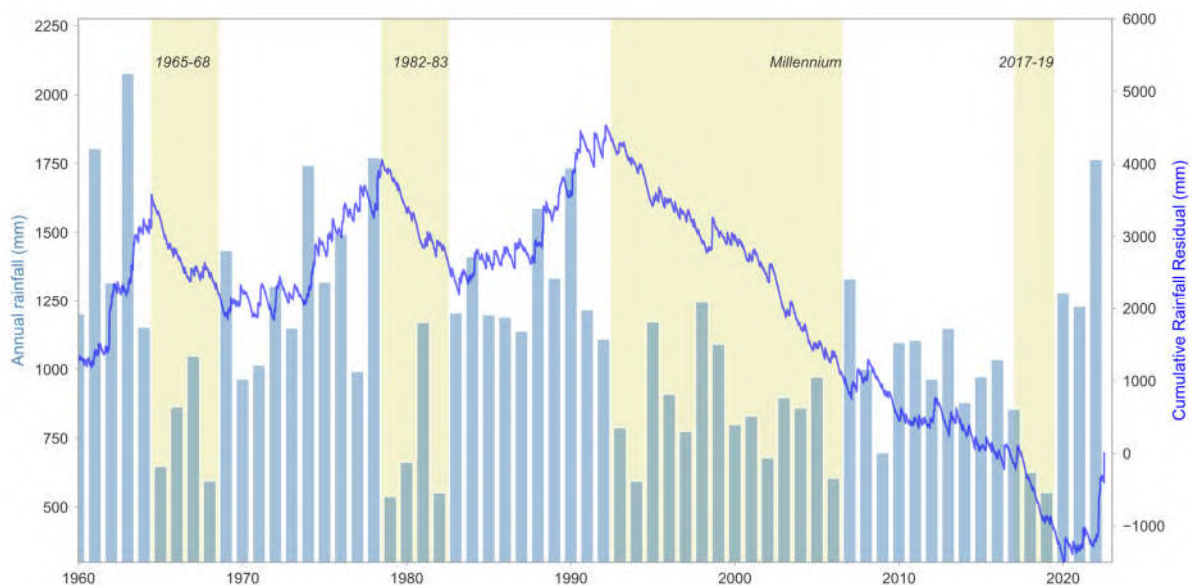


Figure 6. Long-term rainfall trends at Dendrobium (1960-2021)

3. Longwall subsidence effects

Figure 7 presents the total subsidence predicted by MSEC (2020) above Area 3B longwalls. This shows that Wongawilli Creek and Lake Avon are outside the main area of subsidence, although tributaries such as WC21 and WC15 lie directly across the area of predicted maximum subsidence from recent or future longwalls (See Figure 1 for tributary locations). Lake Avon tributaries LA2 to LA5 partially overlap some Longwalls 9-18. Longwalls in Area 3B are set back at least 300 m from the Full Supply Level (FSL) of Lake Avon.

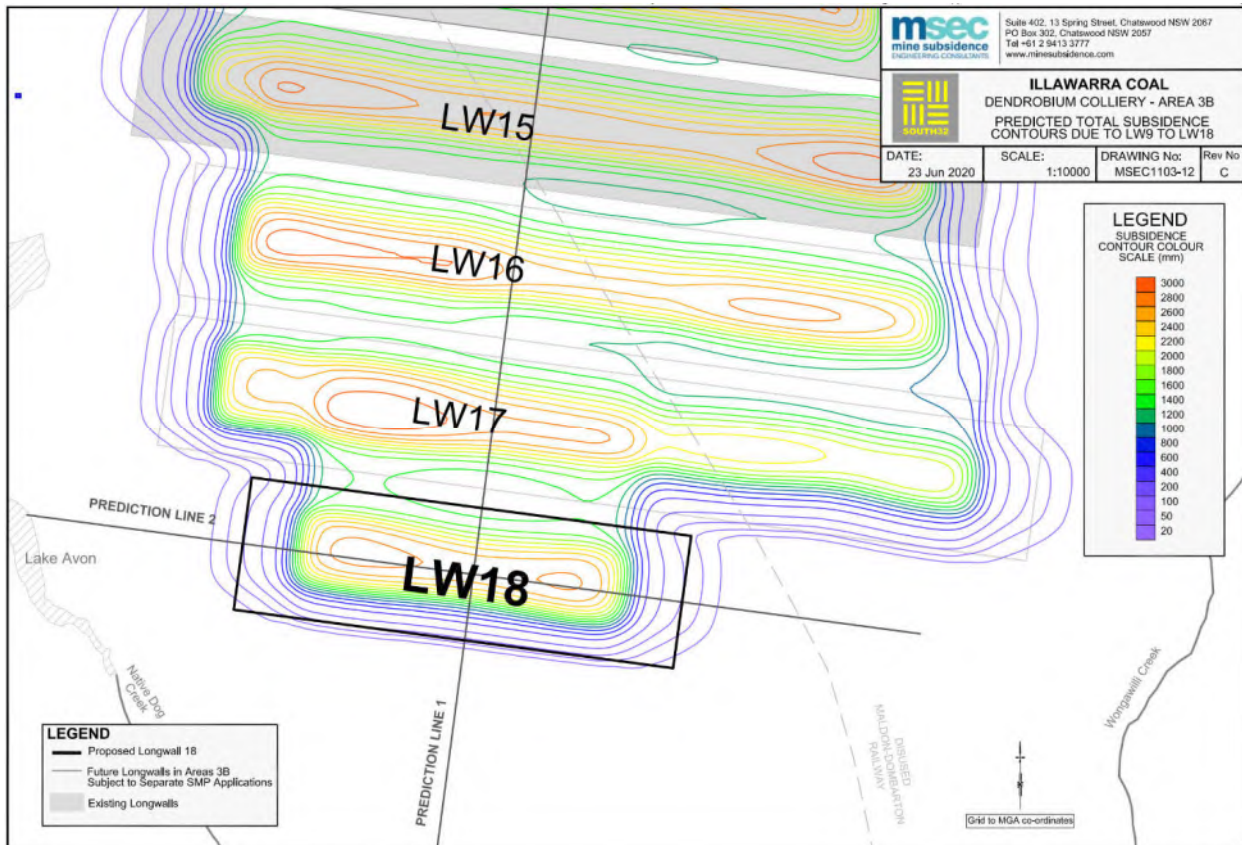


Figure 7. Predicted Subsidence above Area 3B (from MSEC, 2020)

3.1 Measured subsidence

Observed mine subsidence movements due to the extraction of Longwall 18 were reviewed by MSEC (2022). Mine subsidence effects were measured using the Wongawilli Creek closure lines, Avon Dam closure lines, Area 3B and Avon Dam 3D monitoring points, creek and tributary cross lines, swamp cross lines and airborne laser scans. The review concluded that:

- *ground movements measured due to the mining of Longwall 18 are consistent with predictions provided in MSEC Report 1103 (2020) which supported the SMP Application for LW18.*

3.2 Observed surface impacts

Surface watercourses and catchments mined beneath by Longwall 18 are listed in Table 7.

Table 7. Surface water features within area of Longwall 18 influence

| Catchment / location | Approximate dates | Monitoring locations (level and chemistry) | |
|----------------------|--|--|---|
| | | Upstream | Downstream |
| Lake Avon | Longwall 18 commenced at a distance of 303 m from Lake Avon FSL shoreline, including first workings | - | LA, LA4S2 |
| LA2 | Longwall 18 overlapped the southern 20% of the LA2 catchment. The watercourse itself was mined beneath previously by Longwall 17 between 13/12/2020 and 16/3/2021. | - | LA2_Pool 5, LA2S1 |
| Native Dog Creek ND1 | Longwall 18 mined beneath the upper reaches of ND1C in early April 2022. Commenced at a distance of 215 m from gauge NDT1 and 265 m from water chemistry site ND_Pool2 | - | NDT1, ND1_Pool 2, Lake Avon LA1 |
| WC15 tributary | Longwall 18 is outside the WC15 catchment. Longwall 17 passed beneath the upper reaches of WC15 (and Swamp 14) between 18/6/2021 and 27/6/2021. | - | WC15_Pool 34, WC15_Pool 28, WC15_Pool 24, WC15_S1 |
| WC12 tributary | Longwall 18 is outside the WC12 catchment. Longwall 17 passed beneath the upper reaches of WC12 on 12/9/2021. | - | WC12_Rockbar 18, WC12_Pool 12, WC12_Pool 1 |
| Swamp 14 | Longwall 18 is passed 280 m south of Swamp 14 at its closest point. The swamp was previously mined beneath by Longwalls 17, 16 and 15. | Piezometers 14_01, 14_02 Soil moisture sensors | |
| Swamp 149 | Longwall 18 passed immediately to the south of Swamp 149 on 10/2/2022. Longwall 17 passed beneath the swamp in late March 2021. | No piezometers installed due to shallow soil profile 1 New soil moisture sensor (for Longwall 18 SMP) | |
| Swamp 35a | Longwall 18 mined beneath the northern fringe of Swamp 35a between 17/4/2022 and 17/5/2022. | Piezometer 35a_01 + Soil moisture sensor | |
| Swamp 35b | Outside Longwall 18 footprint. Longwall 18 passed Swamp 35b at a distance of 90 m on 20/3/2022. | Piezometer 35b_01 + Soil moisture sensor | |
| Swamp 150 | Outside Longwall 18 footprint. Longwall 18 passed 232 m from Swamp 150 at its closest point. | Piezometer 150_01 + Soil moisture sensor | |
| Swamp 151 | Outside Longwall 18 footprint. Longwall 18 passed 330 m from Swamp 151 at its closest point. | Piezometer 151_01 + Soil moisture sensor | |

Observed subsidence impacts on the landscape, including surface fracturing and iron staining are monitored by the IMCEFT and reported separately in the EoP Landscape Report (South32, 2021). A total of 25 new ground surface impacts attributed to the extraction of Longwall 18 were recorded (Figure 8). There was no fracturing or diversion of surface water flows identified along the streams. However, iron staining was identified in one location along LA3 to the west of LW16 (Table 8). The iron staining is most likely related to increased groundwater discharge via fractures associated with earlier longwalls and in response to higher-than-average rainfall since 2020.

Table 8. Reported subsidence impacts to stream beds during Longwall 18

| Site ID | Watercourse | Date Observed | Reported | Description | Tarp Level |
|----------|-------------|---------------|-----------|--|------------|
| LW18_024 | LA3 | 16/8/2022 | 16/8/2022 | Increase in Iron staining in Tributary LA3 | 1 |

MSEC (2022) reviewed surface impacts associated with Longwall 18 and reported that:

- *It is considered that the observed impacts on the natural features due to the mining of LW18 are consistent with the MSEC assessments provided in Report No. MSEC1103 (MSEC, 2020) which supported the SMP Application for LW18.*

3.3 Specialist advice in relation to observed impacts

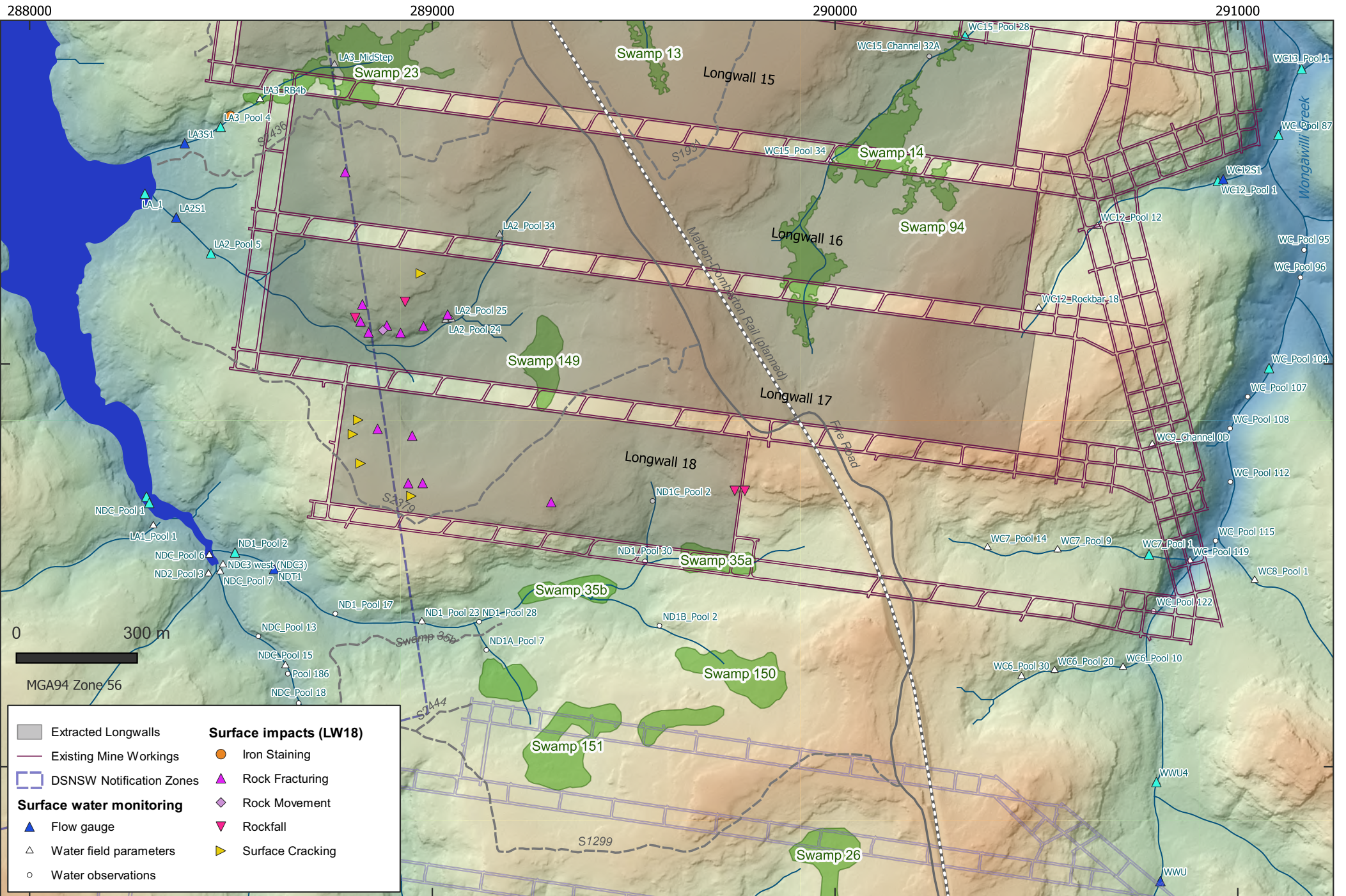
Subsidence impacts of TARP Level 2 or above require specialist advice in relation to possible Corrective Management Actions (CMAs), reporting and/or monitoring. No specialist advice reports issued in relating to subsidence impacts to watercourses of TARP level 2 or above for Longwall 18.

3.4 Valley closure at Waterfall 54

The mine subsidence effects for Waterfall 54 (WF54) along Wongawilli Creek have been measured by IMC using 2D survey techniques. Survey lines were established prior to the commencement of Longwall 16. MSEC (2022) notes a period of valley closure movement between August and September 2021, towards the end of Longwall 17. As a precautionary measure, Longwall 17 finished approximately 105 m short of the approved finishing location to minimise further closure at the waterfall. As of August 2022, total closure remains within predictions.

A rockfall was identified at Waterfall 54 along Wongawilli Creek after the completion of Longwall 18. A review of historical photographs found that the rockfall occurred between 6 and 28 October 2021 during (or soon after) the mining of Longwall 17. The impact is therefore not associated with Longwall 18. The rockfall was not identified earlier because it occurred behind a densely vegetated area and safety reasons prevented detailed inspections.

A Time Domain Reflectometry (TDR) cable is installed at monitoring bore S2478, located between Longwall 17 and Waterfall 54. TDR can identify small movements in geological strata such as basal shearing associated with valley closure. The most recent monitoring report (HGEO, 2022) identified: *a very subtle reflectance anomaly ... approximately 2.4 m above the base of the Newport Formation. The anomaly is first apparent after 18/8/2021 when Longwall 17 was 665 m west of Waterfall 54 and 470 m west of S2478. The anomaly reaches a maximum at ~1/10/2021 and remains unchanged thereafter.* In addition: *No further anomalies are apparent during or following Longwall 18 extraction.*



4. Assessment of surface water quality effects

During the 7-month reporting period between the start of Longwall 18 (2/12/2021) and one month after the end of Longwall 18 (17/6/2022), monitoring was carried out at more than 149 surface water sites. Sites were monitored on an approximately weekly or monthly basis, as per the Watercourse Impact Monitoring Management and Contingency Plan (WIMMCP).

Trigger values for water quality field parameters are defined in the WIMMCP Attachment 1 (South32, 2020a). Trigger thresholds (TARPs) have been defined for three locations downstream of the mining area for which there is adequate high-quality baseline information (Wongawilli Creek (at Fire Road 6 [FR6]) and Donalds Castle Creek (at FR6) and Lake Avon (tributary site LA4_S1). The TARPs are based on the field parameters pH, EC and DO and defined by the value three standard deviations (SD) from the baseline mean (mean plus 3SD for EC and mean minus 3SD for pH and Dissolved Oxygen). TARP levels are defined as follows:

- Level 1: One exceedance within six months
- Level 2: Two non-consecutive exceedances within six months
- Level 3: Three exceedances within six months
- Exceeding prediction: Mining results in two consecutive exceedances during the monitoring period. Predicted impacts are summarised in the WIMMCP.

TARP triggers for the monitoring period are summarised in Table 9.

Table 9. Summary of Water Quality TARPs for the monitoring period

| DATE | CATCHMENT / LOCATION | PARAMETER | VALUE | TARP | TRIGGER LEVEL |
|------------|----------------------|----------------------|-------|-------|-----------------------|
| 17/12/2021 | LA4_S1 | EC (µS/cm) | 160 | 129.8 | Exceeding Predictions |
| 9/02/2022 | LA4_S1 | EC (µS/cm) | 138 | 129.8 | |
| 17/12/2021 | LA4_S1 | pH | 4.01 | 4.9 | Exceeding Predictions |
| 9/02/2022 | LA4_S1 | pH | 4.60 | 4.9 | |
| 17/12/2021 | LA4_S1 | Dissolved Oxygen (%) | 67.6 | 69.5 | Level 1 |

Assessment of surface water quality effects, including TARP triggers is presented by catchment (watercourse) in the following subsections.

4.1 Overview of surface water quality

Hydrographs of stream field parameters (EC, pH and DO) are presented in Appendix A for 137 observation sites and hydrographs of dissolved sulfate, Fe, Mn, Al, Si and Zn are presented for 61 sites at which sampling and laboratory analysis are carried out. Due to the large volume of data, water quality trends (qualitative) for the review period are summarised for representative sites and sites at which significant or noteworthy trends are apparent in Table 10. A quantitative analysis of water quality trends is presented in Section 4.2, below.

In general, stream salinity (EC) has decreased over the last three years (and last three longwalls) due to higher-than-average rainfall and significant increase in runoff compared with the preceding two years. The decreasing trend follows slightly more saline conditions at most locations during the 2017-

2019 drought which resulted in low flows and evaporative concentration of salts. Similarly, DO has trended higher or remained stable over the reporting period due to high stream flows.

Anomalous water quality effects are noted in streams that have been directly mined under by previous longwalls (e.g. WC21, SC10C, LA4, DCC). Those effects include transient or persistent increases in EC, increases (or decreases) in pH and increases in dissolved metal concentrations such as Fe, Mn, Al and Zn. Iron staining in creek beds is commonly associated with watercourses that have been directly mined beneath or are within the mining area of influence. Over the last two years, new or recurrent iron staining has been noted on Wongawilli Creek (see Section 4.5), WC21, LA5 and SC10C. Iron staining has also been observed in natural catchments on the Woronora Plateau that are located outside mining influence.

Table 10. Summary of surface water quality observations and trends

| Catchment | Field parameters (EC, pH and DO) | Dissolved metals |
|------------------------------|---|---|
| Wongawilli Creek | <p>Downstream (WC_FR6): Baseline (TARP) mean EC = 99 μS/cm, pH = 6.0; DO = 89.5%. No adverse trends during reporting period compared with upstream control (WWU1).</p> <p>WC_FR6 TARP: None in review period</p> <p>Iron staining: Iron staining reported between Pool 50 and RB12 during Longwall 17 (TARP Level 3). During Longwall 18, iron staining extended to the confluence with Cordeaux River. Iron not elevated at WC_FR6.</p> | <p>Downstream (WC_FR6): Baseline median Fe = 0.28, Mn = 0.04, Al = 0.04, Zn = 0.004 mg/L</p> <p>WC_FR6: No adverse trends</p> <p>WC_Pool38,49: Dissolved Fe, Mn declined from peak in 2021 (associated with iron staining event), but still slightly elevated (See Section 4.5)</p> |
| Wongawilli Creek tributaries | <p>WC7: No adverse trends</p> <p>WC12: No adverse trends</p> <p>WC15: No adverse trends</p> <p>WC21: Increase in EC and increase in pH at Pool 5 after Longwall 10. Fracturing / Loss of flow upstream.</p> | <p>WC7: No adverse trends (decline in metals)</p> <p>WC12: No adverse trends</p> <p>WC15: Increasing Zn trend in Pools 2 and 9</p> <p>WC21_Pool5: Increasing Fe, Mn and SO₄ since early 2020; remains elevated compared with baseline.</p> |
| Donalds Castle Creek | <p>Downstream (DCC_FR6): Baseline (TARP) mean EC = 116 μS/cm, pH = 5.4; DO = 85.6%. Decline in EC and pH and decrease in DO from 2020. No significant trends further downstream (at DCL3).</p> <p>Upstream at DC13_Pool2B return to baseline EC, pH and DO after elevated EC and declined in pH, DO from 2018-2022.</p> <p>DCC_FR6 TARP: None in review period</p> | <p>DCC_FR6: Increase in Sulfate, Zn, Al and Mn after Longwall 14; Trends in EC, pH and Zn evident (Section 4.5). Trends not evident further downstream at DCL3.</p> <p>Upstream sites: DC13_Pool2B and DC_Pool22; Transient increases in Fe, Mn, Al, Zn after Longwall 13; Declined to near baseline levels from 2020.</p> |
| Lake Avon tributaries | <p>LA4 Baseline (TARP) mean EC = 91 μS/cm, pH = 5.4; DO = 89.9%. Fracturing / loss of flow after Longwall 13; EC slightly higher and pH, DO lower than baseline since flow returned in 2020. Latest EC and pH within baseline range. LA4_S1 TARP: Exceeding Predictions for EC & pH; Level 1 for DO.</p> <p>LA2: No adverse trends.</p> <p>LA3: iron staining observed following completion of Longwall 18. Field parameters returned to baseline levels following 2017-19.</p> | <p>LA4: Dissolved Fe, Mn, Al, Zn and Si remain elevated above baseline after flow returned in 2020.</p> <p>LA2: No adverse trends.</p> <p>LA3: No adverse trends; Metal concentrations have declined since 2019.</p> |
| | Native Dog Creek NDT1 (Pools 2, 23): No adverse trends | ND1_Pool2: No adverse trends |
| Avon River | No adverse trends | Decrease in dissolved metals Fe, Mn, Al and Zn since 2018 (AR19, AR32); Apparent increase in Fe, Al at AR_S1 and AR_S2 (2020-21). |
| Sandy Creek | SCK_Rockbar5: Pre-3A Baseline mean EC = 91 μ S/cm, pH = 5.1; DO = 79.2%. Slight increase in EC from 2017; no other adverse trends. | SCK_Rockbar5: Increase in Fe, Mn from 2020 (to ~2.0 and 0.8 mg/L); small increase in Zn from 2016 (to ~0.05 mg/L). |

| | | |
|----------------|---|--|
| | <p>SC10: Increase in EC from 2017; increase in pH to ~6.5 from 2020. EC declined to within baseline range during period; No adverse trend in DO.</p> <p>SC10C: Increase in EC (to ~260 µS/cm), decrease in pH (to ~3.8) and DO in Pool 1 after Longwall 8 mined under tributary. pH returned to ~6.2 from 2019. DO remains slightly below baseline; EC continues to decline but remains above baseline. See Section 4.6.</p> <p>SCK_Rockbar5 TARP: None in review period</p> | <p>SC10_Rockbar3: Small increase in Fe, Mn from 2019; small increase in Zn from 2016. Most recent samples within baseline range.</p> <p>SC10C_Pool1: Increase in Fe, Mn, Al, Zn, Si and sulfate following Longwall 8. Declining trends since 2020.</p> |
| Cordeaux River | No Adverse trends | CR_S1 and CR_S2: Slight increase in Fe, Mn, Al and Si from 2020-22. |
| Reservoirs | <p>Lake Avon (LA5_S2): No adverse trends.</p> <p>Lake Cordeaux (SANDY CREEK ARM): No adverse trends</p> <p>Lake Cordeaux (Sandy Creek Arm) TARP: None in review period</p> | <p>Lake Avon (LA5_S2): No adverse trends.</p> <p>SANDY_CREEK_ARM: Small spike in concentrations of Fe and Mn associated with 2017-2019 drought. No other adverse trends.</p> |

4.2 Quantitative assessment of water quality trends

WaterNSW endorsed the recommendation of the Independent Advisory Panel for Underground Mining (IAPUM) that “A method of quantifying and reporting trends in key water quality indicators (both concentrations and loads) should be trialled in addition to applying the proposed water quality TARPs.”

A methodology for trend analysis was developed and trialled at two monitoring locations on Wongawilli Creek and Donalds Castle Creek (HGEO, 2021b). The methodology was reviewed and considered appropriate by WaterNSW. Trend analysis is carried out as follows:

- Generate a flow-corrected residual timeseries by applying the LOWESS smoother to concentration versus stream discharge data.
- For each specified time period, calculate the Mann-Kendall test statistic for significance at the 5% significance level; the Theil-Sen slope; and compare the mean concentration during the period with the baseline period using the non-parametric Mann-Whitney U rank sum test statistic.
- Trend analysis should be carried out on field EC, pH and DO, and sulphate, dissolved Fe, Mn, Al, Zn.
- Tabulate and discuss significant trends, including comparison with control site(s). Trend analysis should be carried out for monitoring sites with associated flow gauges on the major 3rd order streams WC_FR6, DCC_FR6), SCK_Rockbar5 and an appropriate control site (O’Hares Creek, or WWU4).

4.2.1 Trend analysis results

Flow-corrected water quality time series and tabulated results are included in Appendix A2. A summary table, highlighting results of statistical significance is provided in Table 11. The trend analysis results reflect the qualitative assessment presented in the previous section, with the following being statistically significant:

- At WC_FR6, trends in EC and sulfate, iron and manganese are identified in flow-corrected data. Except for sulfate, the mean flow-corrected concentrations are not significantly higher than the baseline. With reference to the time-series plots in Appendix A it is noted that similar trends in sulfate are apparent in the upstream control site (WWU4).
- At DCC_FR6, the most recent 1-year and 3-year mean pH is lower than the baseline mean (at the 95% level). A step-change is apparent after mid-2019 in flow-corrected time-series (Figure 9) and in the non-corrected time-series plots in Appendix A (pH, SO₄, Mn, Al, Zn and Si); but with no

significant trend over the last 3 years according to the Mann-Kendal test. Dissolved zinc shows an increasing trend over the last 1 and 3 years, with the 1- and 3-year means higher than the baseline (Figure 10). A similar trend is apparent at the downstream location DCL3 for dissolved Zn, but not pH. There is an apparent 1-year increasing EC trend; however the mean flow-corrected EC remains below the baseline mean.

- At SCK_Rockbar5, there are no significant trends identified over the last 1- and 3-year periods. However, with reference to flow-weighted timeseries in Appendix A, statistically significant trends are apparent over the entire post-baseline period (> 9/2/2010) for sulfate and dissolved Fe, Mn and Zn. There is insufficient pre-baseline flow data at the SCL2 gauge site, however if flow data from the upstream site SC10S1 is used, it is apparent that dissolved metals Fe, Mn and Zn and sulfate are above the baseline (flow-corrected). These trends reflect contributions from tributary SC10C which was mined under by Longwall 8.

Table 11. Summary of flow-corrected water quality trends

| Parameter | WC_FR6 | DCC_FR6 | SCK_Rockbar 5 | WWU4 (Control) |
|----------------------|----------------------|-----------------------------------|----------------------|----------------|
| EC (uS/cm) | ↗ (3 yr) ■ | ↗ (1 yr) ■ | →■ | →■ |
| pH (field) | →■ | ▼▼ (1 yr, 3 yr) | →■ | →■ |
| DO (%) | →■ | →■ | →■ | →■ |
| SO4 mg/L) | ↗ (3 yr) ▲ (1 yr) | →■ | → ▲ (3 yr) | ↗ (3 yr) ■ |
| Fe (Dissolved, mg/L) | ↗ (1 yr) ■ | →■ | → ▲ (1 yr) | ↗ (3 yr) ■ |
| Mn (Dissolved, mg/L) | ↗ (3 yr) ■ | →■ | → ▲ (3 yr) | →■ |
| Zn (Dissolved, mg/L) | →■ | ↗↗ (1 yr, 3 yr) ▲▲ (1 yr, 3yr) | → ▲▲ (1 yr, 3 yr) | →■ |
| Al (Dissolved, mg/L) | ↗ (1 yr) ■ | →■ | →■ | →■ |

Note: Mann-Kendall nonparametric test for ordinal trend (flow-corrected concentration); ↗ Increasing trend; ↘ Decreasing trend (pH, DO) → No significant adverse trend; Mann-Whitney U test for difference in means; ▲ Above baseline mean; ▼ Below baseline mean (pH, DO); ■ No significant adverse change in mean (All at 95% significance Level).

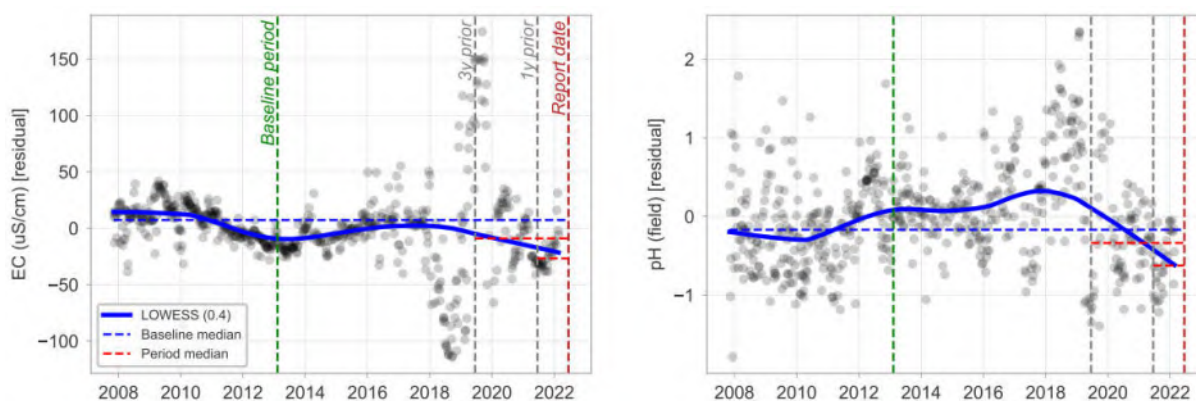


Figure 9. Flow-corrected time-series plots of EC and pH at DCC_FR6

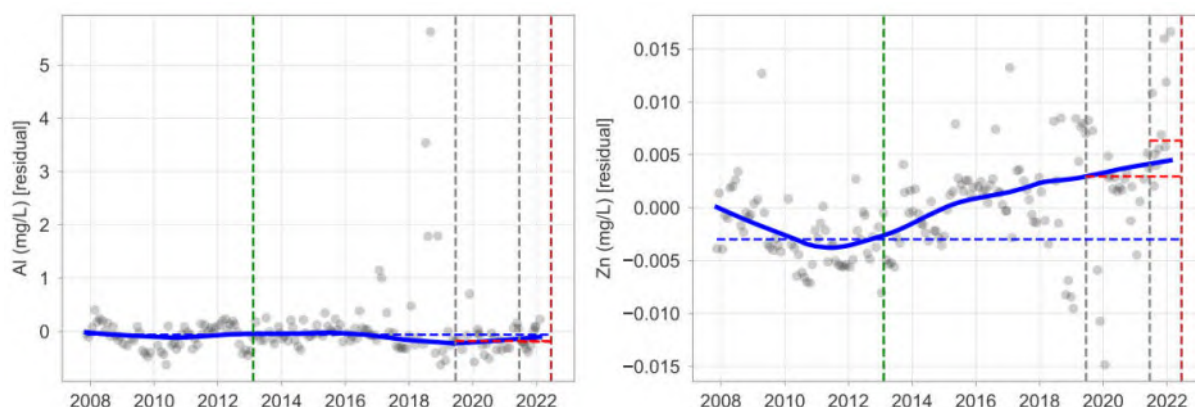


Figure 10. Flow-corrected time-series plots of Al and Zn at DCC_FR6

The plots in Figure 10 and Figure 10 also show how trend analysis can produce apparently conflicting or non-intuitive results, depending on the technique and time-period chosen (e.g. EC at DCC_FR6). The Mann-Kendall trend analysis identifies trends only over the specified 1 year and 3 year periods, whereas the Locally weighted regression trend (LOWESS – blue line) smooths over short- and medium-term fluctuations and reflects the longer-term trends. Results should therefore be interpreted with reference to both the flow-corrected and non-corrected hydrographs.

4.3 Catchments within Longwall 18 area of influence

Native Dog Creek Tributary 1 (ND1). Longwall 18 Commenced at a distance of 215 m from flow gauge NDT1 and 265 m from the water chemistry site ND1_Pool2. The longwall mined beneath the upper reaches of ND1C in early April 2022. There are no significant adverse trends in field parameters or dissolved metals at ND1_Pool2 during the review period. pH was below the baseline mean in early 2022 but remained within the historical range.

Longwall 18 overlapped the southern 20% of the LA2 catchment. The watercourse itself was mined beneath previously by Longwall 17 between 13/12/2020 and 16/3/2021. No adverse trends in field parameters or dissolved metals are noted.

4.4 TARP triggers at tributary LA4

LA4 is a tributary to Lake Avon. Its catchment overlaps with Longwalls 12 and 13 in Area 3B. Longwall 12 passed beneath the upper reaches of LA4B towards the end of April 2016. Longwalls 13 and 14 commenced at distances of 275 m and 295 m from LA4 monitoring site LA4_S1 on 4/3/2017 and 27/5/2018, respectively.

Impacts to the LA4 watercourse, including fracturing of the creek bed and diversion of flow were reported in End of Panel Reports for Longwalls 12, 13 and 15. At the gauging site LA4_S1, pooling was insufficient to collect samples between mid-2017 and mid-2020 due to a combination of mining related flow-diversions and low rainfall conditions. Sampling of the site resumed in mid-2020 following higher rainfall and more frequent ponding and flow since that time. Water quality TARP triggers were recorded on 17/12/2021 (EC, pH and DO) and 9/2/2022 (EC, pH) which reflects sustained change in water quality since the return of flow to the site. Two or more consecutive exceedances of the TARP triggers results in a TARP level of Exceeding Predictions for EC and pH.

4.5 Iron-staining in Wongawilli Creek

In August 2021, an increase in iron staining was observed along reaches of Wongawilli Creek adjacent to Areas 3A and 3B during routine monitoring. The most noticeable iron-staining effects extended over a 2.9 km reach from WC_Pool 50 to WC_Rockbar 12. The source of the iron staining was identified as a (pre-existing but reactivated) spring located on the valley slope of Wongawilli Creek, approximately 35m to the east and upslope from WC_Pool 50. The observations corresponded to a Level 3 trigger according to the Dendrobium Area 3B Watercourse Impact Monitoring, Management and Contingency Plan (WIMMCP) and were reported by IMC on 9/8/2021.

An assessment was carried out to determine the root cause of the impact and recommend appropriate management responses (HGEO, 2021a). The assessment included review of surface and groundwater monitoring and analysis of additional samples of creek and spring water and iron deposits (Figure 11). The assessment concluded that the appearance of iron staining along Wongawilli Creek in August 2021 was caused by increasing groundwater levels and reactivation of slope springs in response to high rainfall and groundwater recharge events in March and May 2021 (and generally high rainfall since 2020). Evidence of groundwater recovery adjacent to Wongawilli Creek is shown in Figure 28. The groundwater discharge is naturally elevated in dissolved iron, resulting in visible iron staining and accumulation of iron floc (iron oxyhydroxides and iron bacteria) on exposure to air or on mixing with oxygenated water. Spring discharge is likely facilitated by fracturing associated with mine subsidence; however discharge via natural fractures is also possible (and was observed during baseline monitoring). It should also be noted that increased Iron staining has been observed in natural catchments on the Woronora Plateau that are located outside mining influence. It is expected that spring flow and iron staining will continue to decline over time as groundwater levels stabilise or decrease, and as the fractures weather and age.

During 2022 visible iron staining was observed by the IMCEFT to extend further downstream on Wongawilli Creek as far as Pool 2 (immediately upstream of the confluence with Cordeaux River). No iron staining is visible in Cordeaux River. With reference to Figure 11 and as noted in Sections 4.1 and 4.2, dissolved iron concentrations in Pools 50 and 38 have declined from their peak in August 2021 when the iron staining appeared, but remain above baseline levels. Dissolved iron at Pool 20, WC_FR6 and locations downstream remain within the baseline range. The movement of iron staining downstream of WC_FR6 is therefore due to disaggregation and transport of iron floc downstream during flood events rather than an increase in dissolved iron concentrations in the watercourse.

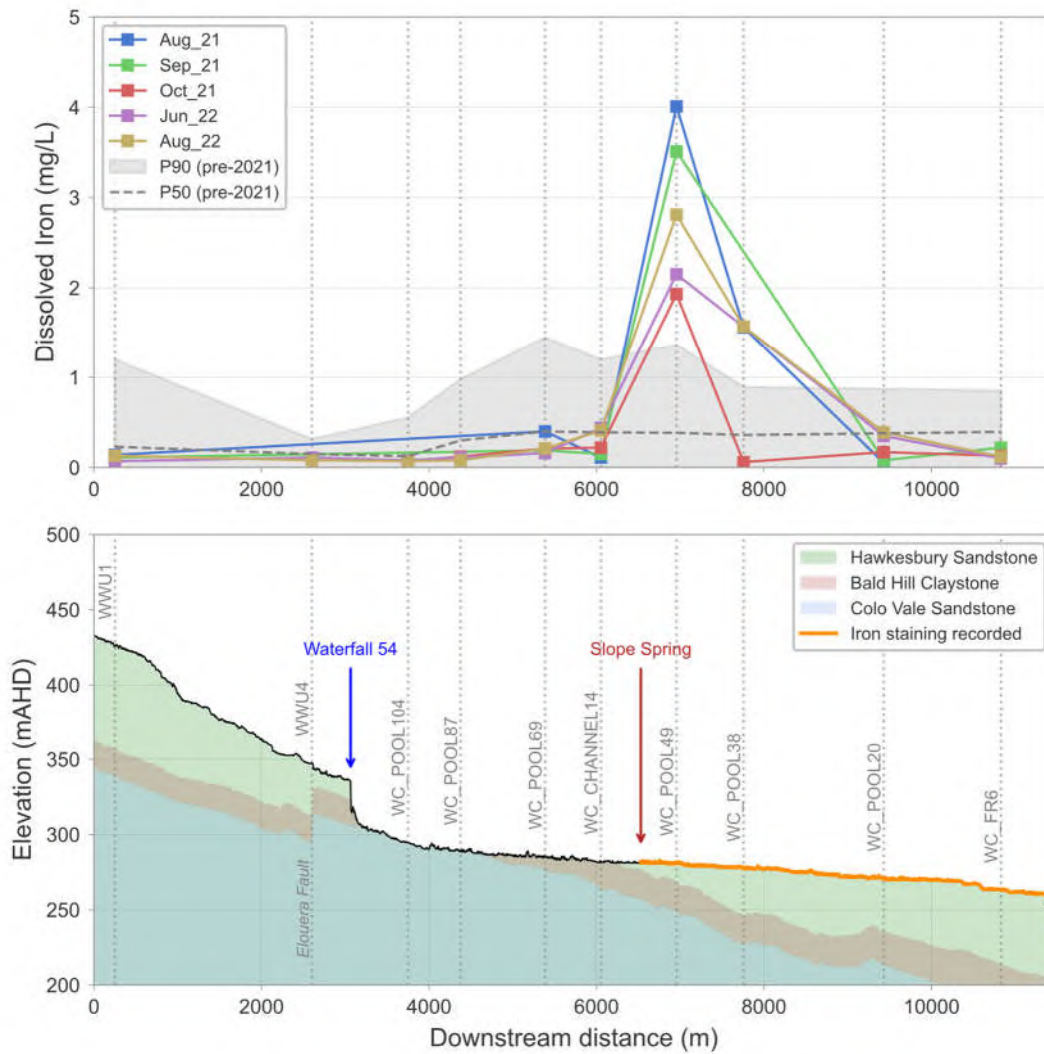


Figure 11. Dissolved iron concentration in Wongawilli Creek

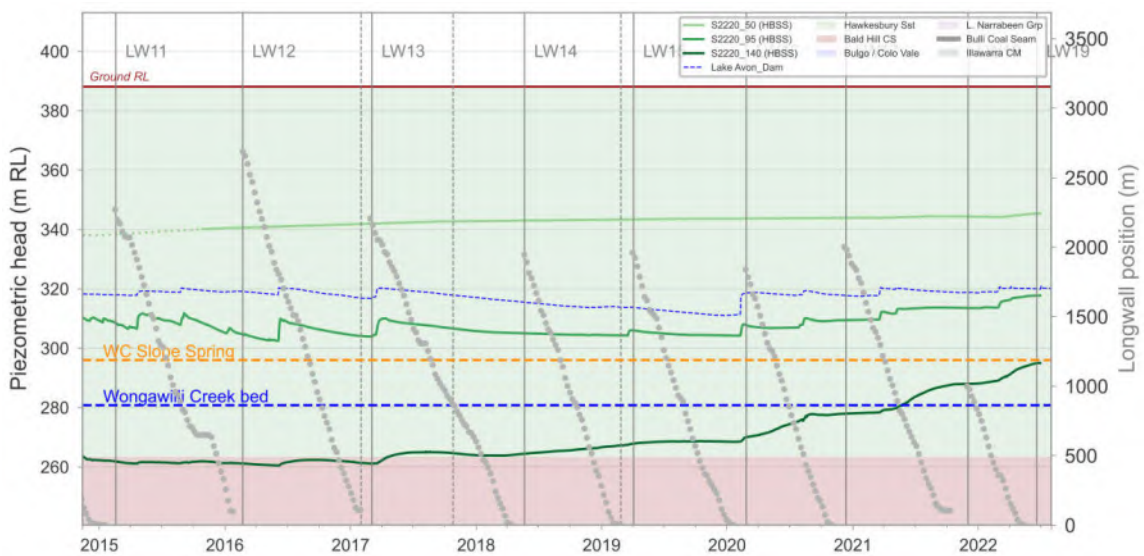


Figure 12. Groundwater hydrograph at monitoring bore S2220 (total piezometric head)

4.6 Sandy Creek

In May 2022, WaterNSW (via DPE) requested that IMC *Investigate a potential link between anomalous water quality results observed in the tributary of Banksia Ck (SC10C_Pool1) and increasing 3-year trends in sulphate and dissolved metal (manganese, zinc) concentrations at Sandy Creek (SCK_Rockbar5).*

In response to this request, IMC initiated a longitudinal study of water quality from SC10C_Pool 3 to Sandy Creek (SCK_ROCKBAR5), including additional sample locations and upstream control sites on SC10_Pool11 and Sck_Pool7. The sites will be sampled monthly and analysed for major ions and dissolved metals. Findings will be reported in a separate report after 6 months of sampling has been completed.

Based on the existing monitoring data as described in Section 4.1, water quality in SC10C was affected after Longwall 8 mined beneath the watercourse (Longwall 8 was completed on 29/12/2012). At SC10C_Pool1, EC increased to ~260 $\mu\text{S}/\text{cm}$ and there was a corresponding decrease in pH (to ~3.8). DO declined slightly from about 2015. Since 2020, EC has trended downwards, and pH has increased to above baseline values (average > 6.0). DO remains slightly below the baseline. Dissolved metals Fe and Mn trended higher following Longwall 8, peaking between 2017 and 2020 at ~25 mg/L and 4.5 mg/L respectively. Sulfate and silicon show similar trends. Dissolved Al and Zn spiked several times since 2014, with peaks diminishing over time. All dissolved metals have trended lower since 2019-2020; however, Sulfate, Fe and Mn remain above the baseline.

In Sandy Creek, downstream of the confluence with SC10 (SCK_Rockbar5), EC trended higher during the 2017-2019 drought and has since returned to within the baseline range. Water pH has trended slightly higher since 2020 (~ pH 6.0) and DO shows no significant change. Dissolved metals Fe, Mn and Zn have trended higher since approximately 2016, and have remained elevated since 2020 (above baseline).

The increase in dissolved metals at SCK_Rockbar5 would appear to be related to upstream inputs from tributary SC10C; however, further analysis will be carried out when results of the longitudinal study become available.

4.6.1 Iron staining

Iron staining was first reported in SC10C (Pool 3) on 11 March 2013 after Longwall 8 mined beneath the watercourse (Impact reference LW8_158). The iron staining corresponds to the first detection of high dissolved Fe (13/3/2013; 15.6 mg/L) and followed two months of high rainfall. Iron staining was also observed downstream of the SC10 and SC10C confluence following the extraction of Longwall 8.

Iron staining was again observed at SC10C_Pool3 on 3/9/2020, following high rainfall in 2020 and reported by IMC as an update to impact LW8_158 (Report date 19/10/2020). During the 2020 event, the staining was observed to have extended downstream into Sandy Creek and to SCK_Rockbar5. No iron staining was observed within Lake Cordeaux (Sandy Creek Arm). The iron staining triggered (and remains) a Level 2 TARP.

5. Assessment of surface water flow effects

5.1 Surface Water Flow TARPs

The surface water flow assessment and relevant TARPs have been modified from those used previously for Area 3B End of Panel reports. Consultation with agencies during 2018-2019 led to final agreement in early 2020 of new TARPs, as outlined in the WIMMCP (South32, 2020a).

This assessment of surface water flow in this End of Panel report relies on comparison against flows at Reference Sites, as recommended by the IEPMC (IEPMC, 2019, 2018). The revised and agreed assessment methods are described in more detail in (Watershed Hydrogeo, 2019a). Other recommendations of the IEPMC are addressed in this assessment, as listed in Table 12.

Table 12. Recommendations of the IEPMC (2018) (revised as IEPMC, 2019a)

| RECOMMENDATION / COMMENT | RESPONSE / ACTION |
|---|--|
| Assessment of impacts to be made against the full post-mining period, not longwall by longwall. | Implemented. Assessment of effects is now reported for the complete post-mining period at each site, rather than for each longwall. This provides an assessment of cumulative effects. |
| The EOP reports to provide more information on the data sources for rainfall, evaporation and the monitoring sites. | Implemented. This is presented Appendix B and C, noting that rainfall and evaporation are not required for assessment against Reference Sites. |
| Document the specific sources of rainfall and evaporation data used in the rainfall-runoff modelling. | Implemented. The rainfall and evaporation data sources are documented in Appendix B, noting that rainfall and evaporation are not required for assessment against Reference Sites. |
| Discussion of flow monitoring errors and their impact on assessing compliance should be published and peer reviewed. | Implemented (On-going). IMC are progressing with a review of gauging station accuracy. This report has been issued (EnviroMon, 2019) but will be extended in the near future to include other sites. |
| Use techniques to supplement the rainfall-runoff modelling. This has been done in some EOP reports, including for LW11 but has been excluded from the LW12 and LW13 EOP reports | Implemented. The use of Reference Sites, as documented in WatershedHG (2019a) and agreed by agencies, is now adopted for this assessment. |
| There is no validation on flow measurements from outside the calibration period. | Not yet implemented: The use of Reference Sites as the basis for assessment was agreed by agencies in the approved WIMCCP. More recent feedback by IAPUM recommended that comparison against rainfall-runoff modelling should be re-instated (Section 1.3). This has been done so for a limited number of sites (Section 5.3.1), and if deemed necessary can be expanded in future. |
| Given the criticality of low flows for this project, attempts to improve the low flow modelling should continue, and should be reported and peer reviewed. | Not yet implemented: As above. Peer review to occur early 2022. |

The agreed Assessments A, B, and C are respectively focussed on assessing:

- general hydrological behaviour compared with Reference Sites,
- the frequency and duration of ecologically-significant cease-to-flow events compared with Reference Sites; and
- changes to median flow compared with Reference Sites which is now the agreed measure of the water resource availability in each sub-catchment.

A further assessment, Assessment D is specific to Wongawilli Creek, relies on comparison of qualitative flow data from gauging stations and semi-quantitative field observations by IMCEFT along the “middle reach” of Wongawilli Creek, which has been shown in the recent past (e.g. in Watershed HydroGeo, (2018)) to be subject to baseflow loss due to depressurisation of groundwater systems as a result of mining activity.

5.2 Performance Measures

Performance Measures have also been agreed and are documented in the WIMMCP. These are outlined in Table 13. The assessment of these is presented in Section 5.4.

Table 13. Area 3B Surface flow Performance Measures

| PERFORMANCE MEASURE | AGREED MEASURE |
|---|---|
| Wongawilli Creek – minor environmental consequences | Assessment Methods C and D, to be compared against predictions made in contemporary groundwater modelling conducted to the satisfaction of the Secretary to assess whether effects that cannot be explained by natural variability “exceed prediction”. |
| Donalds Castle Creek – minor environmental consequences | Assessment Method C to be compared against predictions made in contemporary groundwater modelling conducted to the satisfaction of the Secretary to assess whether effects that cannot be explained by natural variability “exceed prediction”. |
| Lake Avon – negligible reduction in the quantity of surface water inflows to Lake Avon | Surface water inflows calculation = [Impacts at gauged catchments (LA2 + LA3 + LA4 + NDT1) + estimated impacts at ungauged but undermined catchments (e.g. LA5)] / [total inflow to LA]. |
| Cordeaux River – negligible reduction in the quantity of surface water inflow to the Cordeaux River at its confluence with Wongawilli Creek | Flow reduction as determined from measured flow gauging station WWL_A (or WWL, whichever gauge is being used). |

5.3 Assessment for Longwall 18

The following sections present the analysis and results of the agreed Assessments A, B, and C for each sub-catchment relevant to Area 3B. This is followed by Assessment D for the mid-reach of Wongawilli Creek and then followed by assessment against the agreed Performance Measures. The detail and criteria for each assessment are outlined in the WIMMCP (South32, 2020a), as described in Section 5.1.

At the gauged sub-catchments around Area 3B, the assessment consists of a three-step approach (A, B, and C as listed below) to identify and assess any changes in hydrology at the assessment sites in relation to the agreed reference sites. A fourth assessment (D) is carried out for Wongawilli Creek.

The four assessment methods are as follows:

- Change in flow exceedance (“Q%ile”) behaviour compared to Reference Sites. In essence, this aims to quantify an otherwise visual or qualitative assessment of flow behaviour (compared to normalised Reference Site flow). This test is considered a broad indicator of hydrological behaviour.

| Greater proportion of time with lower flow than expected based on Reference Q% | Trigger level (Inference) |
|--|---|
| Proportion of time increased by < 10% | Not triggered – no evidence of impact (or impact below detection) |
| Proportion of time increased by >= 10% | Level 1 |
| Proportion of time increased by >= 15% | Level 2 |
| Proportion of time increased by >= 20% | Level 3 |

- Relative change in the frequency of cease-to-flow days compared to that at Reference Sites. This assessment is focussed on changes that are likely to be significant to ecological values.

| Greater proportion (%) of time that cease-to-flow conditions occur | Inference |
|--|---|
| <= "natural variability" + 5% | No evidence of impact (or impact below detection) |
| > "natural variability" + 5% | Level 1 |
| > "natural variability" + 10% | Level 2 |
| > "natural variability" + 20% | Level 3 |

- Relative change in median flow (“Q50”) compared to Reference Site flows. This assessment is focussed on a measure of the water resource potential of each sub-catchment, noting that ‘average’ flow is not used due to the high uncertainty associated with high flows. The uncertainty is typically less at moderate flows (EnviroMon, 2020) – see charts in Appendix C5, and the calculation of median flow is much less sensitive to uncertainties; and

| Relative change in Q50 | Inference |
|--------------------------------|---|
| <= "natural variability" + 10% | No evidence of impact (or impact below detection) |
| > "natural variability" + 10% | Level 1 |
| > "natural variability" + 15% | Level 2 |
| > "natural variability" + 20% | Level 3 |

Note that this is calculated as a % reduction compared to measured pre-mining Q50 at the assessment site. It is proposed that this be changed to % reduction from ‘expected Q50’.

- Assess whether observed dry pools and ‘cease-to-flow’ conditions along Wongawilli Creek between WWU and WWL gauging stations are anomalies, and indicative of mining-related drawdown along that valley (as described in Watershed HydroGeo, 2018).

| Observations of no flow | Inference |
|---|-----------------------------------|
| Observation that the subject Creek has ceased to flow at spatially consecutive observation sites. | Level 2 → Carry out Assessment D. |

If any of these indicate an impact is likely to have occurred, then the EOP will describe the Impact as it relates to one or more of the broad hydrological behaviours, a reduction in the water resource Indicator, or an effect that could modify or impact upon the ecological values of the stream.

5.4 Assessment against surface water flow TARPs

The following sub-sections (Sections 5.4.1 to 5.4.11) summarise the TARP Assessments A, B and C for each relevant sub-catchment using the criteria outlined in the previous section. A secondary check for sites that do not Trigger Level 3 for Assessment C is presented in Section 5.4.16.

TARP Assessment D for flow conditions along Wongawilli Creek is presented in Section 5.5.

5.4.1 DC13S1 – tributary of Donalds Castle Creek

This tributary lies across the centre of several Area 3B panels. The catchment to DC13S1 was first mined under at the commencement of Longwall 9, and again by Longwalls 10 and 11. Longwalls 12 to 18 did not directly mine under this sub-catchment.

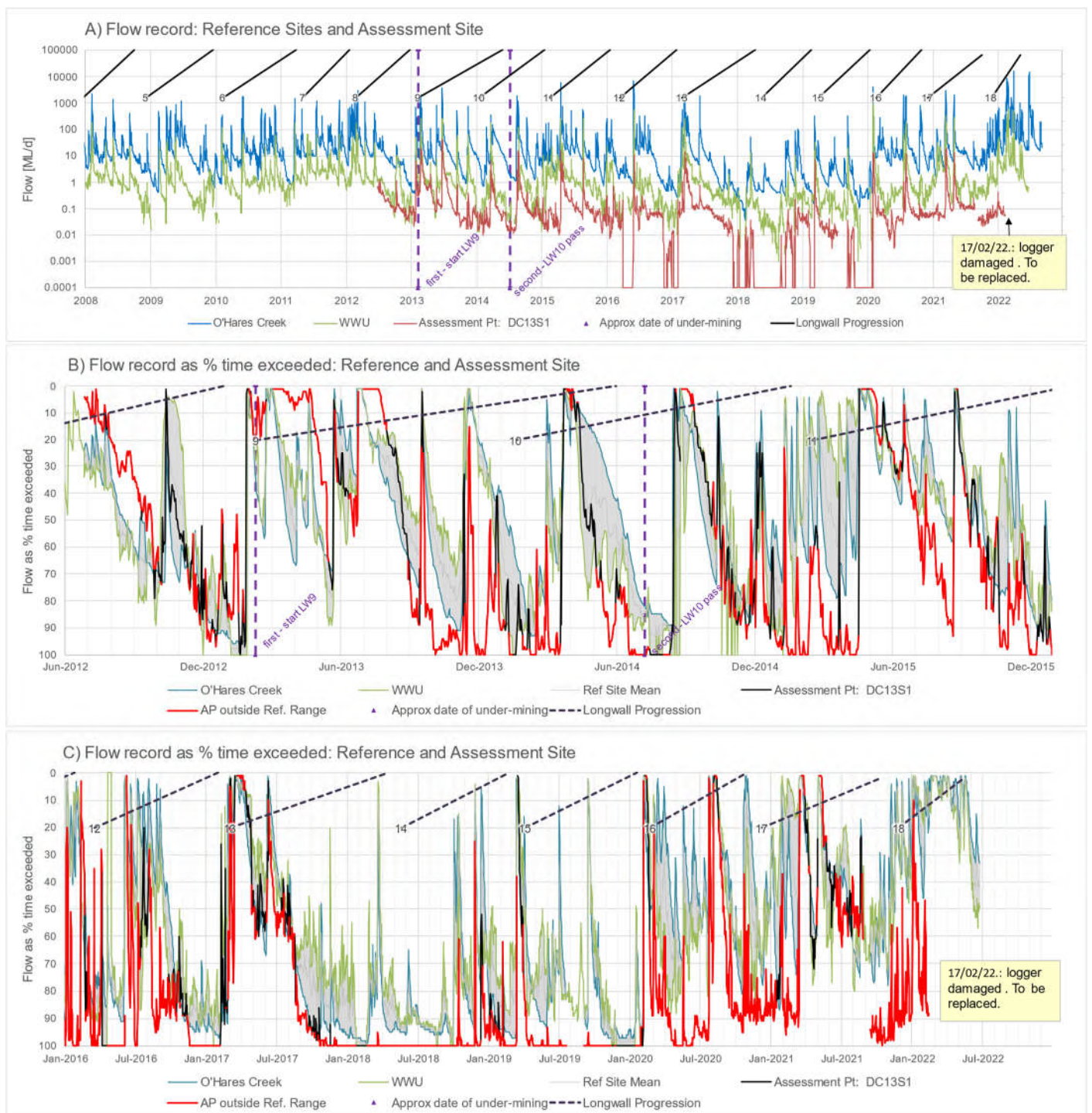


Figure 13. DC13S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the DC13 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 14. Flow assessments A, B and C for the sub-catchment to DC13S1

| DC13S1 | Pre-mining | Post-mining | |
|---|---|---------------------------|----------------|
| | to start LW9 | end LW18 + 30days | |
| | 27/06/2012 | 10/02/2013 | |
| | 9/02/2013 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 14% | 59% | of the time |
| Post-mining | 71% | 10% | of the time |
| Change | 60% | -49% | of the time |
| Assessment A: | | | Level 3 |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 10.1% | 4.6% | -5.5% |
| Average Ref. Site change (= natural variability): | | | -2.7% |
| DC13S1 | 0.0% | 14.5% | 14.5% |
| no. of cease-to-flow days increased: | | | 17.2% |
| Assessment B: | | | Level 2 |
| Method C: Change to median flow (Q50): | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 5.46 | 7.16 | 31% |
| WWU | 0.40 | 0.44 | 10% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | 10% | 20% | 31% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| DC13S1 | 0.126 | 0.055 | -56% |
| 'Expected' post-mining Q50 at DC13s1 | Min | Mean | Max |
| | 0.138 | 0.151 | 0.164 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | -65.7% | -76.5% | -87.3% |
| % change (of 'expected' Q50) | -60% | -64% | -67% |
| ML/d change from natural | -0.083 | -0.096 | -0.109 |
| Assessment C: | | | Level 3 |

5.4.2 DCS2 – Donalds Castle Creek

The upper reach of Donalds Castle Creek lies across several Area 3B panels. This sub-catchment was first mined under by Longwall 9 (July 2013), then by Longwalls 10-12. Longwall 13 passed within 250 m of the creek in May-2017. Longwalls 14-18 have not mined directly under this catchment.

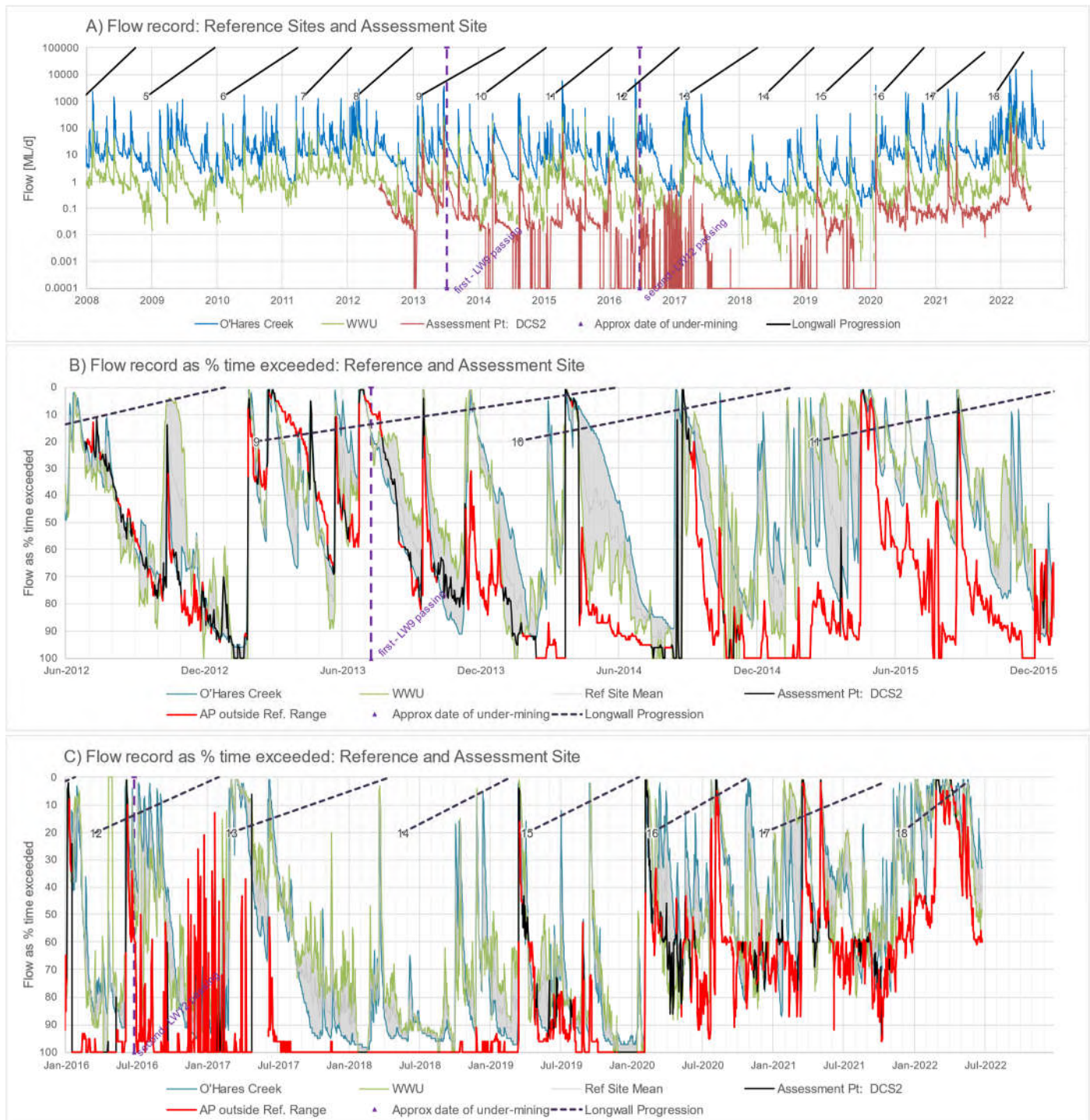


Figure 14. DCS2 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the DCS2 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 15. Flow assessments A, B and C for the sub-catchment to DCS2

| DCS2 | Pre-mining | Post-mining | |
|---|---|---------------------------|----------------|
| | to LW9 passing | end LW18 + 30 days | |
| | 27/06/2012 | 11/07/2013 | |
| | 10/07/2013 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 25% | 35% | of the time |
| Post-mining | 79% | 5% | of the time |
| Change | 54% | -30% | of the time |
| Assessment A: | | | Level 3 |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 6.1% | 4.8% | -1.2% |
| Average Ref. Site change (= natural variability): | | | -0.6% |
| DCS2 | 2.9% | 35.5% | 32.6% |
| no. of cease-to-flow days increased: | | | 33.2% |
| Assessment B: | | | Level 3 |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 8.52 | 7.08 | -17% |
| WWU | 0.82 | 0.42 | -49% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | -49% | -33% | -17% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| DCS2 | 0.164 | 0.024 | -85% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| DCS2 | 0.083 | 0.110 | 0.136 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | -36.0% | -52.0% | -68.5% |
| % change (of 'expected' Q50) | -71% | -78% | -82% |
| ML/d change from natural | -0.059 | -0.086 | -0.112 |
| Assessment C: | | | Level 3 |

5.4.3 DCU – Donalds Castle Creek

This catchment incorporates the headwater sub-catchments DC13 and DCS2 was mined under at the commencement of Longwall 9, and again by Longwalls 10-12, and marginally by Longwall 13. Longwalls 14-18 are beyond it (to the south). About 60% of the DCU catchment is not mined under.

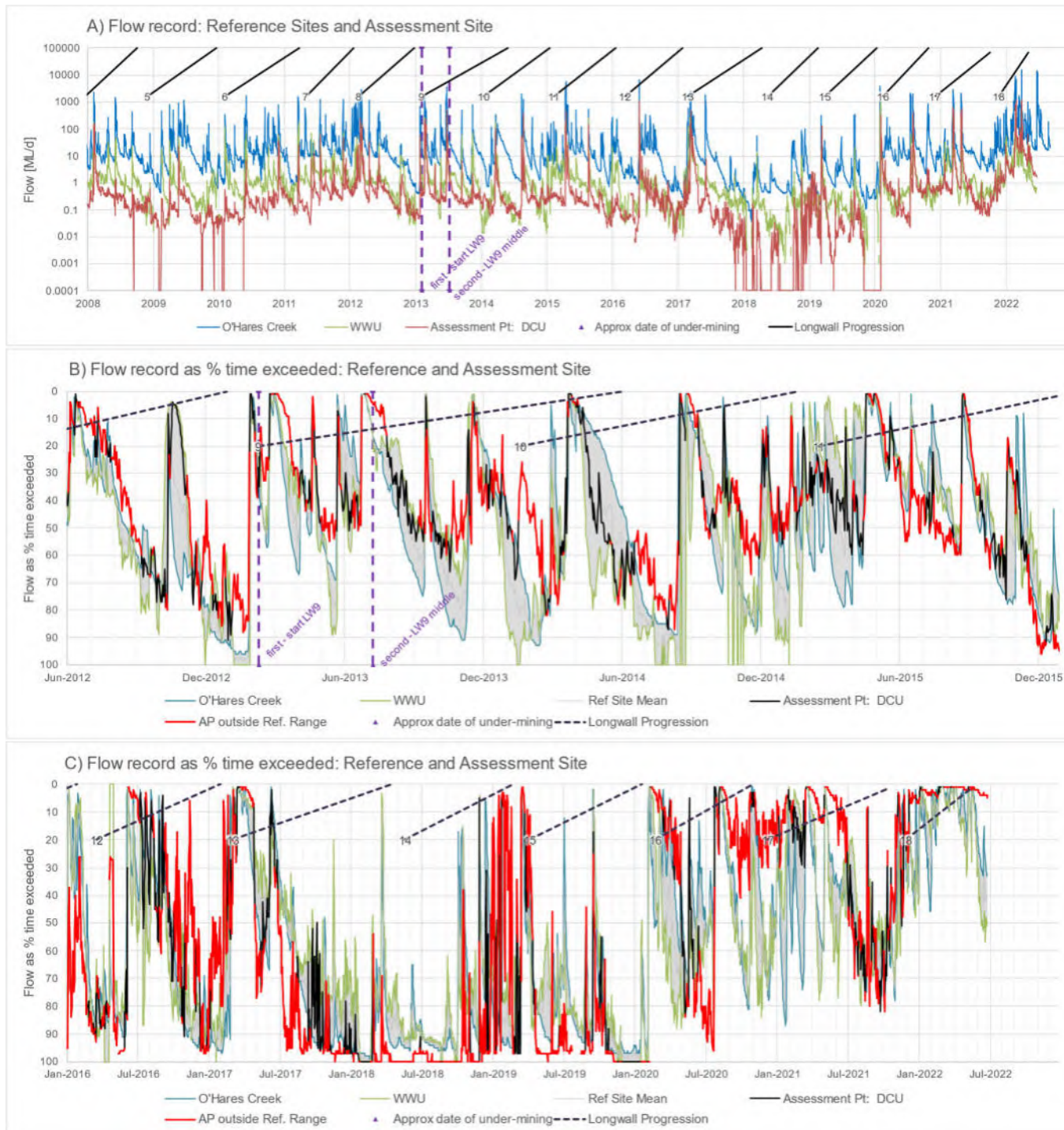


Figure 15. Comparison of DCU against Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the DCU Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 16. Flow assessments A, B and C for the sub-catchment to DCU

| DCU | Pre-mining | Post-mining | |
|---|---|---------------------------|---------------|
| | to LW9 passing | end LW18 + 30 days | |
| | 27/06/2012 | 10/02/2013 | |
| | 9/02/2013 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 45% | 27% | of the time |
| Post-mining | 35% | 38% | of the time |
| Change | -10% | 11% | of the time |
| Assessment A: | | | Not triggered |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 5.2% | 4.6% | -0.6% |
| Average Ref. Site change (= natural variability): | | | -0.3% |
| DCU | 1.8% | 9.2% | 7.4% |
| no. of cease-to-flow days increased: | | | 7.7% |
| Assessment B: | | | Level 1 |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 11.64 | 7.61 | -35% |
| WWU | 1.03 | 0.46 | -55% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | -55% | -45% | -35% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| DCU | 0.217 | 0.204 | -6% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| DCU | 0.097 | 0.120 | 0.142 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | +49.3% | +38.9% | +28.6% |
| % change (of 'expected' Q50) | 110% | 71% | 44% |
| ML/d change from natural | +0.107 | +0.084 | +0.062 |
| Assessment C: | | | Not triggered |

5.4.4 WC12S1 – Wongawilli Creek tributary

The end of Longwall 15 skirted the north-western edge of this sub-catchment and to within 250 m of the watercourse itself. Longwall 16 mined within 40 m of WC12, and Longwall 17 mined under this watercourse. Longwall 18 did not mined under this sub-catchment.

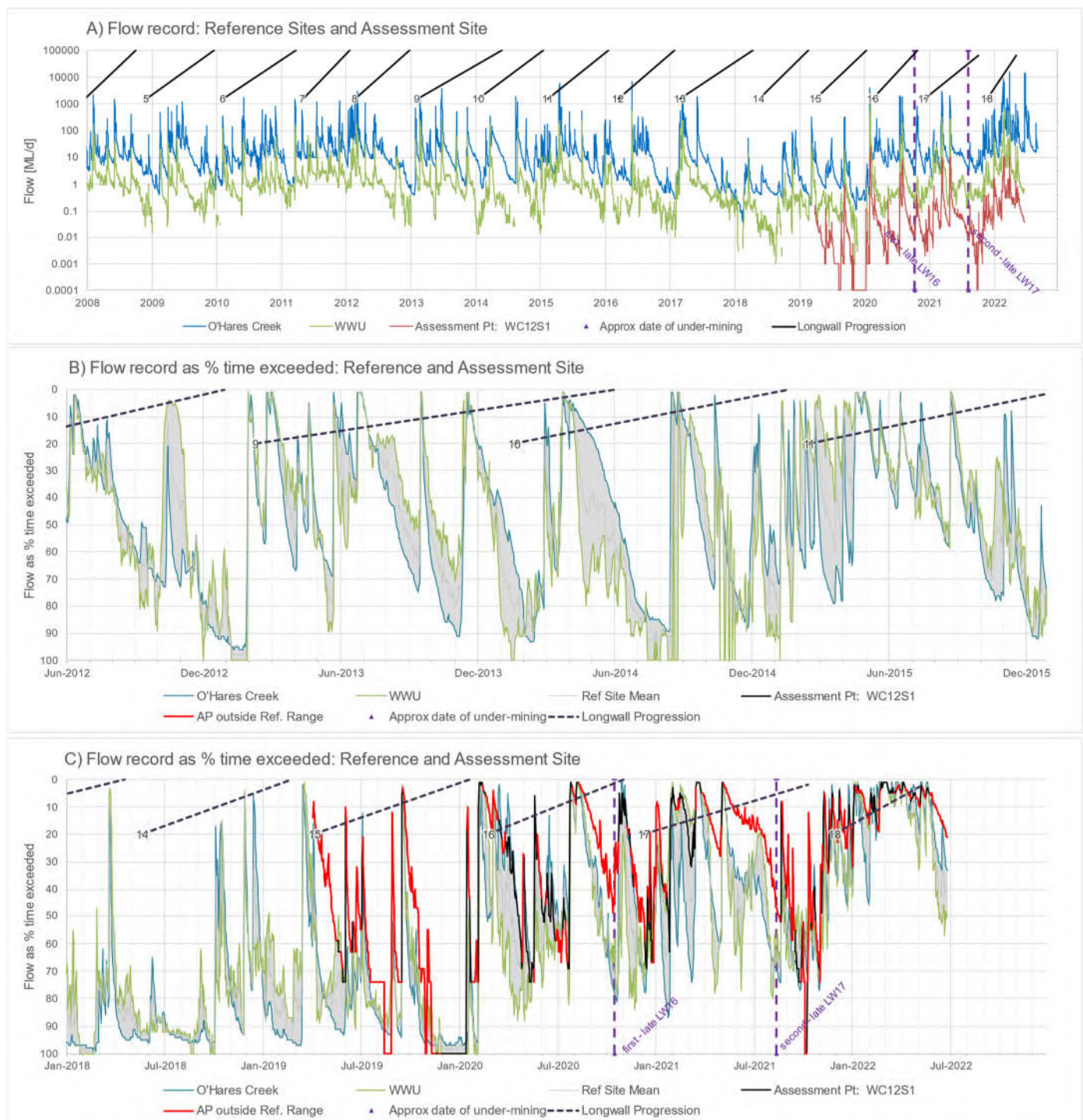


Figure 16. WC12S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the WC12S1 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 17. Flow assessments A, B and C for the sub-catchment to WC12S1

| WC12S1 | Pre-mining | Post-mining | |
|---|---|---------------------------|----------------------|
| | to late LW16 | end LW18 + 30 days | |
| | 5/04/2019 | 19/10/2020 | |
| | 18/10/2020 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 10% | 59% | of the time |
| Post-mining | 6% | 63% | of the time |
| Change | -3% | 4% | of the time |
| Assessment A: | | | Not triggered |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 11.0% | 0.0% | -11.0% |
| Average Ref. Site change (= natural variability): | | | -5.5% |
| WC12S1 | 14.7% | 0.7% | -14.1% |
| no. of cease-to-flow days increased: | | | -8.6% |
| Assessment B: | | | Not triggered |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 5.05 | 16.16 | 220% |
| WWU | 0.23 | 1.19 | 412% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | 220% | 316% | 412% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| WC12S1 | 0.009 | 0.070 | 678% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| WC12S1 | 0.029 | 0.037 | 0.046 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | +457.5% | +361.8% | +266.1% |
| % change (of 'expected' Q50) | 143% | 87% | 52% |
| ML/d change from natural | +0.041 | +0.033 | +0.024 |
| Assessment C: | | | Not triggered |

5.4.5 WC15S1 – Wongawilli Creek tributary

Longwall 12 came within 100 m of the sub-catchment. Longwall 13 mined under this catchment and was within 20 m of this watercourse, and directly under the WC15A tributary). Longwall 14 approached to 40 m. Longwalls 15-17 mined under WC15, while Longwall 18 did not (170 m away).

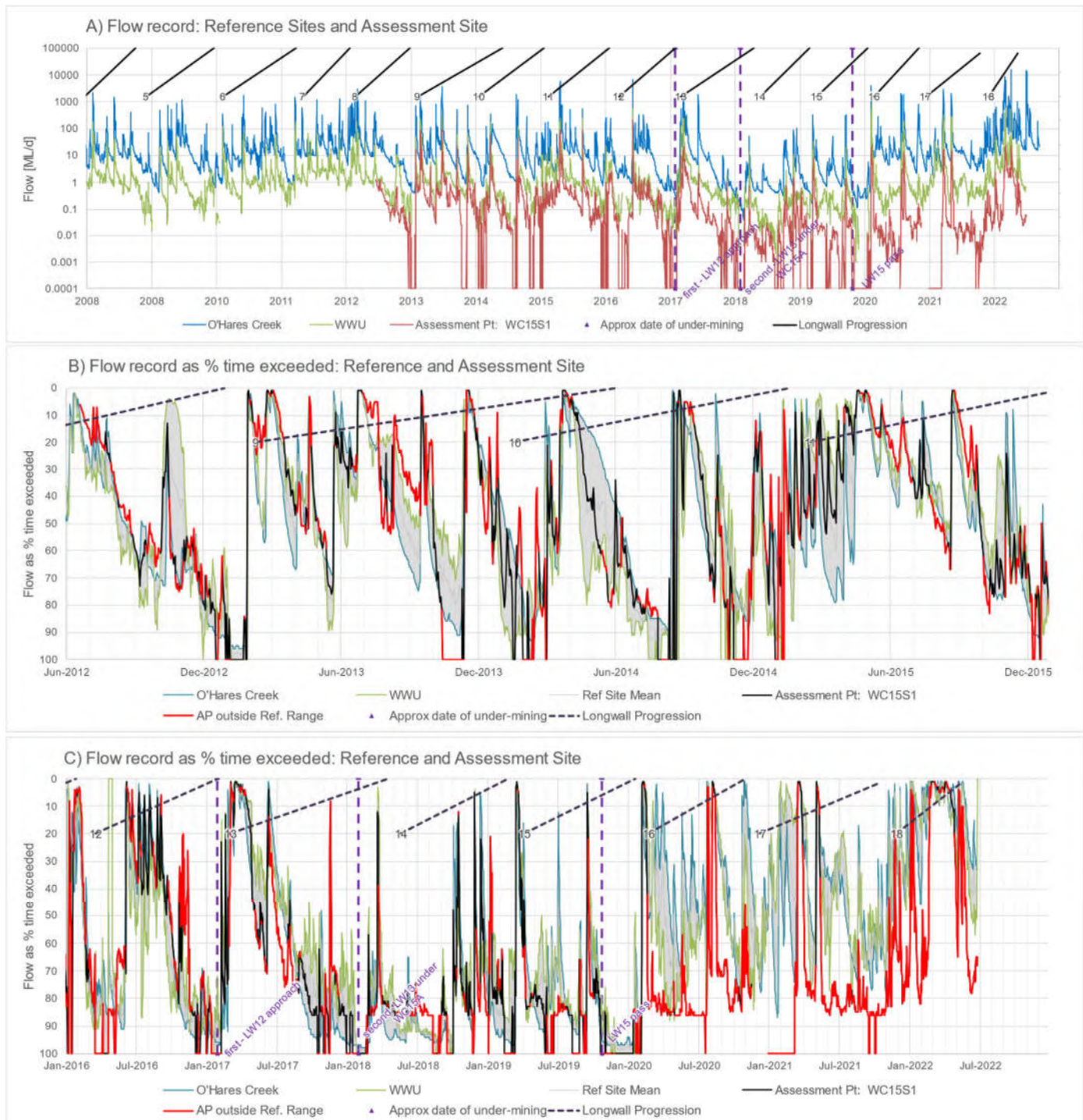


Figure 17. Comparison of WC15S1 against Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the WC15S1 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 18. Flow assessments A, B and C for the sub-catchment to WC15S1

| WC15s1 | Pre-mining | Post-mining | |
|---|---|---------------------------|---------------------|
| | to LW12 approach | end LW18 + 30 days | |
| | 20/06/2012 | 29/01/2017 | |
| | 28/01/2017 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 21% | 38% | of the time |
| Post-mining | 57% | 11% | of the time |
| Change | 35% | -28% | of the time |
| Assessment A: | | | Level 3 |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 5.7% | 5.7% |
| WWU | 5.0% | 8.4% | 3.4% |
| Average Ref. Site change (= natural variability): | | | 4.5% |
| WC15S1 | 13.0% | 34.0% | 21.0% |
| no. of cease-to-flow days increased: | | | 16.4% |
| Assessment B: | | | ↑ to Level 2 |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 8.44 | 4.43 | -47% |
| WWU | 0.58 | 0.33 | -43% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | -47% | -45% | -43% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| WC15s1 | 0.150 | 0.013 | -91% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| WC15s1 | 0.079 | 0.082 | 0.085 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | -43.9% | -45.9% | -48.0% |
| % change (of 'expected' Q50) | -83% | -84% | -85% |
| ML/d change from natural | -0.066 | -0.069 | -0.072 |
| Assessment C: | | | Level 3 |

5.4.6 WC21S1 – Wongawilli Creek tributary

WC21, a tributary to Wongawilli Creek, was mined under late in Longwall 9, and has since been mined under by Longwalls 10-15. Longwalls 16-18 are located south of this sub-catchment. During flooding in early 2022, the logger was damaged and data was unrecoverable. Logger replaced in June-2022.

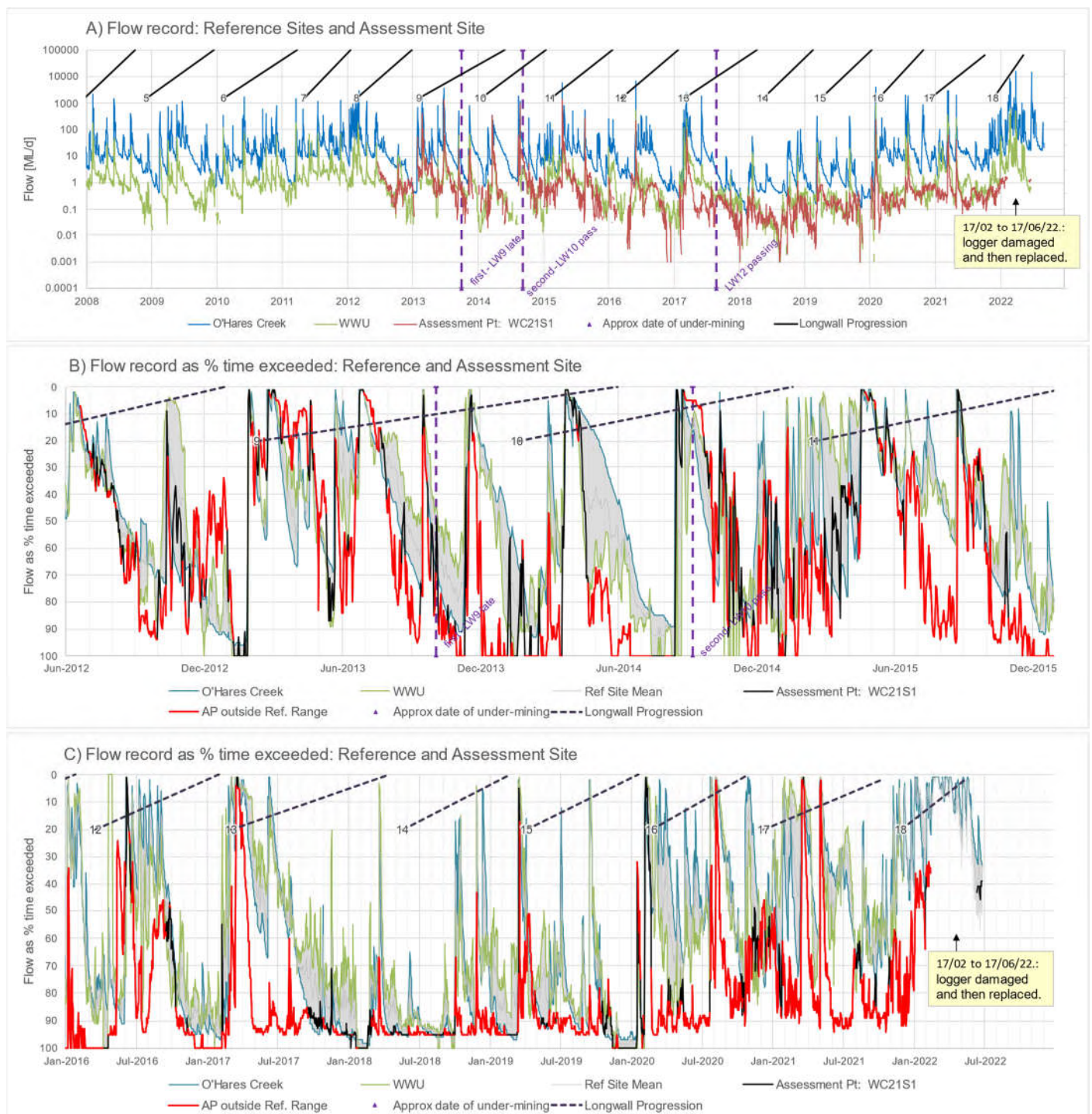


Figure 18. WC21S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C show the Q%ile hydrograph for the WC21S1 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 19. Flow assessments A, B and C for the sub-catchment to WC21S1

| WC21S1 | Pre-mining | Post-mining | |
|---|---|---------------------------|---|
| | to LW9 late | end LW18 + 30 days | |
| | 20/06/2012 | 6/10/2013 | <i>(missing 119 days in early 2022)</i> |
| | 5/10/2013 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 40% | 36% | of the time |
| Post-mining | 73% | 6% | of the time |
| Change | 33% | -30% | of the time |
| Assessment A: | | | Level 3 |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 4.9% | 5.0% | 0.1% |
| Average Ref. Site change (= natural variability): | | | 0.1% |
| WC21S1 | 3.6% | 13.6% | 10.0% |
| no. of cease-to-flow days increased: | | | 10.0% |
| Assessment B: | | | ↓ to Level 1 |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 8.86 | 6.49 | -27% |
| WWU | 0.90 | 0.38 | -58% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | -58% | -42% | -27% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| WC21S1 | 0.960 | 0.209 | -78% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| WC21S1 | 0.407 | 0.555 | 0.703 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | -20.7% | -36.1% | -51.5% |
| % change (of 'expected' Q50) | -49% | -62% | -70% |
| ML/d change from natural | -0.198 | -0.346 | -0.494 |
| Assessment C: | | | Level 3 |

Note: The rating curve at this site has been revised recently, resulting in significantly higher flows (pre- and post-mining), and therefore greater losses (in ML/d) at median flow. Watershed HydroGeo considers the revised flows too high, on a flow per catchment area basis, compared to other sites in this area.

5.4.7 WWL – Wongawilli Creek (lower)

Wongawilli Creek lies between Areas 3A and 3B. The watercourse is not directly mined under by longwalls, but some tributaries (e.g. WC21, WC15 etc.) have been mined under by Area 3A and 3B longwalls 6-17. Watercourse impacts, e.g. cracking at Pool 43a, have been identified in the past. Longwall 18 is outside the catchment to WWL.

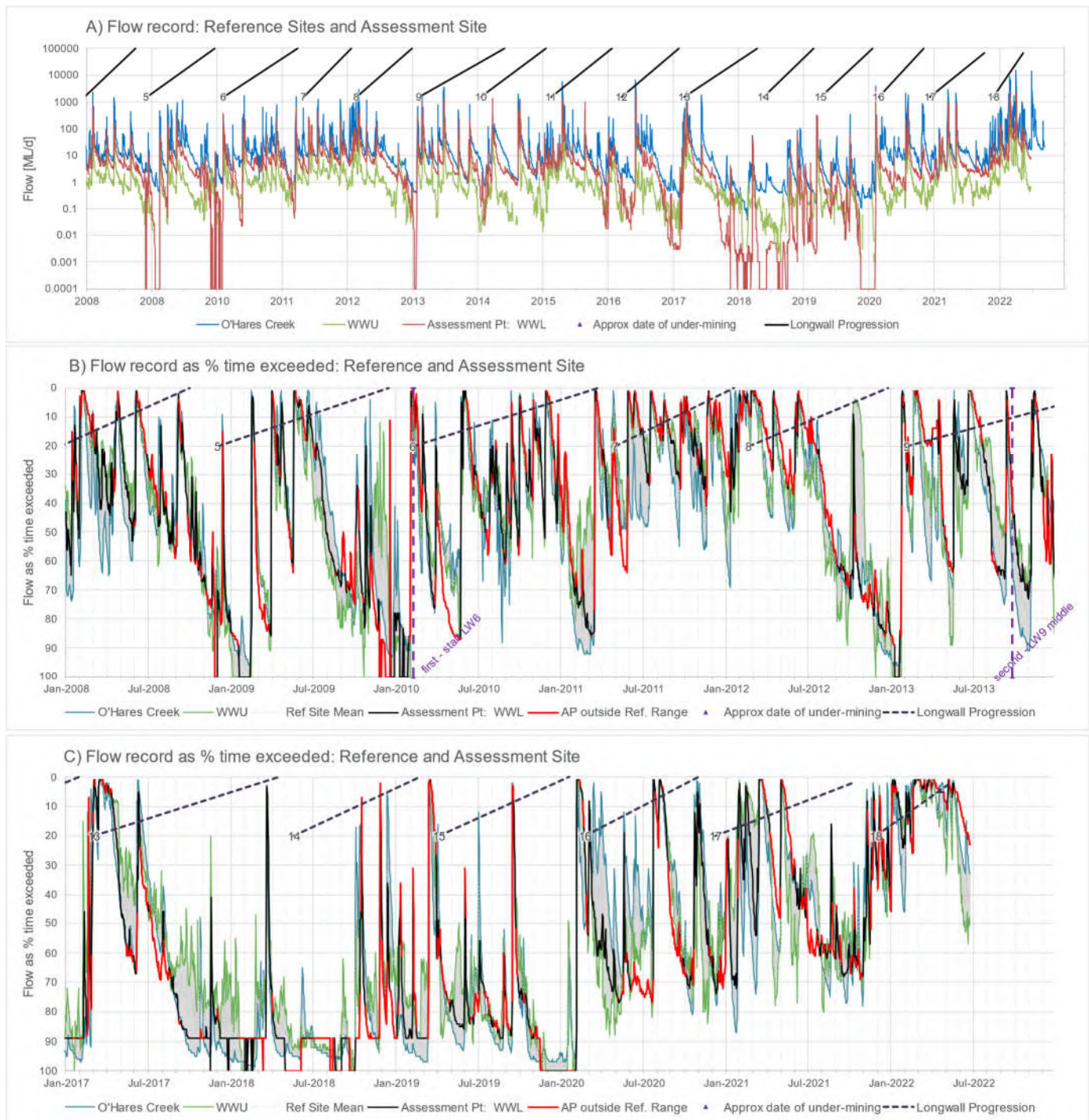


Figure 19. WWL vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the WWL Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 20. Flow assessments A, B and C for the sub-catchment to WWL

| WWL | Pre-mining | Post-mining | |
|---|---|---------------------------|----------------------|
| | to start LW6 | end LW18 + 30 days | |
| | 1/01/2008 | 10/02/2010 | |
| | 9/02/2010 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 22% | 28% | of the time |
| Post-mining | 21% | 34% | of the time |
| Change | -1% | 7% | of the time |
| Assessment A: | | | Not triggered |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 9.7% | 4.0% | -5.7% |
| Average Ref. Site change (= natural variability): | | | -2.9% |
| WWL | 9.1% | 3.8% | -5.2% |
| no. of cease-to-flow days increased: | | | -2.4% |
| Assessment B: | | | Not triggered |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 9.50 | 9.29 | -2% |
| WWU | 0.75 | 0.64 | -15% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | -15% | -8% | -2% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| WWL | 3.372 | 3.171 | -6% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| WWL | 2.875 | 3.087 | 3.298 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | +8.8% | +2.5% | -3.8% |
| % change (of 'expected' Q50) | 10% | 3% | -4% |
| ML/d change from natural | +0.296 | +0.084 | -0.127 |
| Assessment C: | | | Not triggered |

5.4.8 LA4S1 – Lake Avon tributary

LA4, a tributary to Lake Avon, lies above the western ends of Longwalls 11-14, but was not mined under by Longwalls 15-16. The gauging site was directly impacted by Longwall 13 with fractures and flow diversion. Equipment faults mean there was no data Jul-2019 to Apr-2021 – this has been fixed.

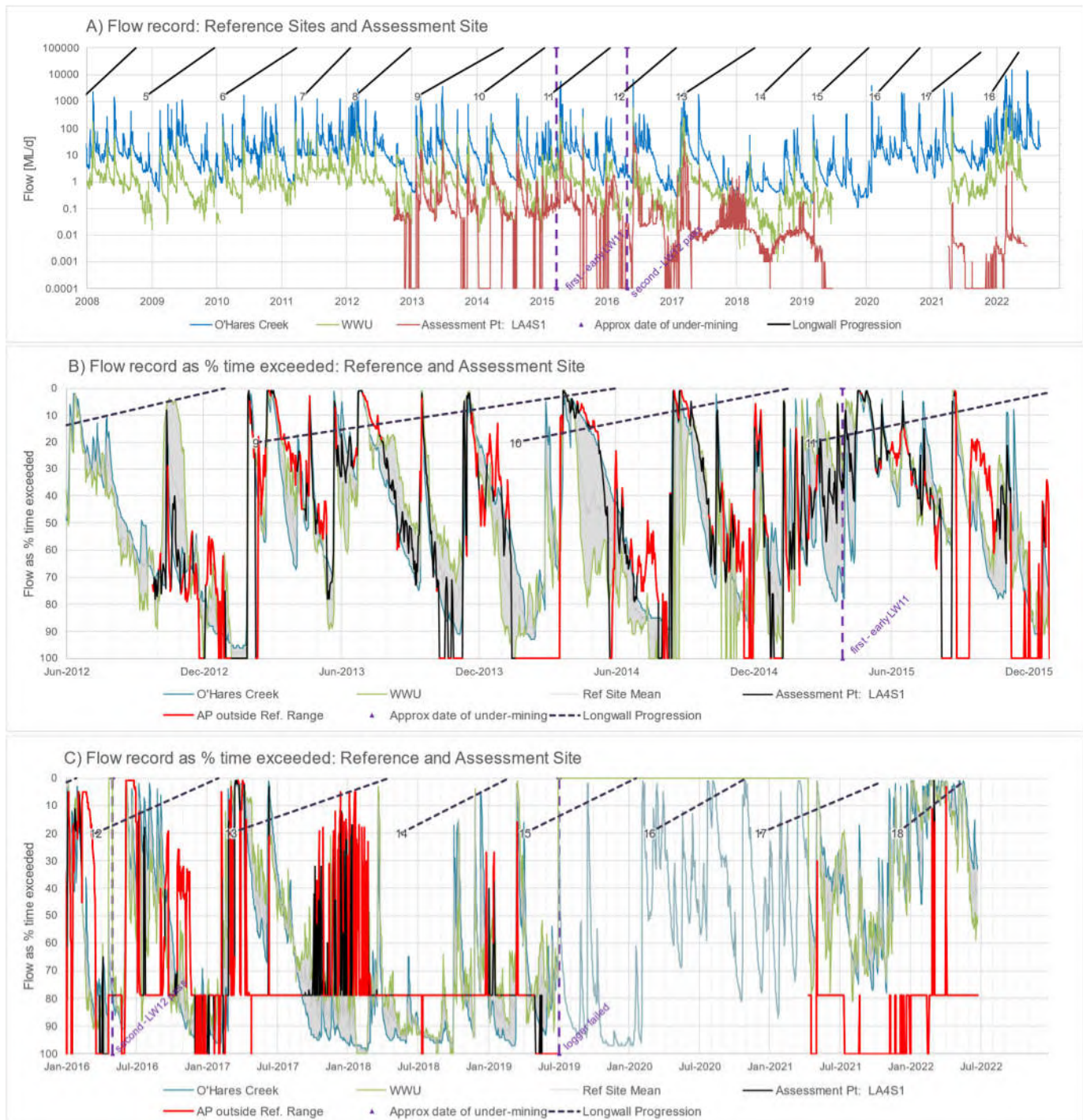


Figure 20. Comparison of LA4S1 against Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C show the Q%ile hydrograph for the LA4S1 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 21. Flow assessments A, B and C for the sub-catchment to LA4S1

| LA4S1 | Pre-mining | Post-mining | |
|---|---|---------------------------|--------------|
| | to early LW11 | end LW18 + 30 days | |
| | 24/09/2012 | 2/04/2015 | |
| | 1/04/2015 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 26% | 32% | of the time |
| Post-mining | 51% | 28% | of the time |
| Change | 25% | -3% | of the time |
| Assessment A: | | | ↑ to Level 3 |
| Method B: Change in cease-to-flow frequency: (this assessment uses 0.02 ML/d as 'cease-to-flow') | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 14.5% | 23.6% | 9.1% |
| WWU | 19.6% | 11.5% | -8.0% |
| Average Ref. Site change (= natural variability): | | | 0.6% |
| LA4S1 | 19.3% | 49.8% | 30.7% |
| no. of cease-to-flow days increased: | | | 30.1% |
| Assessment B: | | | ↑ to Level 3 |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 7.71 | 6.04 | -22% |
| WWU | 0.58 | 0.42 | -27% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | -27% | -24% | -22% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| LA4S1 | 0.081 | 0.010 | -88% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| LA4S1 | 0.059 | 0.061 | 0.063 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | -60.9% | -63.4% | -66.0% |
| % change (of 'expected' Q50) | -83% | -84% | -84% |
| ML/d change from natural | -0.049 | -0.051 | -0.053 |
| Assessment C: | | | Level 3 |

5.4.9 LA3S1 – Lake Avon tributary

LA3 is a tributary to Lake Avon. The upstream end of the watercourse was directly mined under by the western end of Longwall 15, and Longwall 16 approached within 40 m. As shown in Table C4 in Appendix C, the pre-mining baseline period is only 2 months, and so the statistical assessment of impacts is considered somewhat unreliable, however mining effects on flows are obvious.

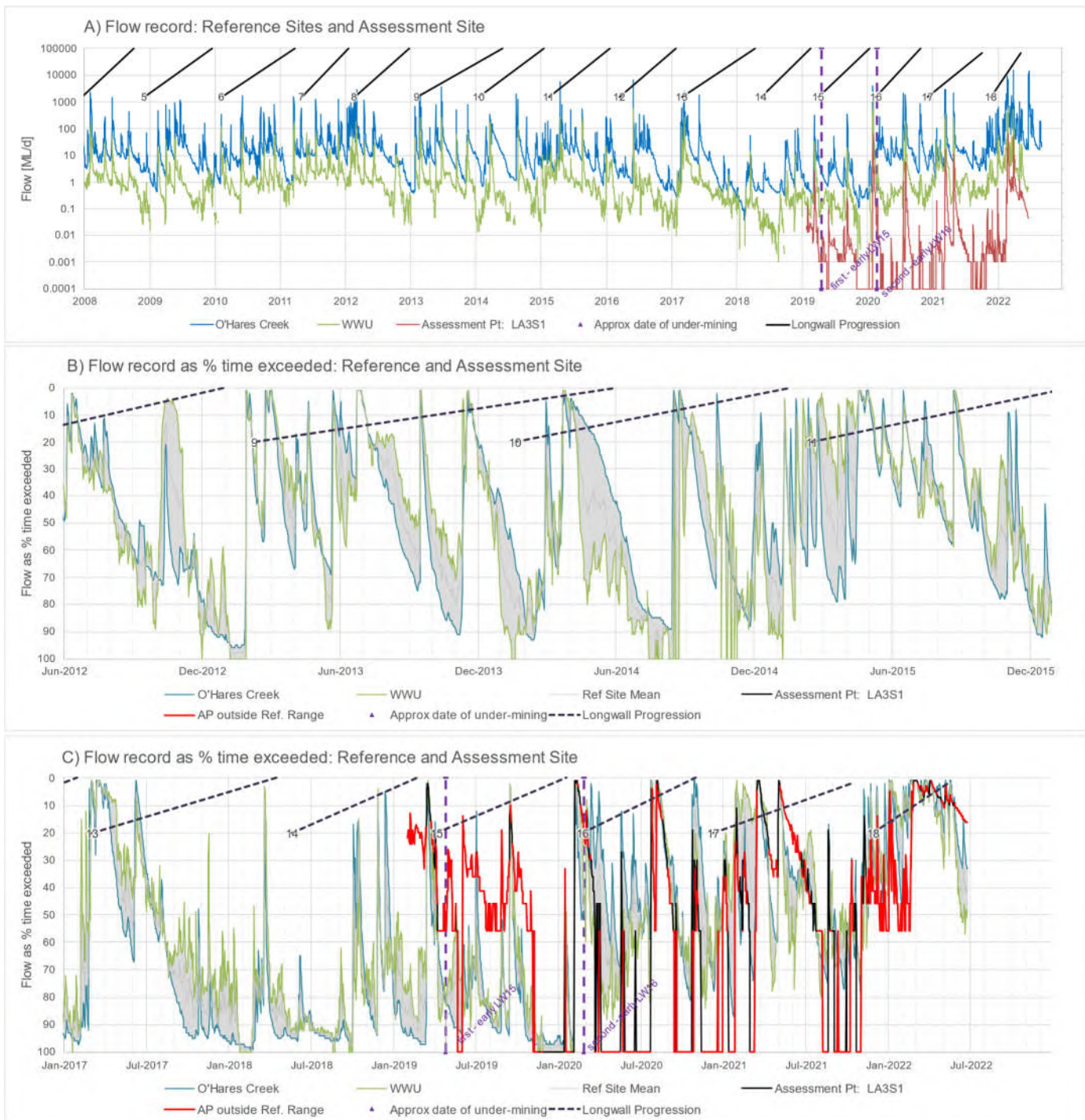


Figure 21. LA3S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the LA3S1 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 22. Flow assessments A, B and C for the sub-catchment to LA3S1

| LA3S1 | Pre-mining | Post-mining | |
|---|---|---------------------------|----------------|
| | to early LW15 | end LW18 + 30 days | |
| | 3/02/2019 | 29/04/2019 | |
| | 28/04/2019 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 0% | 85% | of the time |
| Post-mining | 47% | 30% | of the time |
| Change | 47% | -55% | of the time |
| Assessment A: | | | Level 3 |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 0.0% | 5.4% | 5.4% |
| Average Ref. Site change (= natural variability): | | | 2.7% |
| LA3S1 | 0.0% | 32.4% | 32.4% |
| no. of cease-to-flow days increased: | | | 29.7% |
| Assessment B: | | | Level 3 |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 2.20 | 10.57 | 381% |
| WWU | 0.21 | 0.52 | 149% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | 149% | 265% | 381% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| LA3S1 | 0.015 | 0.002 | -87% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| LA3S1 | 0.037 | 0.055 | 0.072 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | -235.5% | -351.7% | -467.9% |
| % change (of 'expected' Q50) | -95% | -96% | -97% |
| ML/d change from natural | -0.035 | -0.053 | -0.070 |
| Assessment C: | | | Level 3 |

5.4.10 LA2S1 – Lake Avon tributary

Longwall 15 approached within approximately 160 m of LA2 and skirted the northern edge of this sub-catchment. Longwall 16 mined beneath the headwaters of LA2. Longwall 17 mined directly beneath approximately 60% of the length of LA2. Longwall 18 passed approximately 100 m south of LA2.

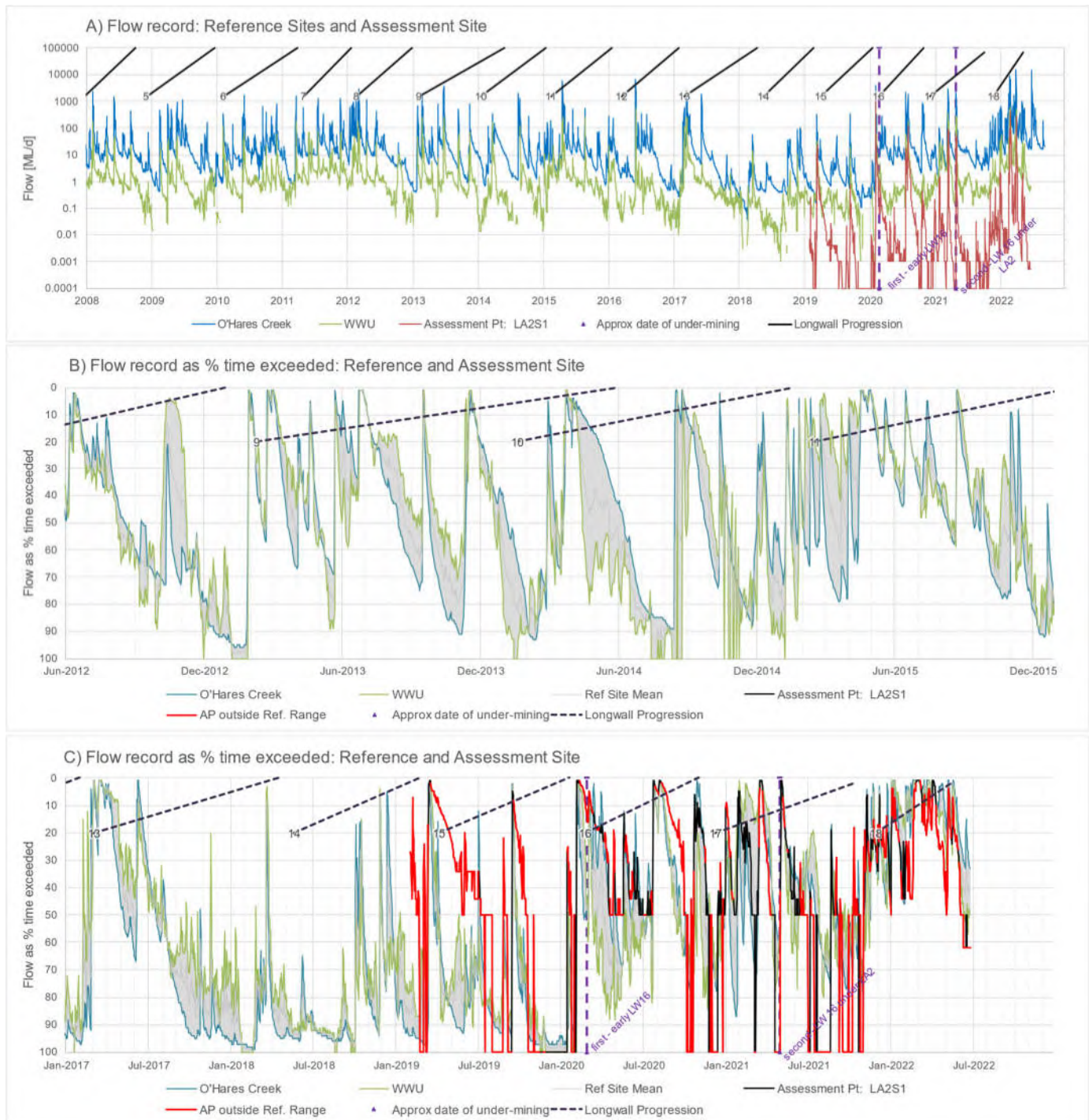


Figure 22. LA2S1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the LA2S1 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 23. Flow assessments A, B and C for the sub-catchment to LA2S1

| LA2S1 | Pre-mining | Post-mining | |
|--|---|---------------------------|--------------------|
| | to early LW16 | end LW18 + 30 days | |
| | 4/02/2019 | 2/03/2020 | |
| | 1/03/2020 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 22% | 59% | of the time |
| Post-mining | 38% | 31% | of the time |
| Change | 16% | -28% | of the time |
| Assessment A: | | | ↑ to Level 2 |
| Method B: Change in cease-to-flow frequency: (this assessment uses 0.002 ML/d as 'cease-to-flow') | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 2.3% | 0.0% | -2.3% |
| WWU | 18.9% | 0.0% | -18.9% |
| Average Ref. Site change (= natural variability): | | | -10.6% |
| LA2S1 | 55.4% | 41.6% | -13.8% |
| no. of cease-to-flow days increased: | | | -3.2% |
| Assessment B: | | | ↓ to not triggered |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 1.25 | 16.02 | 1178% |
| WWU | 0.14 | 0.84 | 508% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | 508% | 843% | 1178% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| LA2S1 | 0.002 | 0.005 | 150% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| LA2S1 | 0.012 | 0.019 | 0.026 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | -357.6% | -692.6% | -1027.6% |
| % change (of 'expected' Q50) | -59% | -73% | -80% |
| ML/d change from natural | -0.007 | -0.014 | -0.021 |
| Assessment C: | | | Level 3 |

5.4.11 ND1 – Native Dog Creek tributary

ND1 is a tributary to Native Dog Creek, which flows into Lake Avon. Elouera Colliery longwalls are within or close to this sub-catchment, but were not directly beneath this watercourse or its tributaries. ND1 is yet to be mined under by Dendrobium longwalls, but Longwall 17 mined under the northern edge of this sub-catchment. Longwall 18 mined under the headwaters of ND1 and its tributary ND1C.

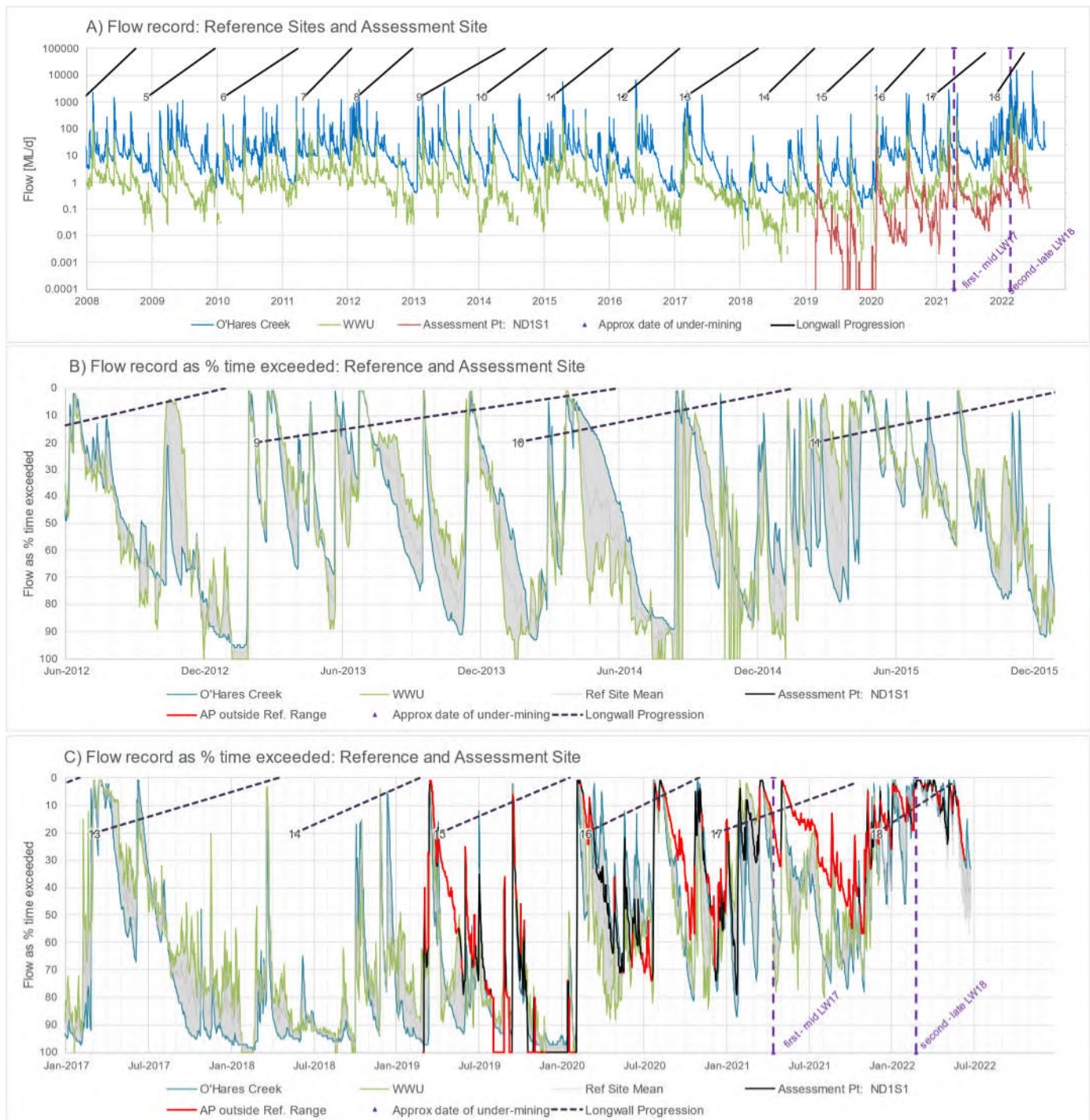


Figure 23. ND1 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the ND1S1 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 24. Flow assessments A, B and C for the sub-catchment to ND1S1

| ND1S1 | Pre-mining | Post-mining | |
|---|---|---------------------------|----------------------|
| | to mid LW17 | end LW18 + 30 days | |
| | 3/03/2019 | 19/04/2021 | |
| | 18/04/2021 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 17% | 38% | of the time |
| Post-mining | 6% | 71% | of the time |
| Change | -11% | 34% | of the time |
| Assessment A: | | | Not triggered |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 8.0% | 0.0% | -8.0% |
| Average Ref. Site change (= natural variability): | | | -4.0% |
| ND1S1 | 17.5% | 0.0% | -17.5% |
| no. of cease-to-flow days increased: | | | -13.5% |
| Assessment B: | | | Not triggered |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 7.04 | 18.49 | 162% |
| WWU | 0.36 | 1.19 | 231% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | 162% | 197% | 231% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| ND1S2 | 0.023 | 0.280 | 1115% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| ND1S2 | 0.060 | 0.068 | 0.076 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | 952.7% | 918.7% | 884.6% |
| % change (of 'expected' Q50) | 363% | 310% | 268% |
| ML/d change from natural | +0.219 | +0.211 | +0.203 |
| Assessment C: | | | Not triggered |

5.4.12 SC10C – Sandy Creek minor tributary

This site is in Area 3A, but included in this report for information. SC10C is a minor tributary to Sandy Creek, which flows into Lake Cordeaux. Area 3A longwalls 7 and 8 mined beneath this watercourse.

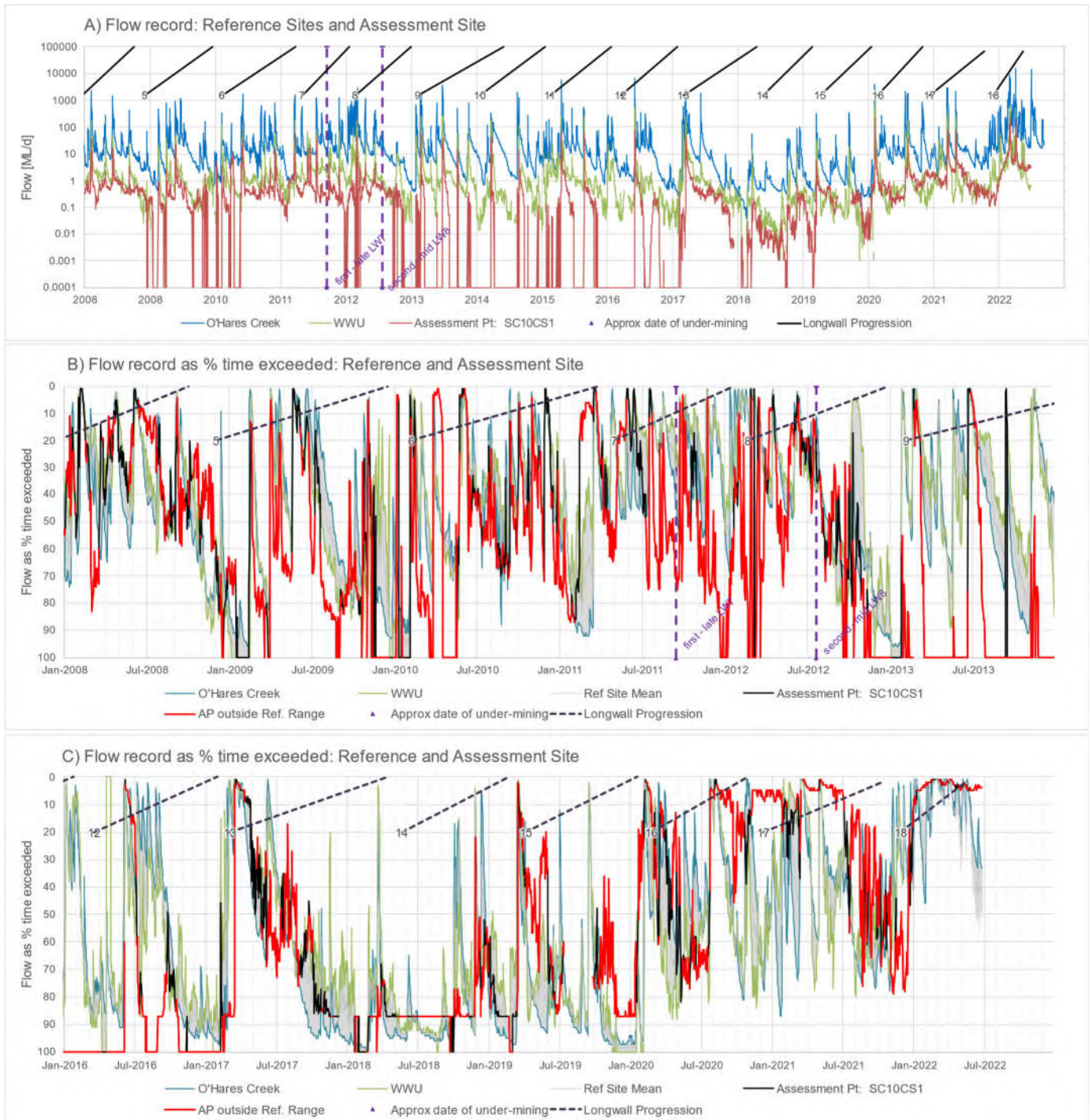


Figure 24. SC10C vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the SC10CS1 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 25. Flow assessments A, B and C for the sub-catchment to SC10C

| SC10CS1 | Pre-mining | Post-mining | |
|---|---|---------------------------|----------------------|
| | to late LW7 | end LW18 + 30 days | |
| | 1/01/2008 | 18/09/2011 | |
| | 17/09/2011 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 49% | 23% | of the time |
| Post-mining | 52% | 28% | of the time |
| Change | 3% | 5% | of the time |
| Assessment A: | | | Not triggered |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 5.5% | 4.6% | -0.9% |
| Average Ref. Site change (= natural variability): | | | -0.5% |
| SC10CS1 | 11.8% | 30.7% | 18.9% |
| no. of cease-to-flow days increased: | | | 19.4% |
| Assessment B: | | | ↓ to Level 2 |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 11.18 | 8.64 | -23% |
| WWU | 0.94 | 0.53 | -43% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | -43% | -33% | -23% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| SC10CS1 | 0.379 | 0.147 | -61% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| SC10CS1 | 0.216 | 0.254 | 0.293 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | -18.2% | -28.3% | -38.5% |
| % change (of 'expected' Q50) | -32% | -42% | -50% |
| ML/d change from natural | -0.069 | -0.107 | -0.146 |
| Assessment C: | | | Level 3 |

Interestingly, at this site, if considering the period from Jan-2017 to now, the Assessments presented above are all 'Not triggered', which is consistent with the hydrograph shown in Figure A above. This is consistent with recovery of groundwater levels and the emergence of iron-staining in this watercourse.

5.4.13 SC10 – Sandy Creek tributary

SC10 is a tributary to Sandy Creek, which flows into Lake Cordeaux. Longwalls 7 and 8 mined beneath this catchment, and the south-eastern corner of Longwall 8 mined beneath the watercourse. Longwall 19 will also mine within 40 m of this watercourse. The control structure changed at this site (Dec-2021) and due to heavy rainfall, has not been completed. Therefore, recent data is unreliable. Longwall 19 will also mine within this sub-catchment.

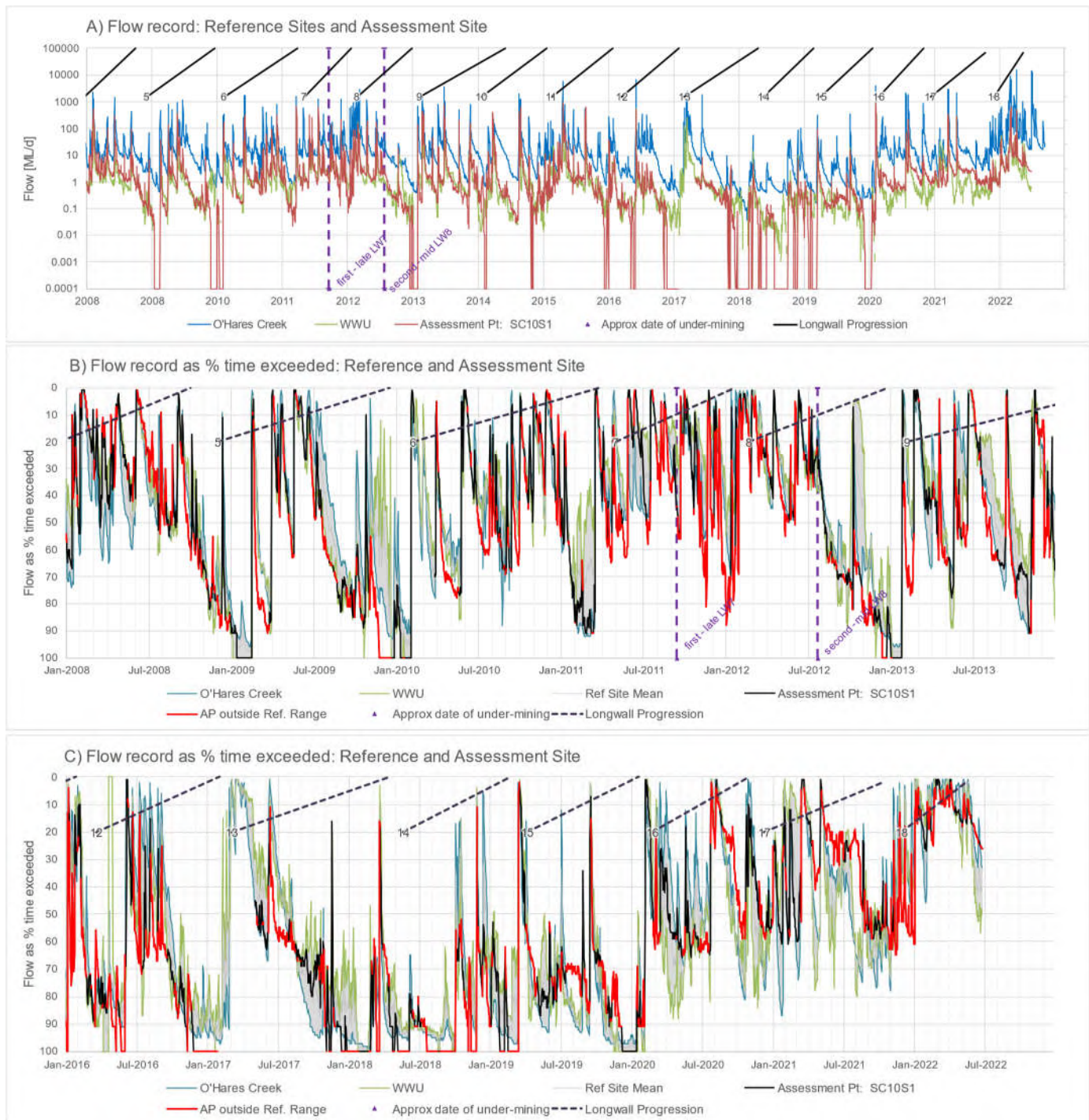


Figure 25. SC10 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures B and C (above) show the Q%ile hydrograph for the SC10S1 Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 26. Flow assessments A, B and C for the sub-catchment SC10

| SC10S1 | Pre-mining | Post-mining | |
|---|---|---------------------------|---------------------------|
| | to late LW7 | end LW18 + 30 days | |
| | 1/01/2008 | 18/09/2011 | |
| | 17/09/2011 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 41% | 16% | of the time |
| Post-mining | 41% | 18% | of the time |
| Change | 1% | 3% | of the time |
| Assessment A: | | | Not triggered |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 5.7% | 4.6% | -1.1% |
| Average Ref. Site change (= natural variability): | | | -0.46 |
| SC10S1 | 7.0% | 11.2% | 4.2% |
| no. of cease-to-flow days increased: | | | 4.8% |
| Assessment B: | | | ↓ to Not Triggered |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 11.27 | 8.10 | -28% |
| WWU | 0.93 | 0.50 | -47% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | -47% | -37% | -28% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| SC10S1 | 1.254 | 0.668 | -47% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| SC10S1S | 0.669 | 0.785 | 0.901 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | -0.1% | -9.3% | -18.6% |
| % change (of 'expected' Q50) | 0% | -15% | -26% |
| ML/d change from natural | -0.001 | -0.117 | -0.233 |
| Assessment C: | | | ↓ to Not Triggered |
| (Assessment C very close to triggering Level 1) | | | |

5.4.14 SCL2 / 2122205 – Sandy Creek

Sandy Creek flows into Lake Cordeaux near Area 3A. Area 2 Longwall 5 mined along the edge of this catchment, while Area 3A Longwalls 7 and 8 mined beneath this catchment. All these longwalls were at least 400 m from the watercourse. Longwall 19 will also mine within this sub-catchment.

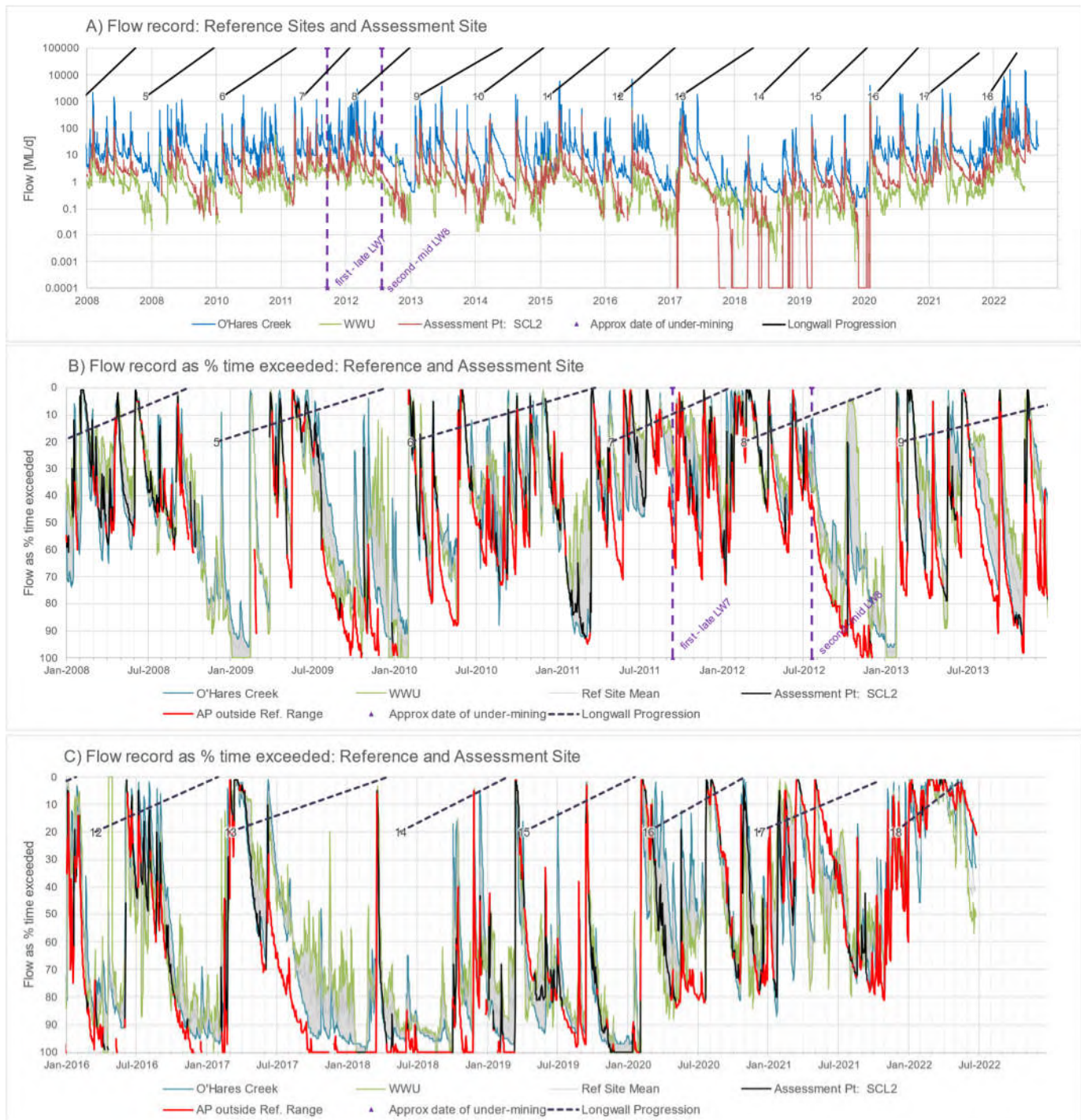


Figure 26. SCL2/2122205 vs Reference Sites A) flows; B) and C) flow duration statistics [Q%iles]

Figures 25B and 25C show the Q%ile hydrograph for the Sandy Creek Assessment Point versus the Q%ile hydrographs for the Reference Sites, O'Hares Creek and WWU. For this assessment 'natural variability' is defined as the range between the Q%ile for the Reference Sites on each day.

Table 27. Flow assessments A, B and C for the sub-catchment to SCL2/2122205

| SCL2 | Pre-mining | Post-mining | |
|---|---|---------------------------|----------------------|
| | to late LW7 | end LW18 + 30 days | |
| | 1/01/2008 | 18/09/2011 | |
| | 17/09/2011 | 16/06/2022 | |
| Method A: Assessment of flow variability: | | | |
| Period | Compared to Reference Sites, gauge is at: Lower flow (higher Q%ile) | Higher flow (lower Q%ile) | |
| Pre-mining | 57% | 9% | of the time |
| Post-mining | 57% | 13% | of the time |
| Change | 0% | 4% | of the time |
| Assessment A: | | | Not triggered |
| Method B: Change in cease-to-flow frequency: | | | |
| Cease to flow as % of daily record during pre- and post- mining periods | | | |
| Site | Pre-mining | Post-mining | Change |
| O'Hares Creek | 0.0% | 0.0% | 0.0% |
| WWU | 6.5% | 4.9% | -1.6% |
| Average Ref. Site change (= natural variability): | | | -0.8% |
| SCL2 | 0.0% | 8.3% | 8.3% |
| no. of cease-to-flow days increased: | | | 6.5% |
| Assessment B: | | | Level 1 |
| Method C: Change to median flow (Q50) | | | |
| Reference Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| O'Hares Creek | 12.68 | 9.06 | -29% |
| WWU | 0.98 | 0.58 | -41% |
| Natural variability | Min | Mean | Max |
| from 2 x Ref. Sites | -41% | -35% | -29% |
| Assessment Site Q50 [ML/d] | Q50 (pre-) | Q50 (post-) | % Change |
| SCL2 | 2.168 | 1.380 | -36% |
| 'Expected' post-mining Q50 at | Min | Mean | Max |
| SCL2 | 1.280 | 1.415 | 1.550 |
| Change beyond natural | Min | Mean | Max |
| % change (of pre-mining Q50) | +4.6% | -1.6% | -7.8% |
| % change (of 'expected' Q50) | 8% | -2% | -11% |
| ML/d change from natural | +0.100 | -0.035 | -0.170 |
| Assessment C: | | | Not triggered |

5.4.15 Discussion of flow assessments A, B, C.

Comments are made here on specific sub-catchments and assessments where the above assessments required some further explanation.

LA4 (Section 5.4.8): there is an apparent modification to the accuracy in the estimation and/or reporting of low flows from 2016, approximately coincidental with pre- and post-mining periods. This includes a number of suspected 'false zeroes' from earlier in the record that we could not confidently 'process' or 'infill'. Our review of visual inspection records ("flow observations") by IMCEFT, such those summarised on the maps presented in Appendix G and considering the changes to the frequency of low flows above cease-to-flow. As such we have modified the 'cease-to-flow' to a very low flow (0.02 ML/d) and used this in Assessment B, giving a Level 2 trigger. A similar behaviour is apparent at:

1. WC15: based on assessment of recorded low flows, we have adopted a low-flow (0.005 ML/d) in place of true 'cease-to-flow' for Assessment B, and as a result a Level 2 TARP was considered more appropriate than 'Not triggered' (consistent with previous End of Panel reporting).
2. LA2: based on assessment of recorded low flows, we have adopted a low-flow (0.002 ML/d) in place of true 'cease-to-flow' for Assessment B, and as a result a Level 1 TARP was considered more appropriate than 'Not triggered'.

5.4.16 Comparison against rainfall-runoff modelling

Up until Longwall 14, effects of surface water flow quantity were assessed via comparison of observed flow against rainfall-runoff modelling, either the RUNOFF-2005 model (used by Ecoengineers up until 2015) and then AWBM (used for Longwalls 13 and 14). As discussed in Section 5.1, this was superseded in consultation with agencies, with comparison against Reference Sites preferred. The IAPUM (Section 1.3) has recently requested that this be re-instated, and the most conservative result (of the new, agreed TARPs and the rainfall-runoff comparison) be adopted as the finding.

While the use of rainfall-runoff modelling itself is valid (especially so if appropriate Reference Sites are not available), the issue is that the pre-Longwall 15 and now-agreed TARPs have multiple differences, including:

1. assessment period (longwall by longwall or cumulatively since mining);
2. the choice of indicator, being changes to 'catchment yield' expressed as a percentage of long-term average rainfall (as per the calculation recommended by Ecoengineers (2011) or changes to other flow indicators such as cease-to-flow frequency and median flow).

Therefore, while the rainfall-runoff method is used here for limited sites, as per IAPUM's request, this is only as a secondary check until further discussion with agencies has occurred.

Appendix H presents a summary of rainfall-runoff modelling using the superseded assessment methods. This is done specifically for those sub-catchments where Assessment C (for median flow) does not already trigger TARP Level 3, and so provides a secondary check on effects.

DCU: Section H1 presents a summary of rainfall-runoff modelling. Hydrographs and ratios for the pre- and post-mining periods do not clearly indicate a systematic or significant change in catchment behaviour during Longwall 18 (or post-mining in general). The calculation of catchment yield did not trigger the former TARP. This finding is consistent with the agreed TARPs.

WC12: Section H2 presents a summary of rainfall-runoff modelling. Hydrographs, ratios and flow duration curves for the pre- and post-mining periods suggest that a mild change in catchment behaviour might have occurred. The calculation of catchment yield triggered the former TARP Level 1. While suggesting a mining effect that was not indicated by TARP Assessments B and C, the Level 1 trigger indicates a very minor potential effect.

WWL: Section H3 presents a summary of rainfall-runoff modelling. Hydrographs, ratios and flow duration curves for the pre- and post-mining periods do not clearly indicate a systematic or significant change in catchment behaviour. Using the calculation of ‘catchment yield’ derived by Ecoengineers (2011) suggests that there could be a mild reduction in flow to the period ending with Longwall 18 but still fall just short of triggering Level 1 (-3% for the full post-mining period).

ND1: Section H4 presents a summary of rainfall-runoff modelling. Hydrographs, ratios and flow duration curves for the pre- and post-mining periods suggest that flows during the second half of Longwall 18 may have declined relative to modelled flows (TARP Level 3). This finding is significantly different to the findings with the agreed TARPs using Reference Sites.

5.5 Assessment D: flow reduction Wongawilli Creek

Surface water flow observations made by IMCEFT are recorded in a semi-qualitative fashion. At each field site (such as at the upstream or downstream end of a pool), an observation of flow conditions is made as follows:

| | |
|---|--------------------------|
| 0 | No flow visible |
| 1 | Subsurface flow observed |
| 2 | Surface seepage observed |
| 3 | Surface trickle observed |
| 4 | Surface flow observed |

Field surveys typically make an observation at each of the nominated sites around Area 3A and 3B over the period of a month. The “Outflow” results of IMCEFT’s surveys are plotted on the maps in Appendix G for each month during the period covering the extraction of Longwall 17. As noted on the maps, observations are limited in two months during Longwall 17 (January and February-2021) due to the heavy rainfall conditions and catchment closures.

Sites along the main channel of Wongawilli Creek are the subject of Assessment D, and these sites are shown with a hollow black circle in Appendix G to minimise confusion with sites on tributaries yet very close to the main branch of Wongawilli Creek.

While there are often “no flow” observations on the tributaries which flow into Wongawilli Creek, there are consistent observations of flow along Wongawilli Creek itself. Of the completed surveys, all months are “Not triggered”. As a result, the further calculation of Assessment D is not required.

Table 28. Assessment D for Wongawilli Creek: Longwall 18

| During Longwall 18 | Assessment D |
|---|------------------|
| Dec-2021, Jan-2022, Feb-2022, June-2022 | Not triggered |
| Mar-2022, April-2022, May-2022 | Catchment closed |

Any inferred loss of flow from Assessment D is then used in assessing compliance against Performance Measures for Wongawilli Creek.

5.6 Assessment against surface water flow Performance Measures

There are four agreed Performance Measures for surface water flows in the Area 3B WIMMCP.

Wongawilli Creek – minor environmental consequences

Agreed measure: *Methods C, D, to be compared against predictions made in contemporary groundwater modelling conducted to the satisfaction of the Secretary to assess whether effects that cannot be explained by natural variability “exceed prediction”.*

Assessment C at WWL does not indicate a discernible reduction beyond natural variability in Q50 (Table 20). Therefore, this Performance Measure is met.

Assessment D for flows along the middle of Wongawilli Creek (Table 28) was not triggered. While loss of baseflow is highly likely to occur during the assessment period, weather conditions mask any effect. Therefore, the estimated losses cannot be assessed.

Donalds Castle Creek – minor environmental consequences

Agreed measure: *Method C to be compared against predictions made in contemporary groundwater modelling conducted to the satisfaction of the Secretary to assess whether effects that cannot be explained by natural variability “exceed prediction”.*

Assessment C at DCU does not indicate a discernible reduction beyond natural variability in Q50 (Table 16). Therefore, this Performance Measure is met.

Cordeaux River – negligible reduction in the quantity of surface water inflow to the Cordeaux River at its confluence with Wongawilli Creek

Agreed measure: - - *Flow reduction as determined from measured at flow gauging station WWL_A.*

Assessment C at WWL does not indicate a reduction in Q50 (Table 20). Therefore, this Performance Measure is met.

Lake Avon – negligible reduction in the quantity of surface water inflows to Lake Avon

Agreed measure: - *Surface water inflows calculation = [Impacts at gauged catchments (LA2 + LA3 + LA4 + NDT1) + estimated impacts at ungauged but undermined catchments (e.g. LA5)] / [total inflow to LA].*

The calculation is presented as follows. In mined-under but un-monitored catchments, “inferred” losses are calculated as the same % reduction as a nearby monitored and mined-under catchment. If not mined under directly, but adjacent or neighbouring mining, then 25% of the % loss in the nearest mined under catchment is applied as the inferred loss.

| Sub-catchment | Gauged? | Mined under? | Catch area [km2] | "measured" loss at Q50 | "inferred" loss at Q50 |
|---------------|---------|--------------|------------------|------------------------|------------------------|
| LA1 | N | N | 0.29* | | 0 |
| LA2 | Y | Neighbour | 0.824 | -0.014 | |
| LA3 | Y | Y | 0.375 | -0.053 | |
| LA4 | Y | Y | 0.817 | -0.051 | |
| LA5 | N | Y | 0.53* | | -0.033 |
| LA6 | N | Neighbour | 0.97* | | -0.015 |
| ND1 | Y | Y | 1.13 | 0.000 | |

| Sub-catchment | Gauged? | Mined under? | Catch area [km ²] | | "measured" loss at Q50 | "inferred" loss at Q50 | |
|--|---------------------|--------------|-------------------------------|------|------------------------|------------------------|-------------|
| ND | N | Elouera | 3.85* | | | | |
| Total for mined-under or neighbouring catchments | | | 2.08 | | | -0.166 | ML/d |
| Lake Avon | | N | 142 [^] | 1.5% | | Q50 | Qmean |
| Inflow from catchment | (WaterNSW estimate) | | | | | 18 | 124 |
| Inferred mining loss as % of | total inflow | | | | | -0.9% | -0.13% |
| * catchment area estimated by WatershedHG from GIS. | | | | | | | |
| [^] catchment area from https://www.waternsw.com.au/supply/visit/avon-dam | | | | | | | |

The sub-catchments where mining effects related to Dendrobium are present or inferred constitute about 2% of the total catchment to Lake Avon. The “measured” + “inferred” reduction in Q50 flow in these LA catchments = 0.118ML/d (43 ML/yr). This is 0.9% of median Lake Avon inflow or 0.13% of average Lake Avon inflow for the period 2015-2020, based on WaterNSW lake inflow data.

The estimated losses are equivalent to:

- 22% of predicted losses for the Lake Avon catchment made by groundwater modelling (281 ML/yr) from the approved Longwall 17 SMP Application); and
- 156% of low-end predicted losses for Lake Avon catchment made by groundwater modelling (39 ML/yr) and 44% of the high-end losses (137 ML/yr) from the approved Longwall 18 SMP Application).
- Therefore, the estimated losses are “within prediction”, and this Performance Measure is met.

5.7 Watercourse pool water levels

5.7.1 Wongawilli Creek Pool 50 (previously Pool43A)

Pool 50 is located on Wongawilli Creek, 348 m east of Longwall 9 in Area 3B (extracted between 9/2/2013 and 2/6/2014) and 315 m northwest of Longwall 6 in Area 3A (9/2/2010 - 28/3/2011). Pool 50 is controlled by a rock bar. On 20/11/2017, it was noted during a site visit that water levels in Pool 50 on Wongawilli Creek were below the baseline (impact number DA3B_LW13_015, dated 28/11/2017). The observation triggered a TARP Level 3 because a previously reported fracture (first observed on 18/12/2013) is present in the sandstone forming the pool base. No significant changes to the downstream control were noted by the IMCEFT at Pool 50.

An assessment was carried out into the cause of the declining water levels in Pool 50 by Watershed (2018). The assessment concluded that the decline in pool levels was likely due to depressurisation of the underlying formations (HBSS and BGSS; Figure 28) due to mining adjacent to the creek, exacerbated by the very low rainfall and flow conditions during the 2017-2019 drought. The decline in pool levels started prior to the formation of the fracture (Figure 27) suggesting that water loss from the pool was not related to the formation of the fracture.

Piezometric levels in the sandstone substrate adjacent to Wongawilli Creek have recovered as mining in Area 3B has moved south and away from Pool 50 (Figure 28). Since 2021, piezometric levels in the HBSS adjacent to the pool have recovered to above the elevation of the creek bed. Water levels in Pool 50 have trended higher since 2020 in response to both higher rainfall conditions and recovering groundwater levels. Pool measurements have not been possible during 2022 due to frequent flooding and catchment closures; however groundwater levels continued to rise during the reporting period.

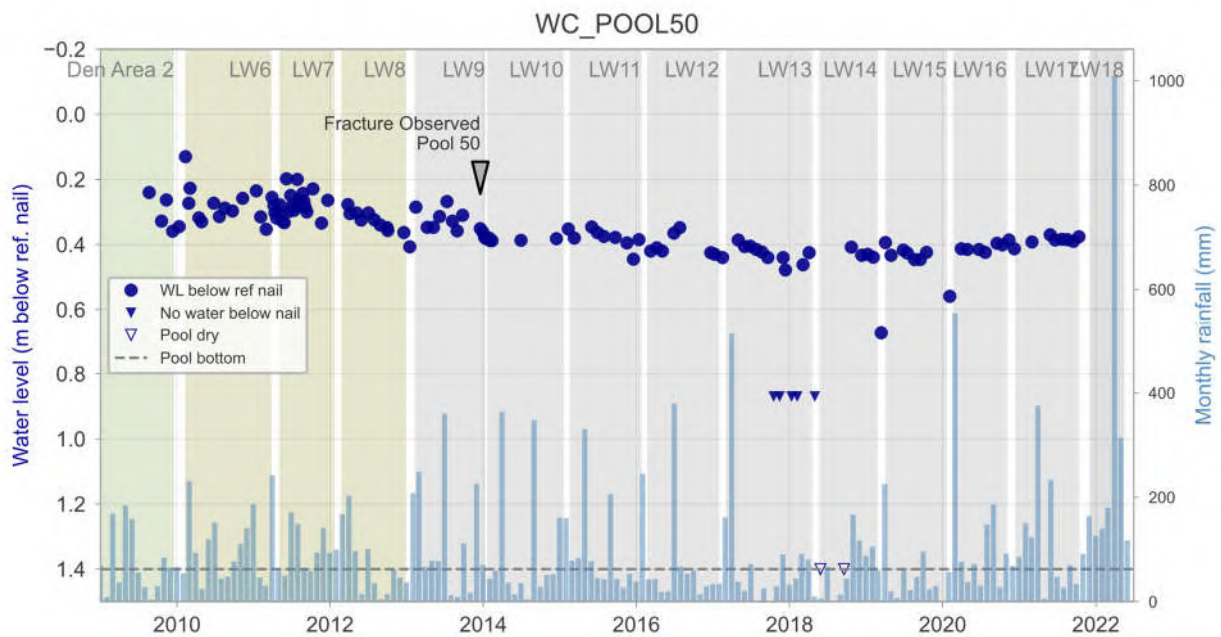


Figure 27. Time series plot of water level observations in Pool 50

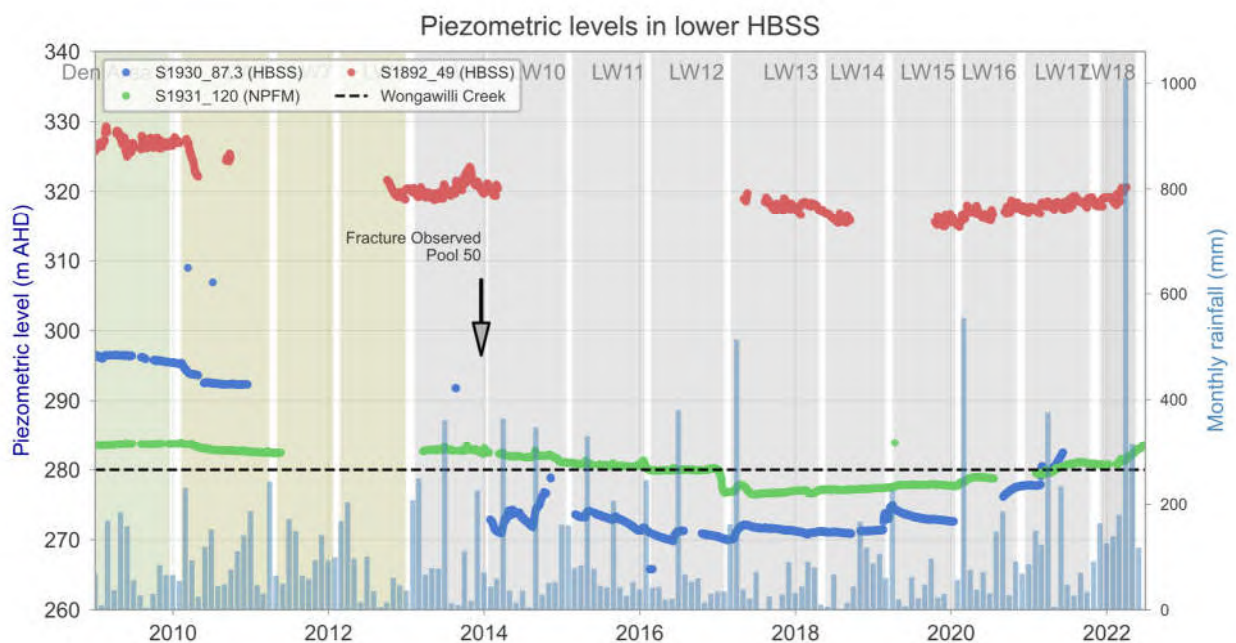


Figure 28. Groundwater hydrographs for lower HBSS adjacent to Wongawilli Creek

5.7.2 Pool level dataloggers in Wongawilli Creek

Pool level dataloggers are installed in seven pools along Wongawilli Creek adjacent to Area 3A and 3B: WWU, Waterfall 54, Pool 45, Pool 49, Pool 50 (WCS2), Pool 41 (WCS1). The dataloggers measure the water level at hourly intervals relative to a surveyed benchmark at the respective sites. Hydrographs for the loggers are included in Appendix F.

Most of the loggers were installed in 2020 and therefore have limited baseline data. Loggers at Pools 50 and 41 were installed in 2018. The hydrographs for those pools show different recession characteristics that are related to the geometry of the pools and their control points:

No adverse trends related to mining are evident as of the end of the current reporting period.

5.8 Pool outflow status

This section reviews the pool outflow status for watercourses that pass within the zone of influence (<400 m) of Longwall 18, as recorded in site field observations.

Pool outflow is summarised using “heatmap” plots showing observed flow status at each pool for monthly monitoring periods, with the passage of longwalls marked as red lines (e.g. Figure 29). Pools are arranged from upstream (bottom of the plot) to downstream (top). Observations of “no water in the pool” are overlain as “-“ symbols. Where more than one monitoring round was carried out in a month, the minimum condition was used in the figure.

5.8.1 Lake Avon tributary LA2

LA2 is a second order tributary to Lake Avon. Longwall 16 passed beneath the headwaters of LA2 and directly beneath LA2_Pool34 on 5/5/2020. Longwall 17 passed beneath 730 m of the main second order watercourse between 13/12/2020 and 2/3/2021, passing directly beneath LA2_Pool24 and LA2_Pool25 on ~18/2/2021. Longwall 18 passed beneath the southern most 17% of the LA2 catchment between 5/12/2021 and 6/3/2022 but did not pass directly between any additional mapped watercourses.

Four pools are routinely monitored along the LA2 watercourse (Figure 29). There is approximately seven years of baseline observations for the downstream site LA2_Pool5 and approximately 2 years baseline for upstream pools (with some gaps). The data indicate that Pools 24 and 25 were affected by the passage of Longwall 17. Both are recorded as having no water following the passage of Longwall 17, in contrast to baseline conditions and despite higher-than-average rainfall since that time. Observations for the upstream Pool 34 are inconclusive since the pool is not recorded as holding water during the baseline period. At the downstream Pool 5, outflow is recorded less frequently since Longwall 17 than prior to the 2017-2019 drought, suggesting a mining effect. While not conclusive, the observation is consistent with the increased cease-to-flow frequency at the downstream gauge LA2S1.

In summary a mining effect on pool outflow is apparent at Pools 24 and 25 and likely at Pool 5. A mining effect at Pool 34 is not clear based on available data.

5.8.2 Native Dog Creek tributary ND1

ND1 is a second-order watercourse that flows west to join Native Dog Creek below the FSL of Lake Avon. When the water level of Lake Avon is above ~318 m AHD, tributary ND1 enters the lake directly. Longwall 18 commenced at a distance of 265 m from ND1_Pool2 and mined directly beneath the upper reaches of ND1C, including ND1C_Pool2 in early April 2022.

Five pools are routinely monitored along the ND1 watercourse with a further three pools monitored on upstream first-order tributaries ND1A, ND1B and ND1C (Figure 30). The data show no evidence for changes to flow status in pools along the main second-order watercourse following Longwall 18. ND1C_Pool2 was reported as dry following the passage of Longwall 18, in contrast to baseline conditions when the pool frequently flowed and is likely a mining effect.

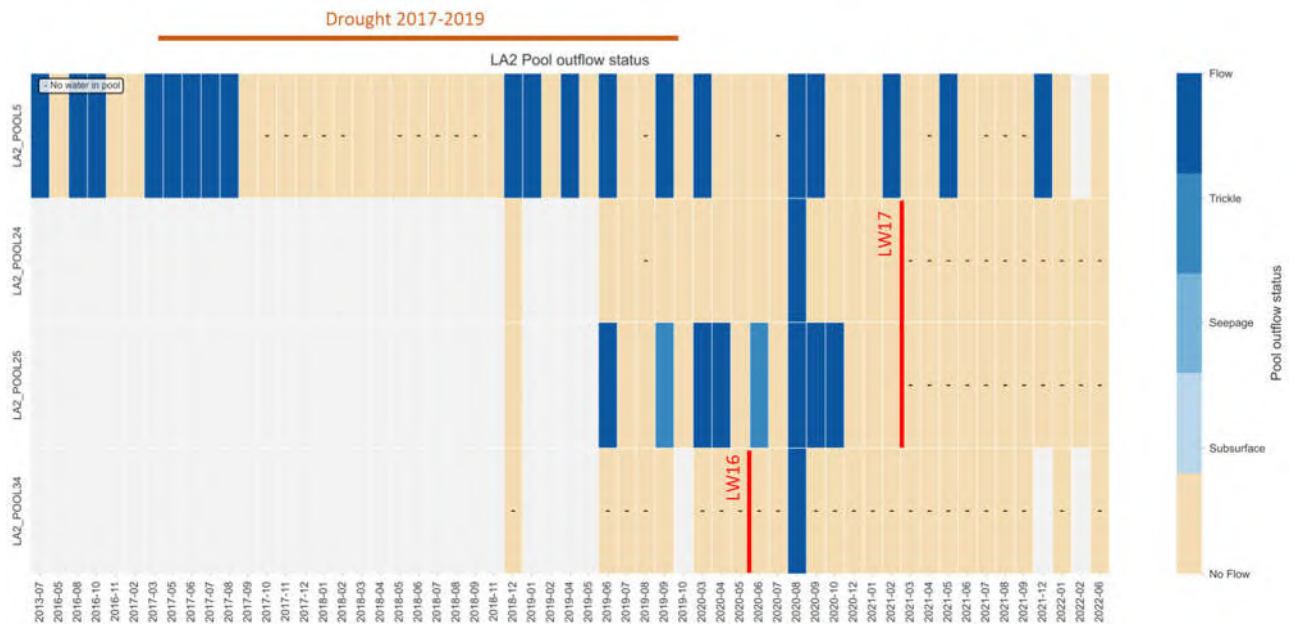


Figure 29. Flow status of pools on the LA2 watercourse

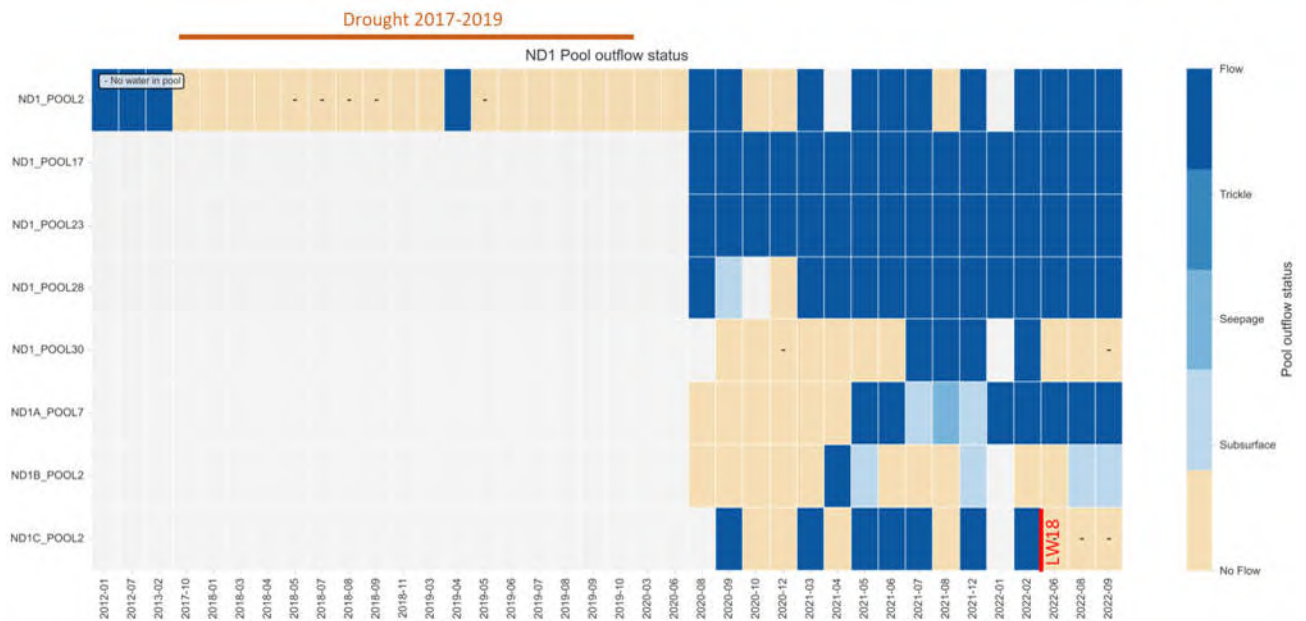


Figure 30. Flow status of pools on the ND1 watercourse

5.8.3 Wongawilli Creek Waterfall 54

There was a requirement for an adaptive management approach to the extraction of Longwall 17, with respect to subsidence and hydrology of WF54 on Wongawilli Creek. Assessments were agreed as part of the Wongawilli Creek and Waterfall 54 Management Strategy (Illawarra Metallurgical Coal, 2021). During the extraction of Longwall 17, frequent analysis and assessment of hydrology at the waterfall was reported to DPIE (Watershed HydroGeo, 2021a).

Further analysis has been commenced following identification of a rockfall (as reported to agencies), and will be reported in full in a separate Impact Report (HCEO, [in-prep]). That analysis uses the same assessment methods to analyse hydrological behaviour to present day (320 days after the end of Longwall 17).

An Orpheus sensor and logger measure water levels at the WF54 pool (above the waterfall) at five-minute intervals, and this has been recorded since 25/03/2020. The same type of equipment is used to measure pool levels upstream at WWU at the same interval (since 23/10/2020). This data is compared in a form of before-after-control-impact (BACI) assessment. Data is also recorded at WWU using a Diver sensor and logger, and that record is longer, but the Diver is not as accurate as the Orpheus, so the data from the Orpheus equipment is relied on. Manual water level measurements are taken regularly by IMCEFT staff at both sites to allow conversion from a measured water depth above each sensor to a mAHD, and a water level above the benchmark, and above the cease-to-flow control. We note that there is some inaccuracy in the manual measurements used to set the datum. During the first 7 months to November 2020, the error was up to 11 mm = 0.011 m, but procedures have improved since and the residual error is approximately 3 mm = 0.003 m.

The “pre-mining” period used for this assessment is from the beginning of the common data period; 23/10/2020 to 26/07/2021. The post-mining period is from 27/07/2021 until the last data available, which for this report is 31/08/2022.

A relationship in water levels at the two sites has been developed by plotting the data against one another for the pre-mining period as a natural and logarithmic relationship (detail on a log-scale is shown in Figure 31).

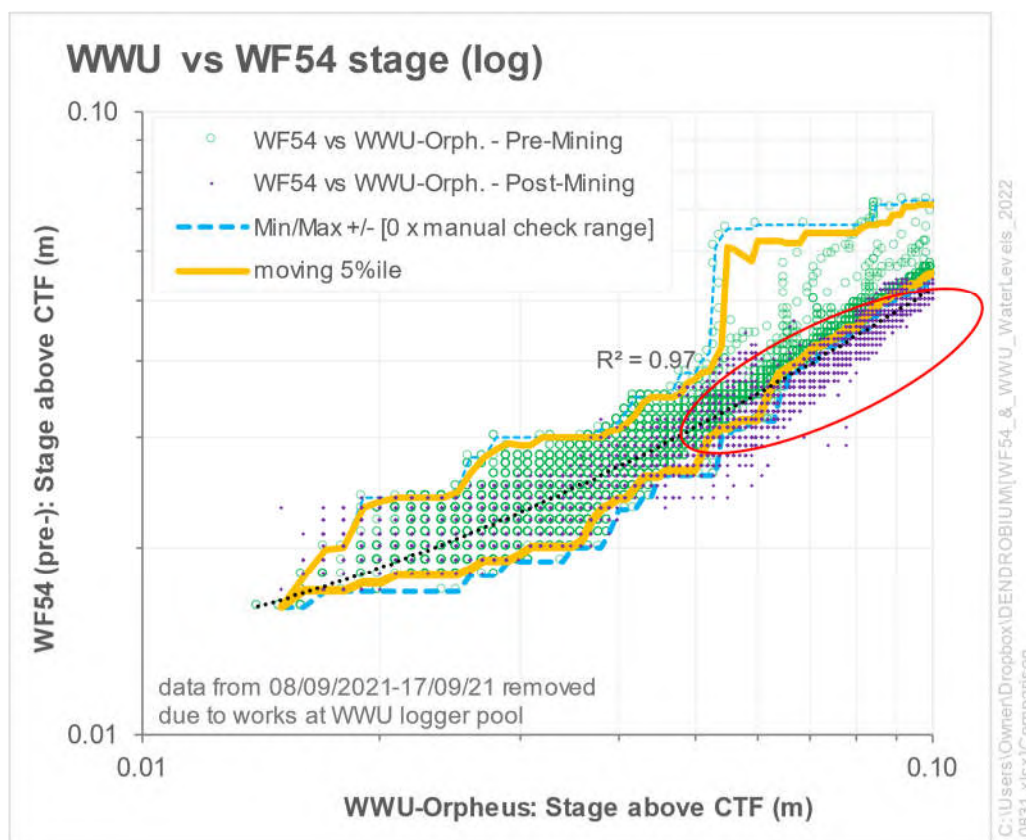


Figure 31. Relationship between water level (stage) at WWU and WF54

On Figure 31, the solid orange lines represent a moving 5%/95%ile (approximately), and the dashed blue lines show the minimum/maximum bounds on this relationship. These ranges do not consider the error in the manual measurements noted earlier. The solid orange lines are used for this analysis to

represent 'expected' pool water level. The pre-mining data falls outside the solid blue lines 3.4% of the time (2,491 of 73,813 readings).

The "post-mining" data is plotted in purple on the charts in Figure 31. Up until mid-December 2021, the post-mining data is in good agreement with the pre-mining relationship, i.e. the purple dots cluster between the orange lines (5%-95%iles). The purple dots fall outside the solid blue lines 3.17% of the time (1,243 out of 39,257 readings). A period of anomalous data was noted in previous reporting (where works were carried out at one of the monitoring site resulting a period of spurious data), and if this is removed, then the period of time that the post-mining data falls outside the expected range is 2.99%. In both cases (3.17% or 2.99%), the frequency of 'extreme' values in the post-mining data is slightly lower than in the pre-mining period (3.4%), suggesting no change in the relationship between WWU and WF54 until December-2021.

However, from that time more of the post-mining observations have fallen outside the (5%-95%iles), circled in red on Figure 31.

An alternative method of presenting the data above is in a timeseries. The upper and lower bounds of the "expected" WF54 stage are calculated from the upstream WWU stage, using the scatter or noise in the pre-mining data (i.e. the 5%-95%ile plot on Figure 31). This timeseries of the range in expected WF54 stage is shown on Figure 32. This shows the pre-mining data (green) is generally within the orange lines but occasionally strays to or outside the orange line as expected.

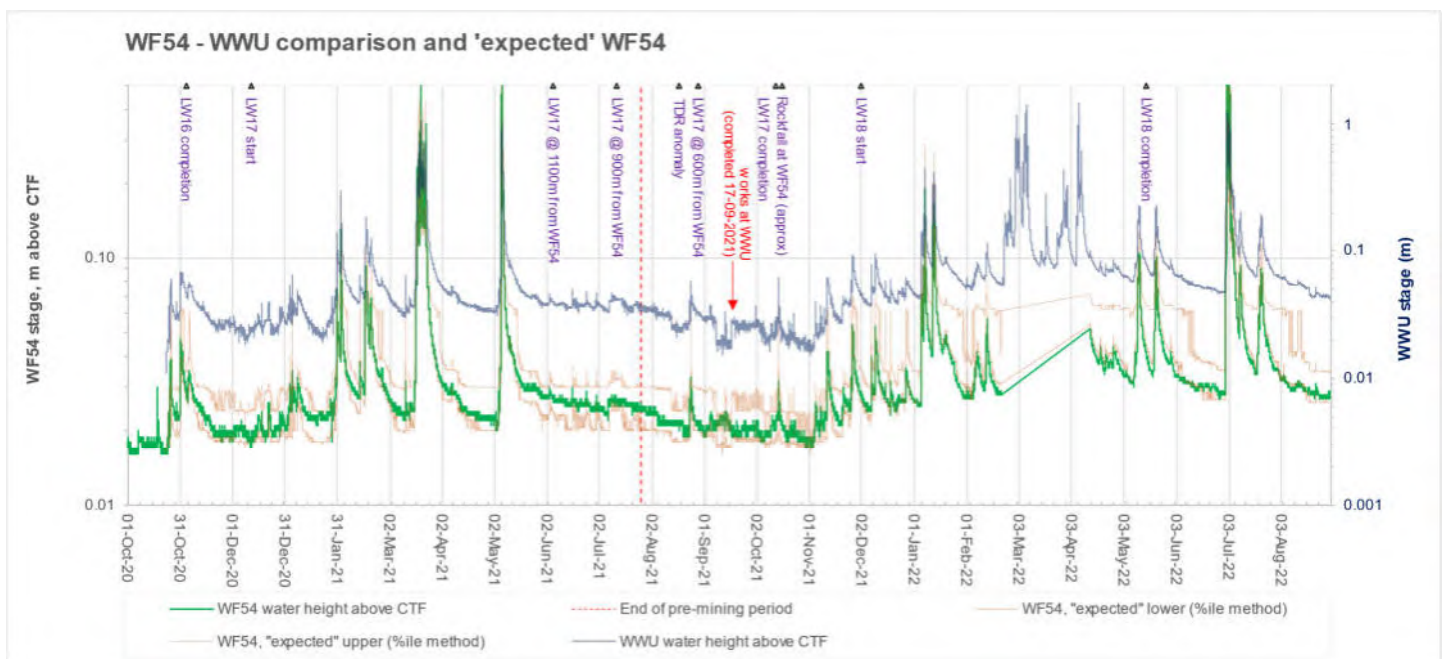


Figure 32. Comparison of WF54 stage with range in "expected" stage

Until mid-December 2021, the post-mining data tracked within the expected range (orange series), although toward the lower bound of expected stage. From that time, the green line has fallen below the expected relationship with greater frequency. This suggests that since December 2021, there has been some change in the relationship between WWU and WF54. Specifically, there is an apparent increase in recession rate at moderate stage heights, but not a reduction at the lowest flow periods (e.g. low flows in mid-June-2022 and late Aug-2022 do not show this effect).

The apparent change does not appear consistent with the conceptual model of far-field losses that we might expect as a result of Longwall 17 extraction. Further assessment will be carried out to determine whether the observed changes are related to mining, or whether they may be (wholly or partly) due to other factors related to very wet conditions in the last 10 months, such stream bed scour or sedimentation. This will be reported on later in October 2022.

6. Assessment of shallow groundwater (swamps)

6.1 Shallow groundwater levels

Trigger values for subsidence-induced decreases in groundwater levels, at surface and near-surface monitoring sites at Area 3B swamps, have been established within the Swamp Impact Monitoring Management and Contingency Plan (SIMMCP) for Area 3B (South32, 2020b). Shallow groundwater level has been identified as an indicator of potential changes in ecosystem functionality of the swamps. TARPS are defined as follows:

Table 29. Performance criteria related to shallow groundwater levels at swamp monitoring sites

| TARP Level | Criteria | Response |
|------------|---|---|
| 1 | Groundwater level lower than baseline level at any monitoring site within a swamp (in comparison to reference swamps); and/or; Rate of groundwater level reduction exceeds rate of groundwater level reduction during baseline period at any monitoring site (measured as average mm/ day during the recession curve). | Increased intensity and frequency of vegetation monitoring and/or further investigations of subsidence impacts on bedrock base and rockbars |
| 2 | Groundwater level lower than baseline level at 50% of monitoring sites (within 400 m of mining) within a swamp (in comparison to reference swamps); and/or Rate of groundwater level reduction exceeds rate of groundwater level reduction during baseline period at 50% of monitoring sites (within 400 m of mining) within the swamp. | |
| 3 | Groundwater level lower than baseline level at >80% of monitoring sites (within 400m of mining) within a swamp (in comparison to reference swamps); and/or Rate of groundwater level reduction exceeds rate of groundwater level reduction during baseline period at >80% of monitoring sites (within 400 m of mining) within the swamp. | |

Groundwater level hydrographs for each shallow piezometer are presented in **Appendix D**. The hydrograph is plotted together with ground elevation and the elevation of the piezometer base, longwall timing, groundwater level recession rate (in mm/day), and the dates that longwalls pass under (if relevant) a piezometer. Assessment of mining effects is based on these hydrographs.

A summary of hydrograph responses and cumulative effects at Area 3B swamps is included in Table 30 for Impact Sites and Table 31 for Reference Sites. In accordance with the definition of the TARPs, the sites within 400 m of mining *and* within the mapped swamp areas are assessed for triggers related to mining impacts.

An overview of shallow groundwater levels and cumulative effects is shown in Figure 33 and Figure 34 as the monthly median % saturation at each reference and impact swamp piezometer. The % saturation is calculated as the level of groundwater within the swamp piezometer relative to the total thickness of the sediments at that location (from base of the piezometer to the ground surface).

6.1.1 Reference swamp sites

IMC maintains shallow groundwater monitoring sites at reference swamps located well outside the mining zone of influence. Those sites provide an important comparison when assessing swamp sites closer to the mine for possible shallow groundwater impacts. Shallow groundwater conditions at reference sites during the assessment period are summarised in Table 31. Shallow groundwater at all

reference sites recovered after the 2017-2019 drought as a result of higher-than-average rainfall from 2020 to 2022.

A review of shallow groundwater hydrographs for reference swamps in Appendix D (and evident in Figure 33) indicate two main hydrological end-member types:

1. Near-continuously saturated swamp sediments. Examples include Swamps 7, 22 and 25. Swamp sediments at these locations remain saturated during periods of prolonged drought. It is assumed that at these locations, groundwater levels within the swamp are sustained by discharge from adjacent and underlying sandstone substrate (groundwater-connected swamps).
2. Intermittently saturated swamp sediments. Examples include Swamps 33, 84, 85, 86 and 88. Swamp sediments at these locations saturate, typically to the ground surface, following large rainfall events and remain saturated for several weeks to months as shallow groundwater levels recede to below the base of the swamp. The duration of saturation and rate of recession vary between locations and likely depend on the characteristics of the swamp substrate, controlling rock-bar and contributions from adjacent or up-gradient perched sandstone aquifers. It is assumed that at these locations, the swamp sediments are likely perched above the water table in the sandstone substrate.

Continuously saturated locations tend to be within deep valleys (valley-fill) where adjacent ridges rise ≥ 50 m above the swamp level. Intermittently saturated swamp locations tend to reside in shallow valleys where the adjacent ridges rise ≤ 20 m above the swamp level (typical of headwater swamps).

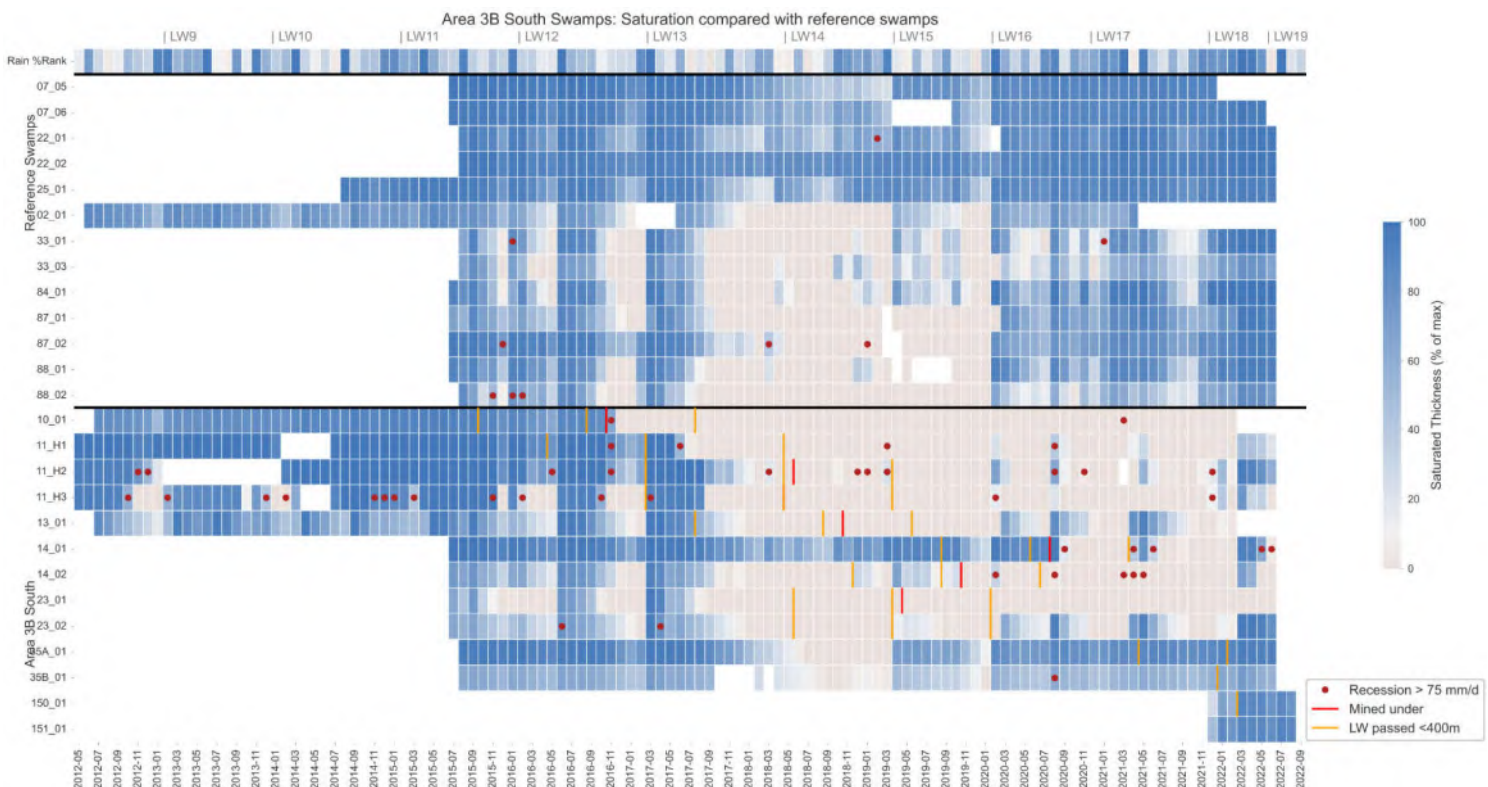


Figure 33. Overview of swamp saturation levels by month, Area 3B (south)

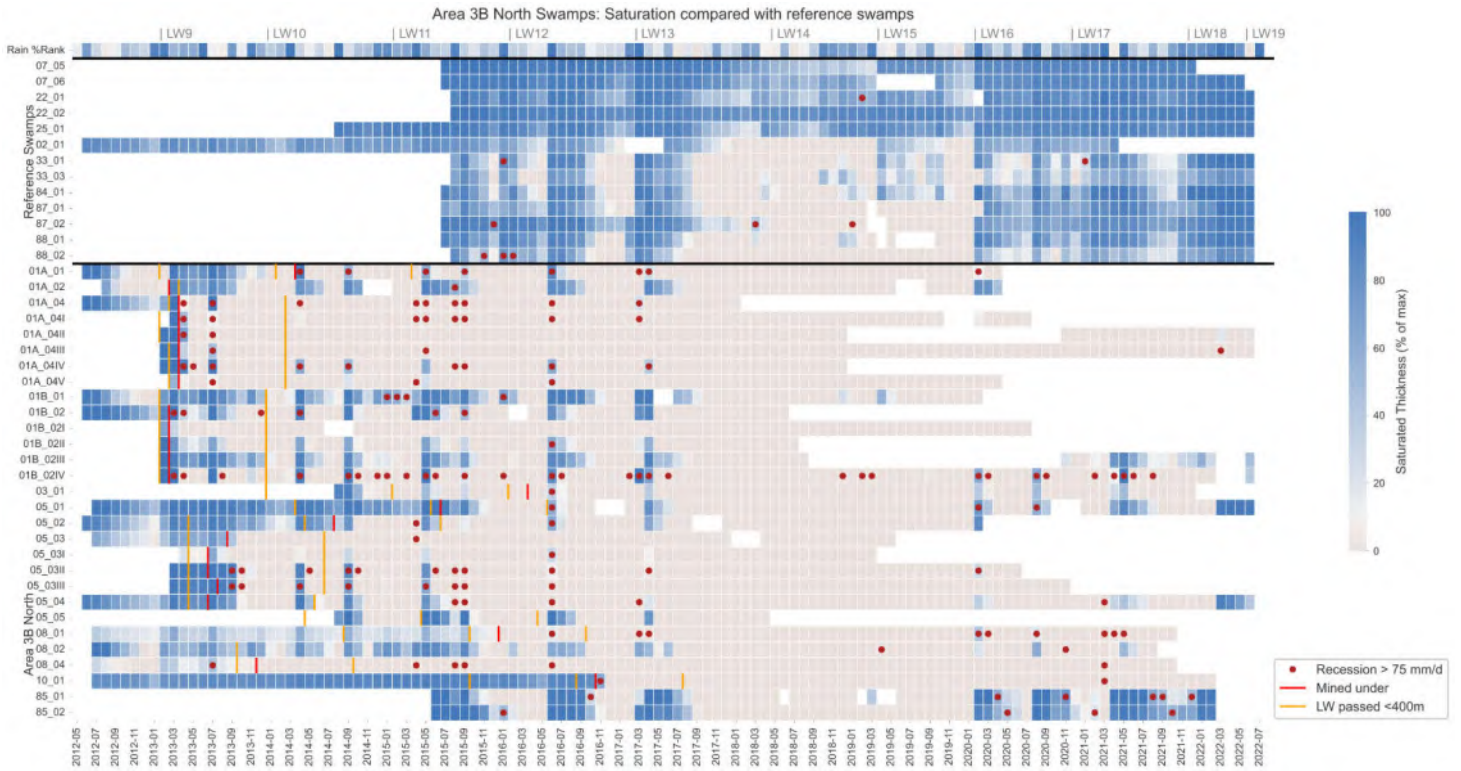


Figure 34. Overview of swamp saturation levels by month, Area 3B (north)

6.1.2 Impact swamp sites

Longwall 18 passed beneath or within 400 m of:

- **Swamp 14:** Longwall 18 was completed on 17/5/2022 at a minimum distance of 265 m from mapped swamp vegetation at Swamp 14. The swamp was previously directly mined beneath by Longwall 17 in June 2021, and by Longwall 16 in August 2020 (including piezometer 14_01). The northern extension of the swamp (including S14_S02) was mined under by Longwall 15 in November/December 2019.
- **Swamp 149:** Longwall 18 passed adjacent to Swamp 149 in February 2022. Previously, Longwall 17 passed directly beneath the swamp in March 2021. There are no piezometers or soil moisture sensors at this swamp.
- **Swamp 35a:** Longwall 18 mined beneath the northern fringe of Swamp 35a between 17/4/2022 and 17/5/2022. Longwall 17 passed within 400m of the swamp in mid-2021. Piezometer 35a_01 is within 10 m of Longwall 18 footprint.
- **Swamp 35b:** Longwall 18 passed the swamp in March 2022 at a closest distance of 90 m. Piezometer 35B_01 is located approximately 115 m from the footprint of Longwall 18.
- **Swamp 150:** Located on the southern side of the ND1 valley. Longwall 18 passed 232 m from the swamp at its closest point. Piezometer 150_01 is located 285 m from Longwall 18.
- **Swamp 151:** Located on the southern side of the ND1 valley. Longwall 18 passed 330 m from the swamp at its closest point. Piezometer 151_01 is located 440 m from Longwall 18.

Observed effects on shallow groundwater levels within these swamps is described below.

6.1.2.1 Swamp 14

Swamp 14 consists of ~5.9 ha of upland swamp vegetation, mostly Teatree and Banksia thicket and Cyperoid heath (sedgeland) surrounding the upper reaches of tributary WC15. The swamp vegetation is discontinuous along WC15 and consists of two main areas with a shallow piezometer located within each area.

Hydrographs for both swamp piezometers show evidence for shallow groundwater impact as a result of being directly mined under by Longwalls 15 to 17 (Figure 35). The effects were reported in previous EoP assessments and included an increase in groundwater recession rates and decreases in durations of saturation following longwall passage, compared with baseline conditions.

For the period of Longwall 18, both hydrographs show saturation of swamp sediments to levels similar to the baseline following high rainfall; however, recession rates remain higher than the baseline. The effects represent a Level 3 TARP trigger at Swamp 14 which remains unchanged in this assessment.

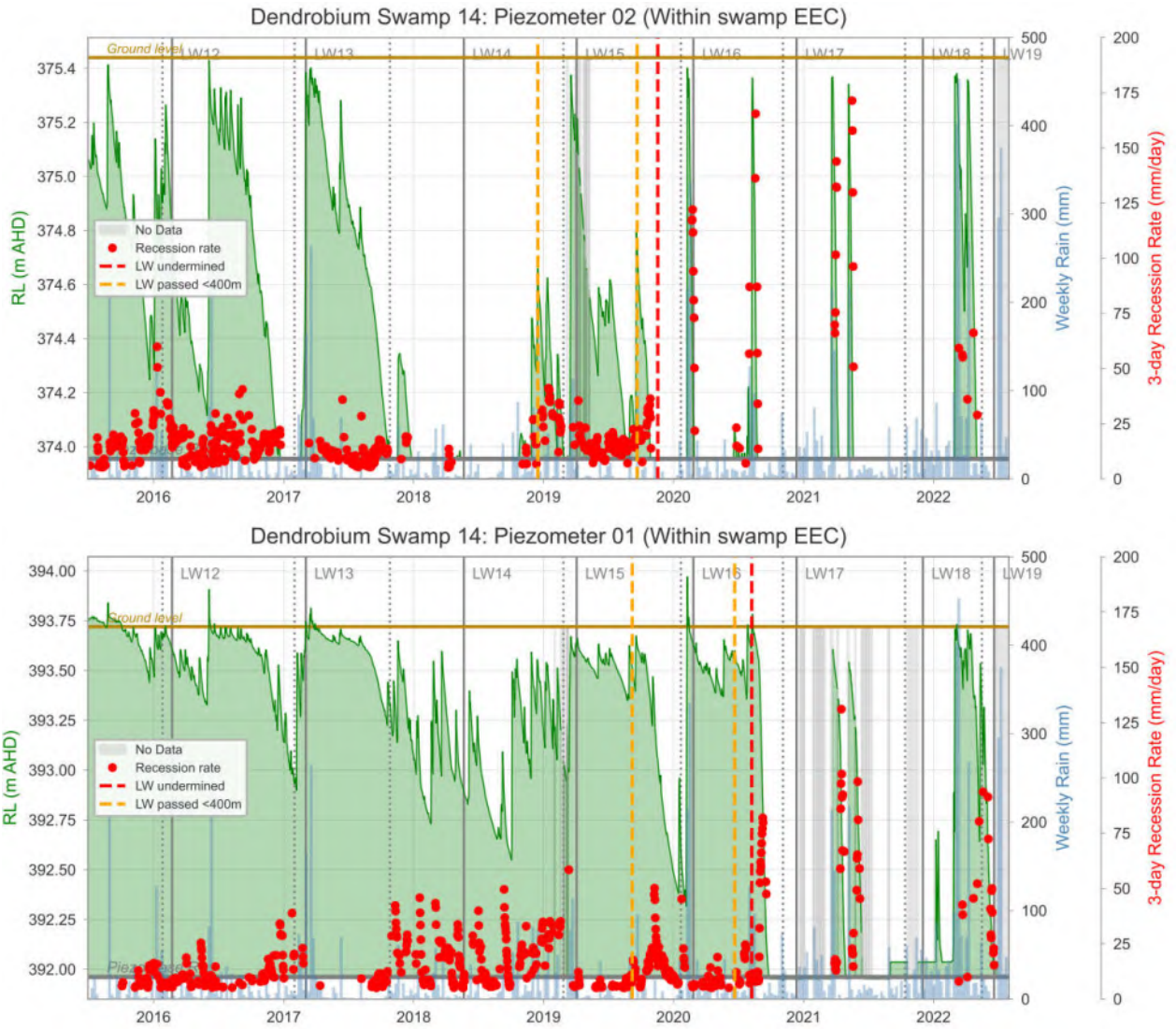


Figure 35. Shallow groundwater hydrographs for Swamp 14

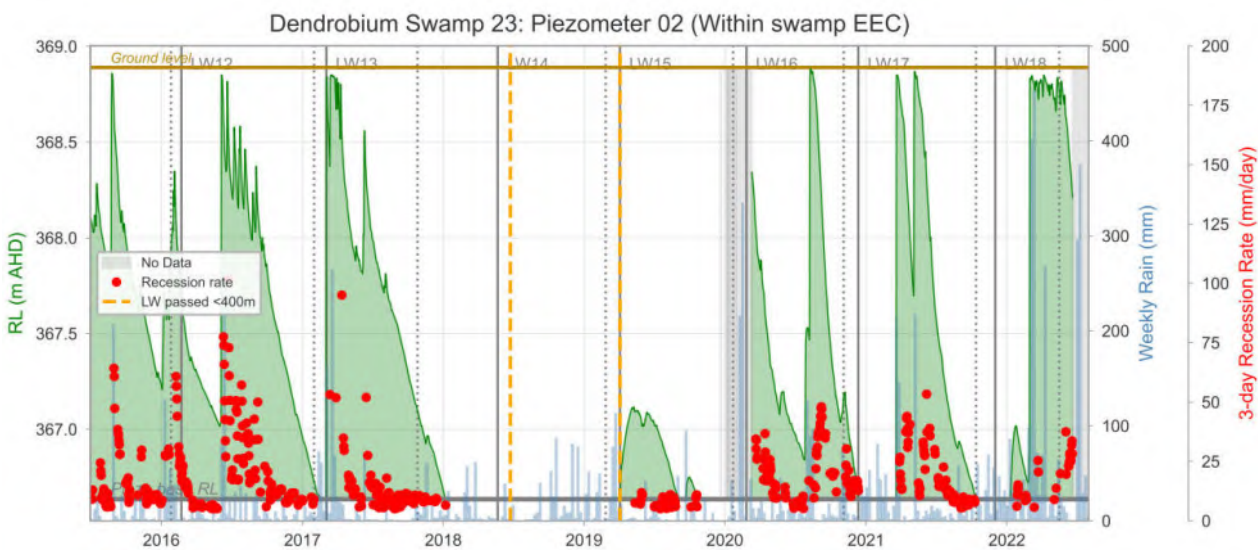


Figure 36. Shallow groundwater hydrograph for Swamp 23, piezometer 02

Swamp 35a

Swamp 35a consists of 1.2 Ha of swamp vegetation along the upper reaches of tributary ND1. The swamp is located on the southern edge of Longwall 18. Longwall 17 passed within 225 m of the swamp in May 2021. One shallow piezometer (35a_01) intersects 2.18 m of swamp sediment ~20 m north of the tributary channel. A hydrograph for piezometer 35a_01 is shown in Figure 37, below. The hydrograph shows that the swamp sediments are saturated most of the time, becoming unsaturated for several months during the 2017-2019 drought. It is likely that the swamp is connected with, and sustained by, groundwater within the underlying and adjacent sandstone substrate, except during times of extreme drought.

The hydrograph for piezometer 35A_01 shows that shallow groundwater levels continue to recover following the 2017-2019 drought. There is no evidence for shallow groundwater effects related to either Longwall 18 or 17; however, this should be reassessed in the next EoP review.

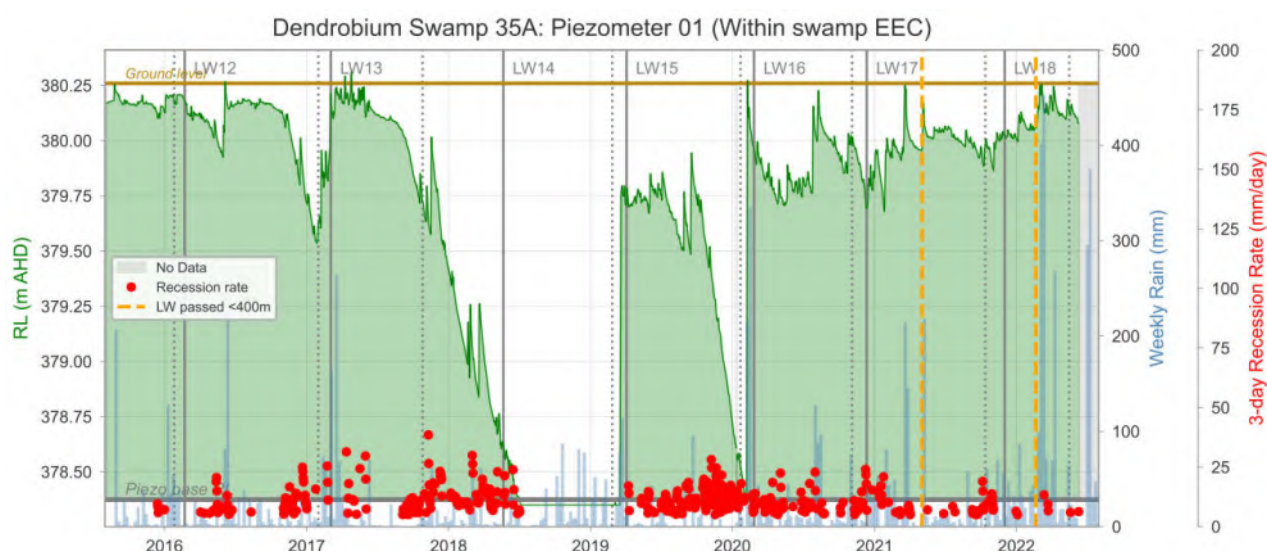


Figure 37. Shallow groundwater hydrograph for Swamp 35a, piezometer 01

Swamp 35B

Swamp 35B consists of 1.1 Ha of swamp vegetation along the middle reaches of tributary ND1. The swamp is located 90 – 140 m south of Longwall 18. One shallow piezometer (35B_01) intersects 3.1 m of swamp sediment ~25 m north of the main tributary channel. A hydrograph for piezometer 35B_01 is shown in Figure 38. The hydrograph shows that the swamp sediments are saturated most of the time, becoming unsaturated briefly in early 2020 at the end of the 2017-2019 drought. As with Swamp 35a, it is likely that the swamp is sustained by groundwater within the underlying and adjacent sandstone substrate, except during times of extreme drought.

The hydrograph for piezometer 35B_01 shows shallow groundwater recovery since 2020, following the 2017-2019 drought. There is no evidence for shallow groundwater effects related to Longwall 18.

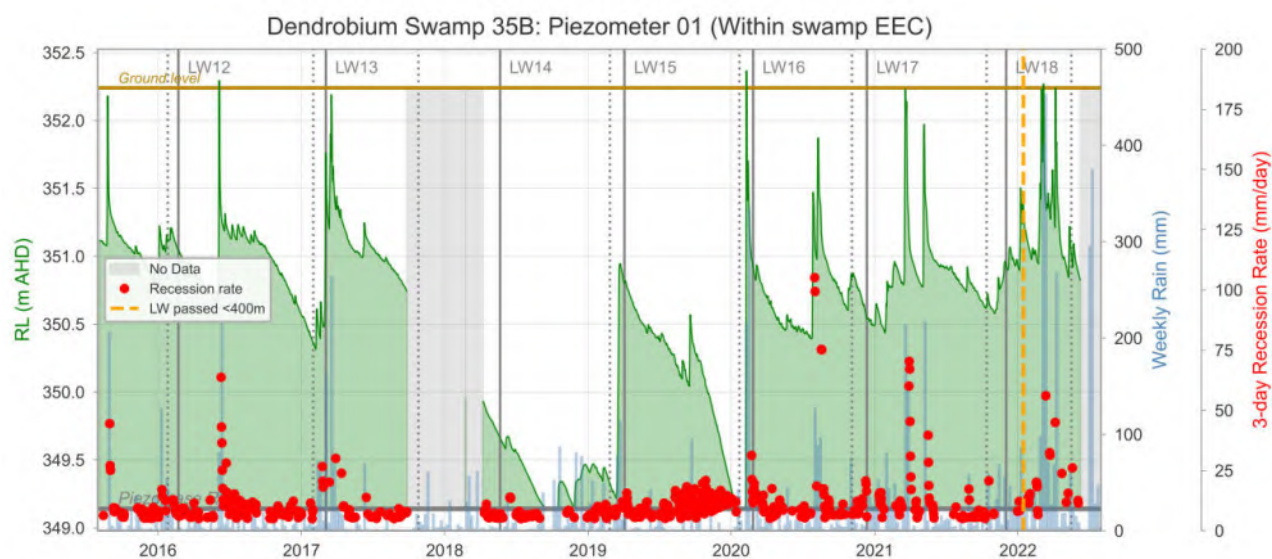


Figure 38. Shallow groundwater hydrograph for Swamp 35b, piezometer 01

Swamp 149

Swamp 149 comprises 1.4 Ha of swamp vegetation located on the mid- to upper slopes of the LA2 tributary catchment. The swamp area does not overlap with mapped tributary channels but likely contributes via seepage to baseflow of the upper reaches of the tributary. A single soil moisture sensor was installed within the mapped swamp area in 2021 and therefore no shallow groundwater data are available for this swamp. However, given its location directly above Longwall 17 it is likely that shallow groundwater levels have been altered at the swamp.

Swamps 150 and 151

These headwater swamps comprise a combined 7.2 Ha of mapped swamp vegetation. A piezometer is installed at each of the two swamp areas (150_01 and 151_01) intersecting 1.1 m and 0.9 m of swamp sediments respectively. Hydrographs for 150_01 and 151_01 are shown in Figure 39 Both hydrographs show high levels of saturation since their installation in late 2021 with no evidence for mining effects during or following Longwall 18.

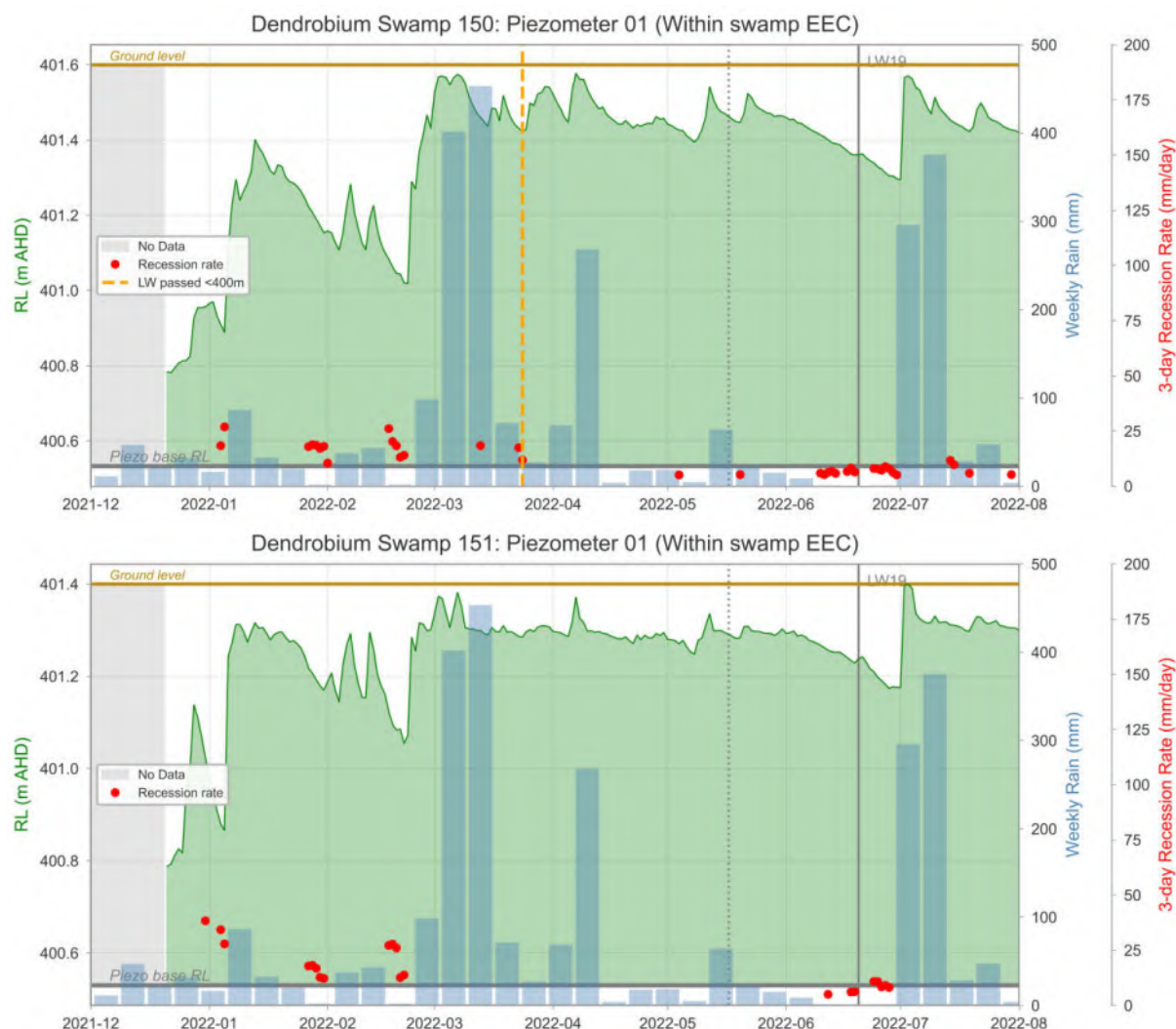


Figure 39. Shallow groundwater hydrograph for swamp piezometers 150_01 and 151_01

6.1.3 Spatial distribution of impacted swamps

Swamps that have been mined under commonly display hydrological changes shortly following the passage of the longwall beneath the monitoring site. An assessment of hydrological change at Upland Swamps was carried out at Dendrobium by Watershed (2019b) and recently updated (Watershed HydroGeo, 2021b). The study concluded that there was no evidence for hydrological change in shallow groundwater piezometers located more than 60 m from the extracted longwall margin. Although it is possible that impacts at greater distances from the mine may be caused by depressurisation along geological structures such as faults (as has been observed at the Springvale Mine in the Western Coalfield; Galvin *et al.* 2016), the Watershed assessment found no evidence for anomalous impacts associated with lineaments at Dendrobium.

Shallow groundwater responses to Longwall 18 discussed in the previous section are consistent with the Watershed review and similarly find no evidence for anomalous impacts at distances greater than 60 m attributable to lineaments.

Table 30. Summary of cumulative shallow groundwater effects and TARP status at *Impact Sites*

| SWAMP | TARP SITES | RELEVANT LONGWALLS | PIEZOMETERS WITH AN OBSERVED RESPONSE | | | OBSERVED BEHAVIOUR | COMMENT | TARP LEVEL |
|-------|------------|--------------------------------------|---------------------------------------|------------|----|---|--|------------|
| | | | YES | UNCLEAR | NO | | | |
| 01a | 6 | Longwall 9, Longwall 10 | 01, 04, 04i, 04ii, 04iii, 04iv, 04v | | 02 | Groundwater levels lower than baseline and recession rate greater than baseline at greater than 50% to 90% of monitoring sites | Limited baseline data for five piezometers. | Level 3 |
| 01b | 5 | Longwall 9 | 02, 02iii | 02ii, 02iv | 01 | Groundwater levels lower than baseline and recession rate greater than baseline at greater than 50% of monitoring sites. | Limited baseline data for five piezometers | Level 2 |
| 03 | 1 | Pillar 11/12 | 01 | | | Possible increase in recession rate and apparently reduced response to rainfall after Longwall 11 passed and Longwall 12 undermined. | Rapid recession after rain during Longwall 13 supports impact at Swamp 3 | Level 3 |
| 05 | 6 | Longwall 9, Longwall 10, Longwall 11 | 01, 02, 03, 03ii, 04 | 05 | | Groundwater levels lower than baseline and recession rate greater than baseline at >80% of monitoring sites | Unclear if piezometer 5_05 impacted by either Longwall 11 or 12 due to limited baseline. | Level 3 |
| 08 | 0 | Longwall 9, Longwall 10 Longwall 11 | 01, 04, 02 | | | Groundwater levels lower than baseline and recession rate greater than baseline at a number of piezometers, not within swamp boundary. | Outside swamp boundary (Not subject to TARP) | n/a |
| 10 | 1 | Longwall 12 | 01 | | | Sharp decline in groundwater levels below base of the piezometer after Longwall 12. Level and rate of decline anomalous compared with baseline. | Mined under by Longwall 12 | Level 3 |
| 11 | 3 | Longwalls 13-14 | H1, H2, H3 | | | All three piezometers show mostly desaturated conditions following the passage of Longwall 14 with only brief periods of saturation following rainfall events. | Partially mined under by Longwall 13 and by Longwall 14 | Level 3 |
| 13 | 1 | Longwalls 13-14 | 01 | | | Groundwater level below the piezometer base since early 2018; Impact apparent as of Longwall 15. Swamp re-saturated 2020-2021 but not to the same level as previously. | Partially mined under by Longwall 13 and by Longwall 14 | Level 3 |
| 14 | 2 | Longwalls 15-18 | 01, 02 | | | Evidence for impact to swamp groundwater levels at 14_01 and 14_02 following Longwalls 16 and 15 respectively. Effects confirmed in post-Longwall 17 assessment. No further effects related to Longwall 18. | Partially mined under by Longwalls 15, 16 and 17 | Level 3 |
| 23 | 2 | Longwalls 15-17 | 01, 02 | | | Evidence for impact to swamp groundwater levels and duration of saturation at 23_01 and 23_02, following passage of Longwalls 15 and 16. | Partially mined under by Longwall 15, passed within 400 m by Longwalls 16 and 17. | Level 3 |

| | | | | | | | |
|---------|---|------------------|--|------------------|---|---|-----|
| 35a | 1 | Longwalls 17,18 | | 01 | No evidence of mining effects from Longwall 17 or 18. | Longwall 18 overlapped the northern fringes of the swamp | n/a |
| 35b | 1 | Longwall 18 | | 01 | No evidence of mining effects from Longwall 18 | Longwall 18 passed ~85 m from the swamp. | n/a |
| 149 | 0 | Longwalls 17, 18 | | - | No shallow groundwater monitoring due to shallow soil profile. Swamp likely to be affected. | Longwall 17 passed directly beneath swamp. | n/a |
| 150/151 | 2 | Longwall 18 | | 150_01 151_01 | Piezometers installed in 2021; No evidence of mining effects from Longwall 18. | Longwall 18 passed within 285 m of Swamp 150_01 and 440m of Swamp 151_01. | n/a |

Note: "i" in site name (e.g. 04i) indicates installation during Longwall 9 extraction. * at these swamps which are located away from active or recent mining areas the data has been logged (recorded) at the piezometer, but not collected since that time.

Table 31. Summary of shallow groundwater level trends at Reference Sites.

| SWAMP | NUMBER OF PIEZOMETERS | PIEZOMETERS | PROXIMITY TO LONGWALLS | COMMENT | OBSERVATIONS |
|-------|-----------------------|-------------|--|------------------------|---|
| 02 | 1 | 01 | 900 m north of Longwall 9 | 900 m from Longwall 9 | Swamp permanently saturated 2013-2017. Water level below piezometer for >60% of time between 2017 and 2019; saturated throughout 2020-2021 with mean levels slightly below pre-2016. |
| 07 | 2 | 05, 06 | 1.2 km from Longwall 6 | 1.2 km from Longwall 6 | Water level above base of piezometers for entire record including dry period between 2017 and 2019. Saturated throughout 2020-2022 with mean levels recovering towards pre-2016 levels. |
| 15a | 3 | 06, 07 | 0.5 km south of Longwall 8, 130 m south of Longwall 19 | 0.5 km from Longwall 8 | n/a |
| 22 | 2 | 01, 02 | Elouera Colliery | Limited baseline data | Water level above base of piezometers for entire record including during the 2017-2019 drought. Saturated throughout 2020 - 2022. |
| 25 | 1 | 01 | 1.4 km from Longwall 5 | 1.4 km from Longwall 5 | Water level above base of piezometers for entire record including during the 2017-2019 drought. Data gap 2012-2014. Saturated 2020-2022. |
| 33 | 2 | 01, 03 | 1 km from Longwall 16 | 1 km from Longwall 16 | Transient but short duration peaks in water level following rain. Shorter duration peaks during 2018-2019 dry period. Saturated most of the time between 2020 and 2022 |
| 84 | 1 | 01 | 500 m from Longwall 5 | 500 m from Longwall 5 | Swamp 84 has typically been saturated 80-90% of time, except for the 2017-2019 dry period (<50%). Saturated throughout 2020 - 2022. |

| | | | | | |
|----|---|--------|-----------------------|-----------------------|--|
| 85 | 2 | 01, 02 | 900 m from Longwall 9 | 900 m from Longwall 9 | Piezometers record saturation of swamp sediments for several months after heavy rainfall events. Swamp unsaturated late 2017 to early 2020 Saturated most of 2020 to 2022 |
| 86 | 2 | 01, 02 | 3 km from Longwall 9 | 3 km from Longwall 9 | Piezometer 01 has similar characteristics to swamp 85. Piezometer 02 typically records saturated conditions except during 2019. Returned to saturated conditions throughout 2020 - 2022. |
| 87 | 2 | 01, 02 | Avon Colliery | Limited baseline data | Piezometer 01 saturated >90% of time and piezometer 02 100% of time prior to 2018. Piezometers dry most of 2018 to 2020. Saturated throughout 2020 - 2022. |
| 88 | 2 | 2 | Huntley Colliery | Limited baseline data | Both piezometers record WL above base for >80% of time prior to 2017. Both piezometers mostly dry 2017 – 2019. Mostly saturated in 2020-2022. |

6.2 Soil moisture

Significant changes in soil moisture characteristics compared with baseline monitoring is identified as an indicator of potential changes in ecosystem functionality of the swamps. Response trigger conditions related to soil moisture at swamp monitoring sites are listed in the SIMMCP (South32, 2020c), and reproduced in Table 32.

Table 32. TARP trigger conditions related to soil moisture at swamp monitoring sites

| TARP Level | Trigger conditions | Response |
|------------|--|---|
| 1 | Soil moisture level lower than baseline level at any monitoring sites (within 400 m of mining) within a swamp (in comparison to reference swamps). | Increased intensity and frequency of vegetation monitoring and/or further investigations of subsidence impacts on bedrock base and rockbars |
| 2 | Soil moisture level lower than baseline level at 50% of monitoring sites (within 400 m of mining) within a swamp (in comparison to reference swamps). | |
| 3 | Soil moisture level lower than baseline level at >80% of monitoring sites (within 400 m of mining) within a swamp (in comparison to reference swamps). | |

The TARP has been assessed by comparing the average moisture content of the soil profile during the longwall assessment period against that of the baseline period. If the average soil moisture level drops below the minimum level recorded during the baseline period, a TARP is triggered. The TARP level increases according to the proportion of monitoring sites that exceed this criterion at each swamp within the area of mine influence (Table 32). This is the same approach used by the IMCEFT for regular impact reporting. The baseline period is the period of monitoring before the site is first mined under or passed within 400 m.

Soil moisture hydrographs for all active monitoring locations are presented in **Appendix E**. Assessment of soil moisture hydrographs for locations within Areas 3A and 3B zone of influence (< 400 m) are presented in Table 33

Soil moisture levels have recovered significantly at reference swamps sites and impact swamps since 2020 in response to above-average rainfall. Average soil moisture declined at all sensors during the 2017-2019 drought, complicating interpretation of longwall subsidence effects over that period. Post-drought recovery resulted in revision of TARP triggers at Swamp 13 and 23. Table 33 summarises observations and cumulative TARP triggers at Area 3B swamps.

In relation to Swamps within the zone of influence for Longwall 18:

- Swamp 35a: Soil moisture stable over the review period. No mining effects.
- Swamp 149: Most sensors stable over the review period. Minimal baseline.
- Swamp 150: Sensors >30 cm depth stable over review period. No mining effect.
- Swamp 151: Sensors >10 cm depth stable over review period. No mining effect.

In summary, no new soil moisture impacts are apparent at swamps within the zone of influence for Longwall 18. Given the high rainfall over the review period, soil moisture at swamps 149, 35a and 35b should be reassessed when more data are available (next EoP review).

Table 33. Cumulative assessment of soil moisture hydrographs in Areas 3A and 3B

| Swamp | Longwall | Sensors and TARP triggers | | | Comment | TARP Level |
|-------|----------|---------------------------|------------------------------------|--|---|------------|
| | | Not triggered | Triggered | Not within mine influence | | |
| 05 | 9-11 | | S05_S05, S05_S01, S05_S02, S05_S08 | | All four sites show soil moisture decline below baseline after LW passed; baseline <2 y). Possible recovery at S05_S08. | 3 |
| 08 | 9-11 | S08_S05 | | | Soil moisture falls below baseline after undermining. <i>Not within mapped swamp boundary.</i> | n/a |
| 11 | 13,14 | | S11_S01, S11_S02, S11_S05 | | Soil moisture at all sensors dropped to lowest levels following LW13 and LW14. Likely mining effect, exacerbated by dry conditions. Some recovery in 2021. | 3 |
| 13 | 13,14 | S13_S03 | S13_S01, S13_S02, | | Soil moisture at all sensors dropped to lowest levels during 2017-2019 drought. Apparent recovery in 2020 and 2021 at S13_S03. Other sensors record lower moisture levels than baseline. | 2 |
| 14 | 15-17 | | S14_S01, S14_S02 | | Soil moisture at S14_S01 below baseline in contrast to recovery at reference swamps 22, 85 and 86. S14_S02 shows lower moisture levels and durations compared with baseline and reference swamps. | 3 |
| 15a | - | | | S15a_Piezo, S15a_S03, S15a_S01, S15a_S04, S15a_S06 | Outside Area 3A Longwalls; Soil moisture in 3 sensors dropped below baseline due to dry conditions | - |
| 23 | 15-17 | S23_S01 S23_S02 | | | No TARP trigger (previously Level 2). Both sensors show recovery in 2020 and 2021 with moisture levels varying within the baseline range. | - |
| 35a | 17,18 | S35a_S01 | | | No TARP trigger | - |
| 149 | 17,18 | S149_01 | | | Installed in 2021, insufficient baseline. No apparent effects | - |
| 150 | 17,18 | S150_01 | | | No TARP trigger | - |
| 151 | 18 | S151_01 | | | No TARP trigger | - |

Note: * Sites for which there are too few data points for a statistically valid assessment (<10)

7. Conclusions

Longwall 18 is the tenth and final panel to be extracted from Dendrobium Area 3B. Extraction of Longwall 18 commenced on 2/12/2021 and was completed on 17/5/2022. Rainfall during Longwall 18 extraction was well above average, totalling 2281 mm in the calendar year to the end of the longwall (18/5/2021 – 17/5/2022). Extremely heavy rainfall was experienced in March 2022 when 1010 mm was recorded in a single month. This follows similarly high rainfall in 2020 (1436 mm) and 2021 (1448 mm) due to sustained La Nina conditions over that period. As a result, there has been a full recovery in stream flow, shallow groundwater levels and soil moisture across all catchments since the severe drought of 2017-2019.

7.1 Effects on surface water quality

In general, stream salinity (EC) has decreased over the last three years (and last three longwalls) due to higher-than-average rainfall and significant increase in runoff compared with the preceding two years. The decreasing trend follows slightly more saline conditions at most locations during the 2017-2019 drought which resulted in low flows and evaporative concentration of salts. Similarly, DO has trended higher or remained stable over the reporting period due to high stream flows.

Anomalous water quality effects are noted in streams that have been directly mined under by previous longwalls (e.g. WC21, SC10C, LA4, DCC). Those effects include transient or persistent increases in EC, increases (or decreases) in pH and increases in dissolved metal concentrations such as Fe, Mn, Al and Zn. Iron staining in creek beds is commonly associated with watercourses that have been directly mined beneath or are within the mining area of influence. Over the last two years, new or recurrent iron staining has been noted on Wongawilli Creek, WC21, LA5 and SC10C. The observations of iron staining are likely related to recovery of groundwater levels and the reactivation of iron-rich springs near creek channels. This is supported by groundwater monitoring data.

Water quality TARPs were triggered at Lake Avon tributary site LA4_S1 for EC, pH and DO. Impacts to the LA4 watercourse, including fracturing of the creek bed and diversion of flow were reported in End of Panel Reports for Longwalls 12, 13 and 15. Dissolved metals Fe, Mn, Zn and Al increased over the same period. The water quality in Lake Avon remains unaffected. Analysis of flow-corrected trends in water quality indicate concentrations of Fe, Mn, Zn and SO₄ are slightly above baseline in Sandy Creek (Rockbar 5). IMC has initiated a longitudinal study to assess the cause of the increase.

7.2 Effects on surface water flow

Surface water flow TARPs were reviewed in 2019 in consultation with relevant government agencies and based on recommendations of the IEPMC (Watershed HydroGeo, 2019a). Key features of the updated TARPs are:

- A. A move to rely on comparison of flows recorded at relevant sub-catchment monitoring sites around the Dendrobium mining area against selected reference sites, rather than relying on rainfall-runoff modelling.
- B. Assessment of sub-catchment hydrology against a number of different indicators that are considered appropriate to identifying and quantifying potential effects on the broad hydrological behaviour within each sub-catchment, effects on cease-to-flow conditions that may be significant to ecological values, and effects on median flow which is a proxy for the water resource potential.
- C. A further assessment has been implemented to analyse the mining effects on low-flows that are known to occur along the “middle reach” of Wongawilli Creek, between Area 3A and 3B.

The results of Assessments A, B and C are summarised on Table 34.

The assessments indicate that sub-catchments in the upper part of the Donalds Castle Creek catchment (i.e. DC13S1 and DCS2) have been and continue to be affected by mining findings for DC13S1, DCS2 (both at Level 3 for all three flow assessments) are similar to those for the EoP report for Longwalls 15-17.

Lake Avon tributaries LA4, LA3 and LA2, have been affected by mining. The effects at LA2 are milder than for LA3 and LA4 (one Level 2 and one Level 3 trigger at LA2 compared to Level 3 for all assessments at the other two sites), and are recent, occurring as a result a Longwall 18.

Similarly, the flow characteristics at WC21S1 and WC15S1 within the Wongawilli Creek catchment have altered as a result of mining with these sites at Levels 2 or 3 for two out of three assessments. As with the sub-catchments above, the effects at WC21 and WC15 are similar to those for the previous End of Panel reports, with an improvement at WC21 in the cease-to-flow assessment. Despite Longwall 16 terminating within 50 m of and the end of Longwall 17 mining under WC12 respectively, no mining-related effects are discernible beyond natural variability/method accuracy, and this has persisted for the Longwall 18 assessment period.

As in recent EoP reports, analysis indicates that mild mining effects are probable at the Donalds Castle Creek downstream monitoring site (DCU). Specifically, the TARP assessments indicate that the general pattern of flow (Assessment A) and the median flows (Assessment C) do not trigger, which suggest that any mining effects or impacts on those indicators are of similar magnitude or less than natural variability. However, the new Assessment B, which examines cease-to-flow duration and frequency, indicates that the watercourse at DCU has been experiencing a mild increase in the number of cease-to-flow days compared to the Reference sites (TARP Level 1). This finding has been consistent for Longwalls 14-18.

Changes to stream flow characteristics are not evident at the downstream gauge on Wongawilli Creek Lower (WWL), despite mining-related effects being clear and significant at upstream tributaries (e.g. WC21, WC15). This suggests that some or all flow lost in headwater catchments is returned downgradient, or that upstream diversions or losses are not significant in relation to the larger catchment water balance given the natural variability and the accuracy of flow measurements. These possible reasons are even more relevant at DCU, where the losses identified in upstream sites DC13S1 and DCS2 are 40-60% of median flow at Q50. Such losses should be clearly apparent at DCU if they were transmitted downstream, but the assessment has not detected a change in median flow at Q50 beyond natural variability (i.e. variability at two Reference sites).

Analysis of available surface water flow observation records for Wongawilli Creek did not trigger TARP Assessment D for any of the months assessed during the Longwall 18 period.

7.2.1 Effects at Waterfall WF54

Comparison of pool water levels measured above WF54 and a reference site at WWU was conducted throughout the latter half of Longwall 17, as part of an Adaptive Management process. The same analysis has been repeated here to include approximately 11 months of data since the end of Longwall 17 and throughout the longwall 18 mining period.

The analysis shows that, initially (and as reported in the Longwall 17 End of Panel report), post-mining behaviour of water levels at WF54 was consistent with pre-mining record, and therefore that Longwall 17 did not have an effect (either no effect or an effect that cannot be discerned beyond natural variability). However, further analysis suggest that the relationship between WF54 pool levels and those at WWU have changed since mid-December 2021. This suggests a possible mining effect; however the manner of the effect and the distance from active longwalls means that further investigation of whether there is a change in hydrology is warranted, and this is being carried out (but not ready for incorporation in this report).

Table 34. Area 3B watercourse flow assessment summary

| Site | Water-course | Area | Date mining occurred under sub-catchment | A) Low flow Q%ile outside Reference Site Q%ile | | B) Change in cease-to-flow frequency (beyond natural) | | C) Change Q50 (beyond natural) as % of pre-mining Q50 | | | Rainfall-runoff model comparison | Comment |
|--------|----------------------|---------|--|--|---------------|---|---------------|---|----------|---------------|----------------------------------|---|
| | | | | Change % | TARP Level | Change % | TARP Level | Change ML/d | Change % | TARP Level | | |
| DC13S1 | DC13 | A3B | 09/02/2013 | 60% | L3 | 14% | L2 | -0.10 | -76% | L3 | n/a | Similar results to LW14-17. |
| DCS2 | Donalds Castle Creek | A3B | 10/07/2013 | 54% | L3 | 33% | L3 | -0.09 | -52% | L3 | n/a | Similar results to LW14-17. |
| DCU | Donalds Castle Creek | A3B | 09/02/2013 | -10% | Not triggered | 7% | L1 | 0.08 | 39% | Not triggered | Not triggered | Effects are similar to those in LW14-17. This is consistent with findings from rainfall-runoff modelling. |
| WC21S1 | WC21 | A3B | 05/10/2013 | 33% | L3 | 10% | ↓ L1 | -0.35 | -36% | L3 | n/a | Similar results to LW14-17, with improvement in (B). |
| WC15S1 | WC15 | A3B | 28/01/2017 | 35% | L3 | 21% | L2 | -0.07 | -46% | L3 | n/a | Similar results to LW15-17 |
| WC12S1 | WC12 | A3B | 18/10/2020 | -3.5% | Not triggered | -14% | Not triggered | 0.033 | 362% | Not triggered | Level 1 | Still no discernible effect. Rainfall-runoff modelling suggests a Level 1 trigger. |
| WWL | Wongawilli Creek | d/s A3B | 09/02/2010 | 1% | Not triggered | -5% | Not triggered | 0.08 | 2% | Not triggered | Not triggered | Effects are similar to those in LW14-17. This is consistent with findings from rainfall-runoff modelling. |
| WWLA | Wongawilli Creek | d/s A3B | 09/02/2010 | | | | | | | | | No pre-mining baseline record. Recommended to be assessed in future mining areas (e.g. A3C). |
| LA4S1 | LA4 | A3B | 01/04/2015 | 25% | ↑ L3 | 31% | ↑ L3 | -0.05 | -63% | L3 | n/a | Improved data availability means that effects are now more confidently quantified (likely similar to those for LW16-17) |
| LA3S1 | LA3 | A3B | 28/04/2019 | 47% | L3 | 32% | L3 | -0.05 | -352% | L3 | n/a | Effects are similar to those following LW16-17. |
| LA2S1 | LA2 | A3B | 01/03/2020 | 16% | ↑ L2 | -14% | ↓ NT | -0.014 | -693% | L3 | n/a | Change to pattern of flow due to Longwall 18, although cease to flow-frequency has improved.. |
| NDS1 | ND1 | A3B | 18/04/2021 | -11% | Not triggered | -17% | Not triggered | 0.211 | 919% | Not triggered | Level 3 | Second panel to mine beneath this sub-catchment. No discernible effect. However rainfall runoff modelling suggests a significant decline in flow from March-2022. |

Bold indicates a change from previous assessment, and **↑↓** = direction of change.

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7.2.2 Performance Measures

The four Performance Measures for surface water flow were assessed (Section 5.6) as follows:

Wongawilli Creek – minor environmental consequences

This Performance Measure is met.

Donalds Castle Creek – minor environmental consequences

This Performance Measure is met.

Lake Avon – negligible reduction in the quantity of surface water inflows to Lake Avon

This Performance Measure is met.

Cordeaux River – negligible reduction in the quantity of surface water inflow to the Cordeaux River at its confluence with Wongawilli Creek

This Performance Measure is met.

7.3 Effects on swamps

It was predicted that Swamps 01a, 01b, 03, 04, 05, 08, 10, 11, 13, 14, 23, 35a, 35b and 149 would be affected by mine subsidence due to mining in Area 3B (South32, 2020c). Longwall 18 passed beneath, or within 400 m of Swamps 149 and 35a, 35b, 150 and 151. No shallow groundwater TARPs were triggered in monitored swamps within the zone of influence of Longwall 18; however, given the high rainfall during the review period, these should be reassessed when more data are available. The cumulative TARP triggers at Swamps in Area 3A are as follows:

- Swamp 01a Level 3
- Swamp 01b Level 2
- Swamp 03 Level 3 (because the only piezometer is affected)
- Swamp 05 Level 3
- Swamp 10 Level 3 (because the only piezometer is affected)
- Swamp 11 Level 3
- Swamp 13 Level 3 (because the only piezometer is affected)
- Swamp 14 Level 3
- Swamp 23 Level 3

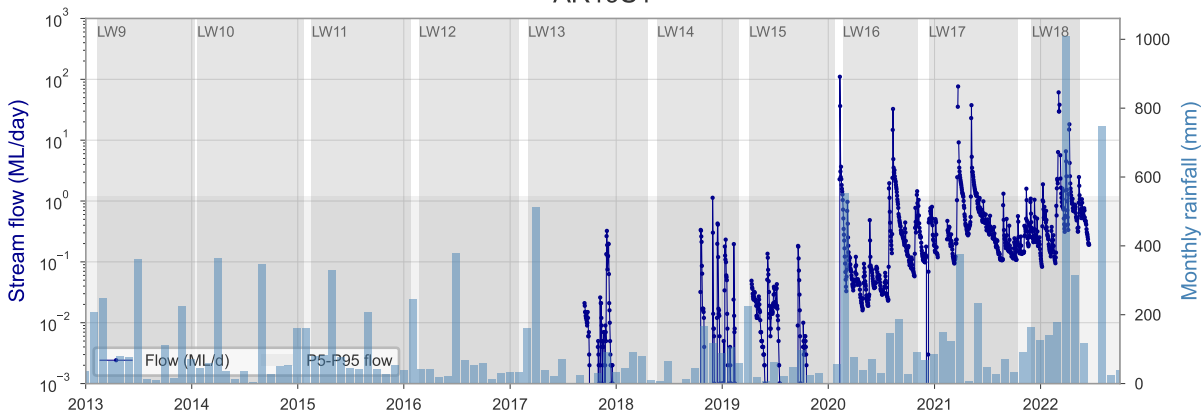
Both shallow groundwater levels and soil moisture levels in reference swamps have recovered since 2020 in response to high rainfall following the 2017-2019 drought period. No soil moisture effects are apparent at monitoring sites within the Longwall 18 zone of influence.

8. References

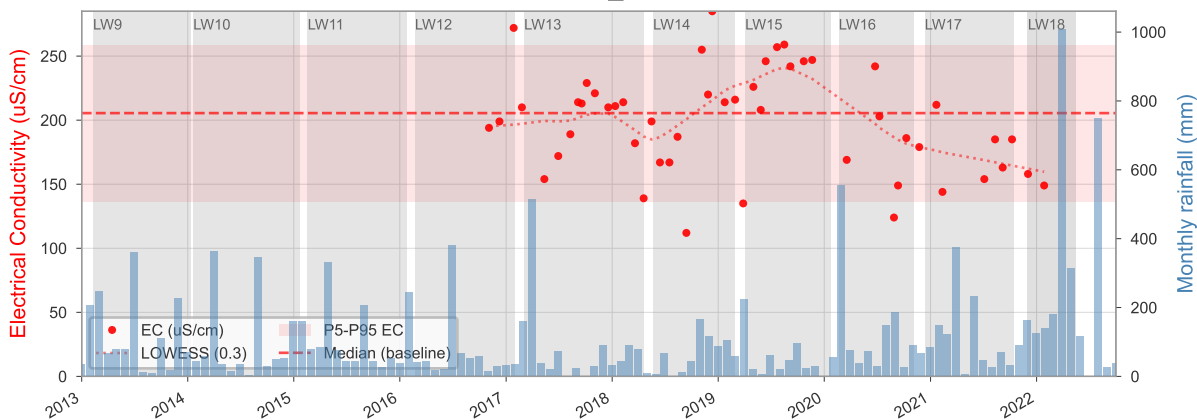
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Appendix A I: Water quality hydrographs

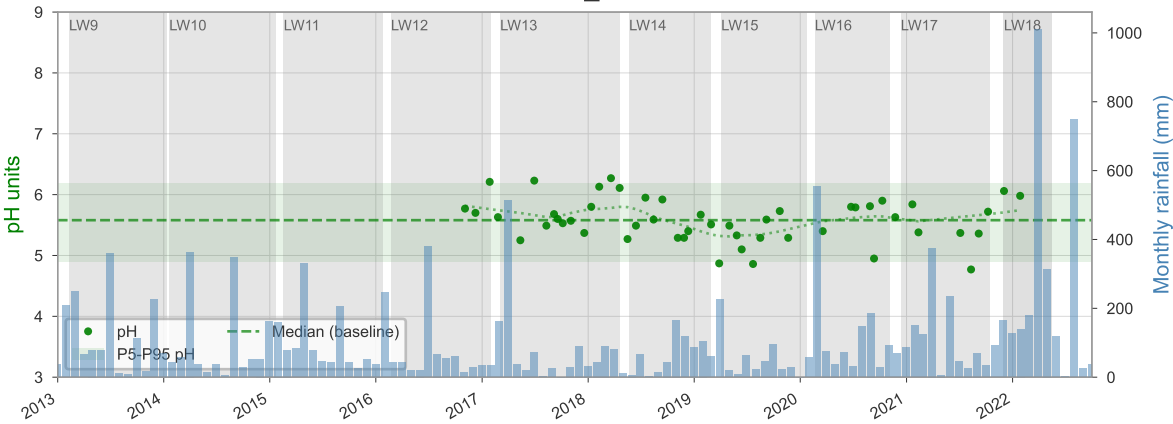
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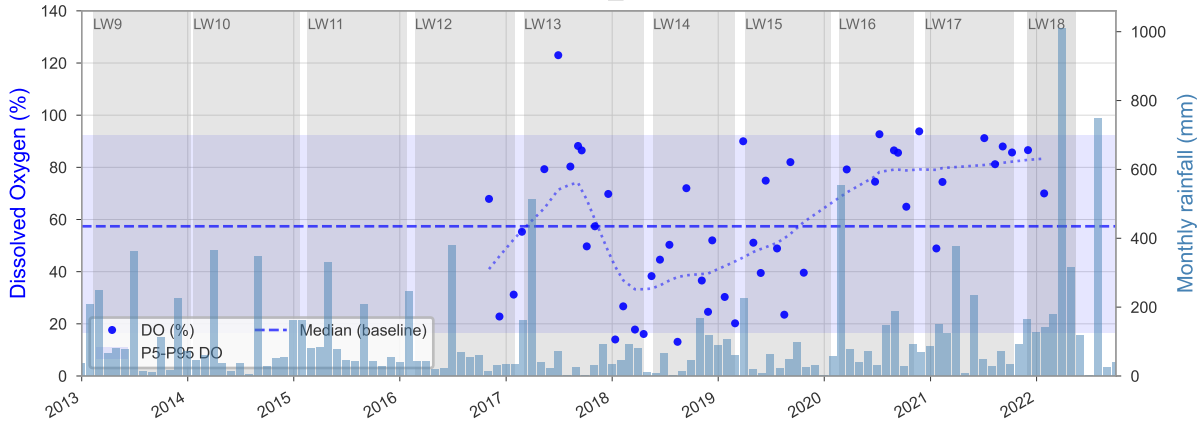
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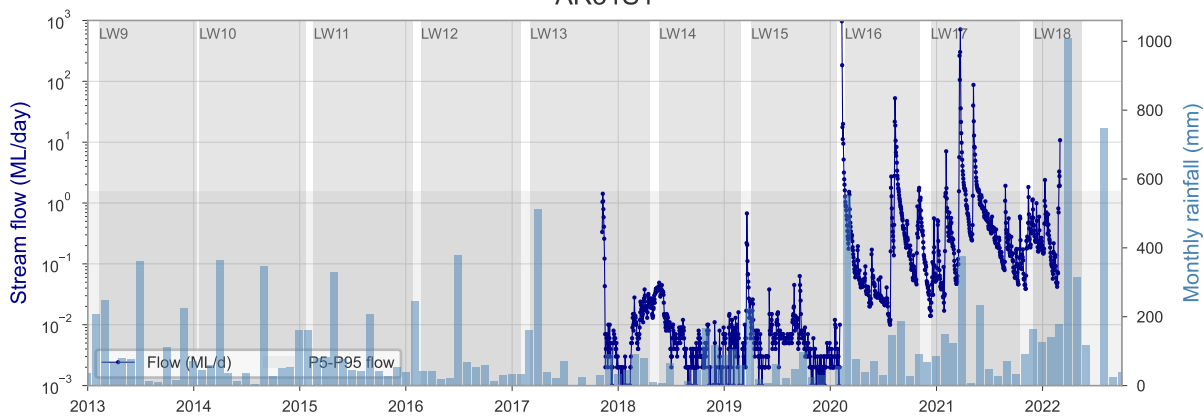
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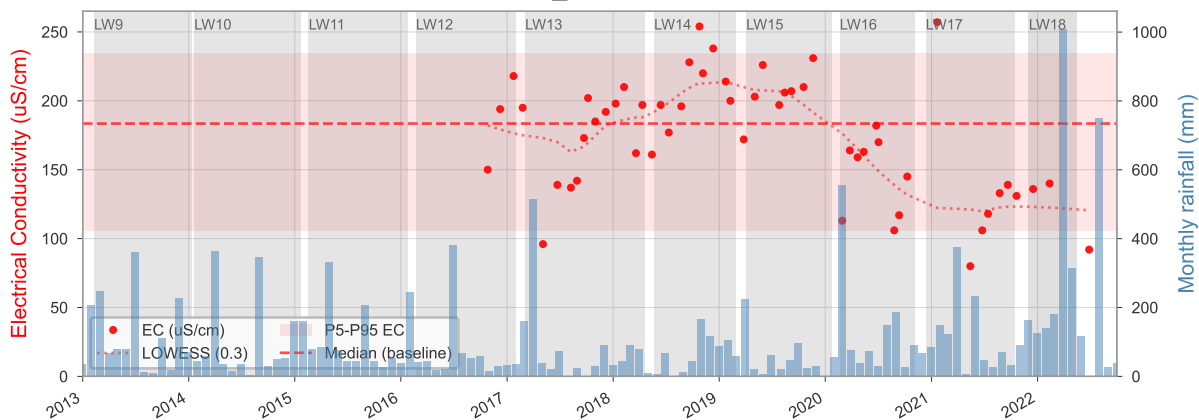
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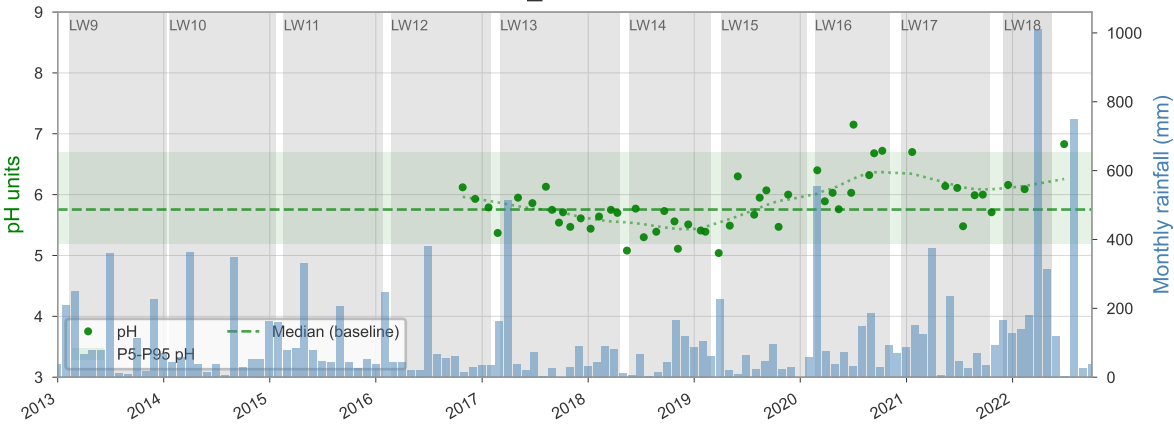
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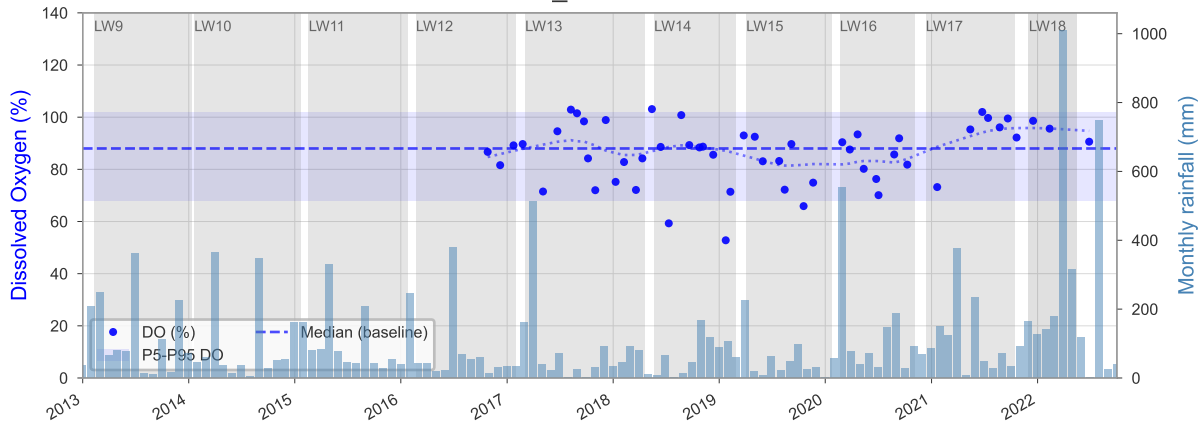
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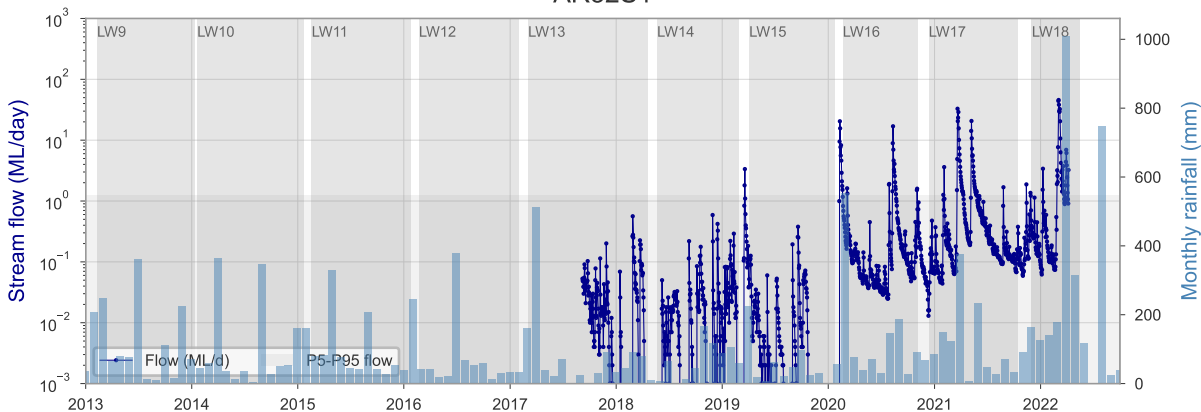
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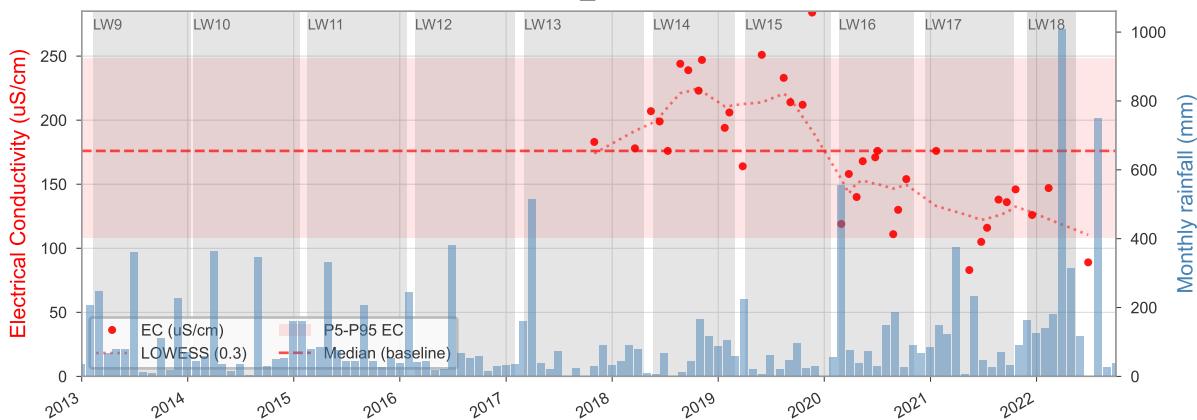
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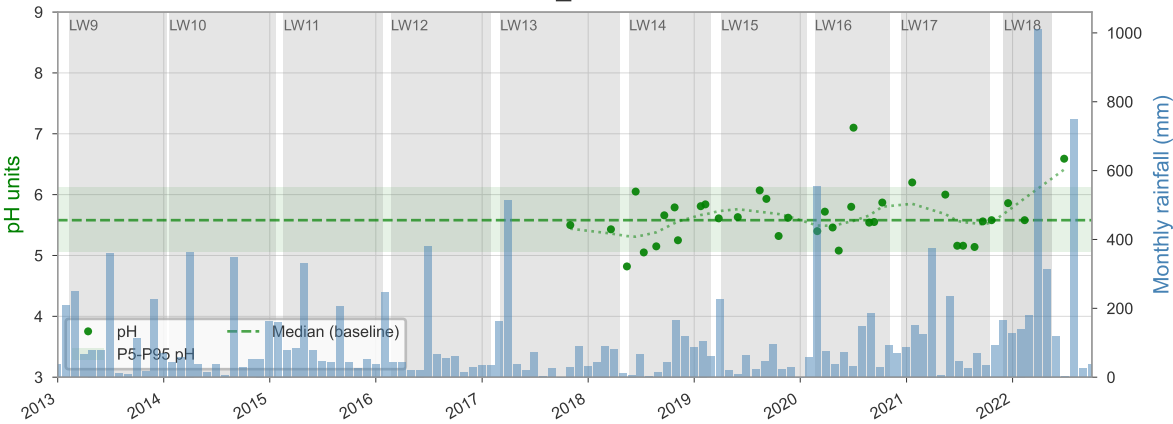
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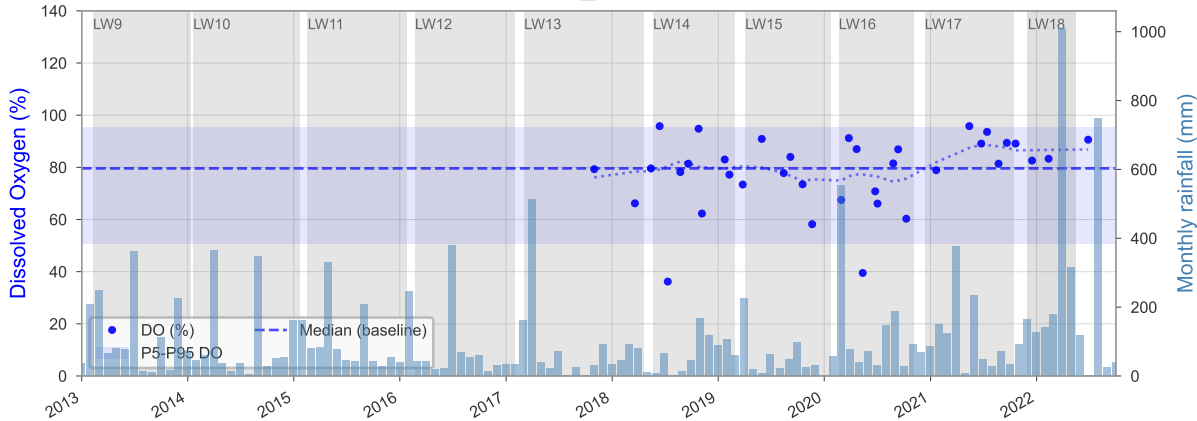
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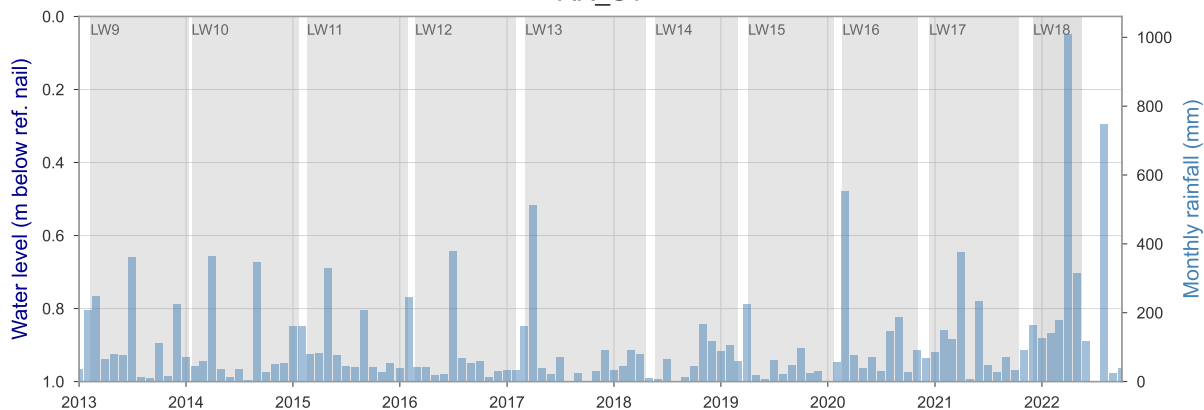
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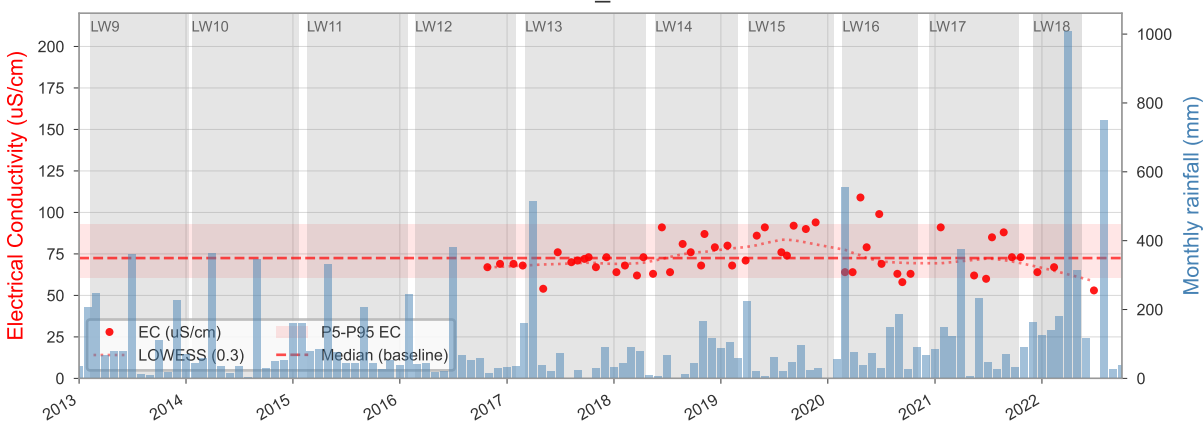
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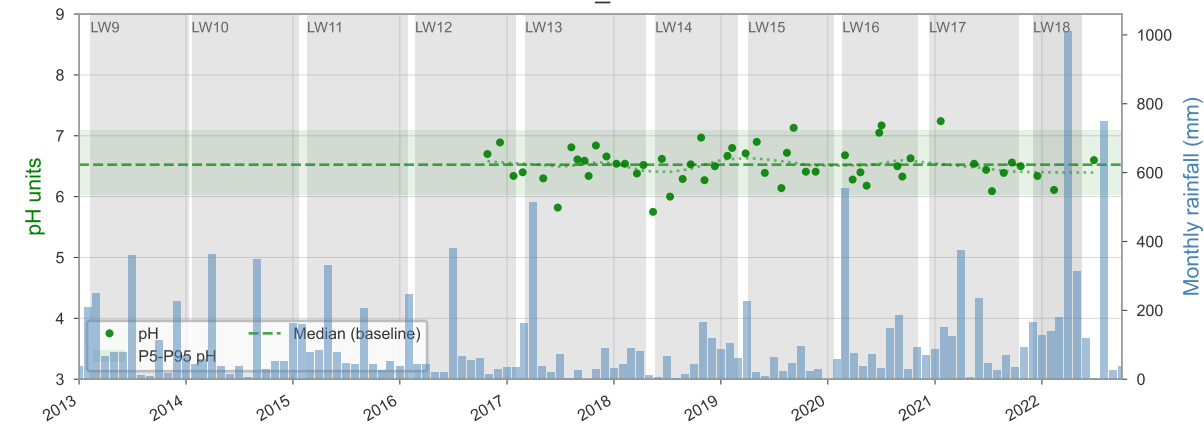
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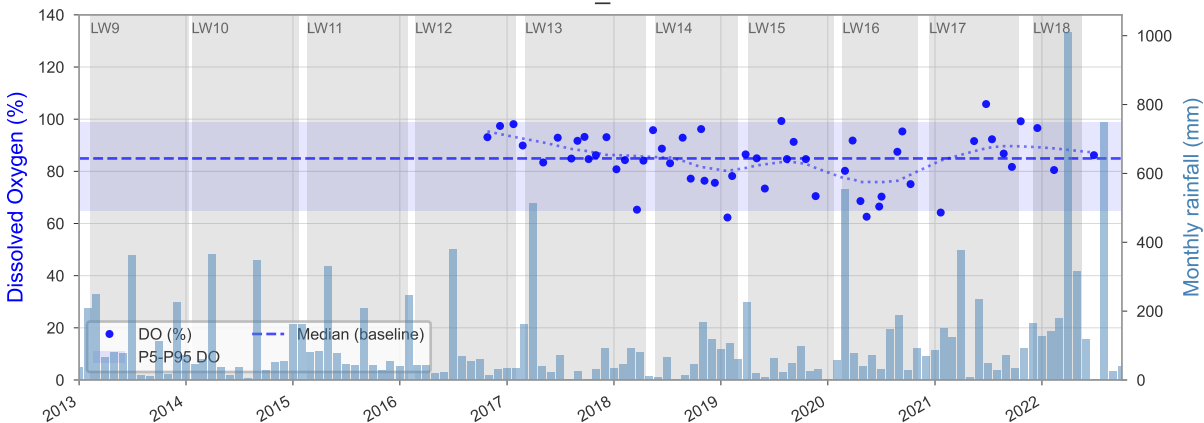
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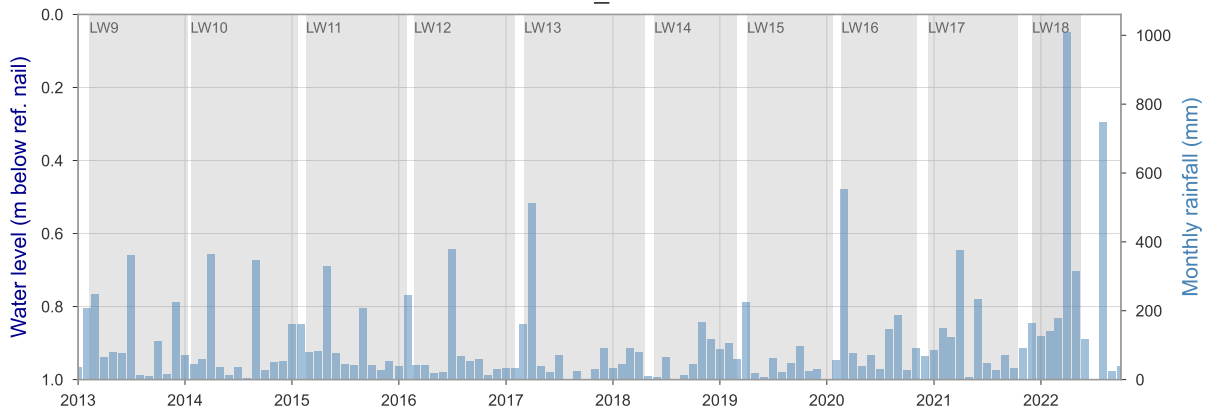
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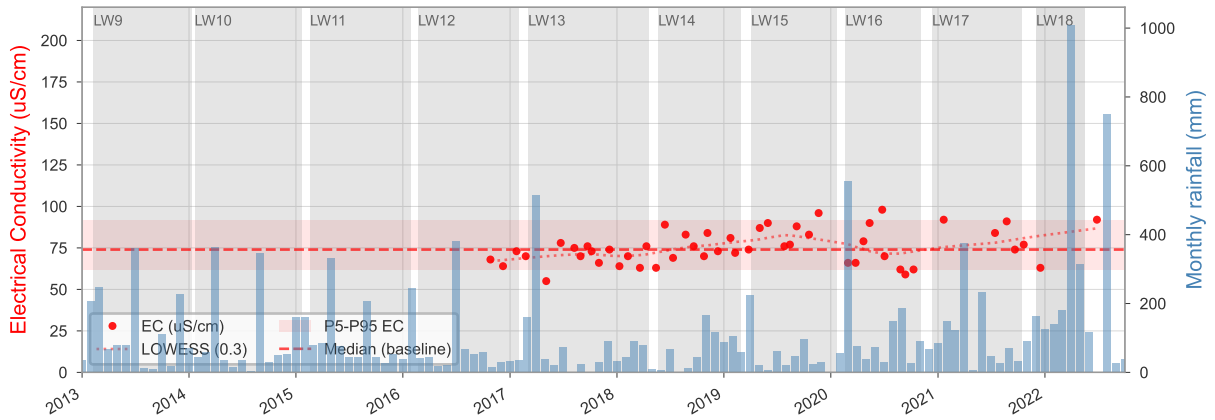
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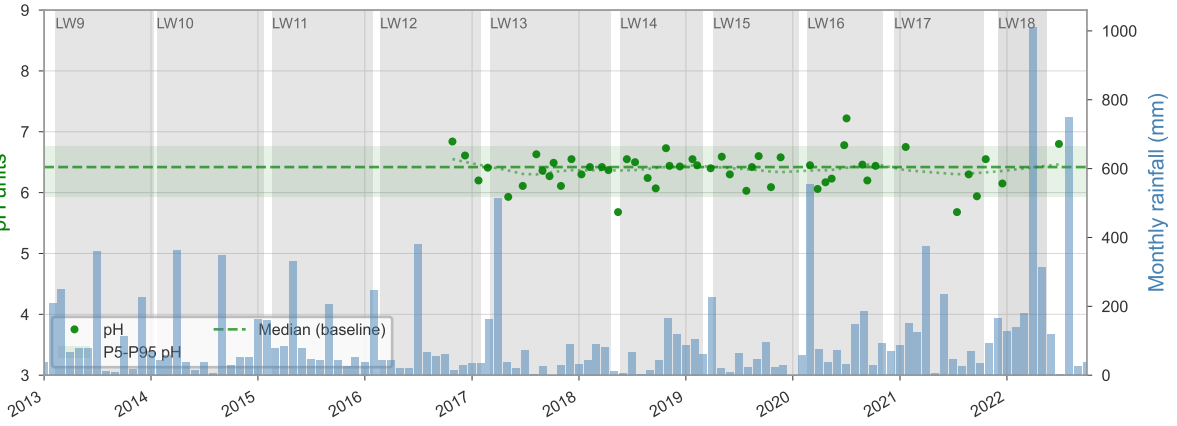
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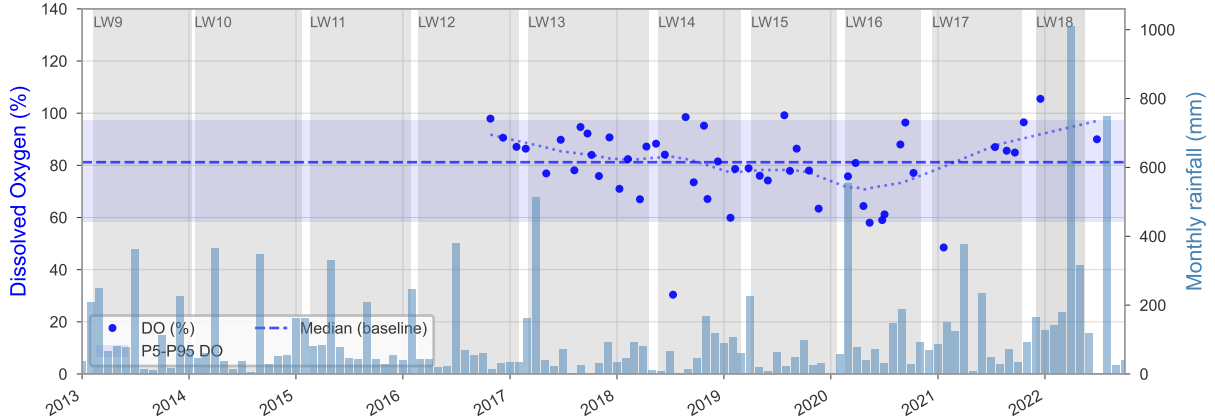
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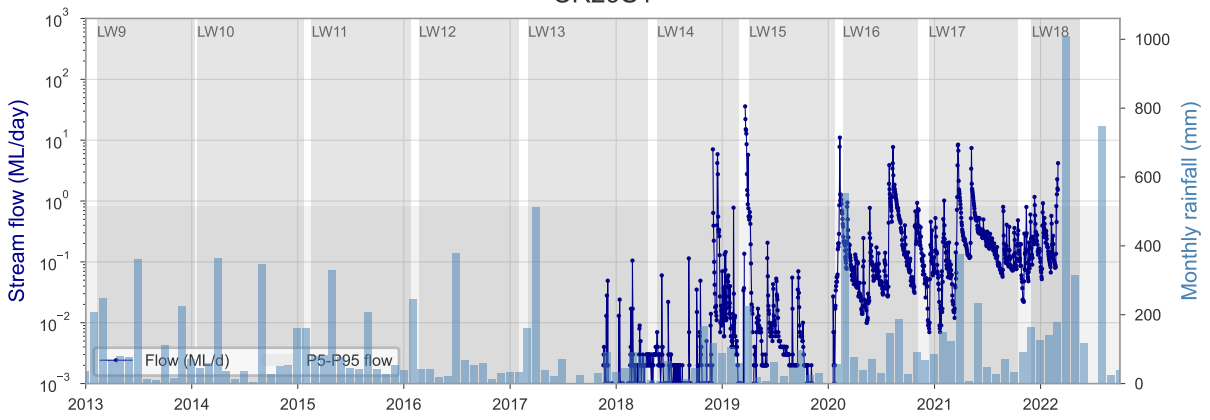
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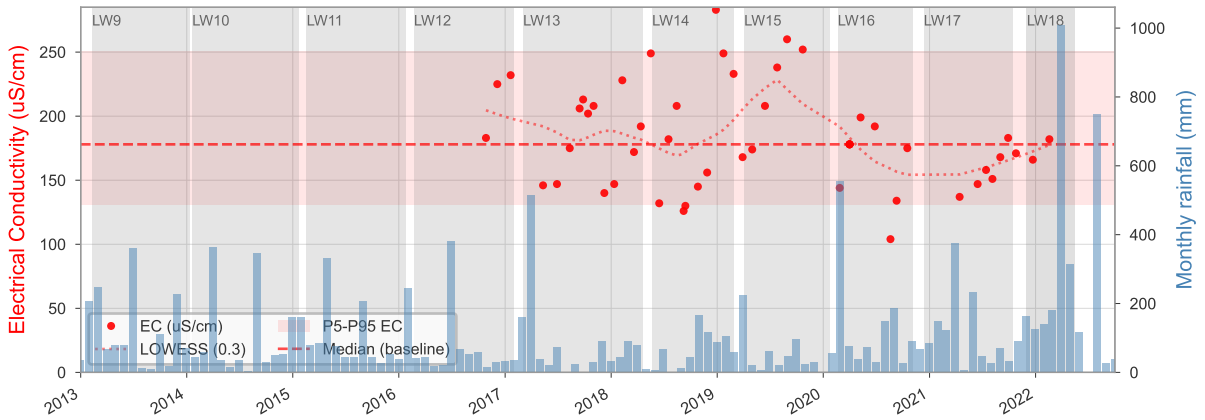
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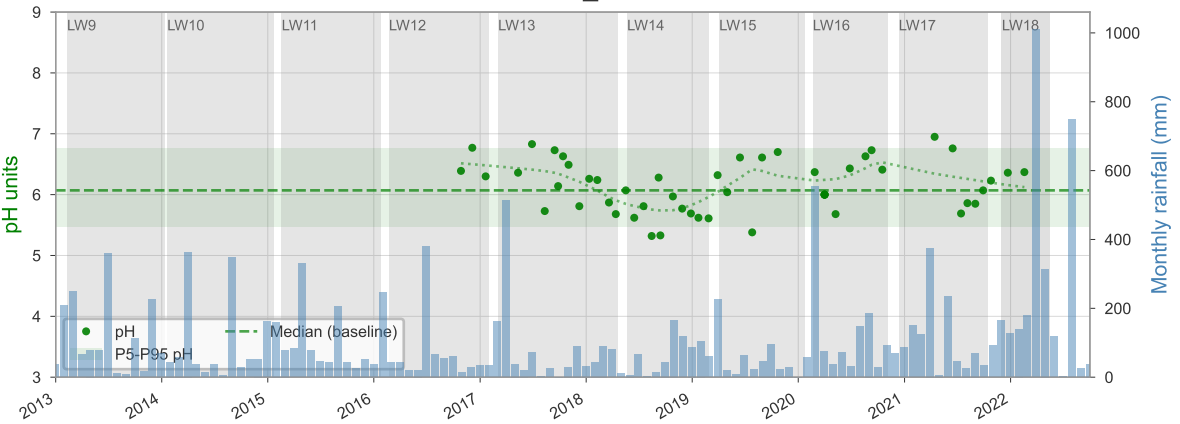
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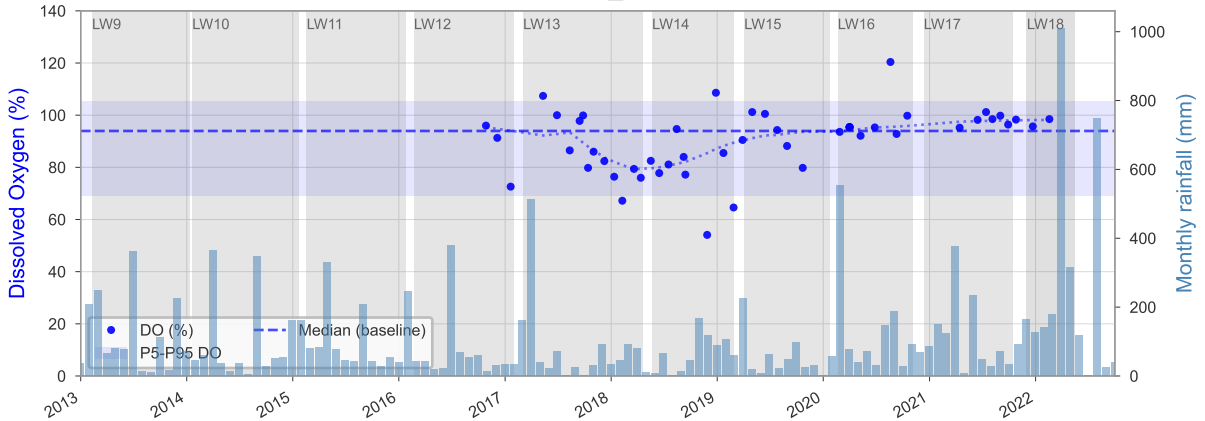
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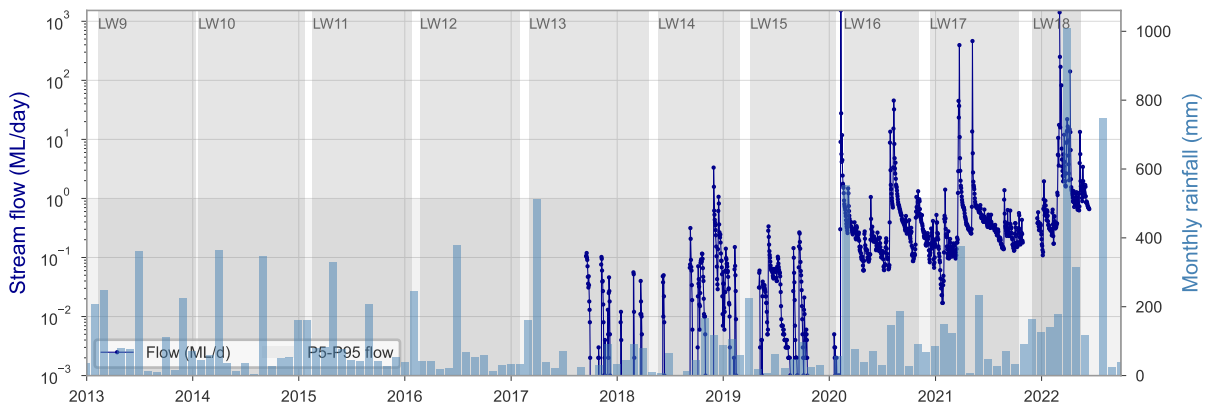
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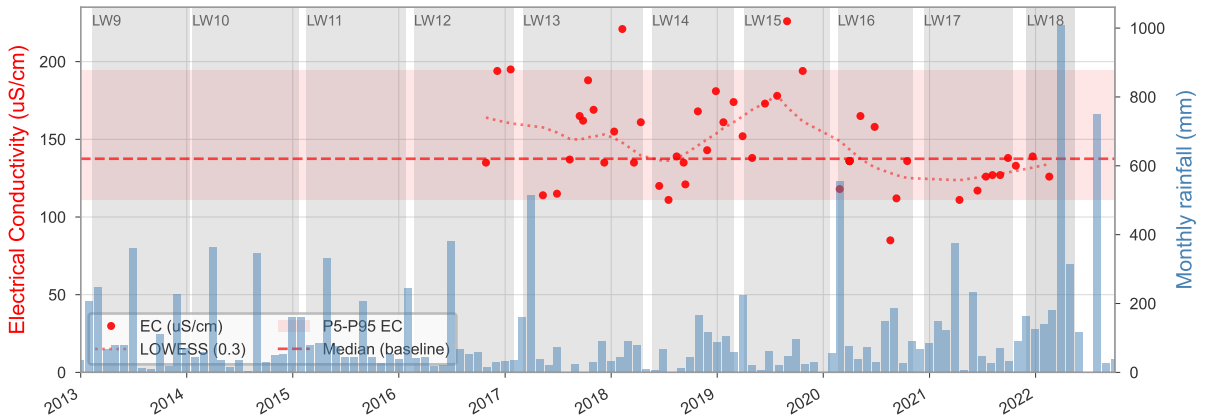
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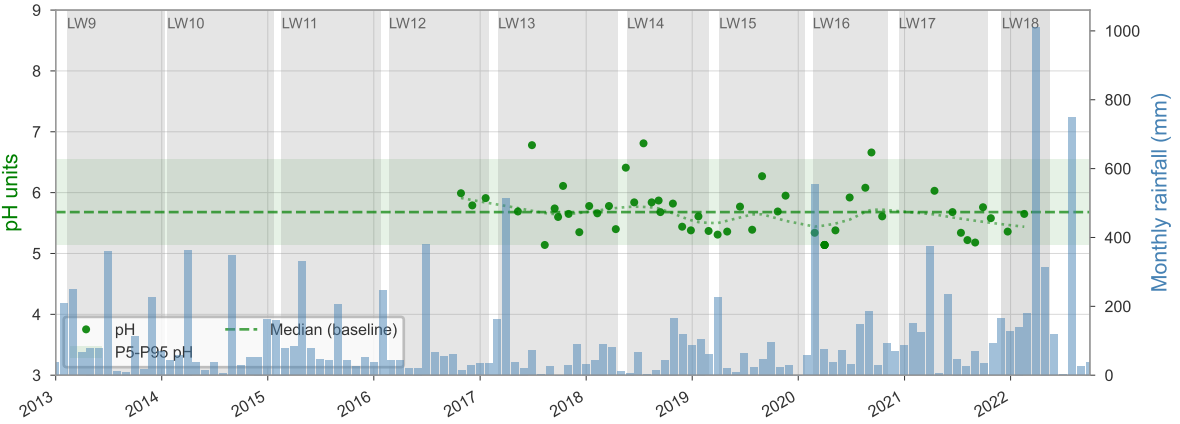
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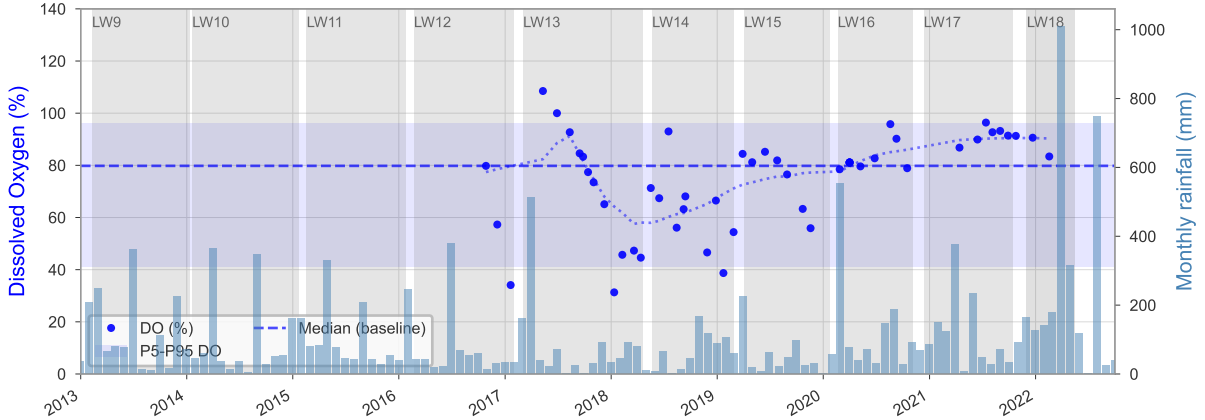
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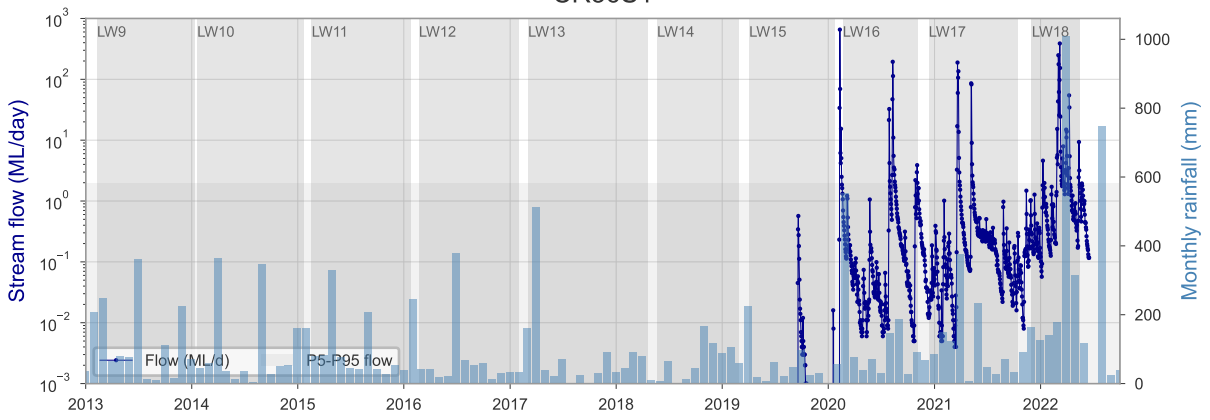
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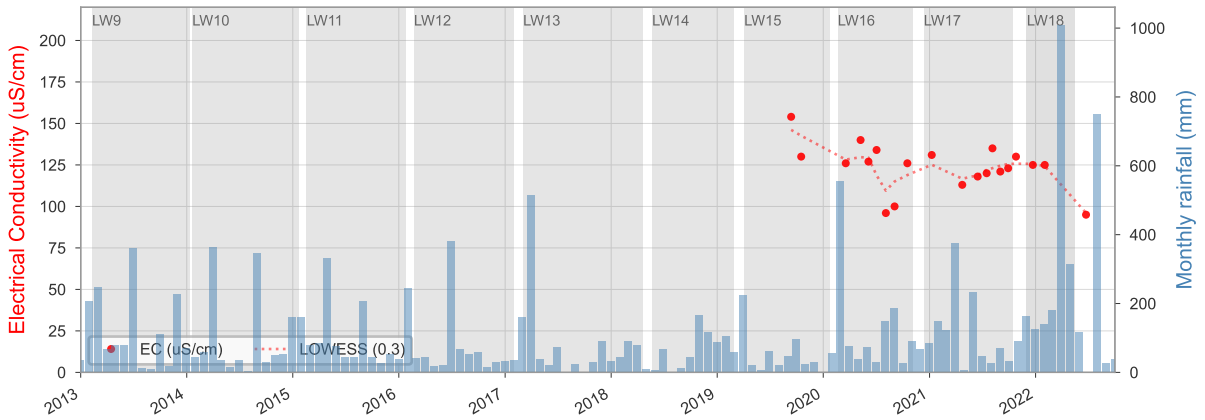
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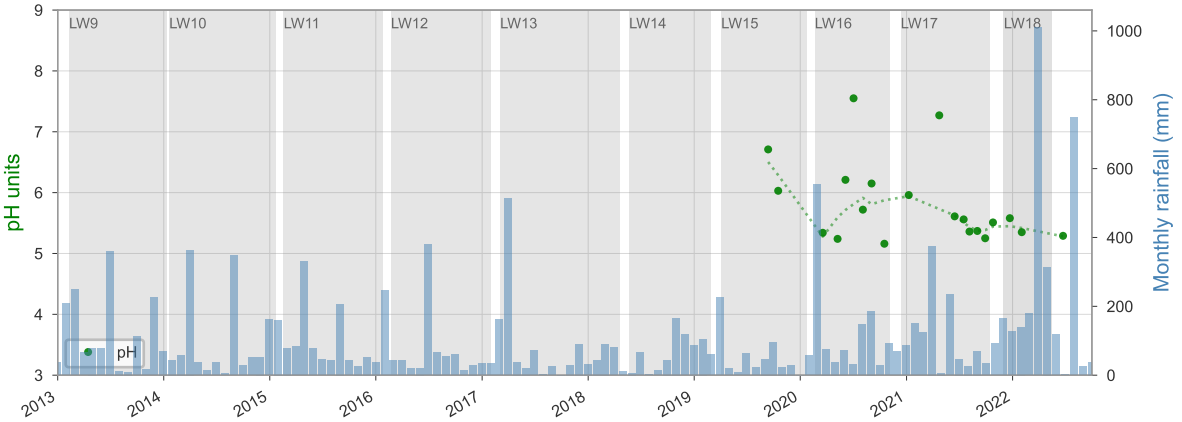
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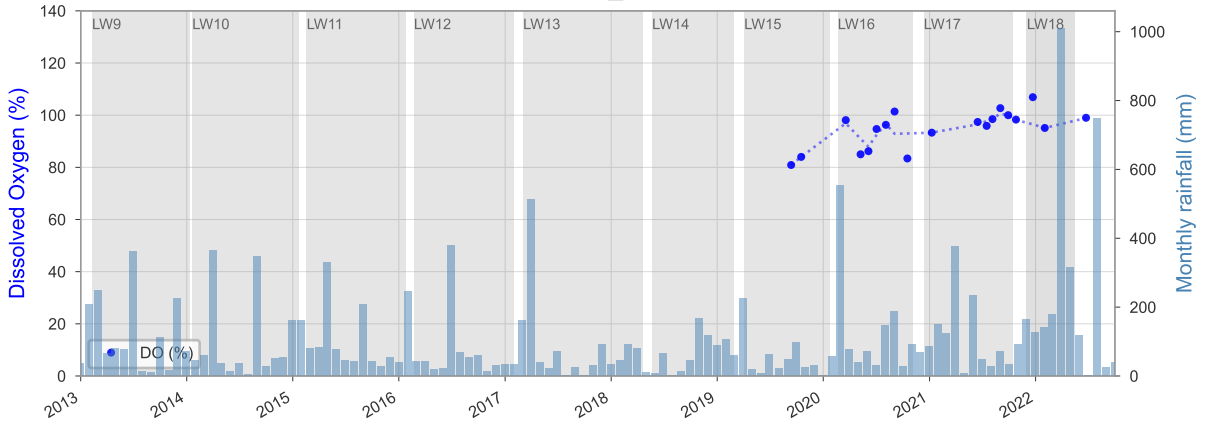
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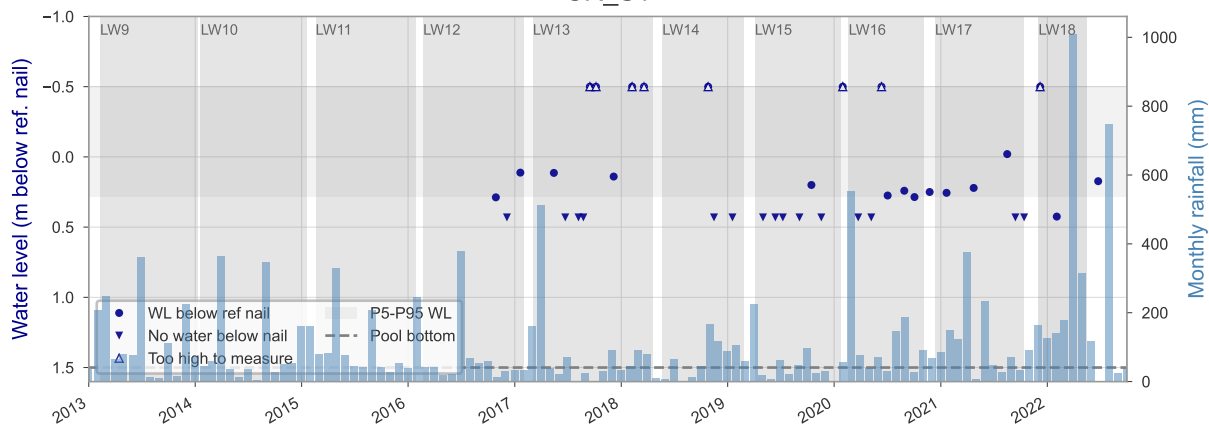
CR36_S1



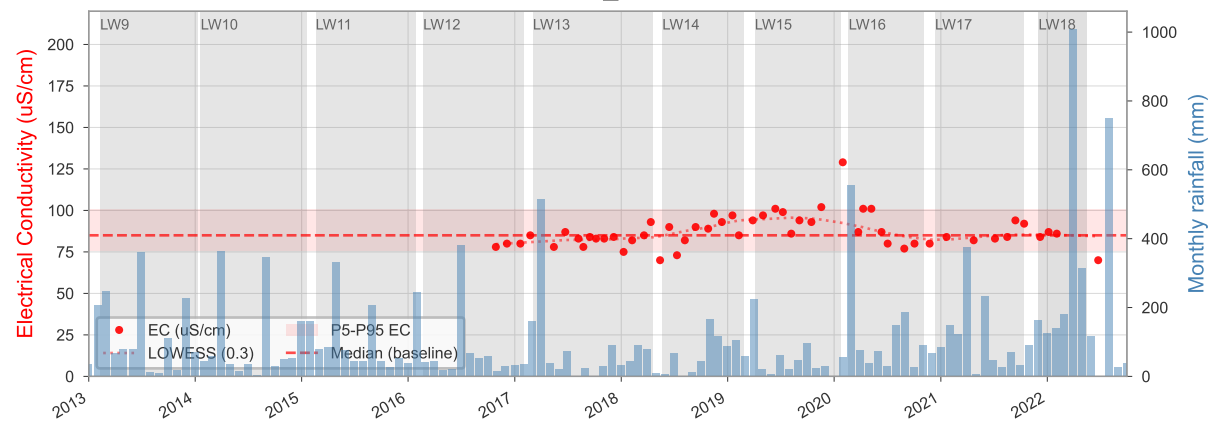
CR36_S1



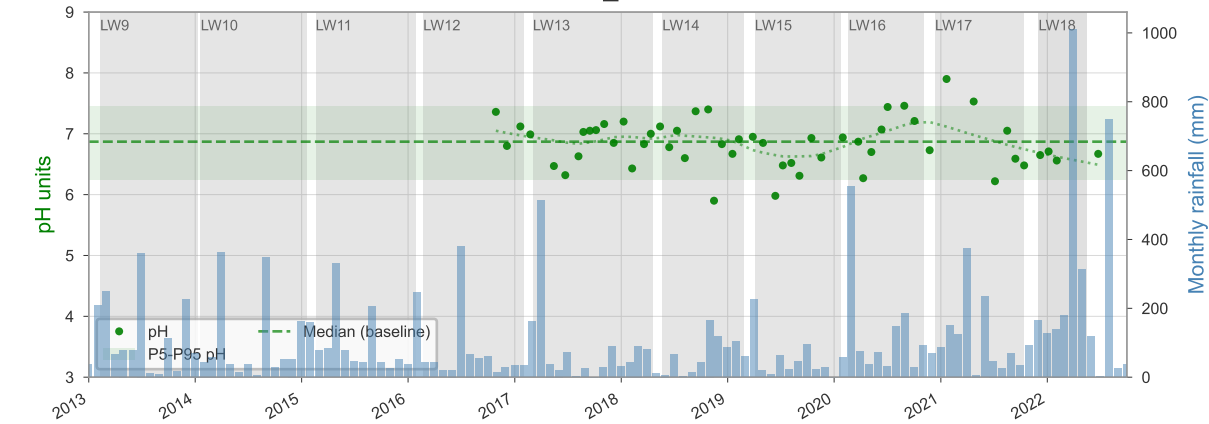
CR_S1



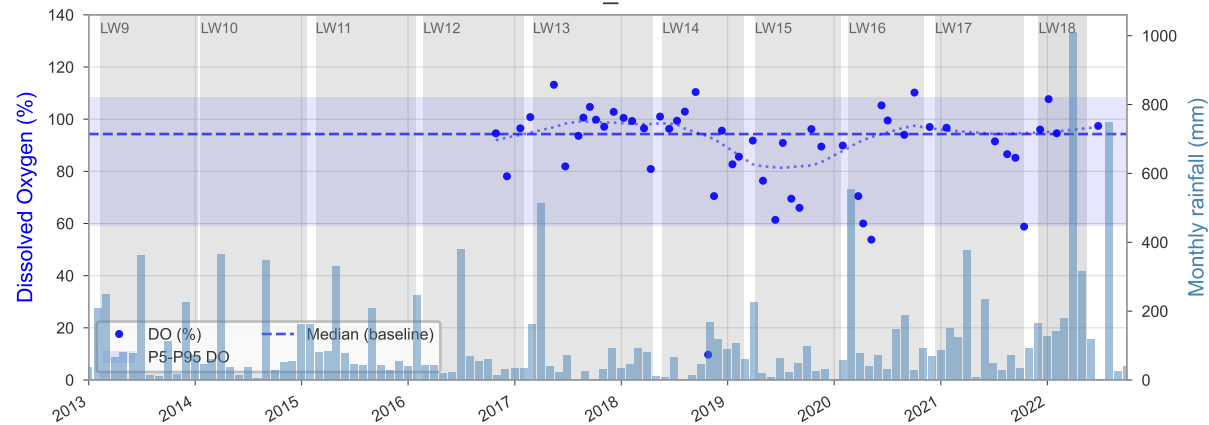
CR_S1



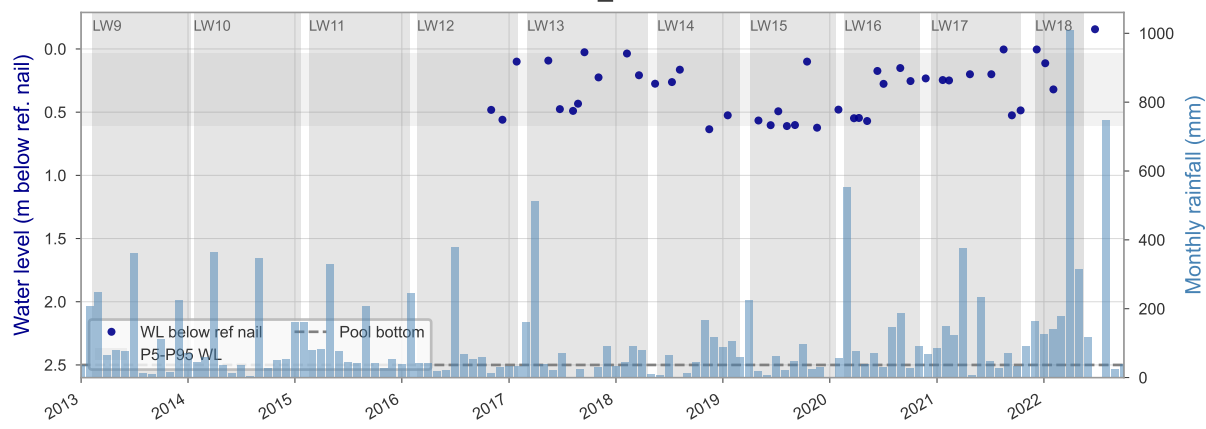
CR_S1



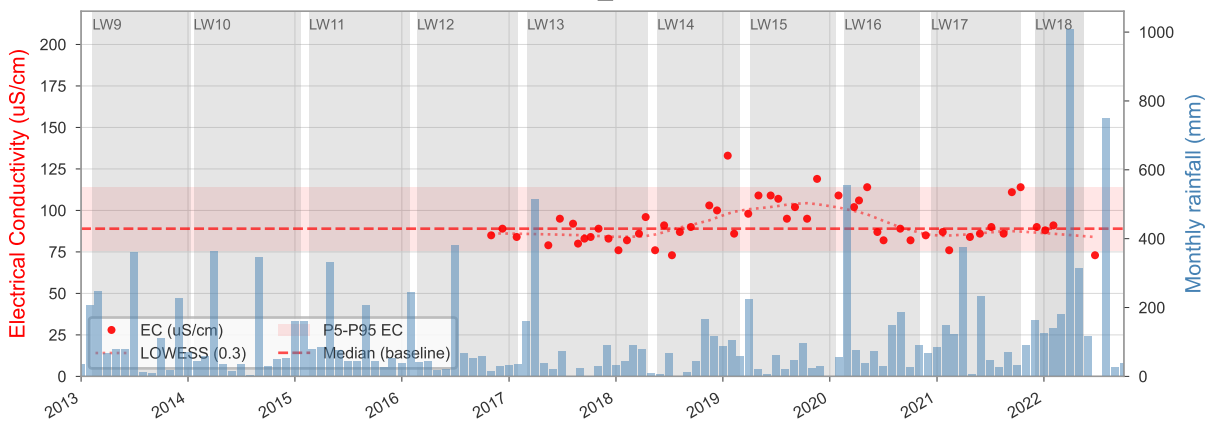
CR_S1



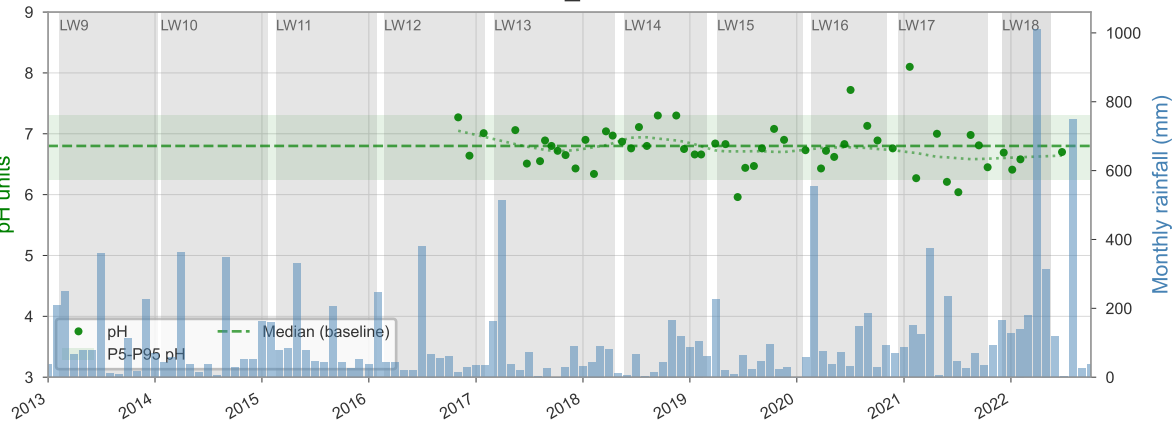
CR_S2



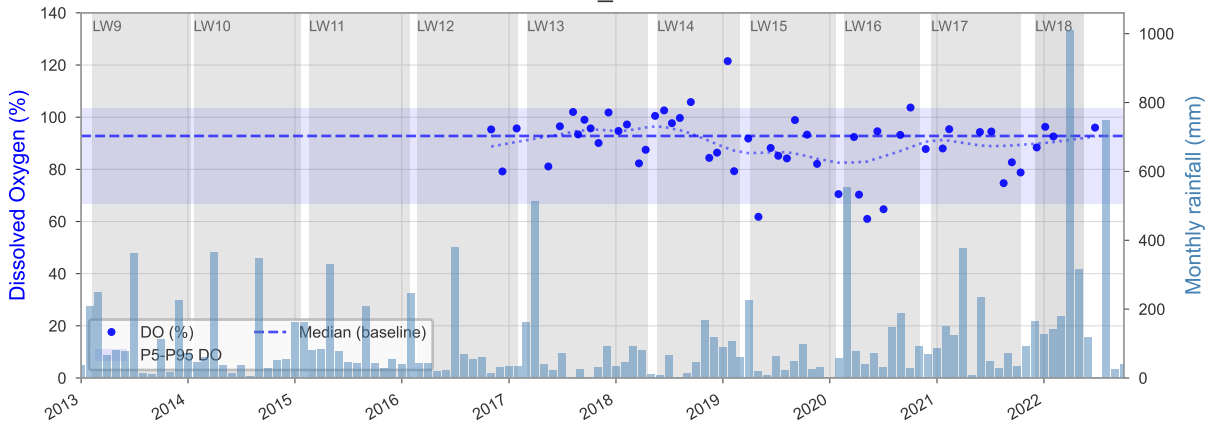
CR_S2



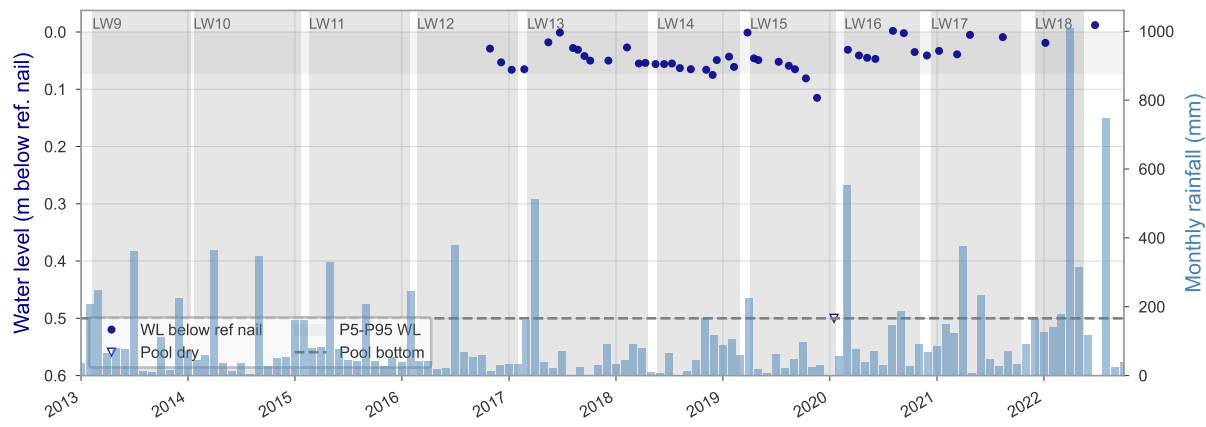
CR_S2



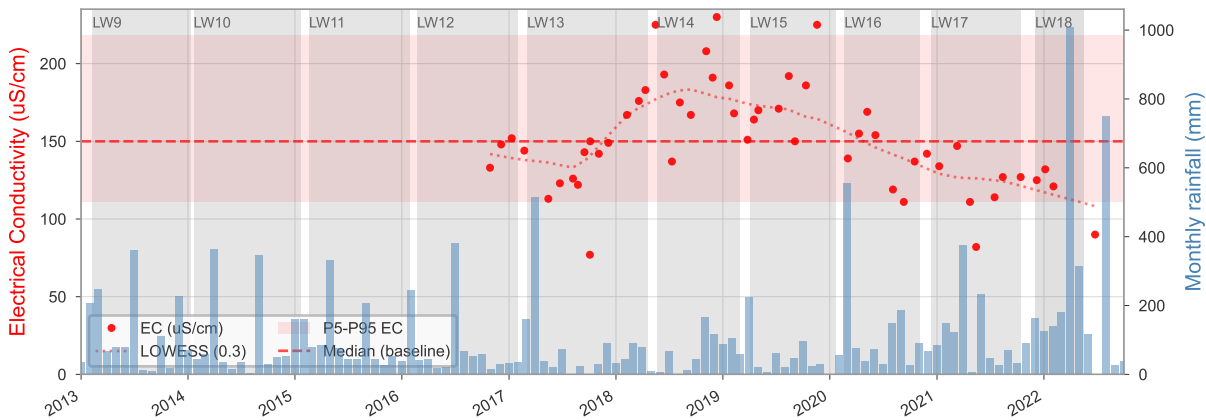
CR_S2



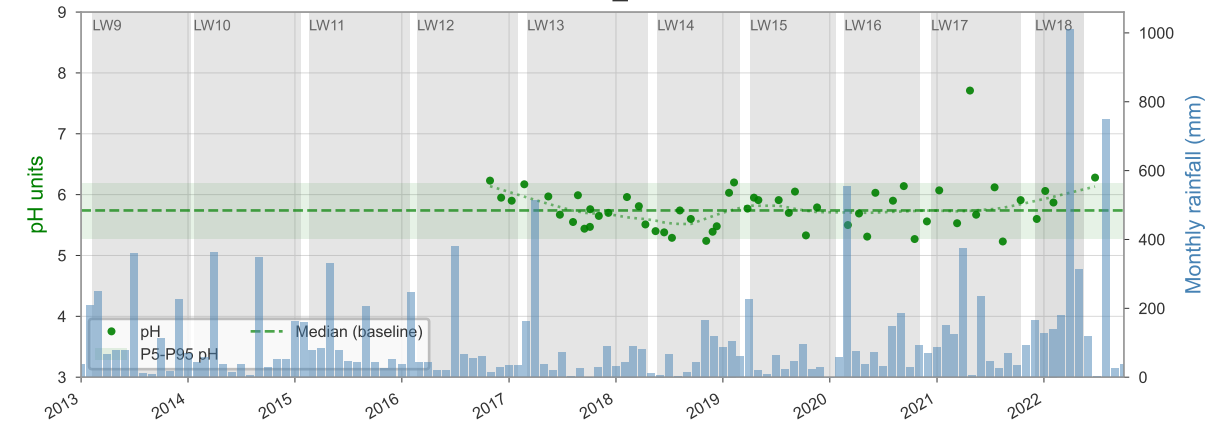
DC10_S1



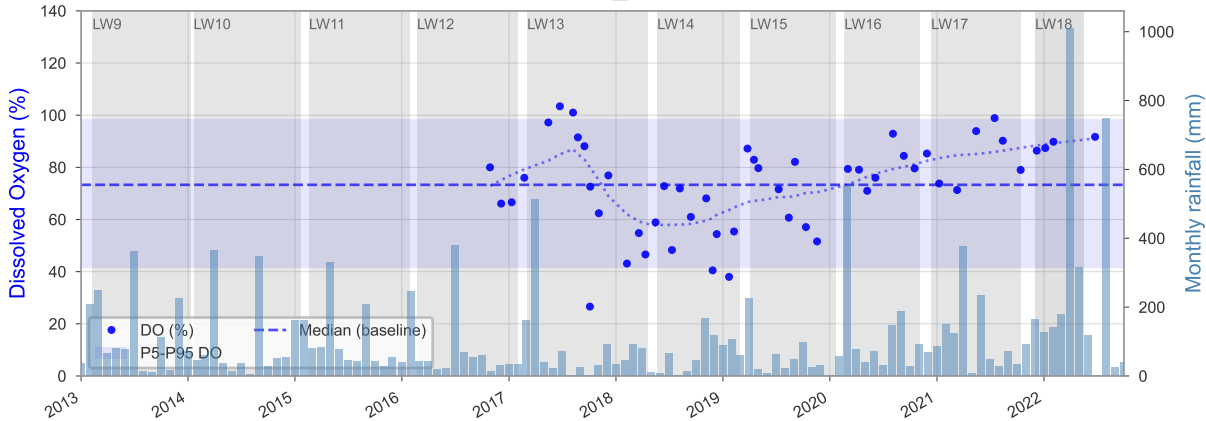
DC10_S1



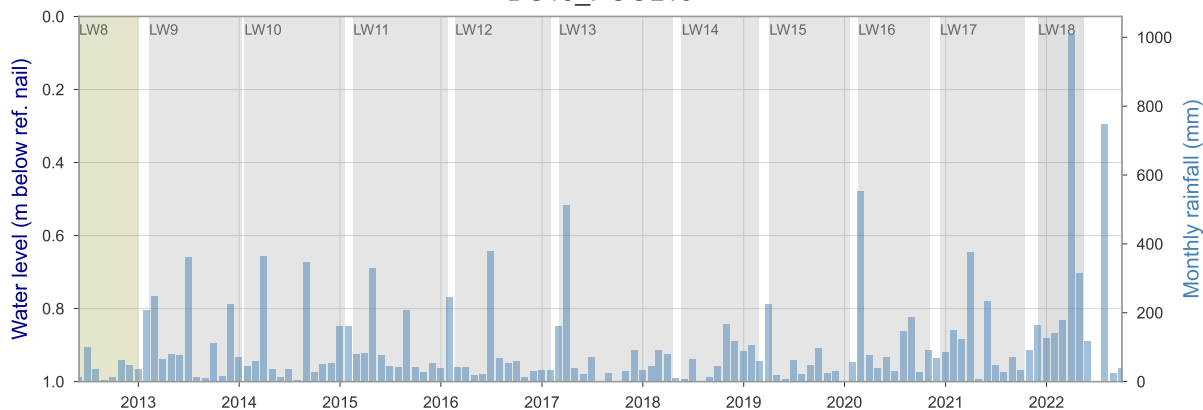
DC10_S1



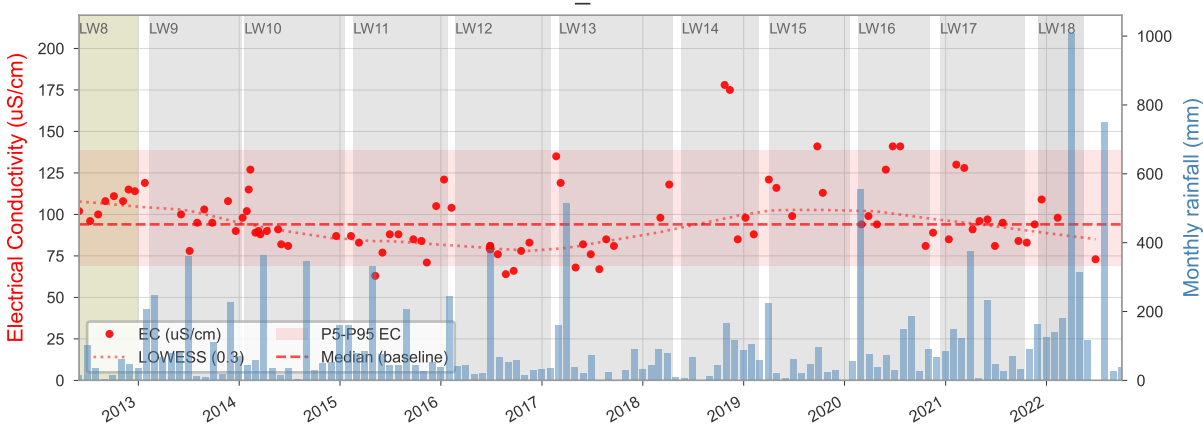
DC10_S1



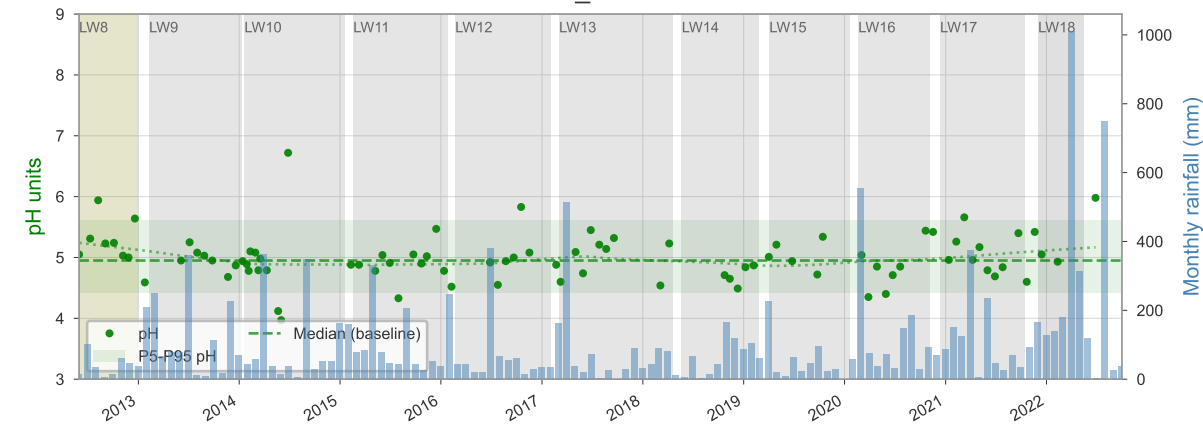
DC13_POOL10



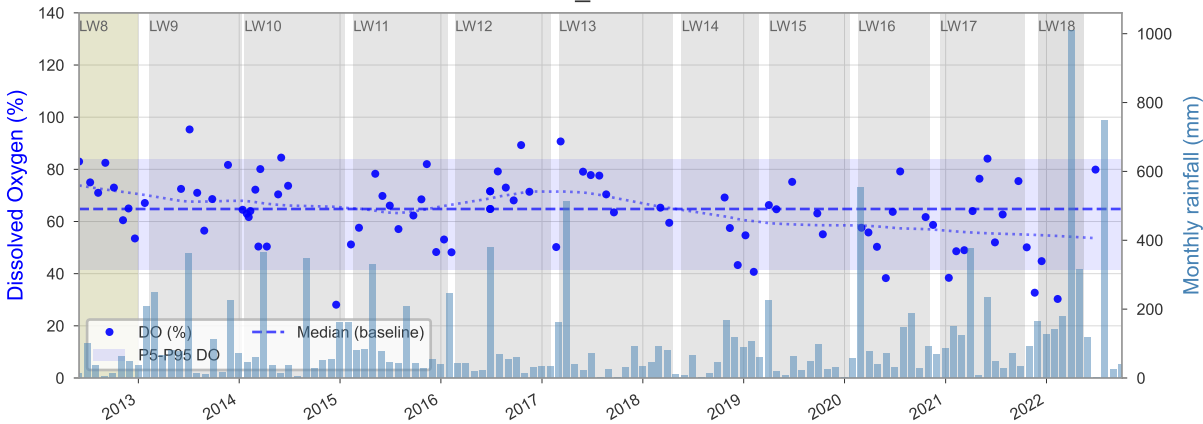
DC13_POOL10



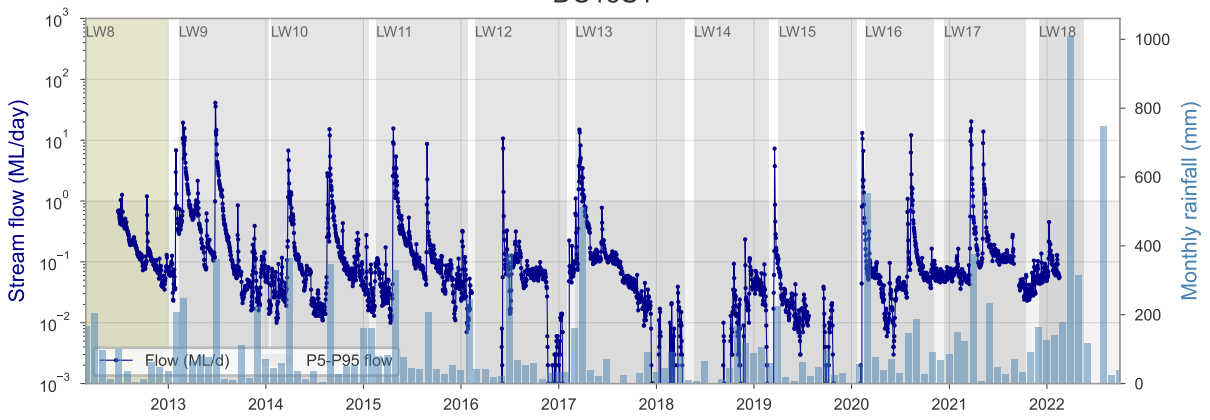
DC13_POOL10



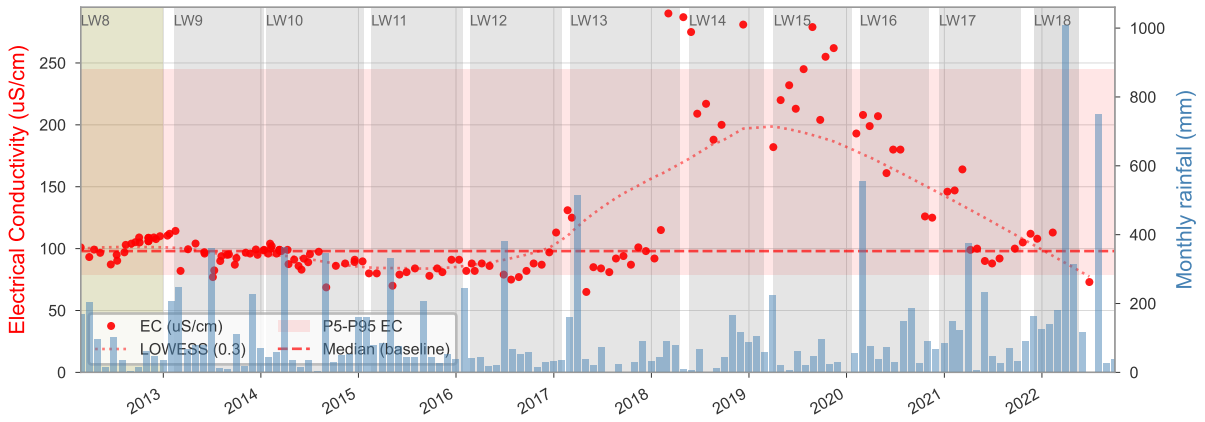
DC13_POOL10



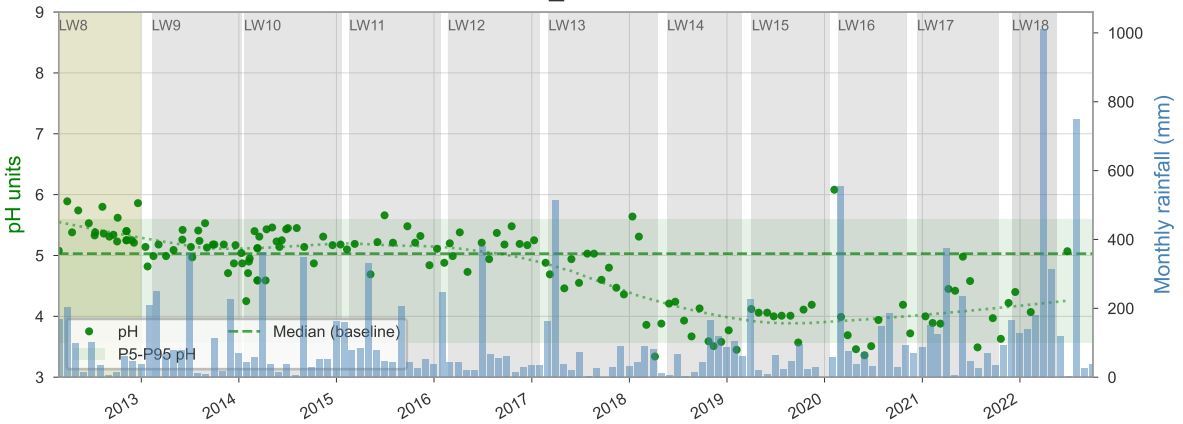
DC13S1



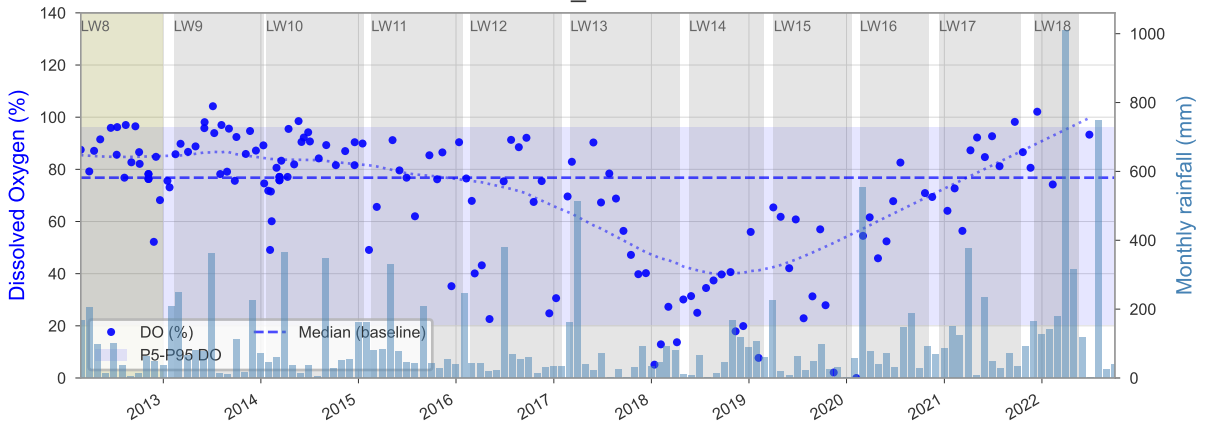
DC13_POOL2B



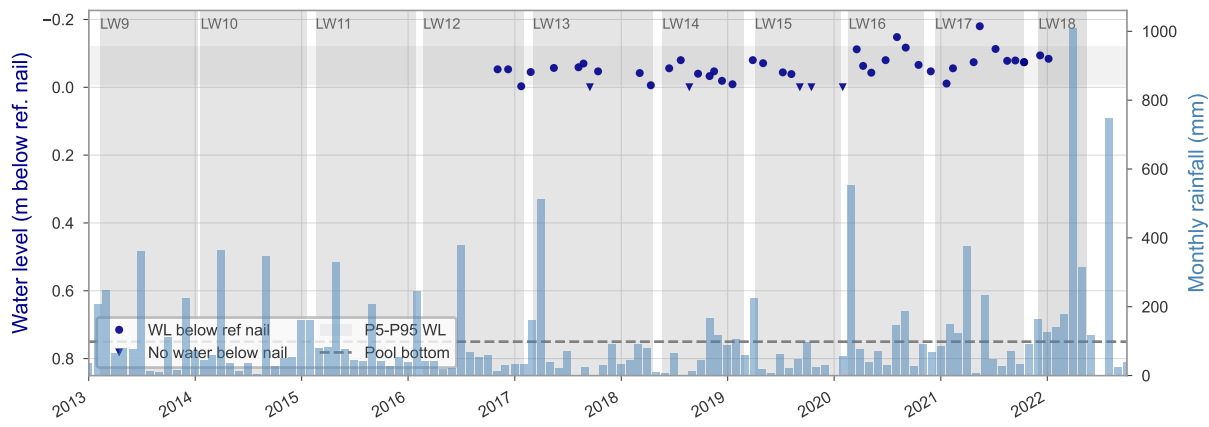
DC13_POOL2B



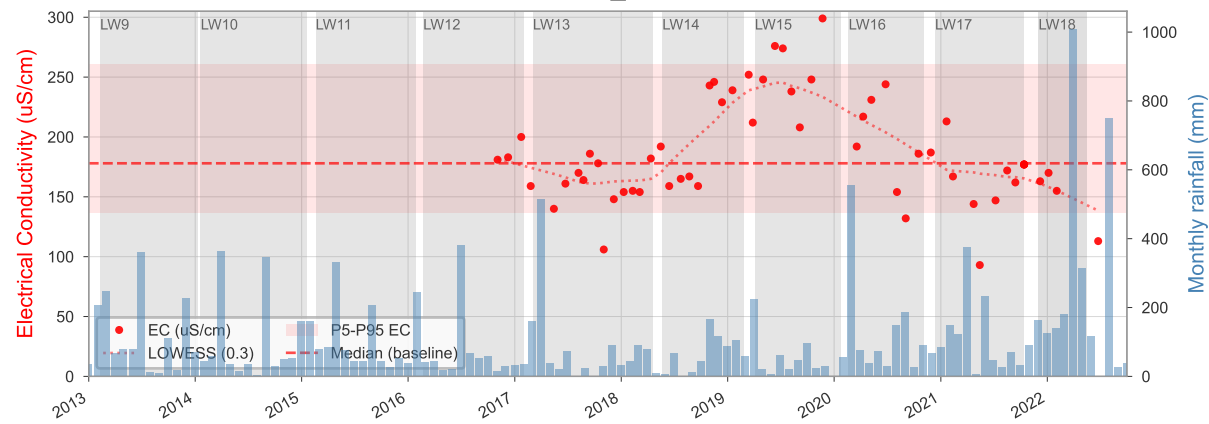
DC13_POOL2B



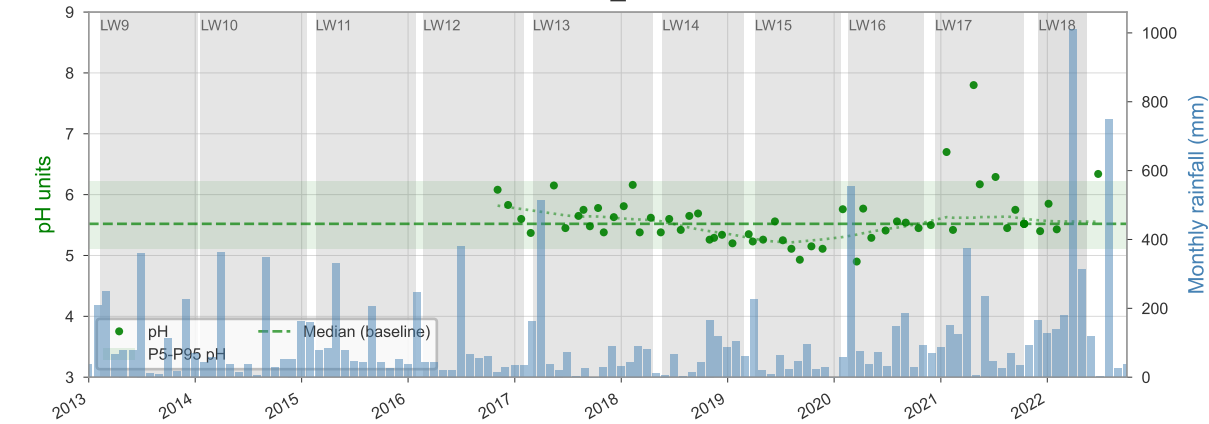
DC8_S1



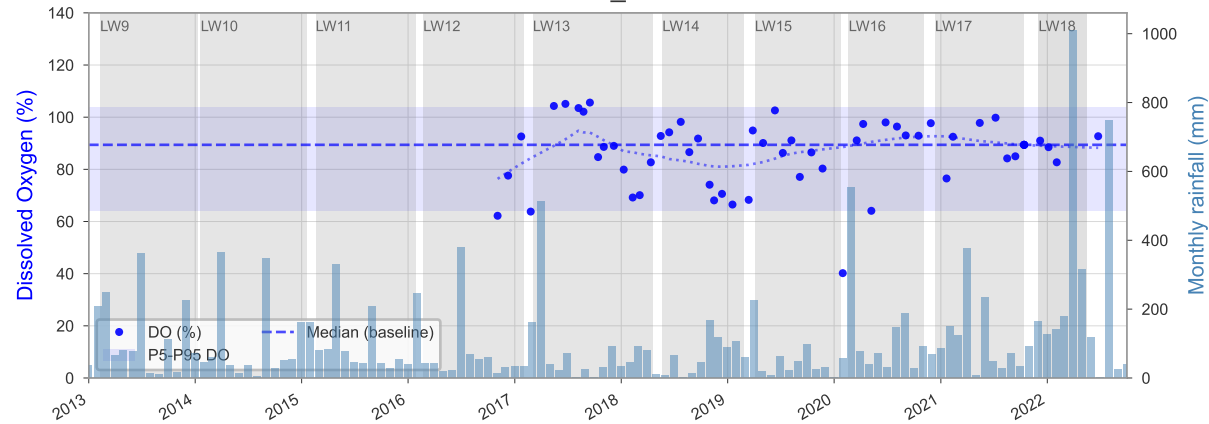
DC8_S1



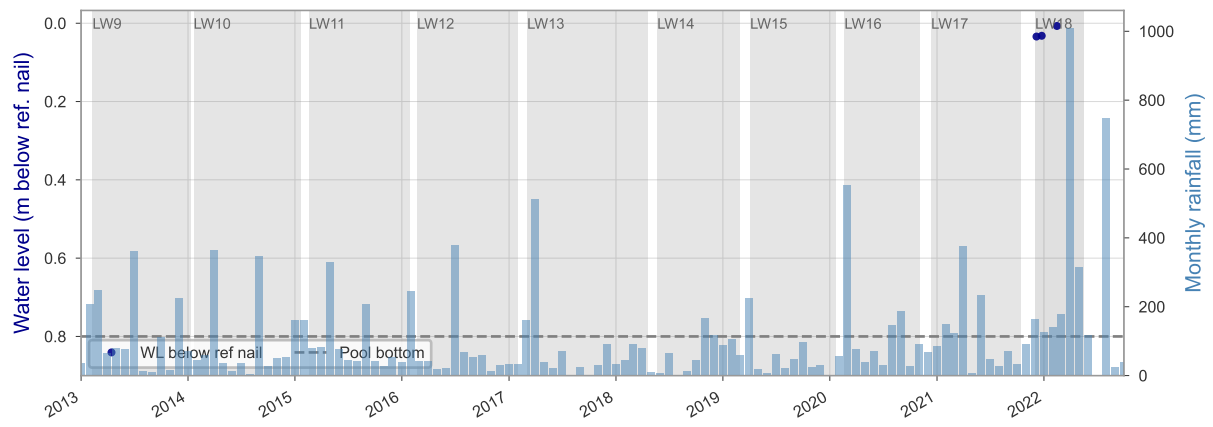
DC8_S1



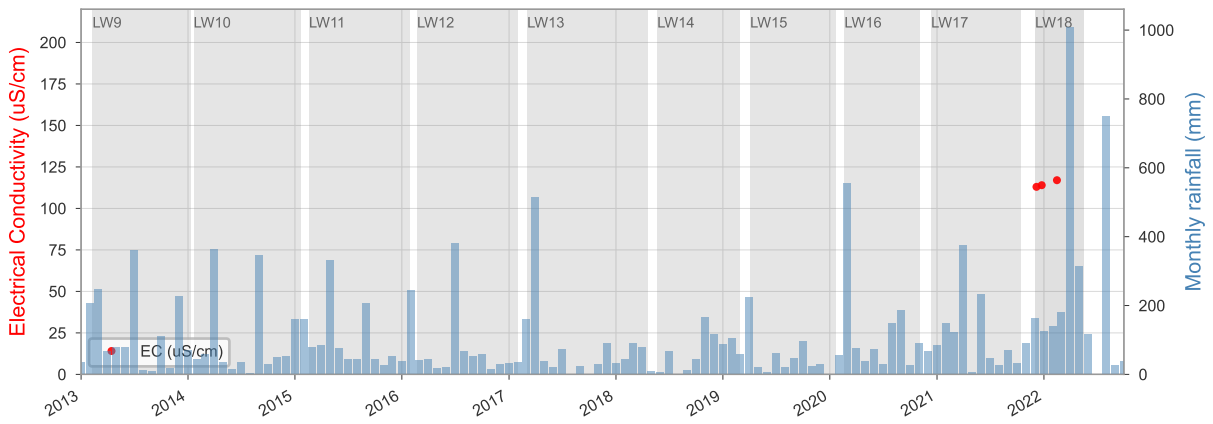
DC8_S1



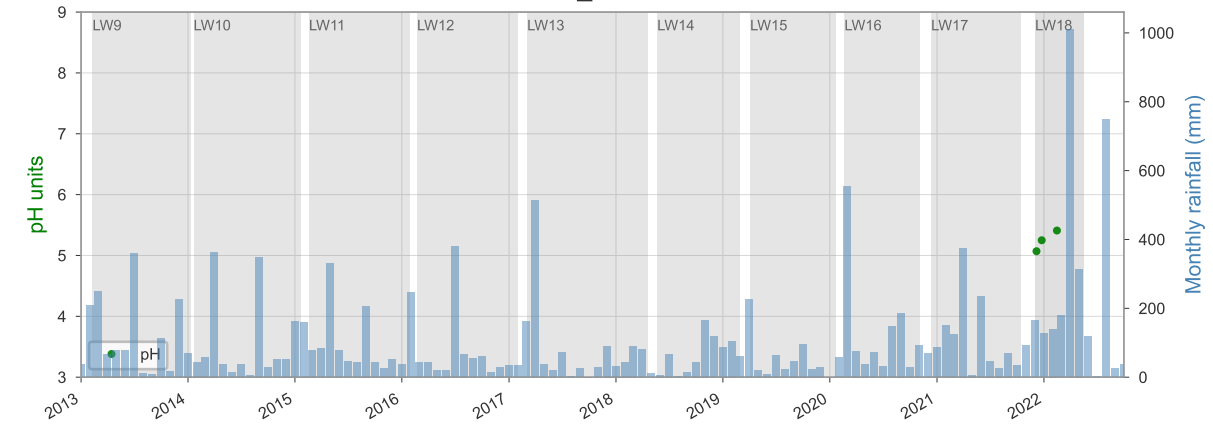
DC9_POOL6



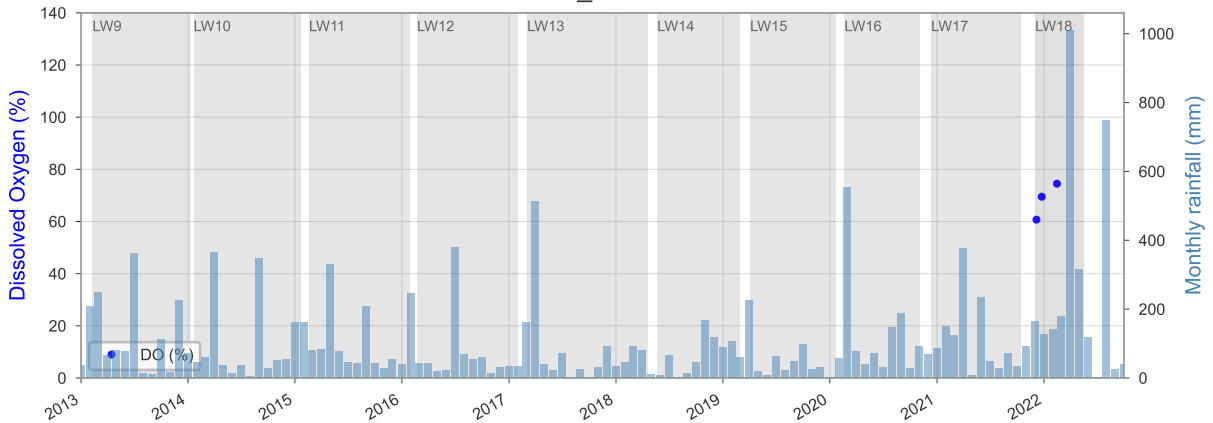
DC9_POOL6



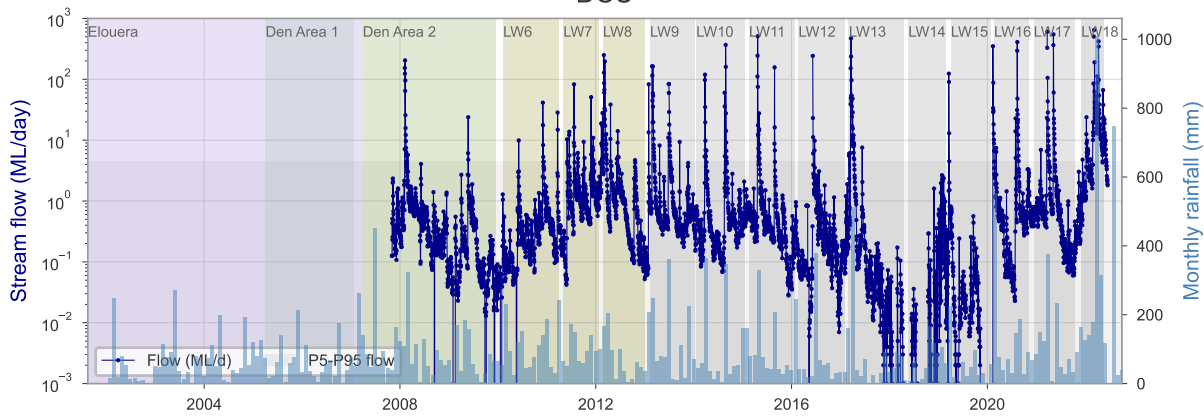
DC9_POOL6



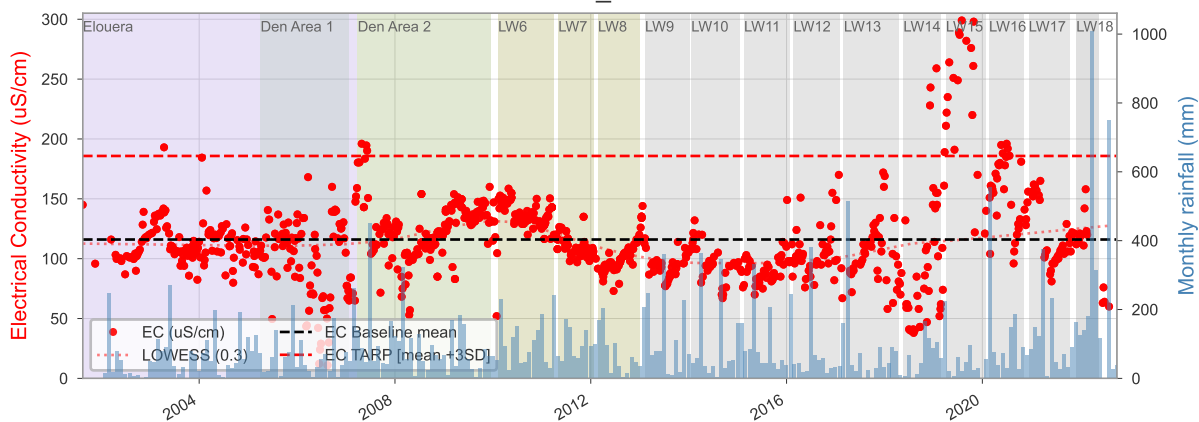
DC9_POOL6



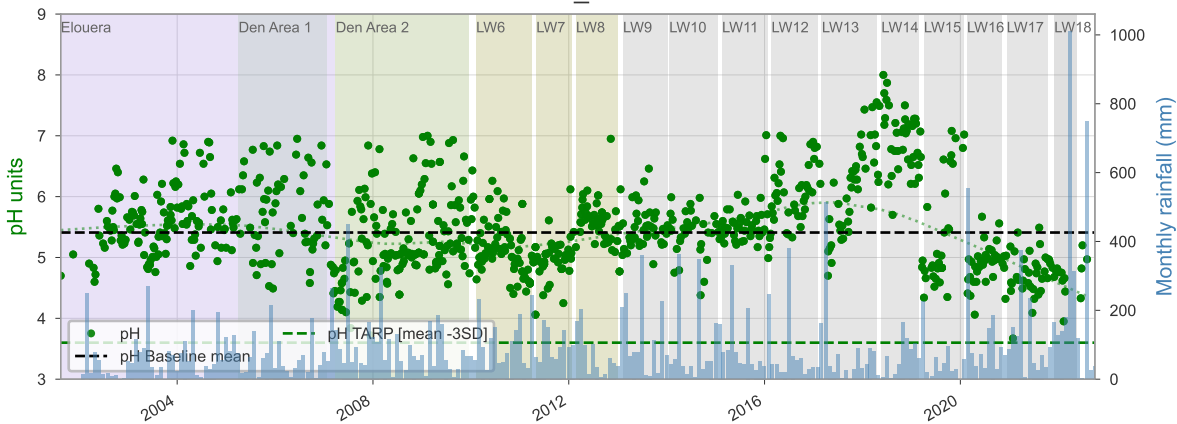
DCU



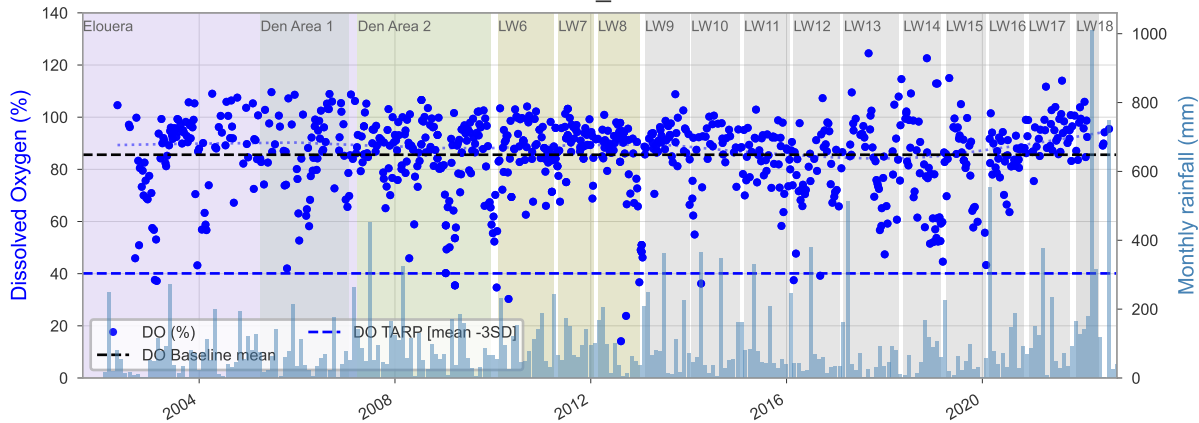
DCC_FR6



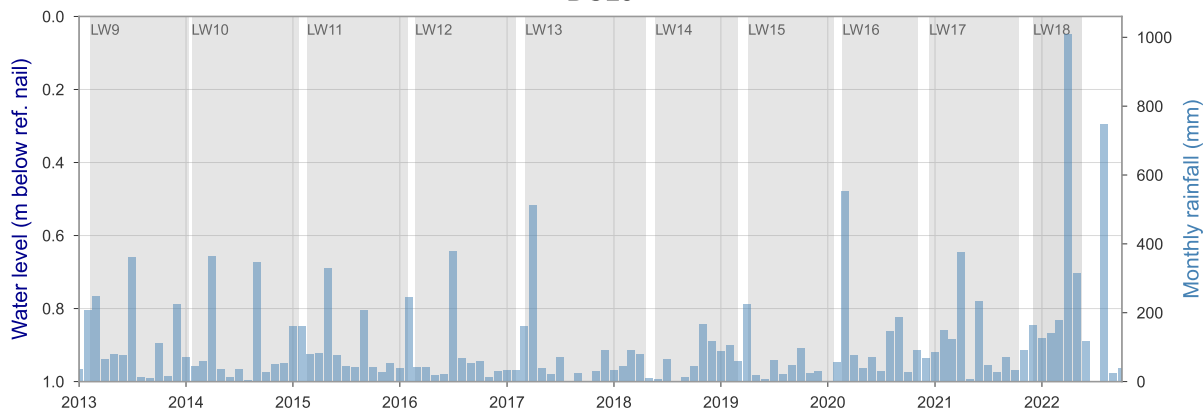
DCC_FR6



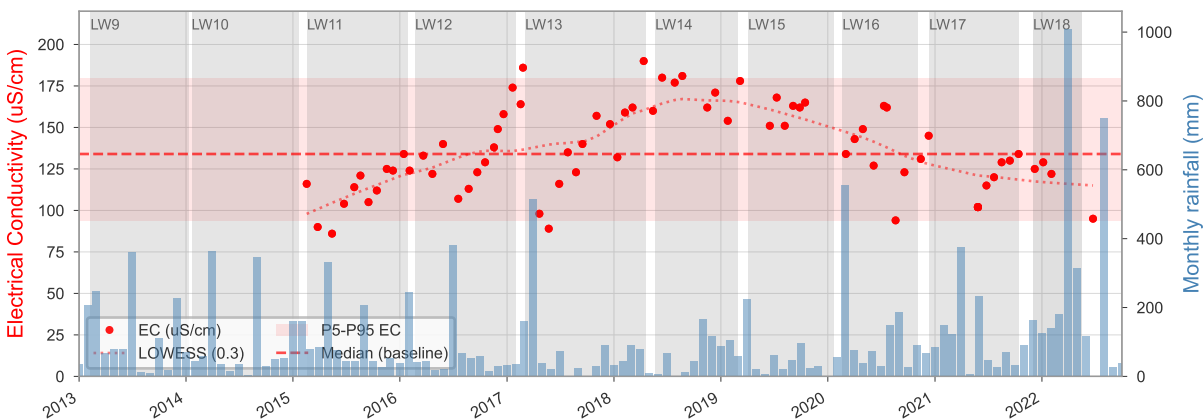
DCC_FR6



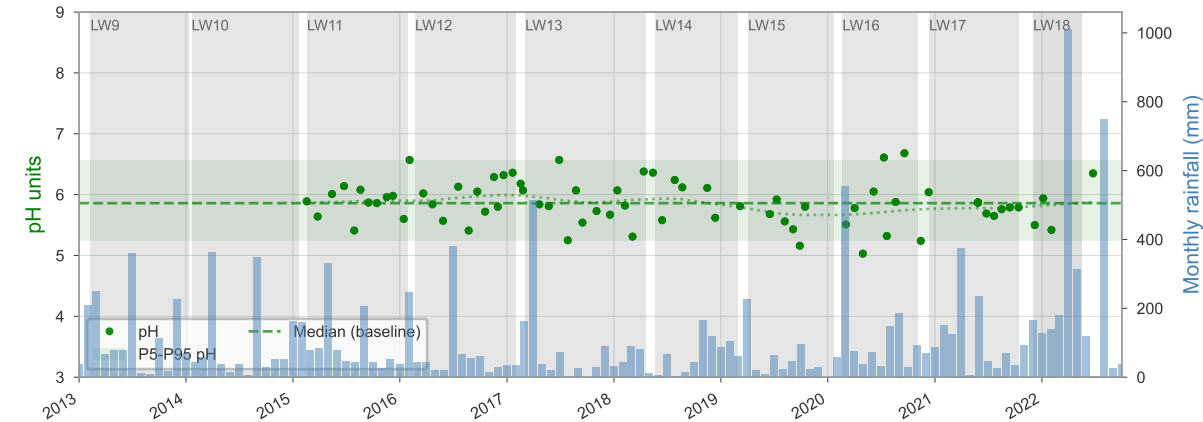
DCL3



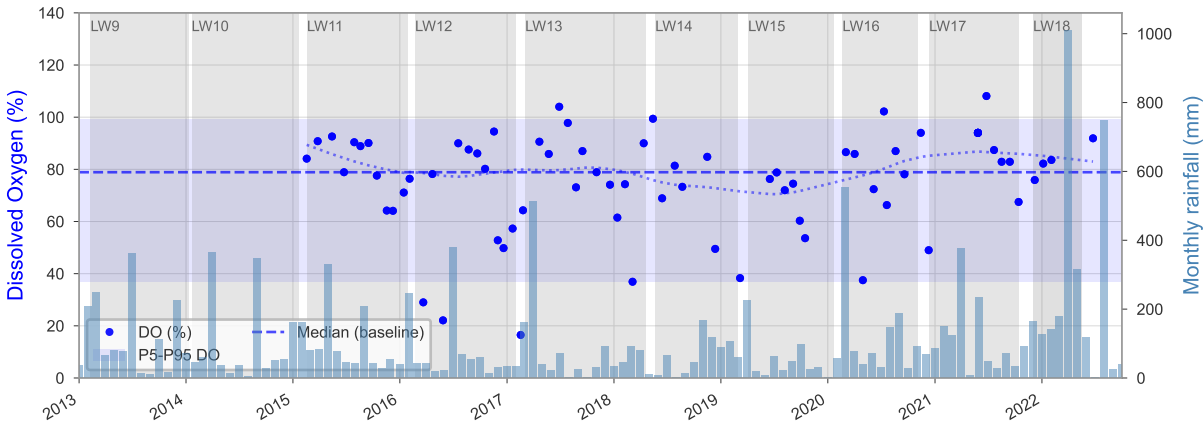
DCL3



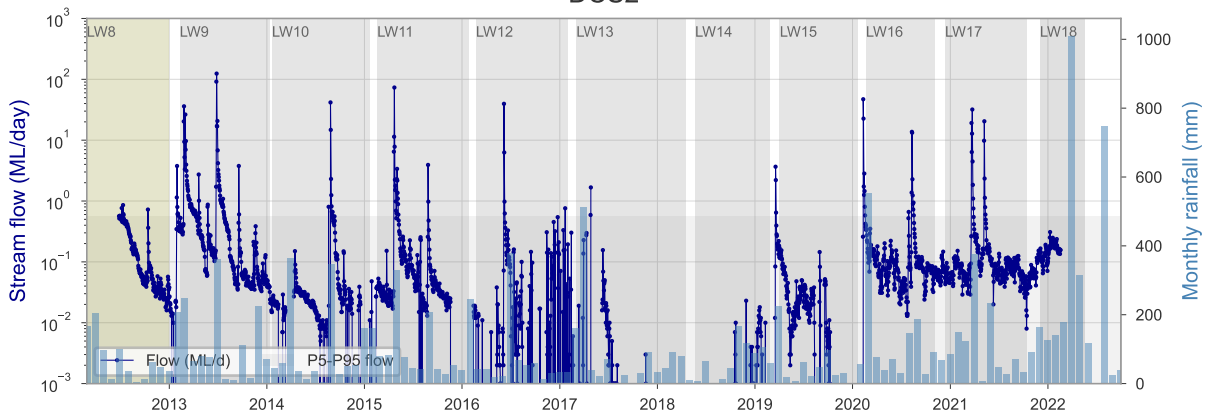
DCL3



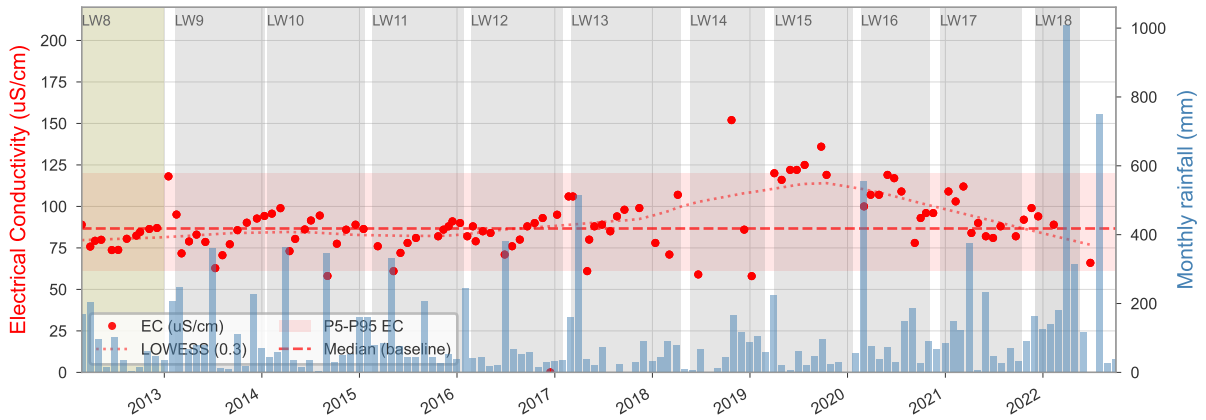
DCL3



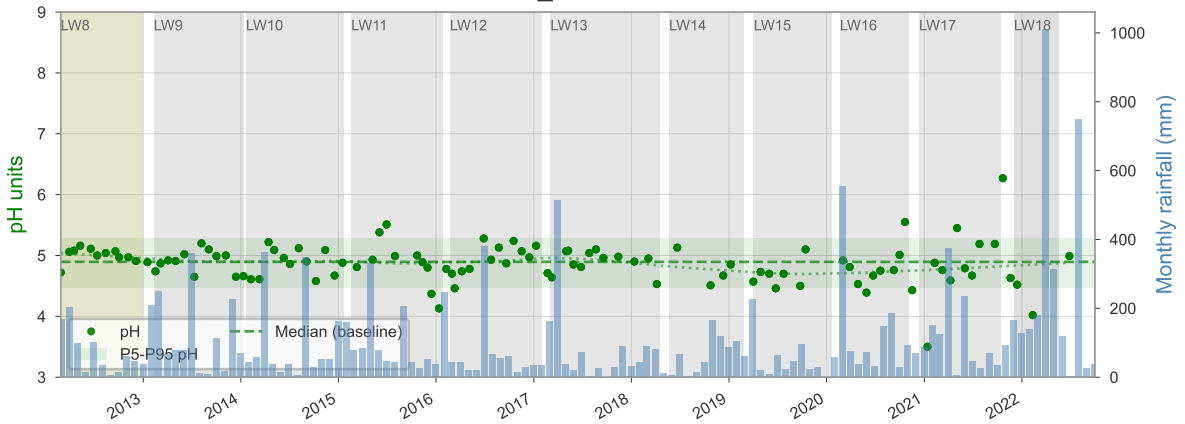
DCS2



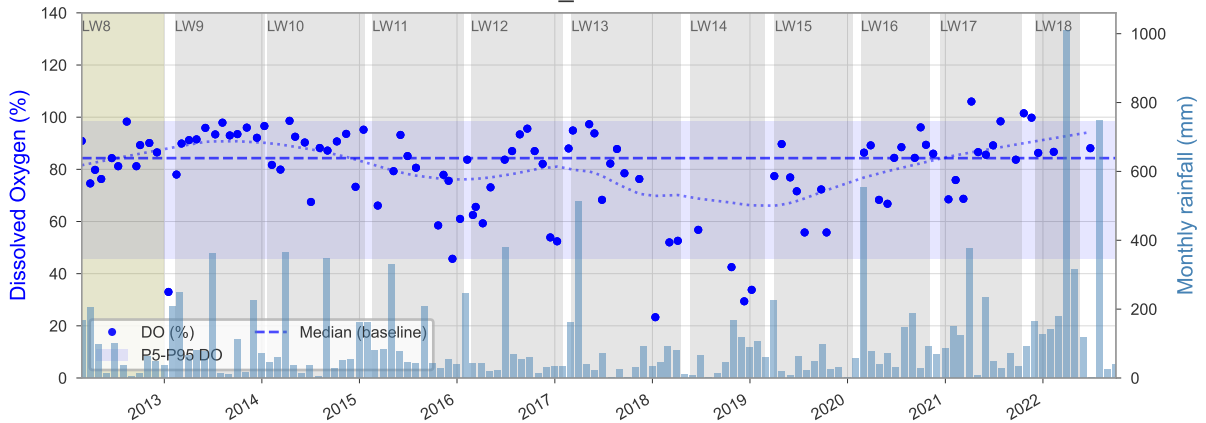
DC_POOL22



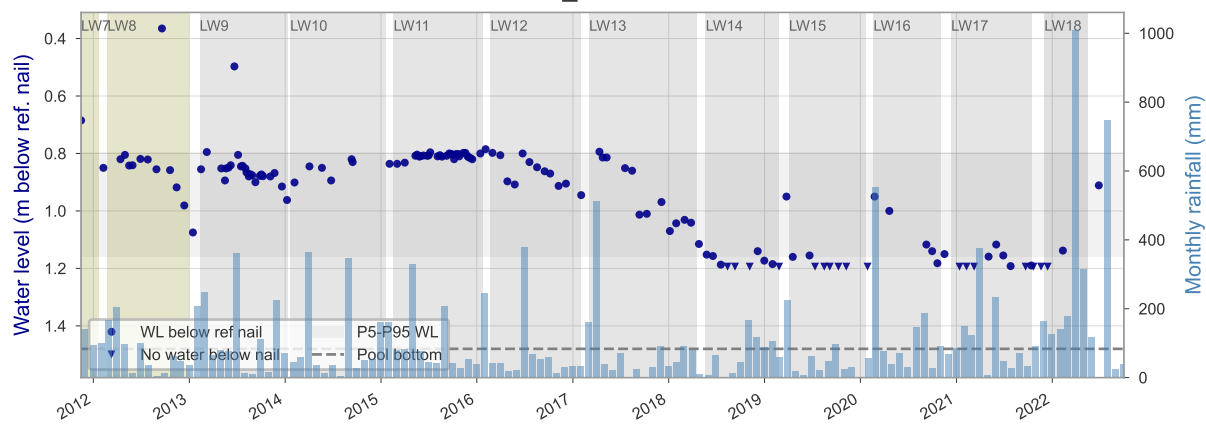
DC_POOL22



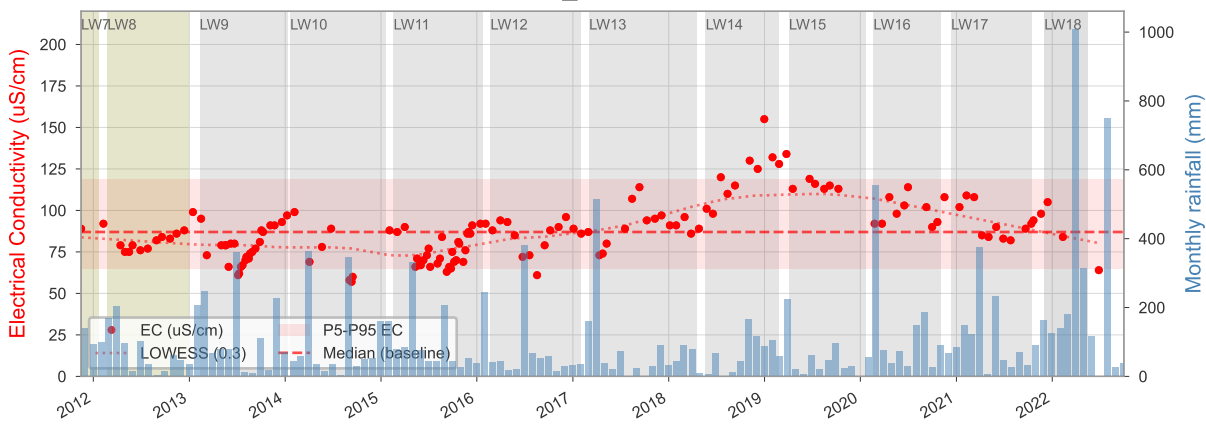
DC_POOL22



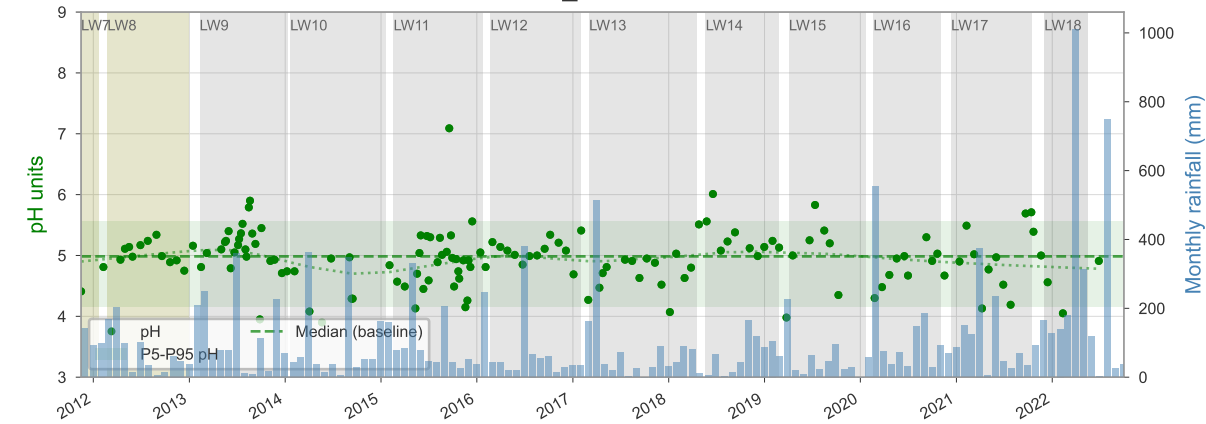
DC_POOL29



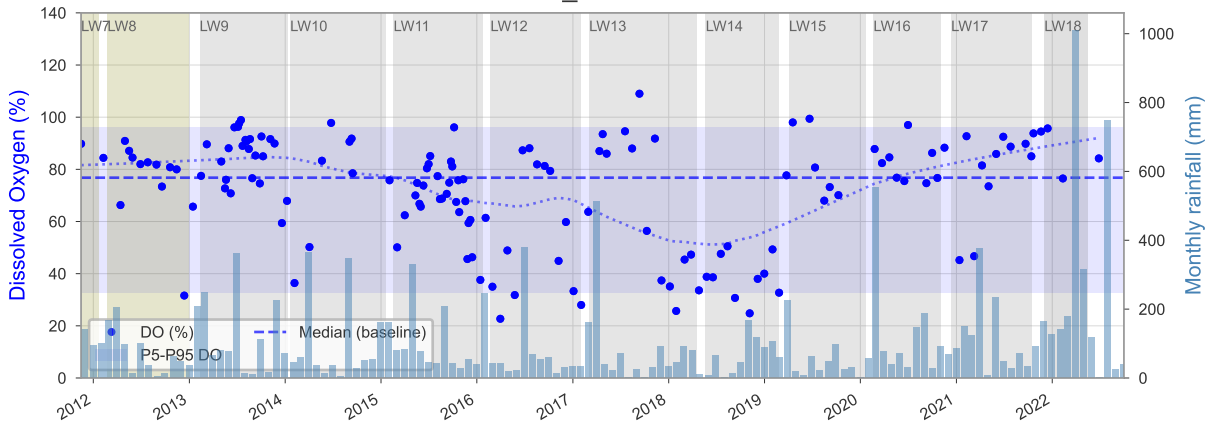
DC_POOL29



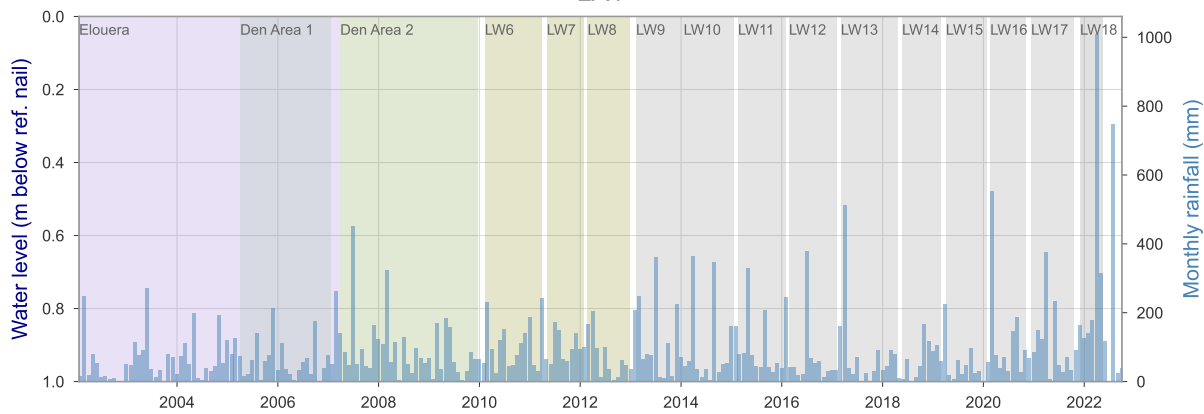
DC_POOL29



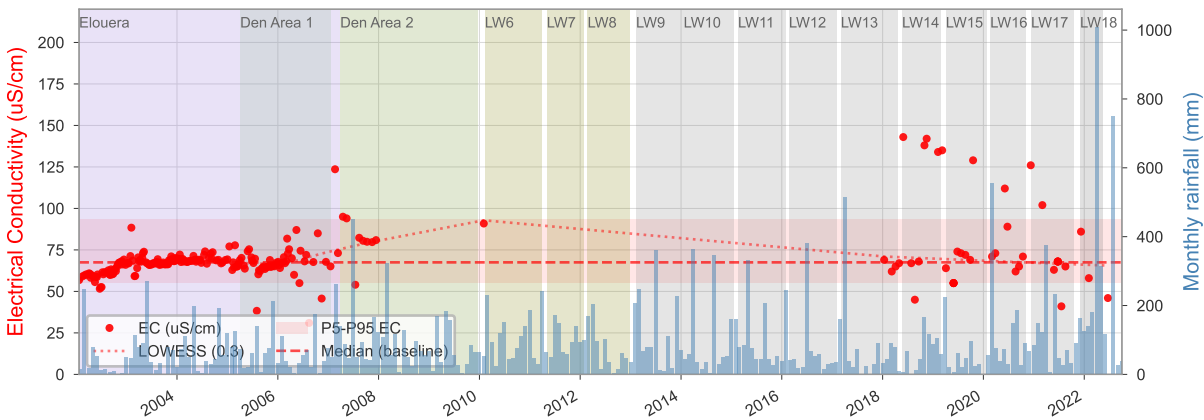
DC_POOL29



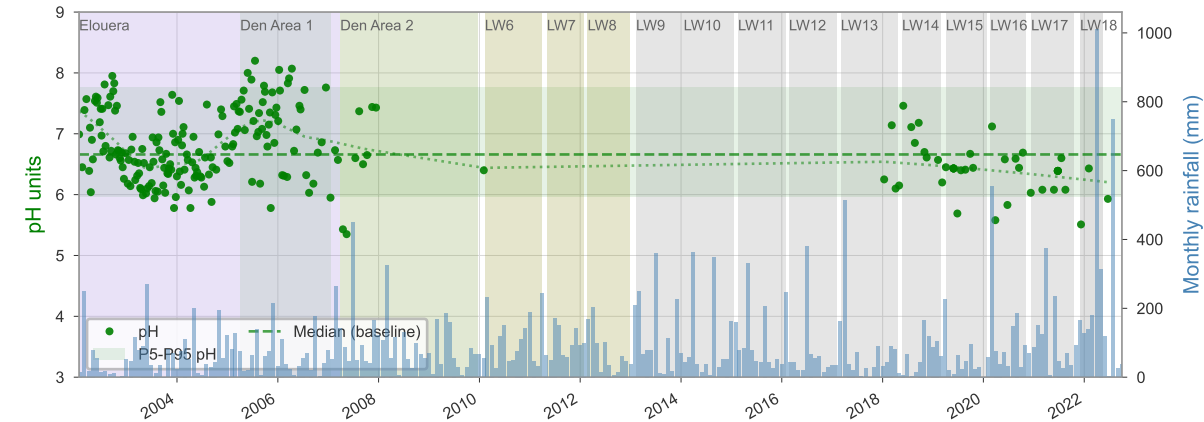
LA1



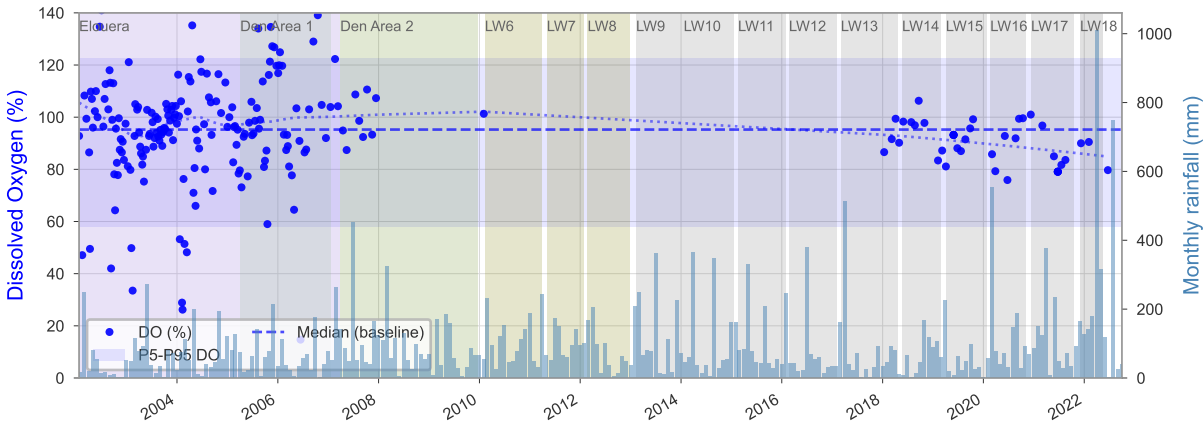
LA1



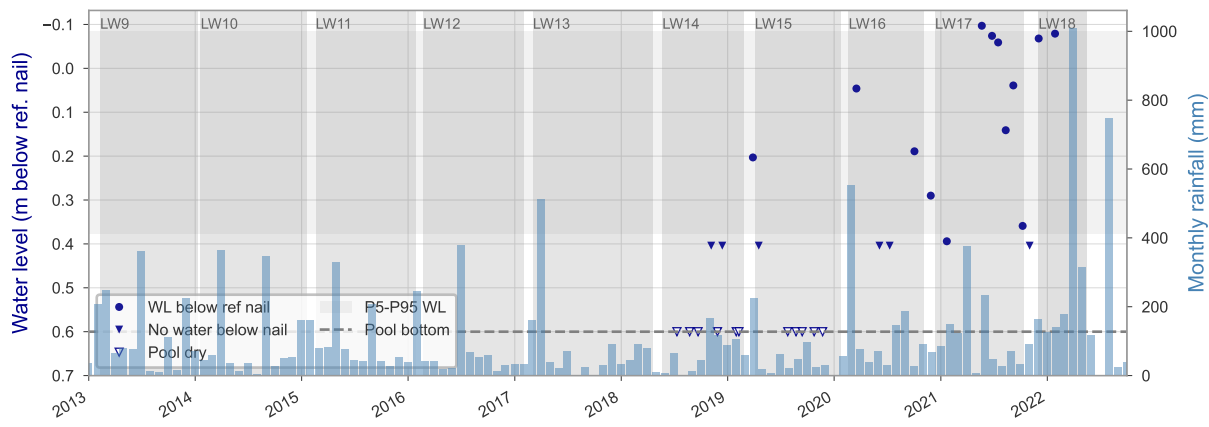
LA1



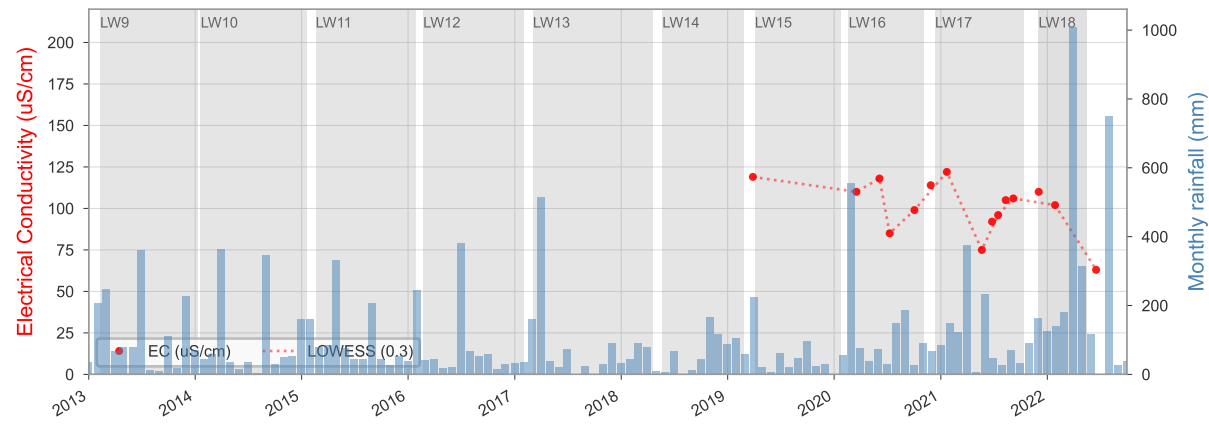
LA1



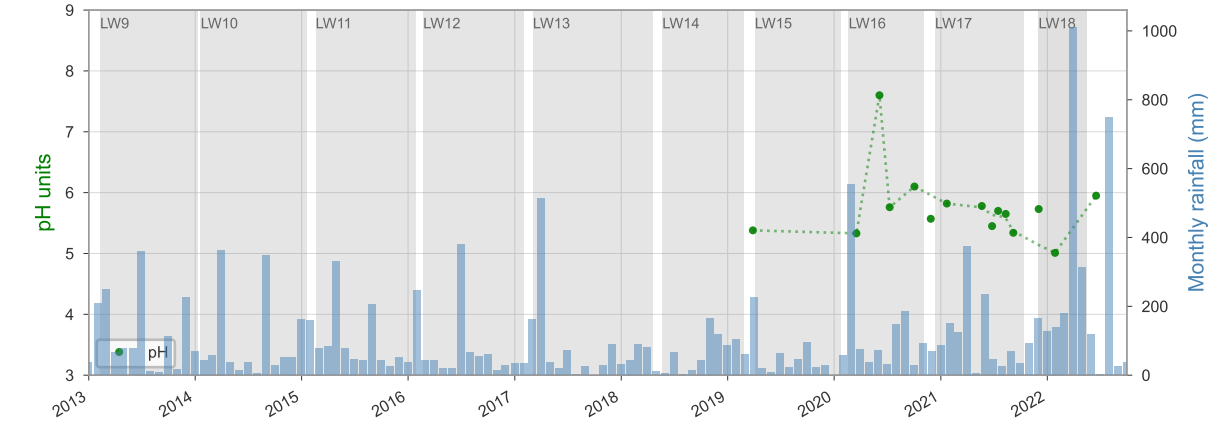
LA12_POOL1



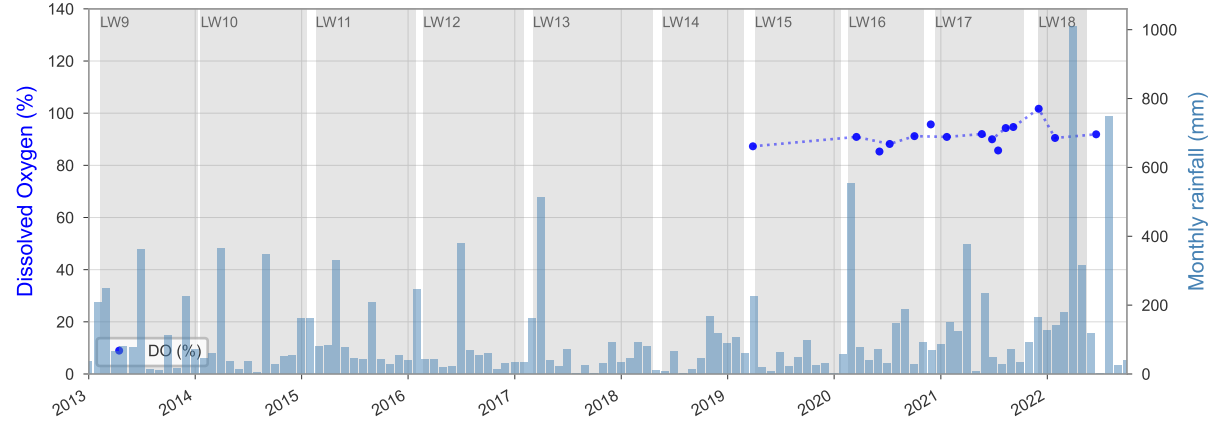
LA12_POOL1



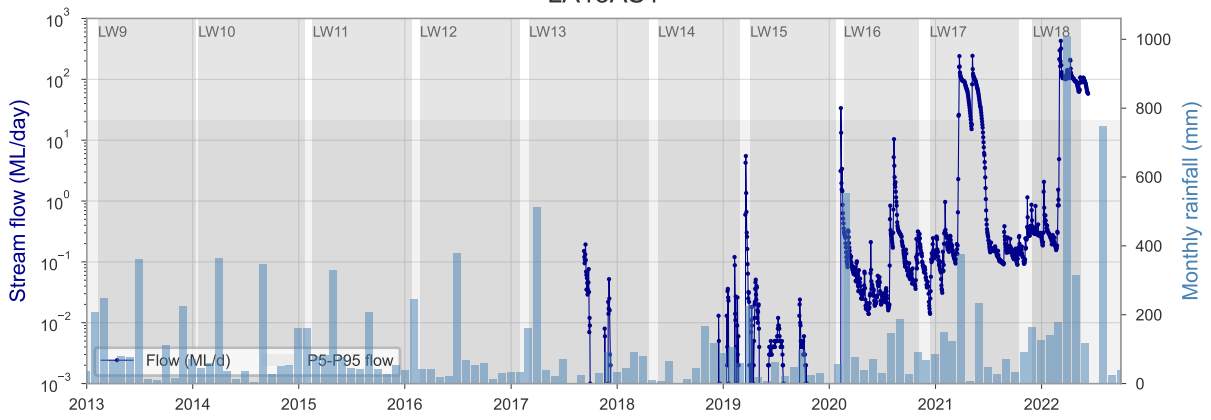
LA12_POOL1



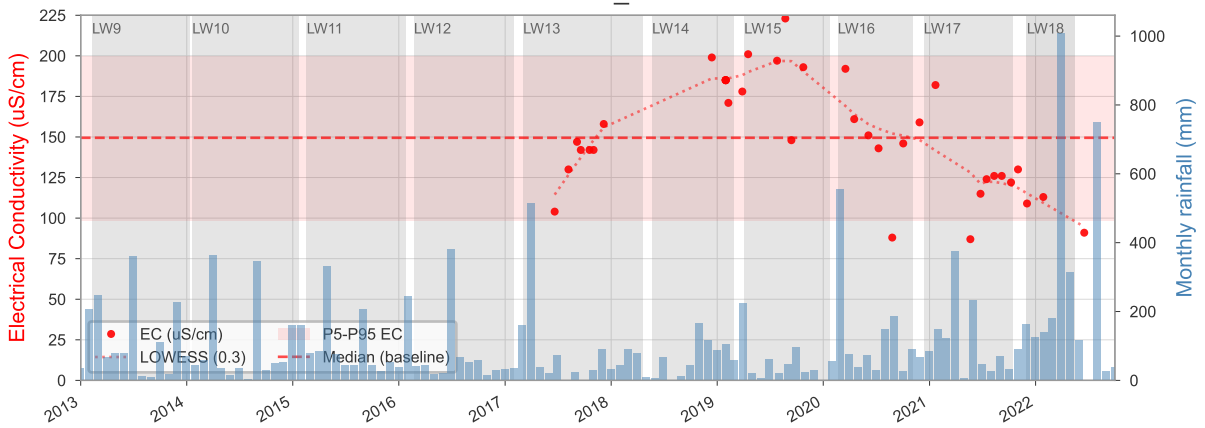
LA12_POOL1



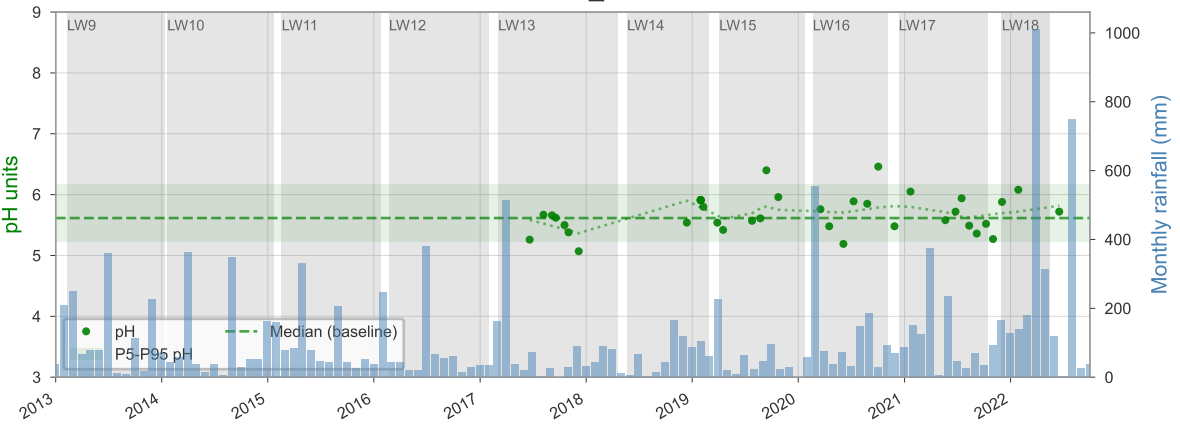
LA13AS1



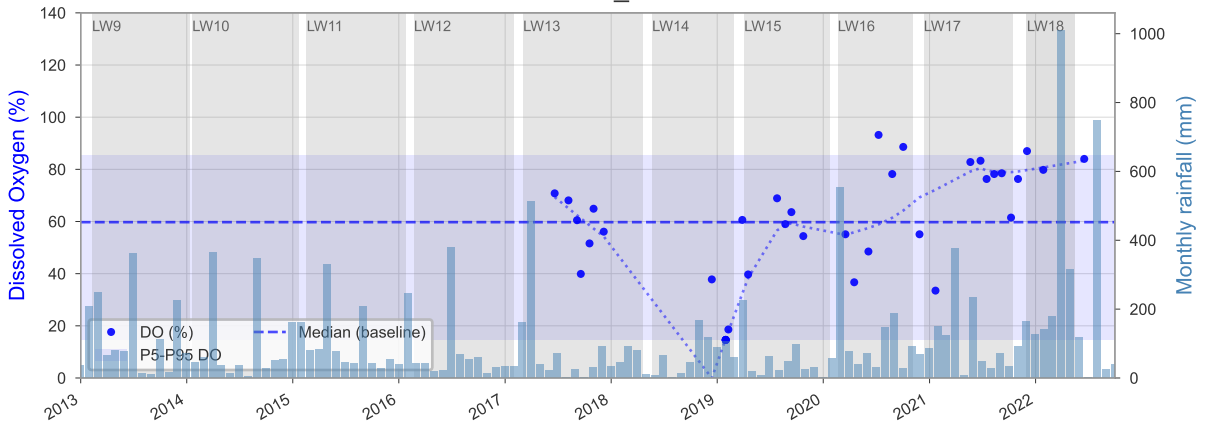
LA13A_S1



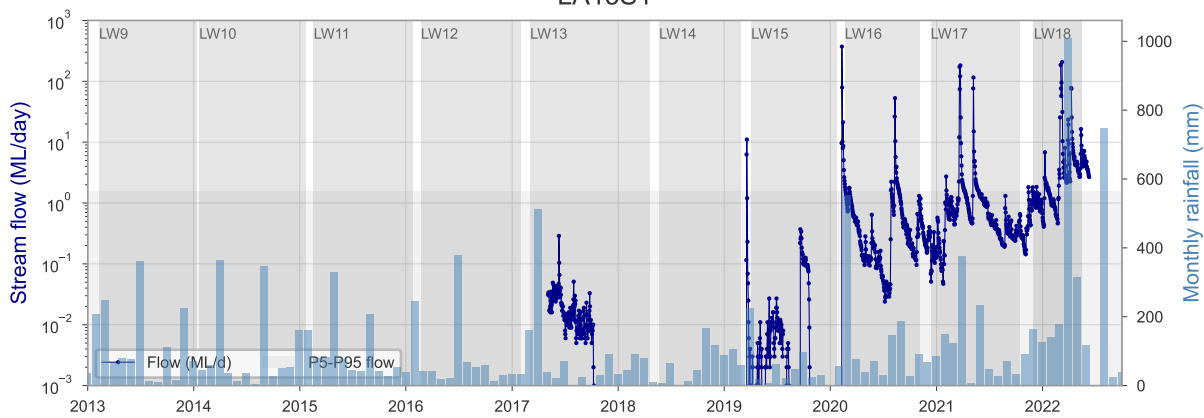
LA13A_S1



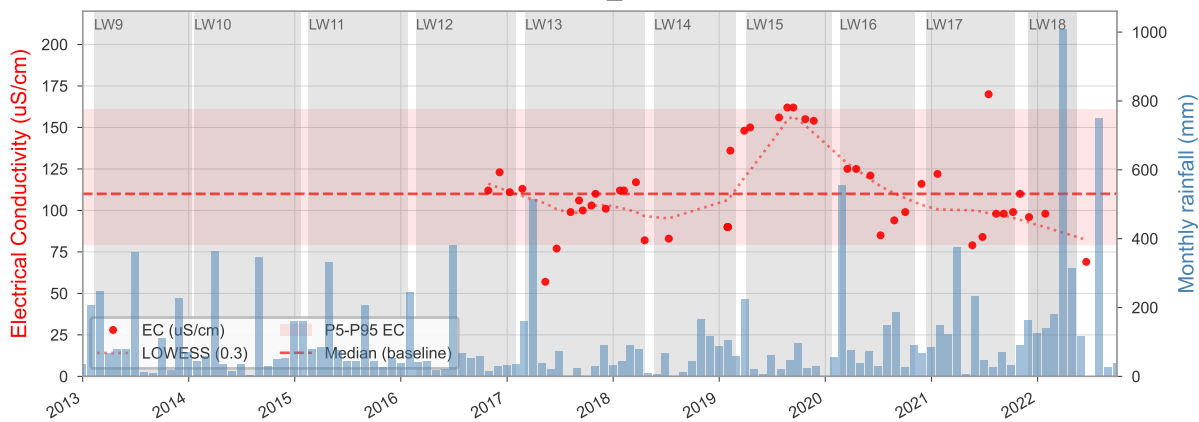
LA13A_S1



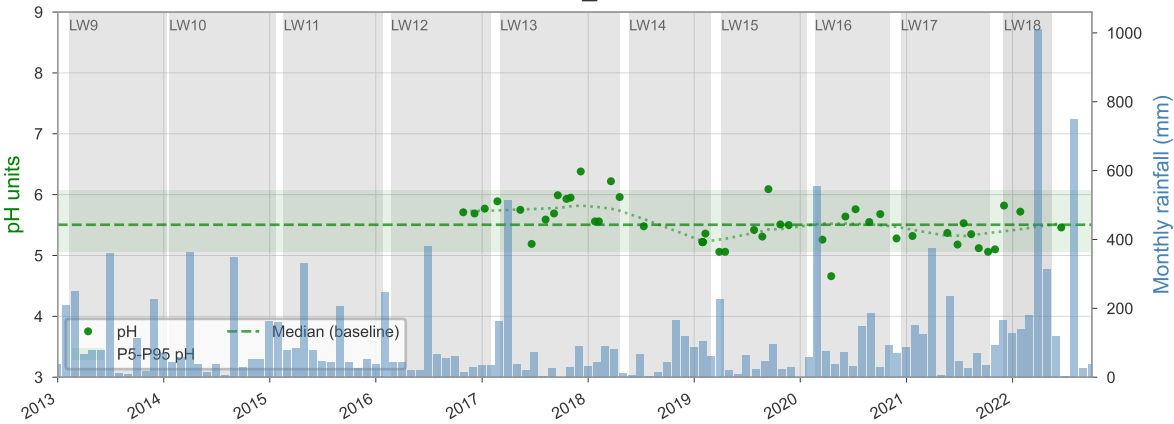
LA13S1



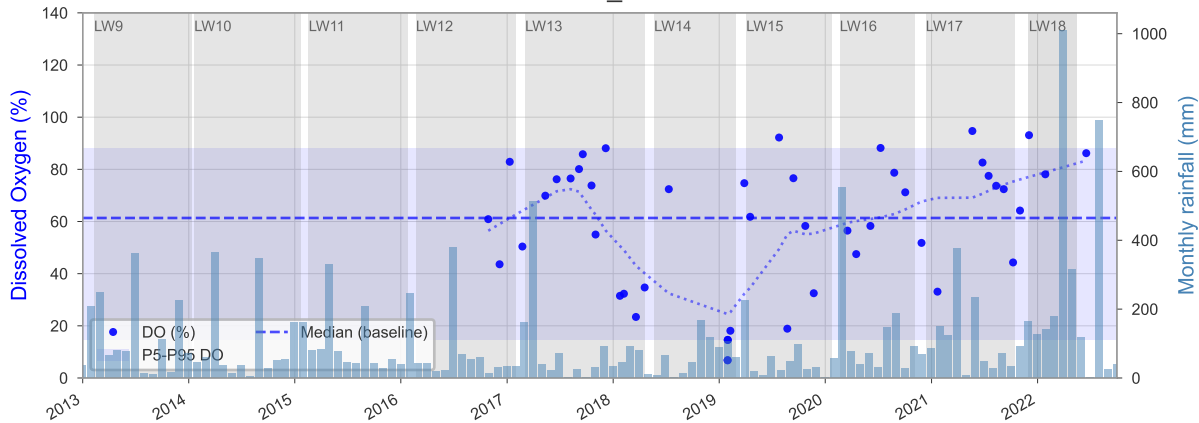
LA13_S1



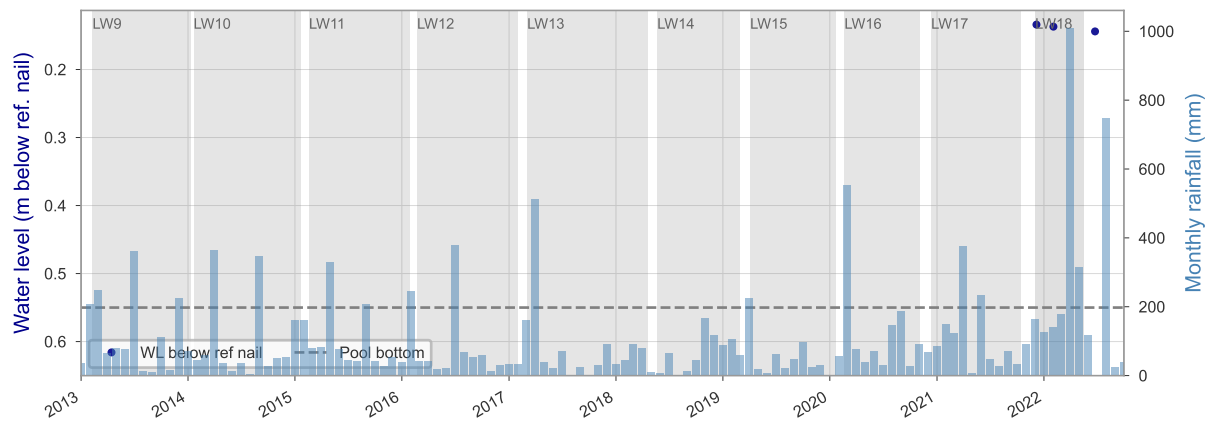
LA13_S1



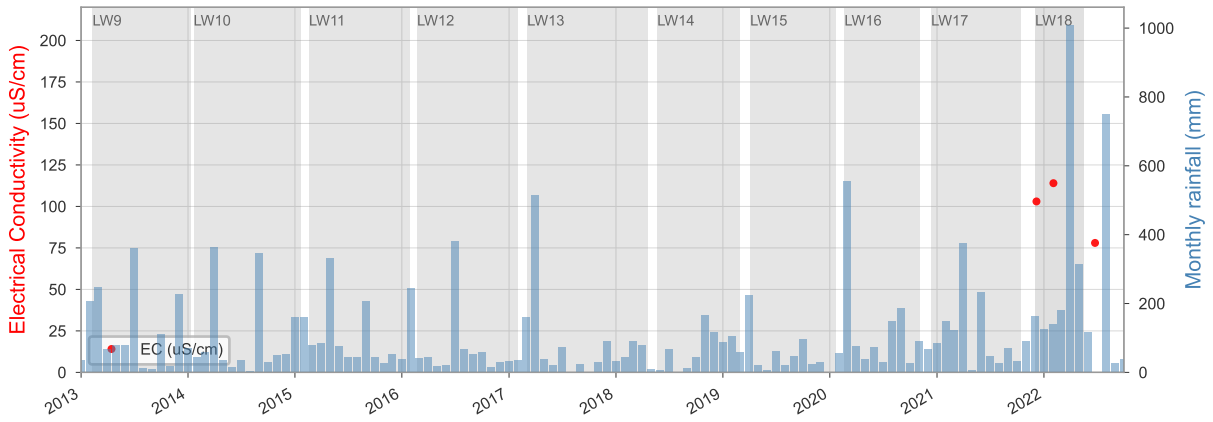
LA13_S1



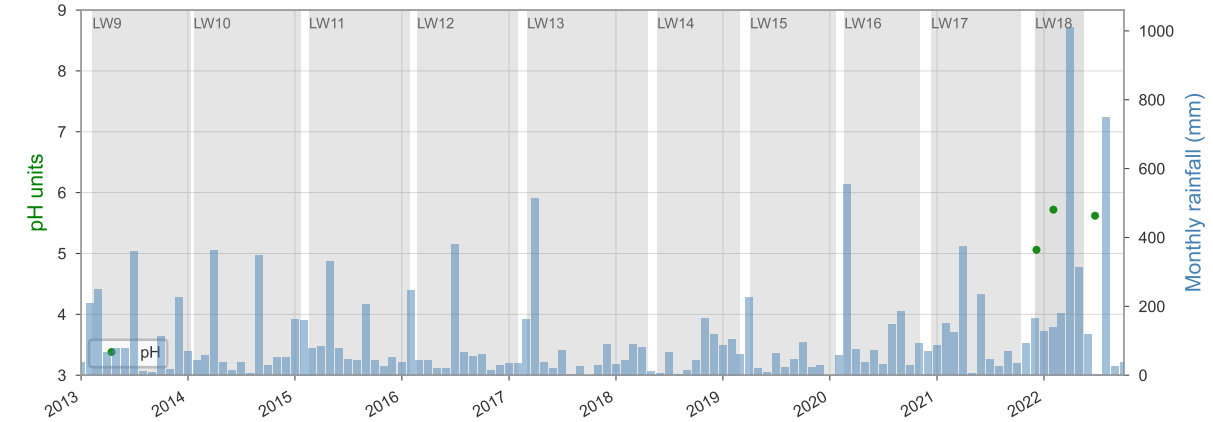
LA14_CHANNEL2



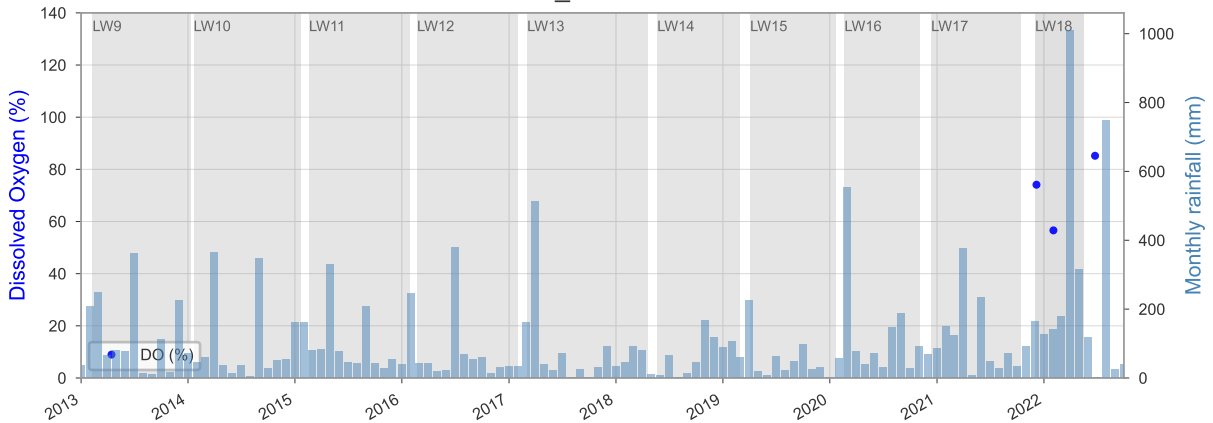
LA14_CHANNEL2



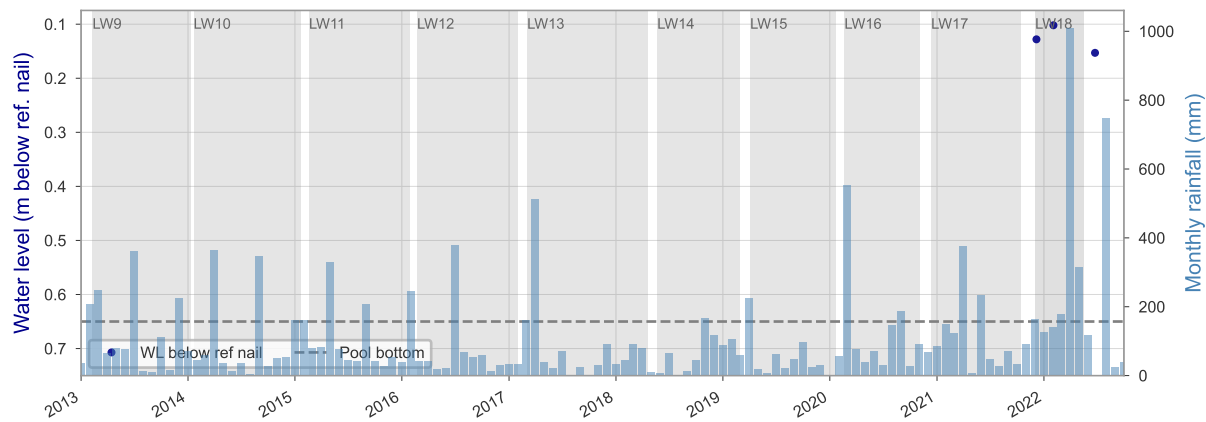
LA14_CHANNEL2



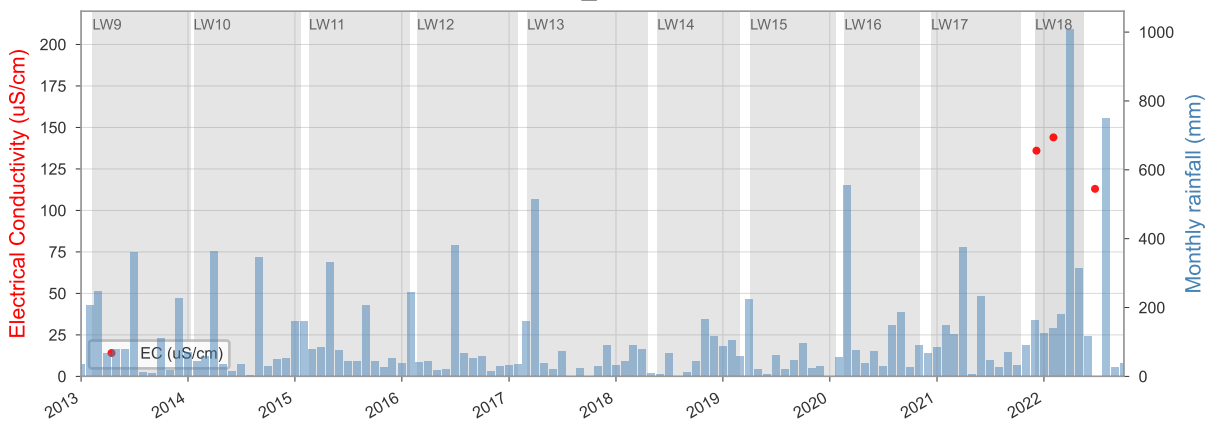
LA14_CHANNEL2



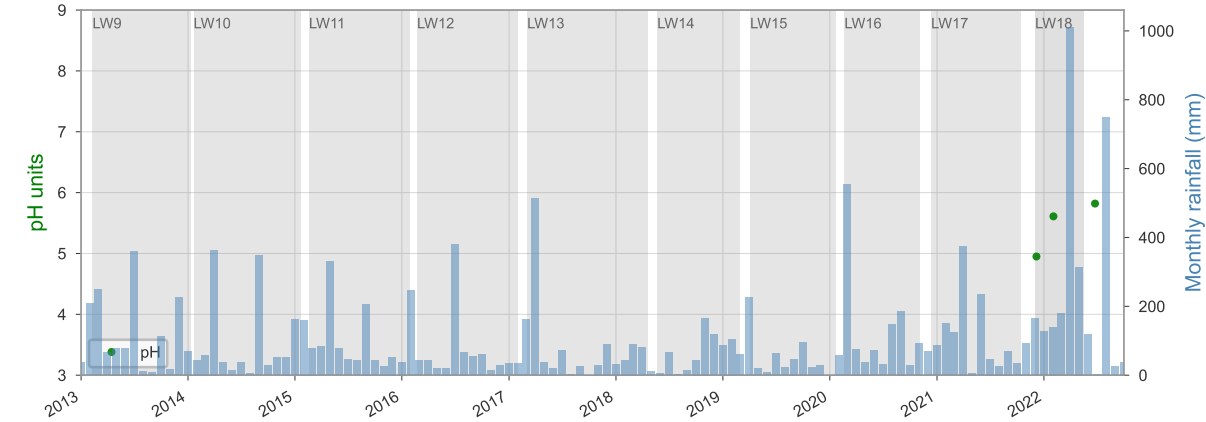
LA15_POOL1



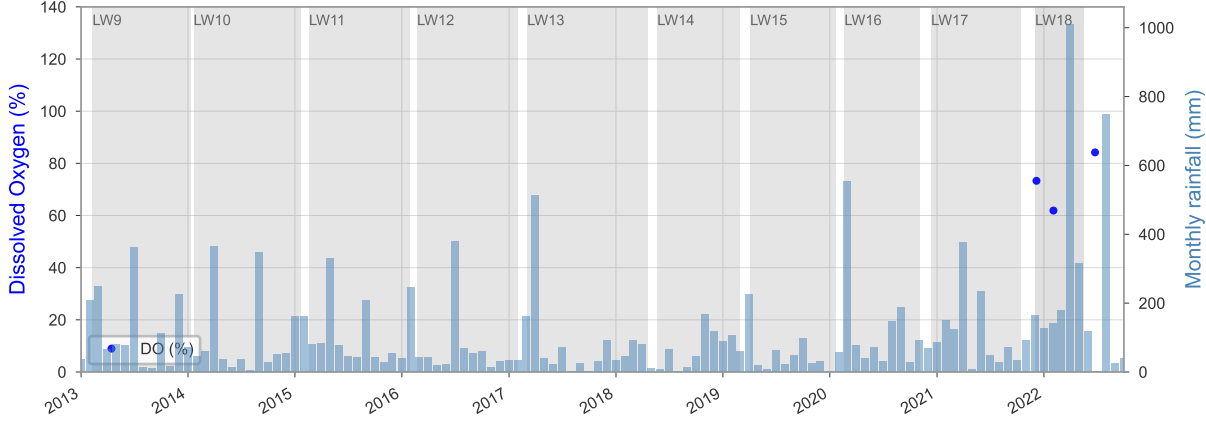
LA15_POOL1



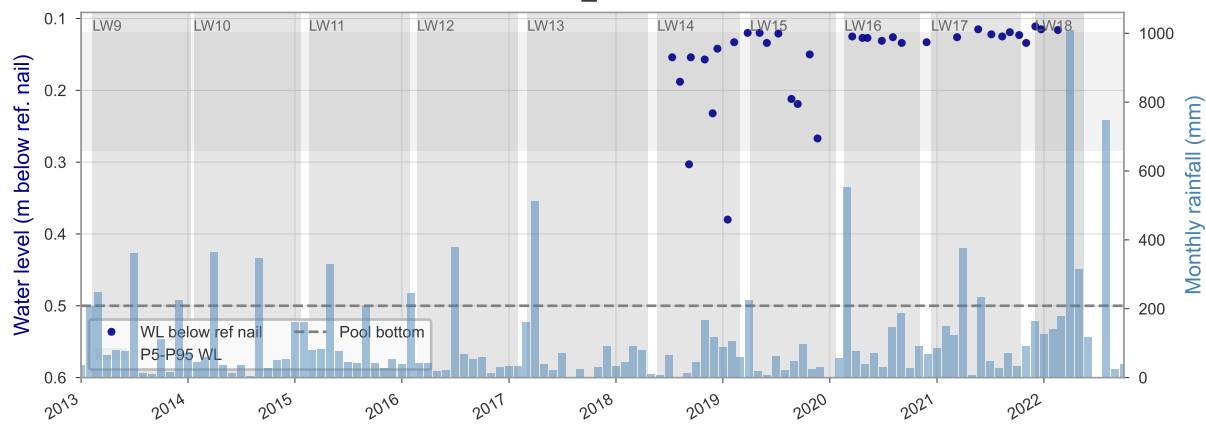
LA15_POOL1



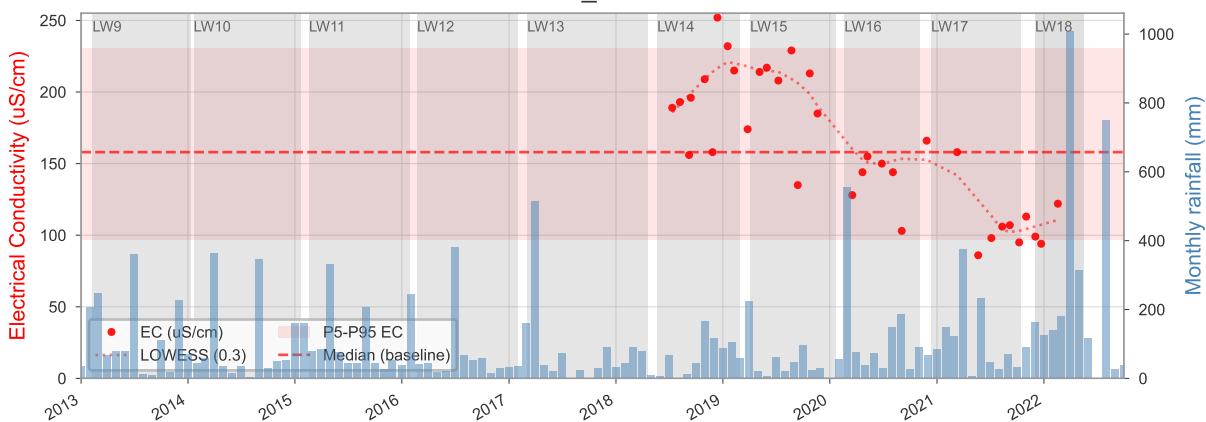
LA15_POOL1



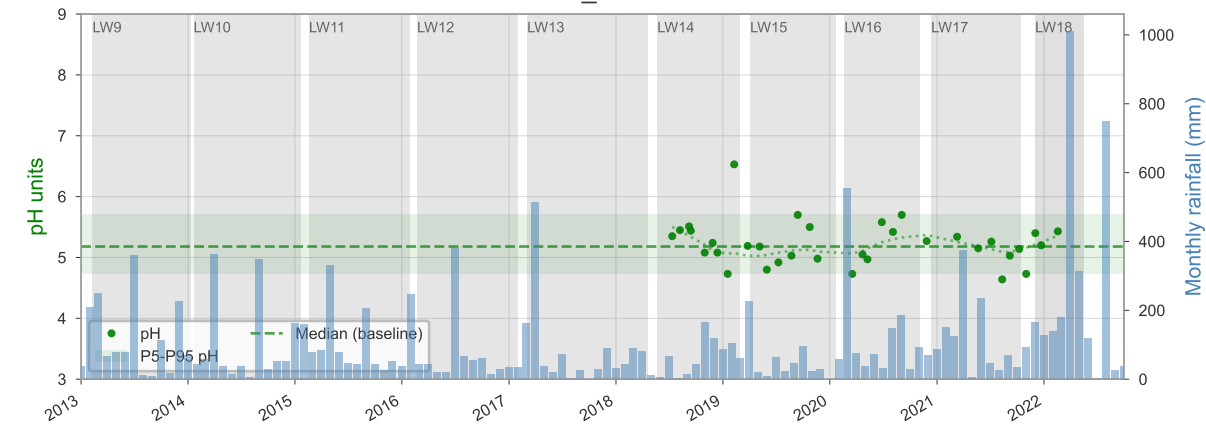
LA17_POOL0



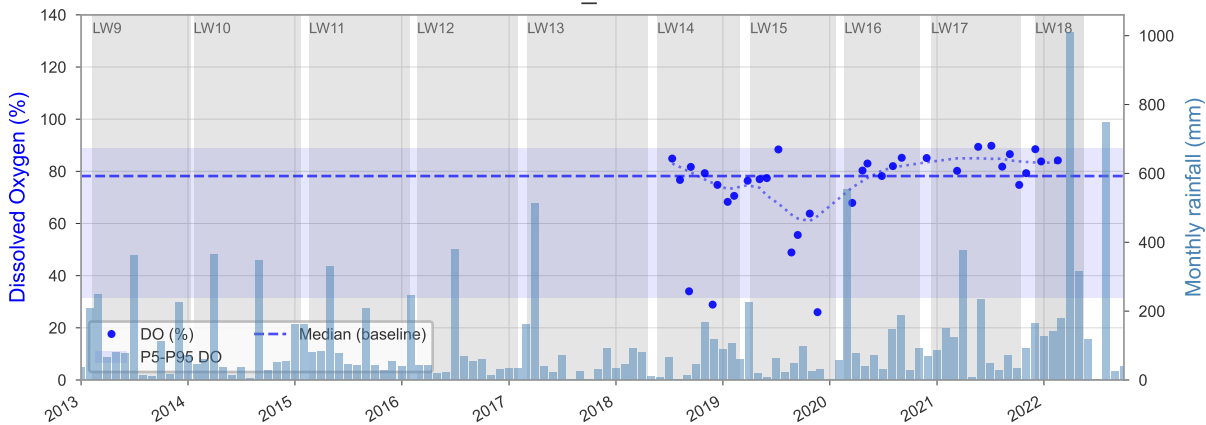
LA17_POOL0



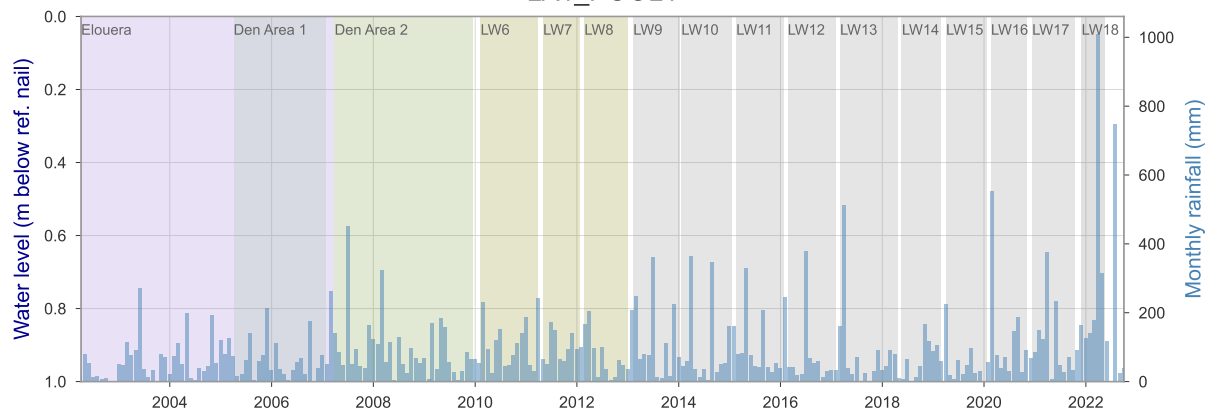
LA17_POOL0



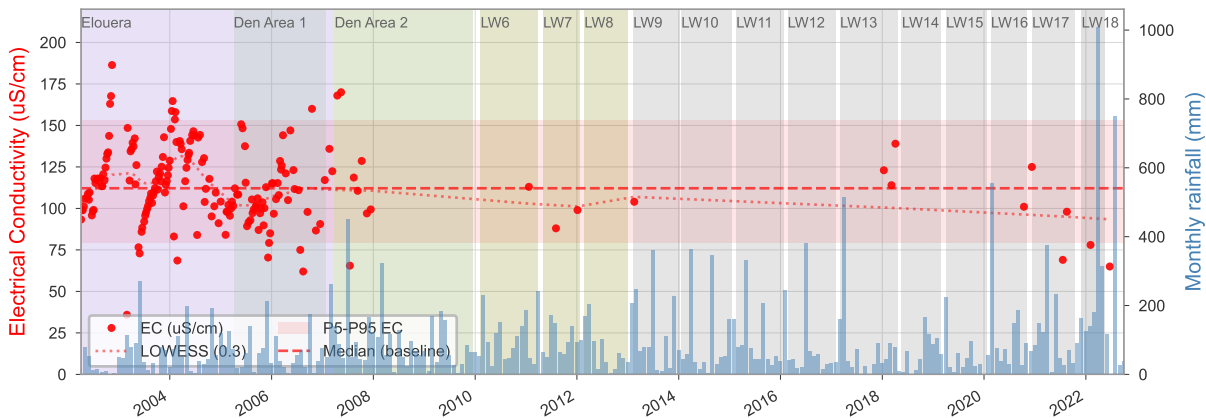
LA17_POOL0



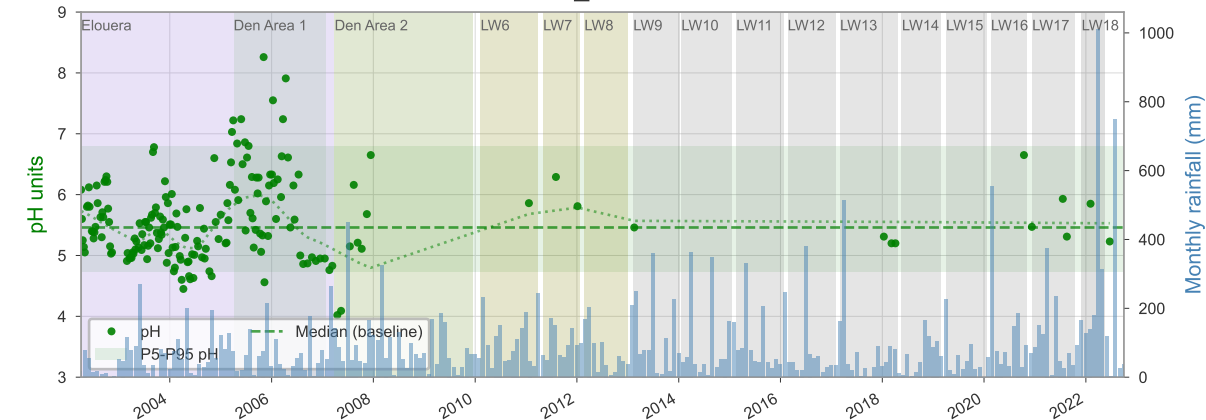
LA1_POOL1



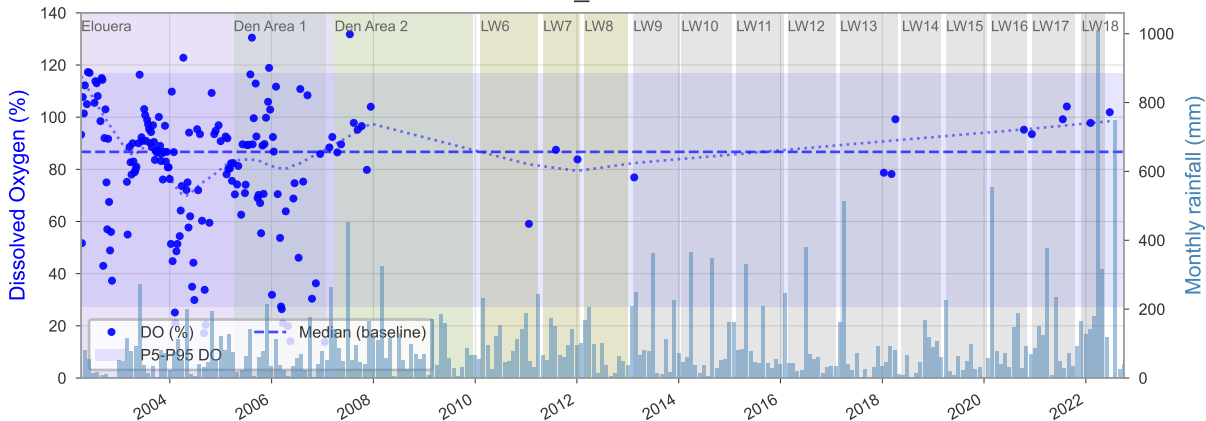
LA1_POOL1



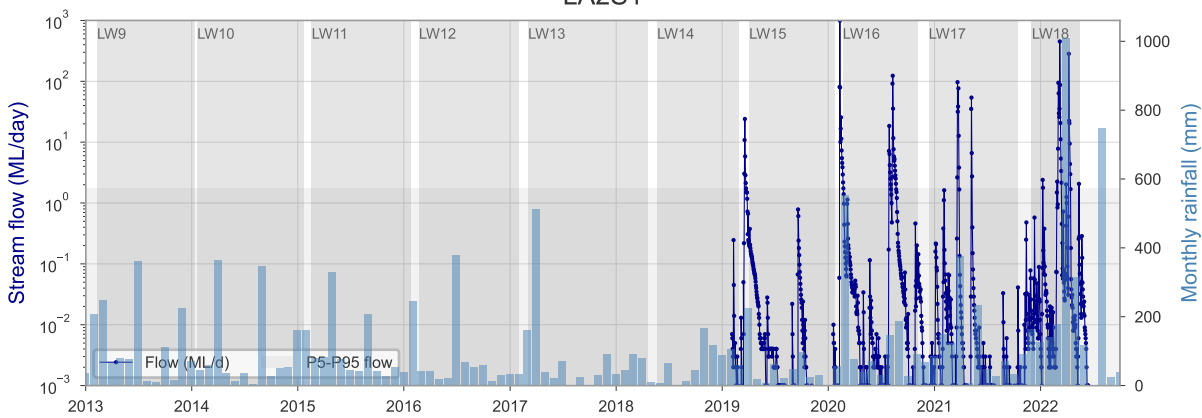
LA1_POOL1



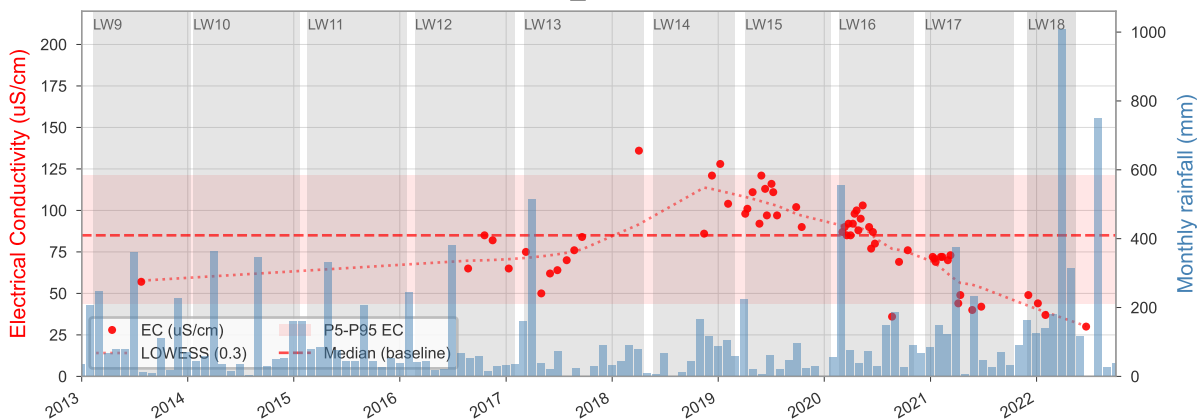
LA1_POOL1



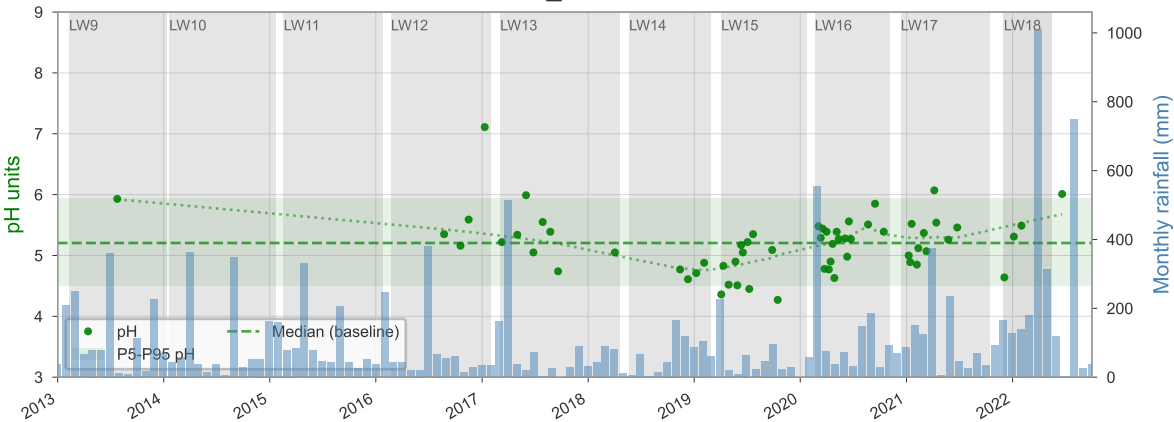
LA2S1



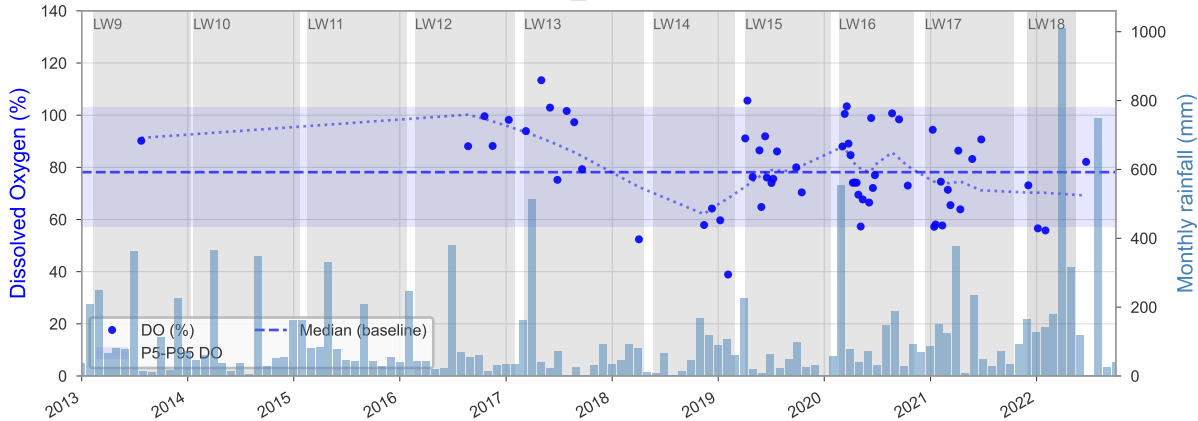
LA2_POOL5



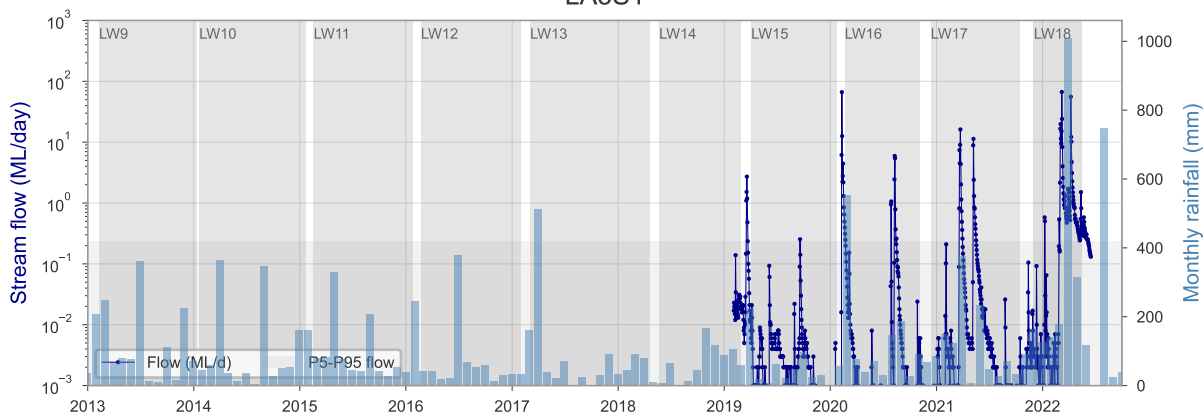
LA2_POOL5



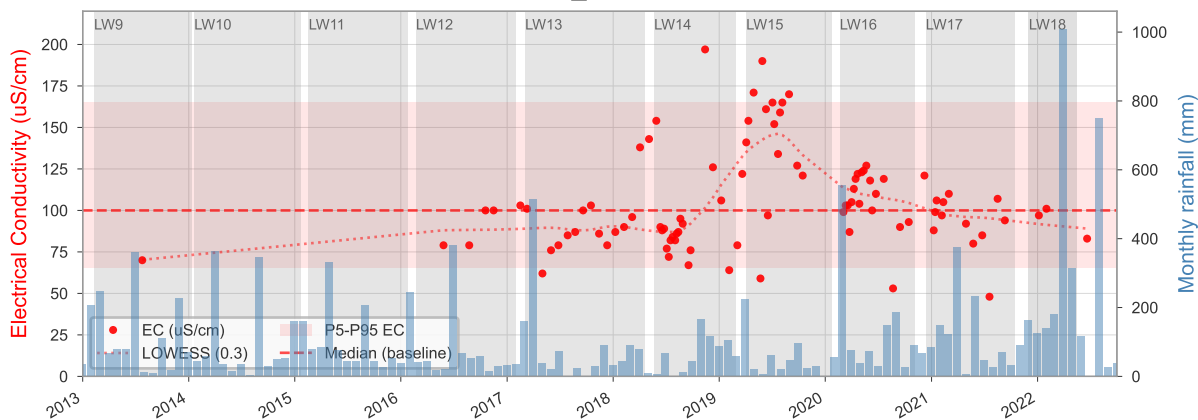
LA2_POOL5



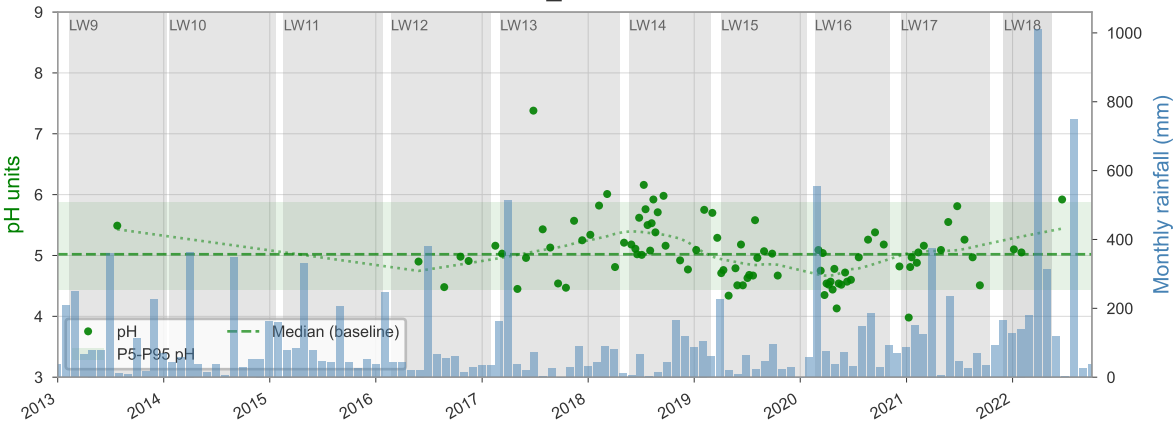
LA3S1



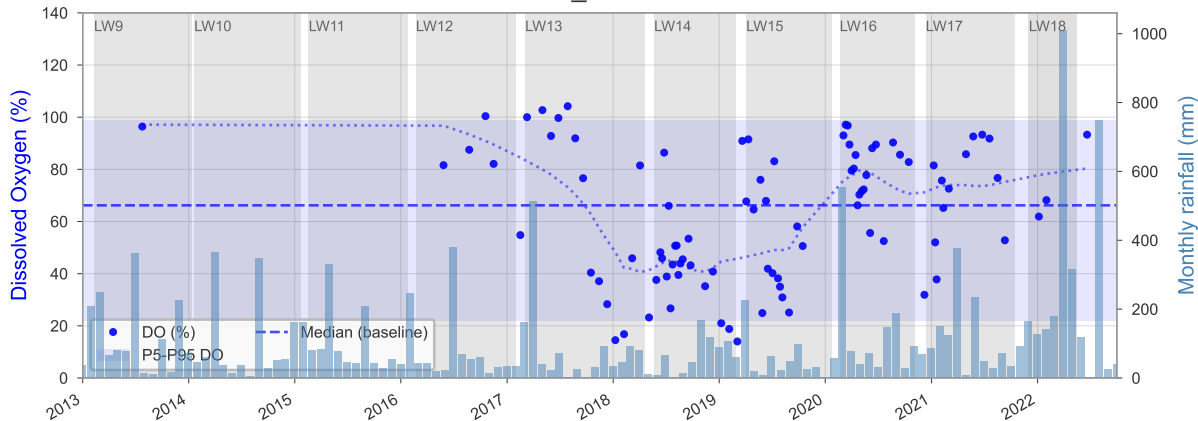
LA3_POOL4



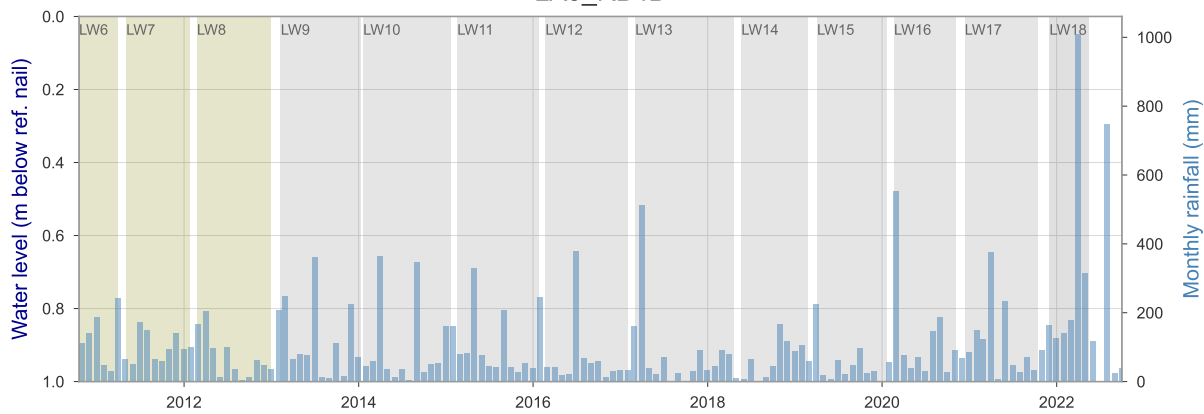
LA3_POOL4



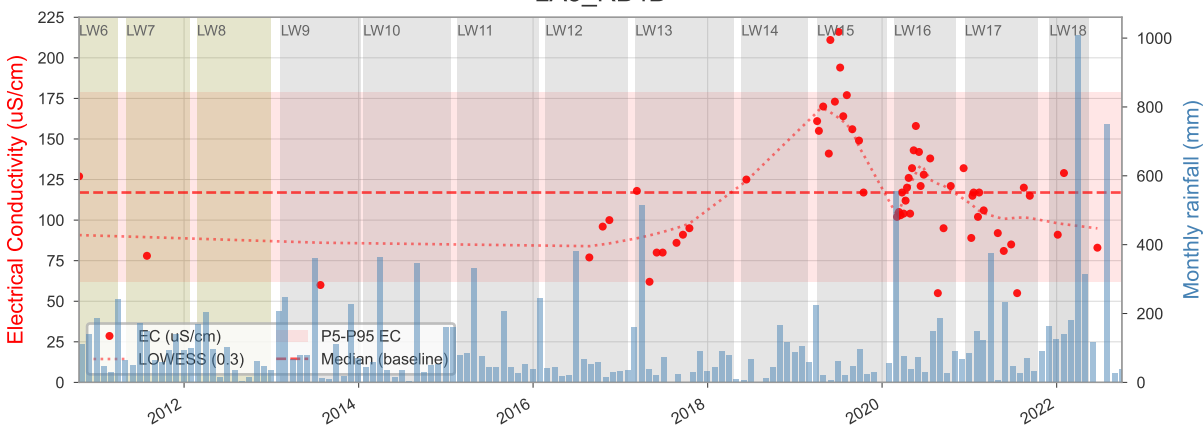
LA3_POOL4



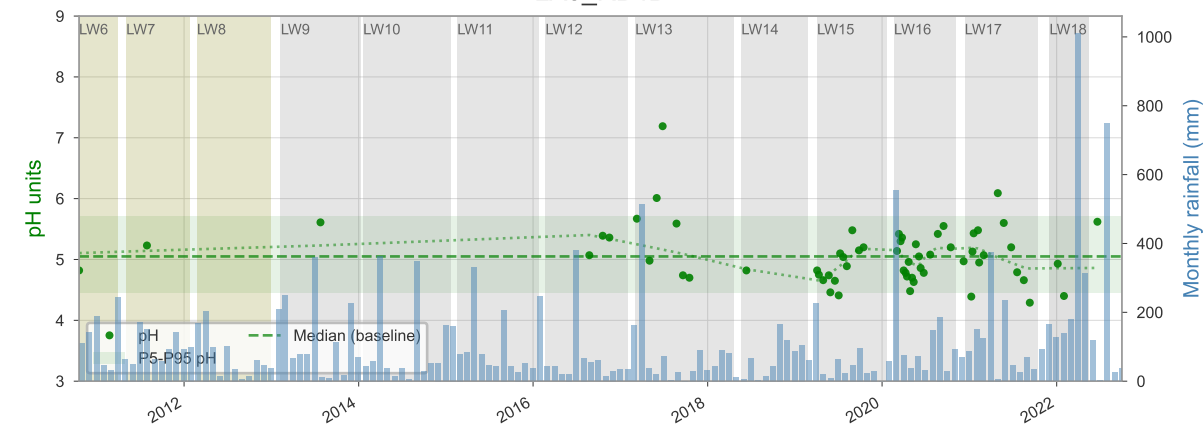
LA3_RB4B



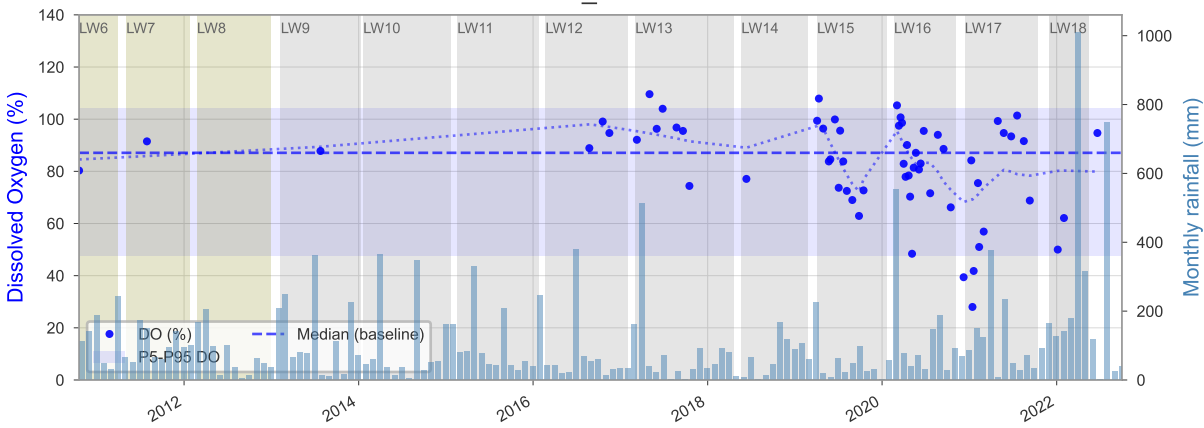
LA3_RB4B



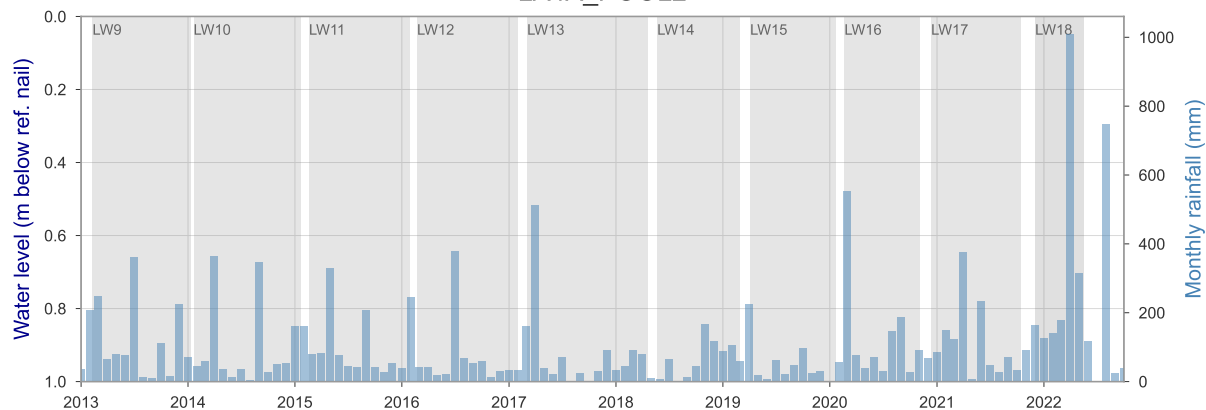
LA3_RB4B



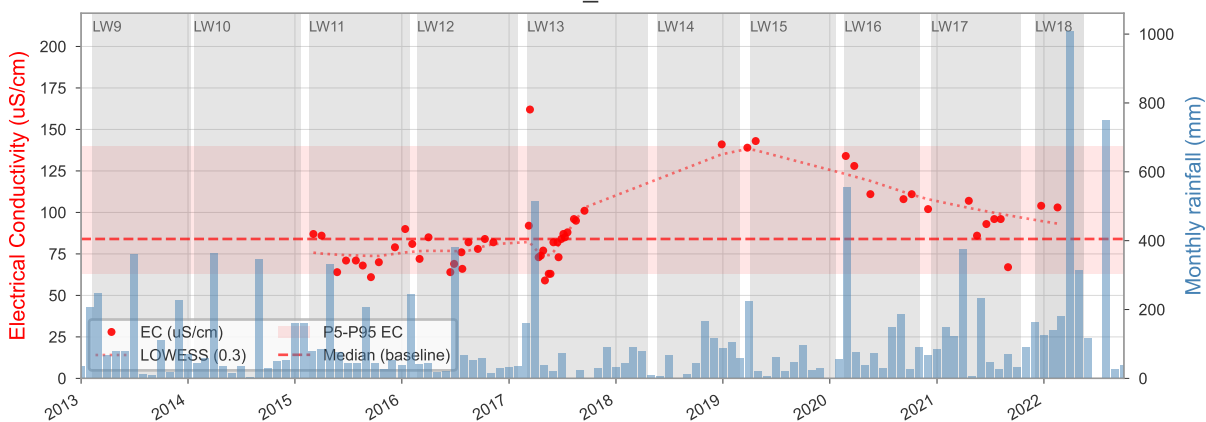
LA3_RB4B



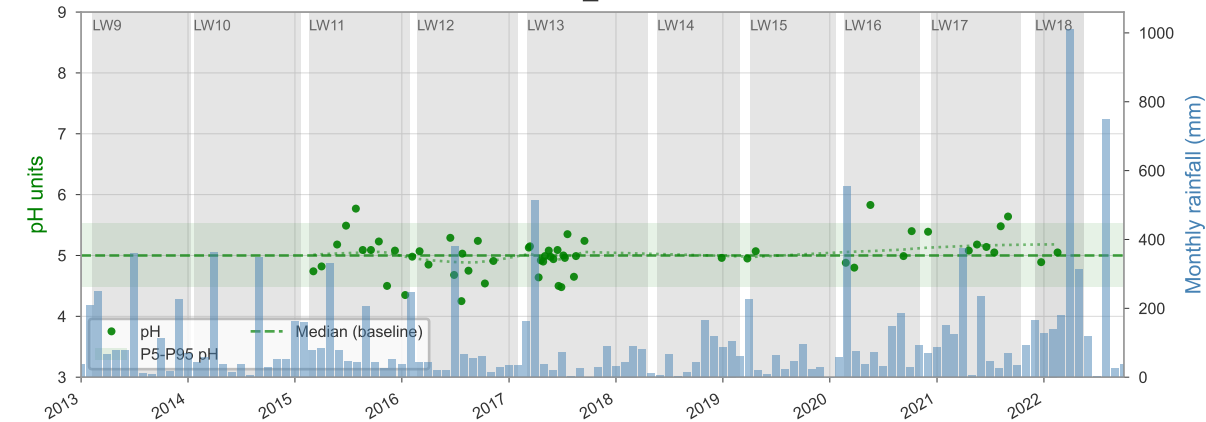
LA4A_POOL2



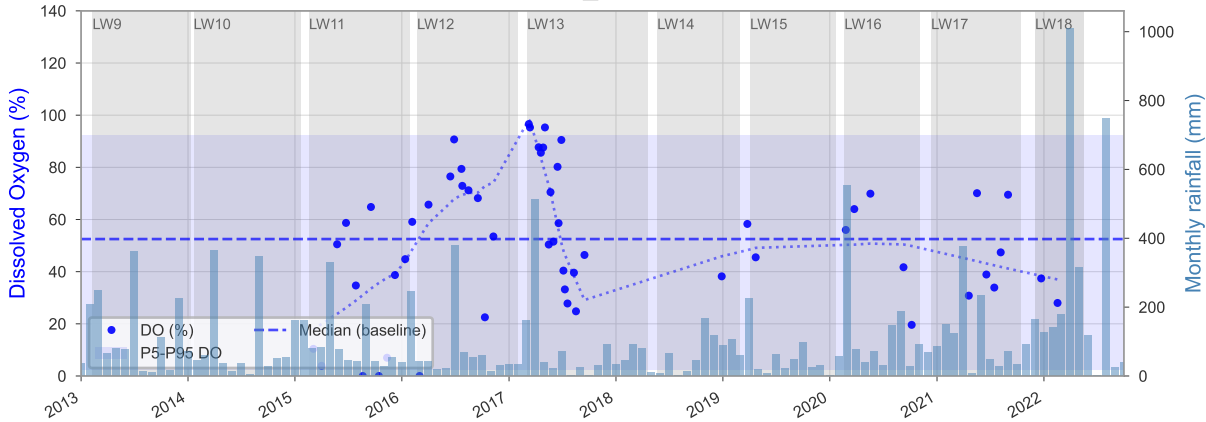
LA4A_POOL2



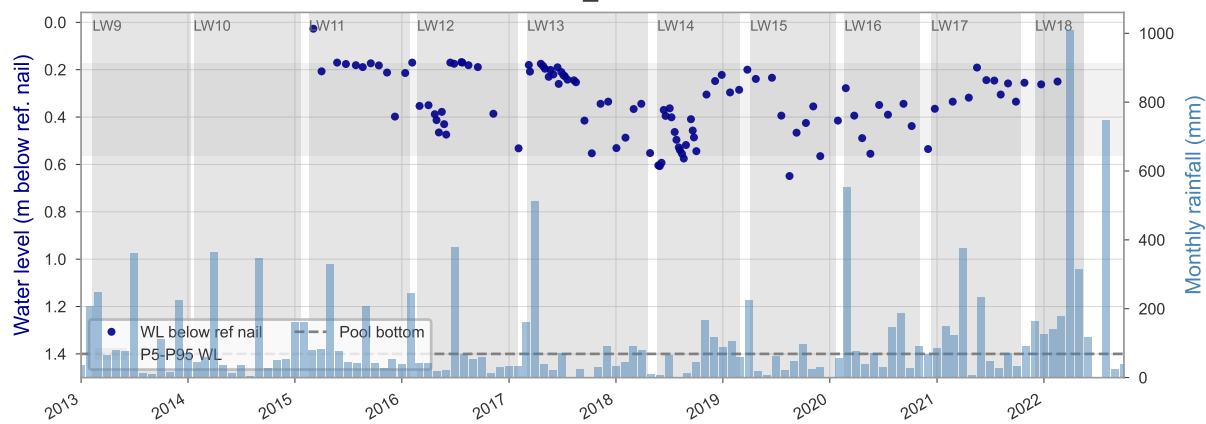
LA4A_POOL2



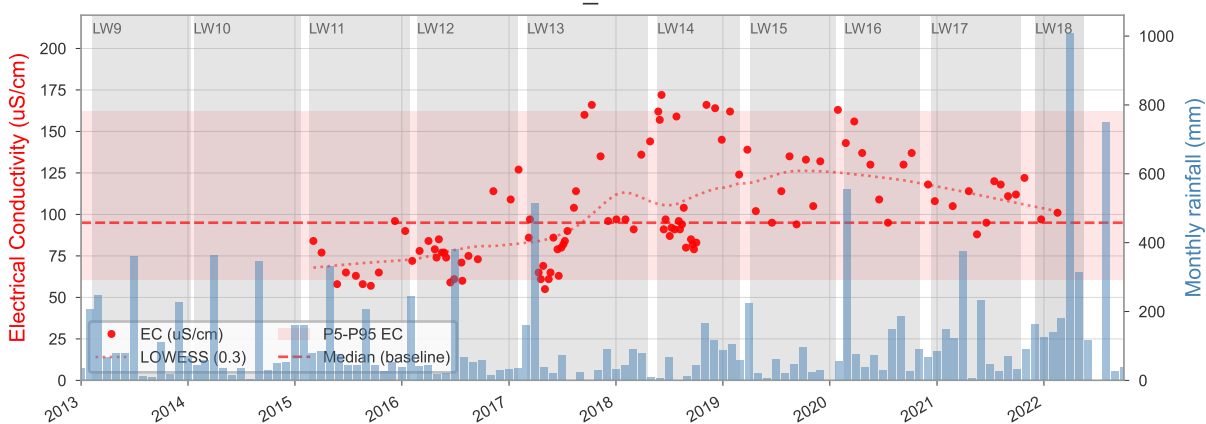
LA4A_POOL2



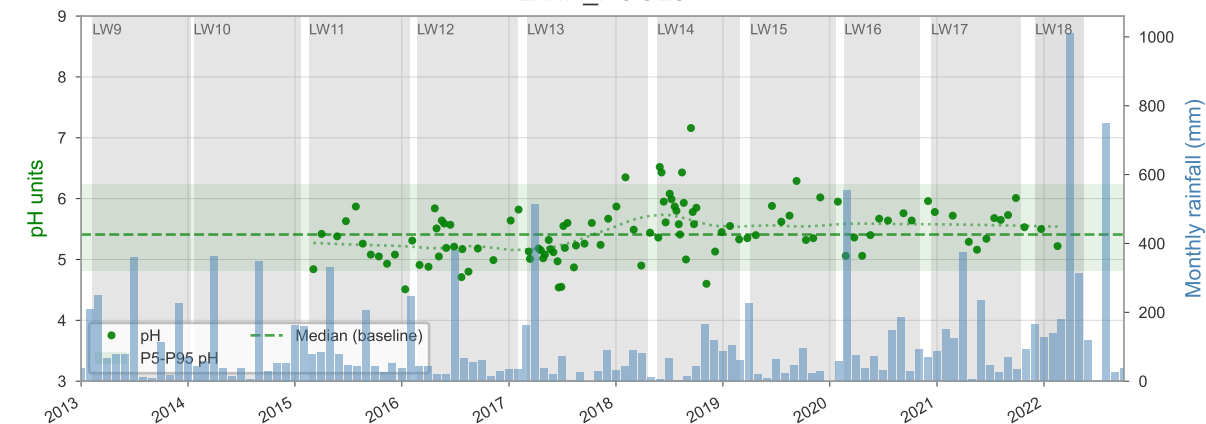
LA4A_POOL3



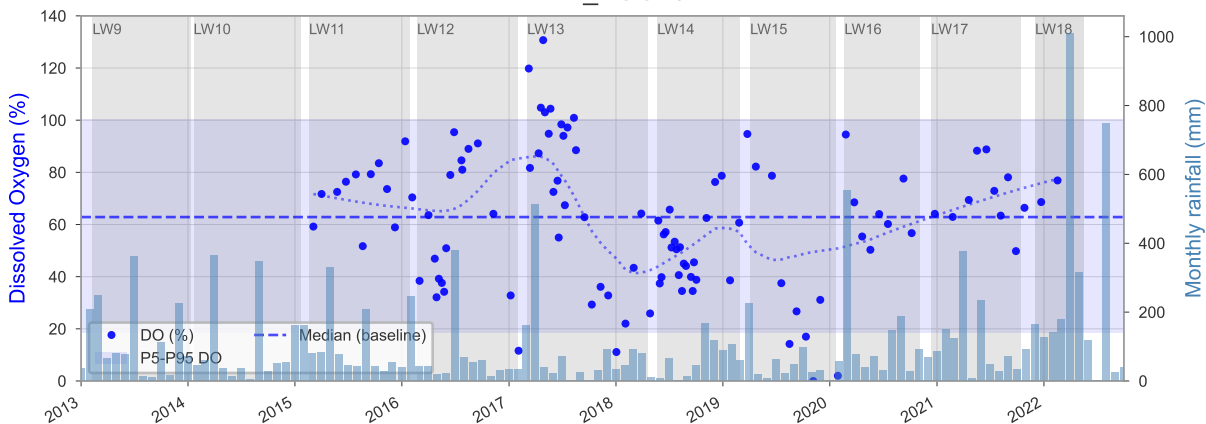
LA4A_POOL3



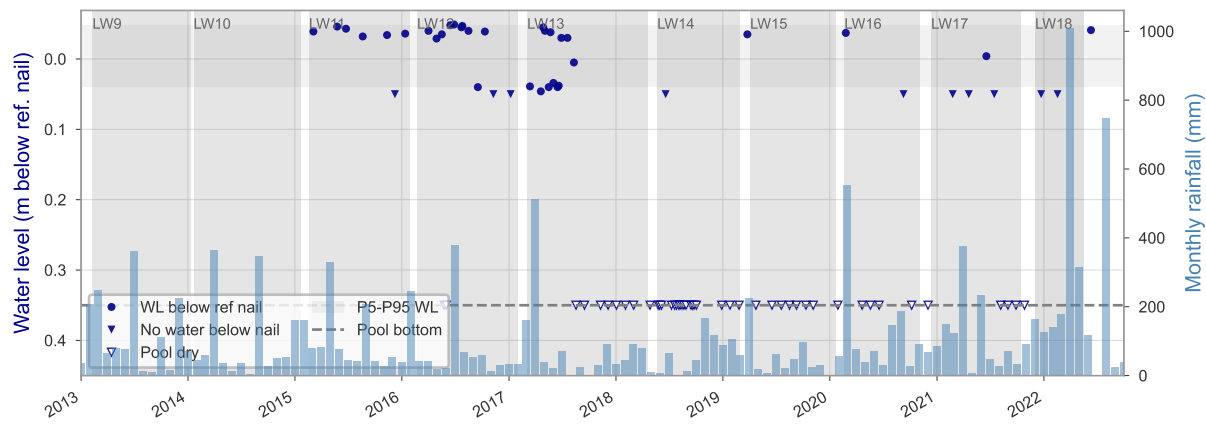
LA4A_POOL3



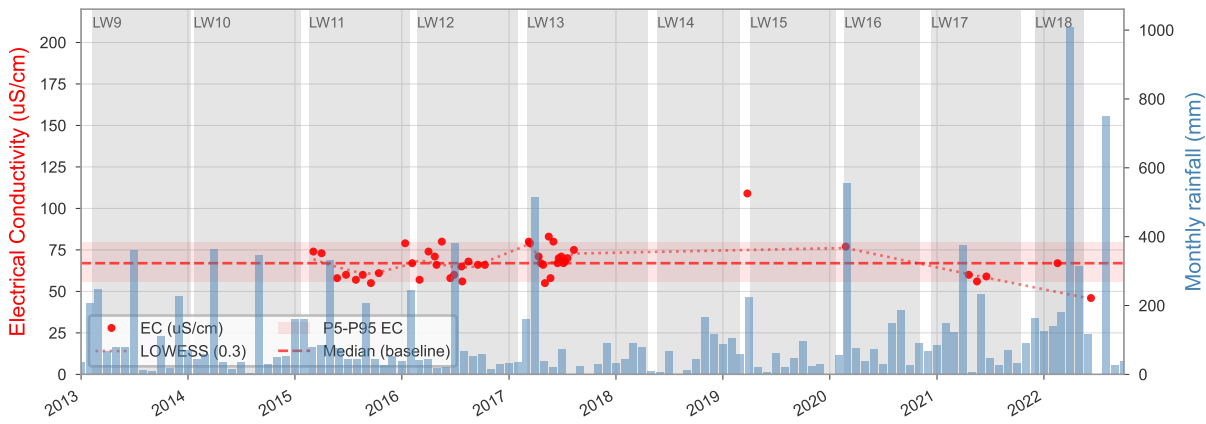
LA4A_POOL3



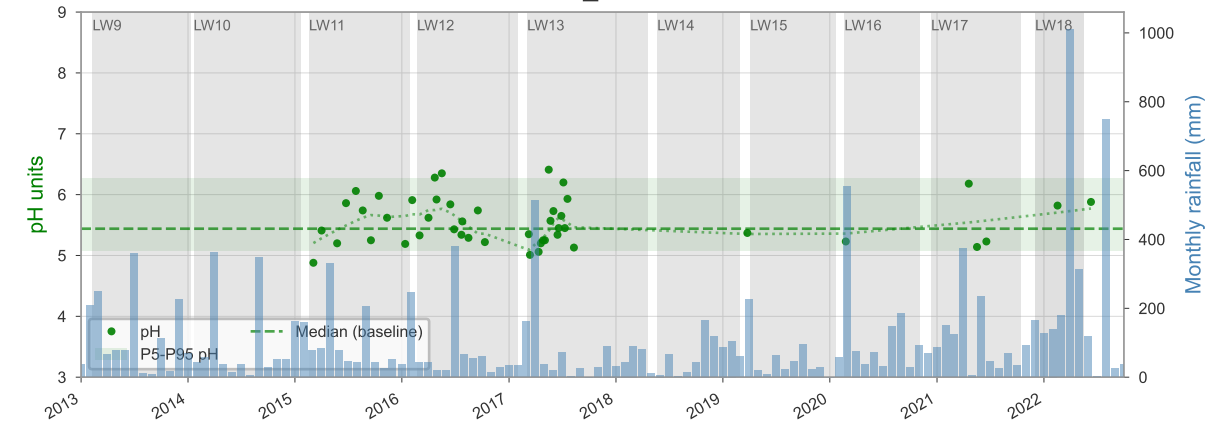
LA4A_POOL4



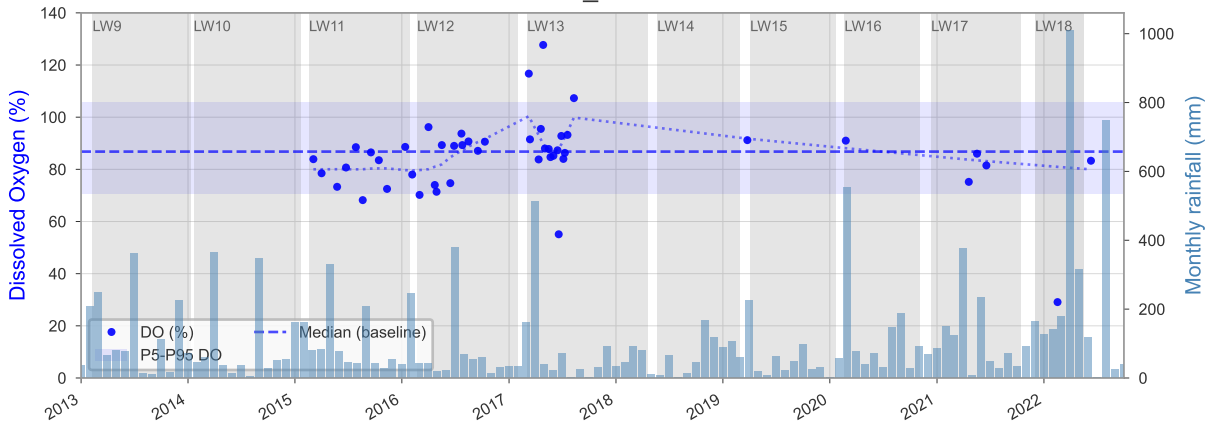
LA4A_POOL4



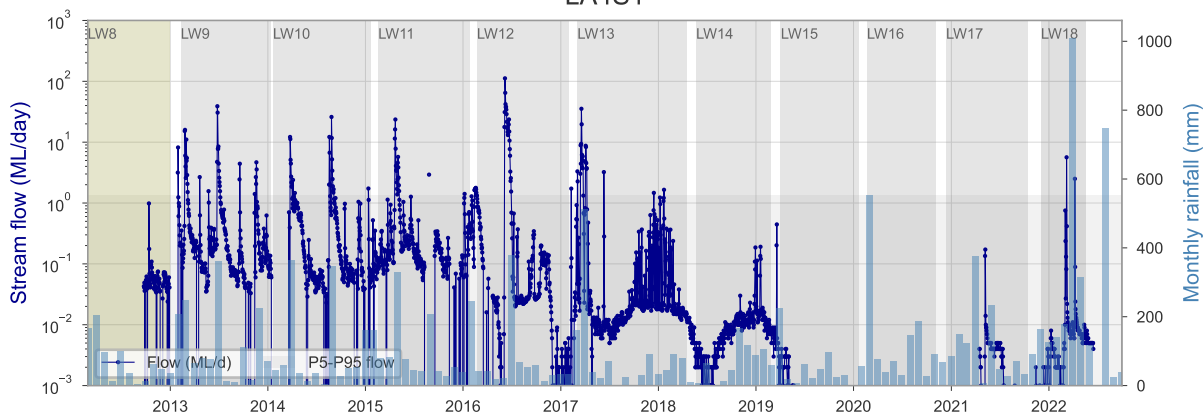
LA4A_POOL4



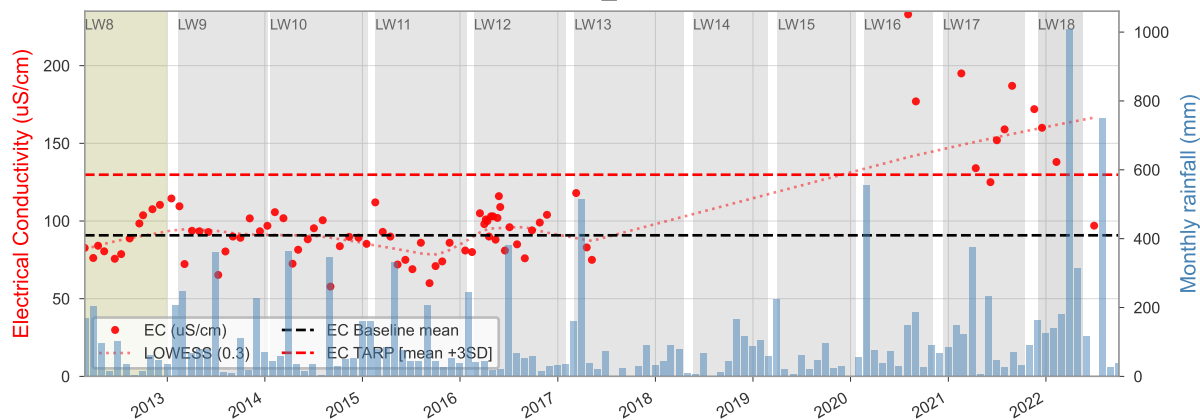
LA4A_POOL4



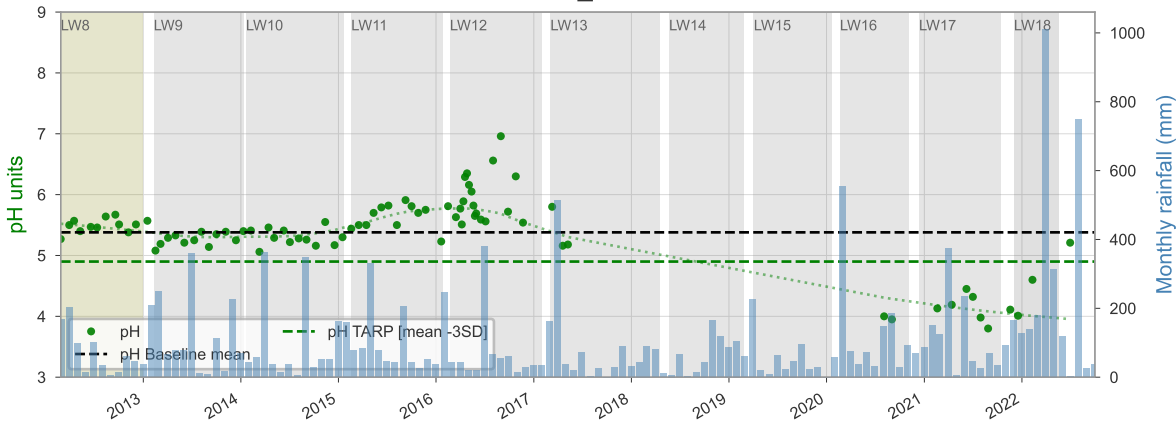
LA4S1



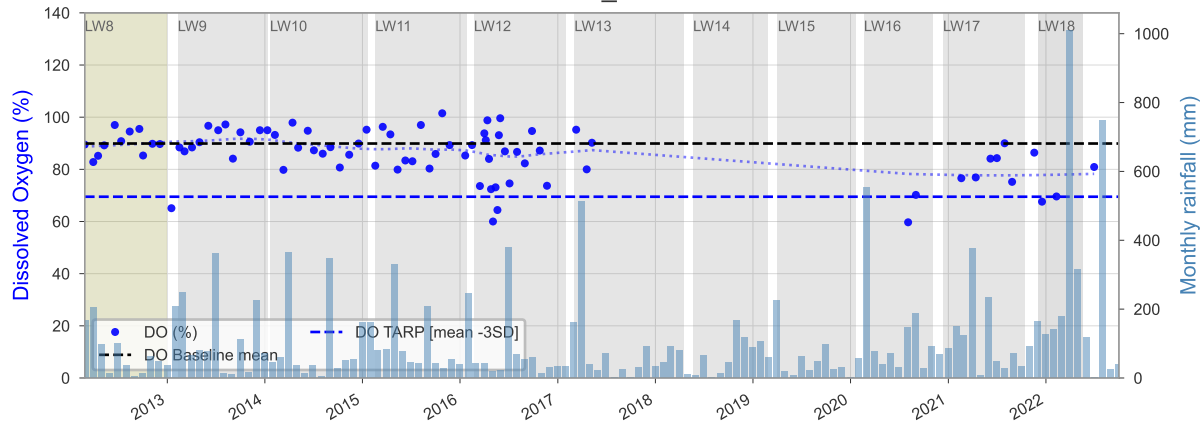
LA4_S1



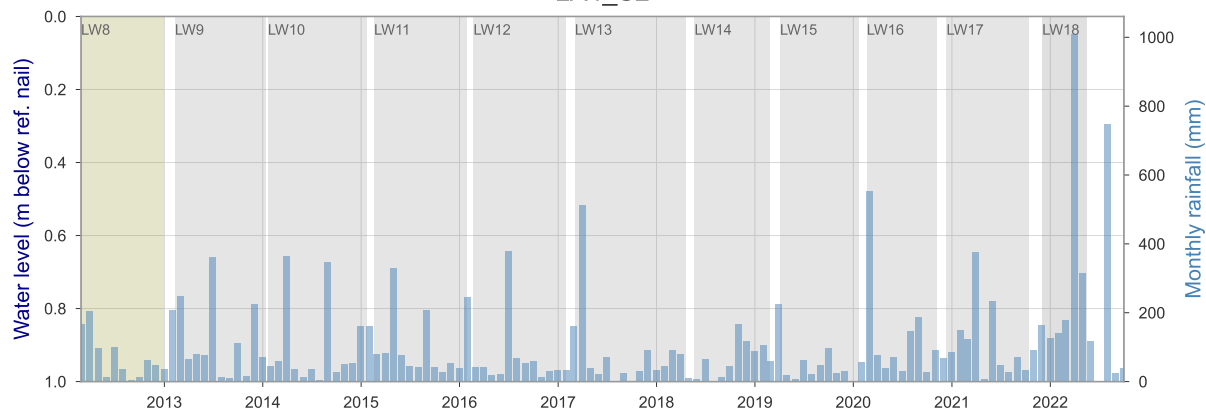
LA4_S1



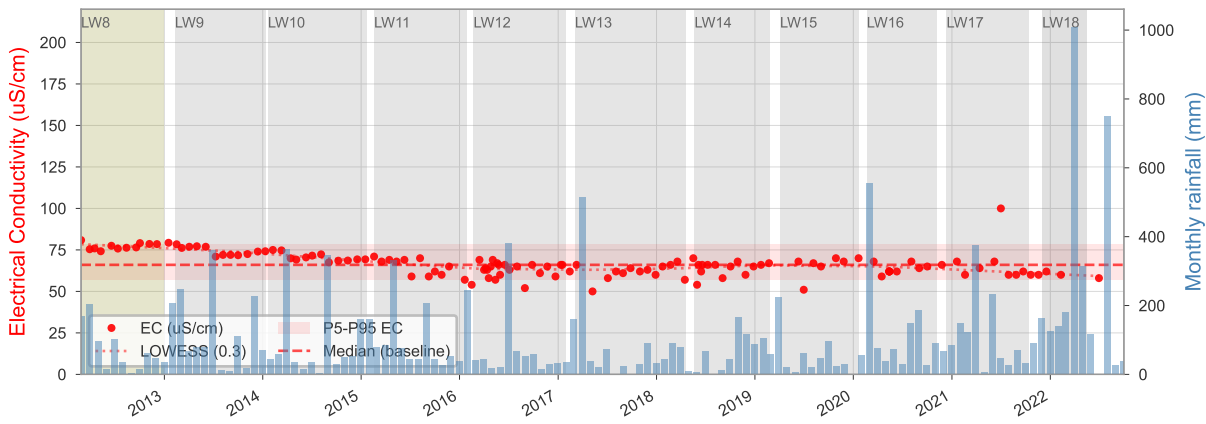
LA4_S1



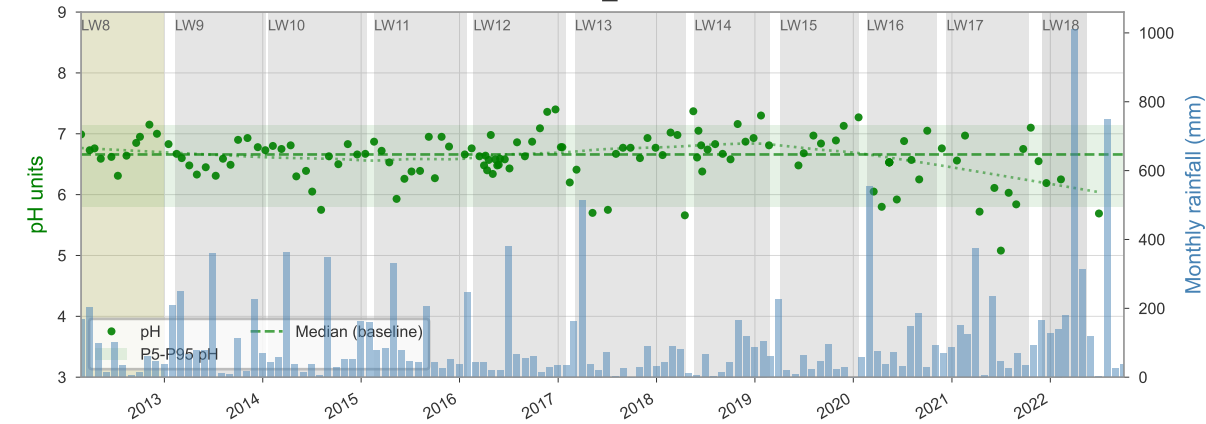
LA4_S2



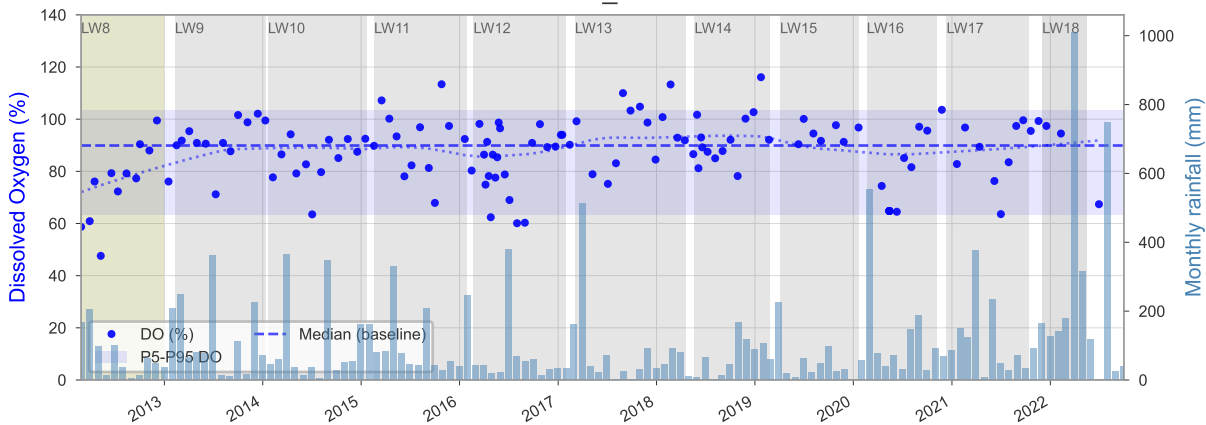
LA4_S2



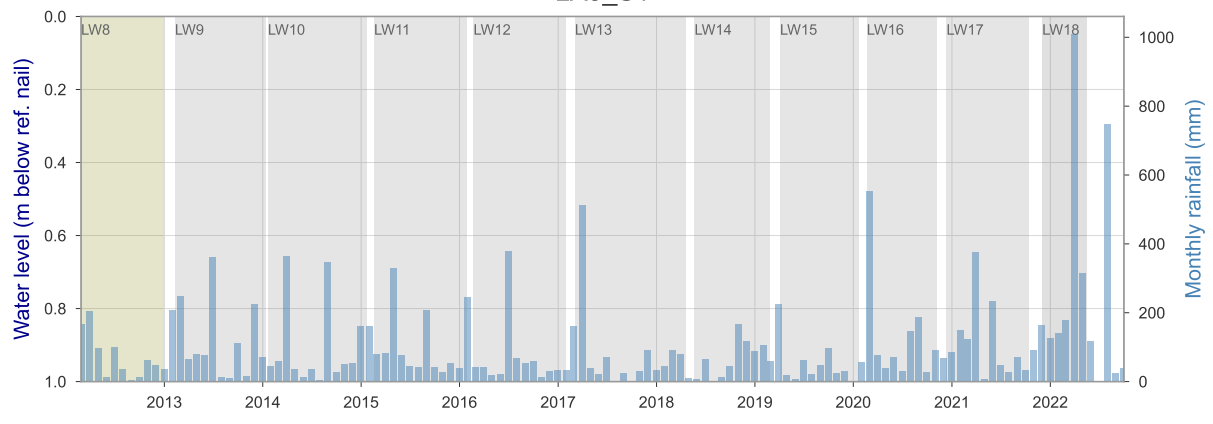
LA4_S2



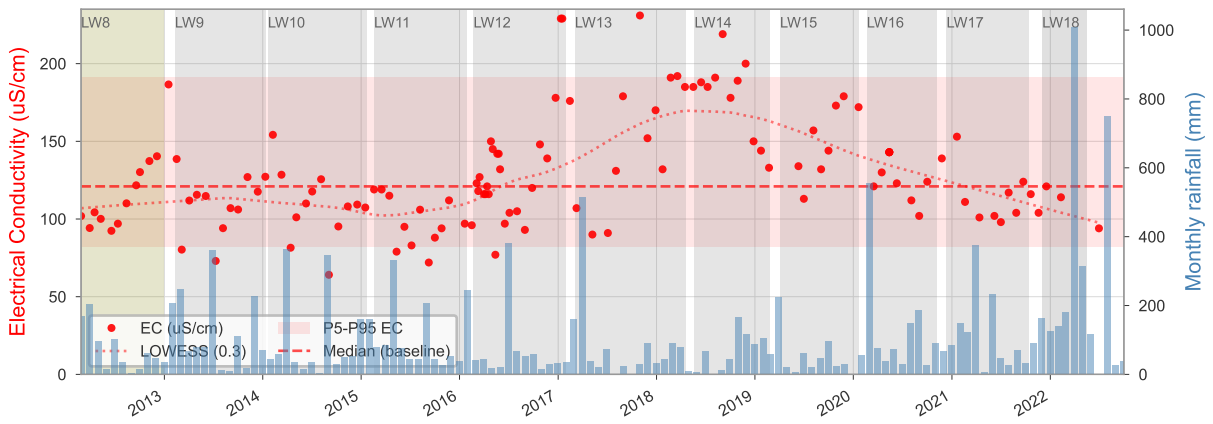
LA4_S2



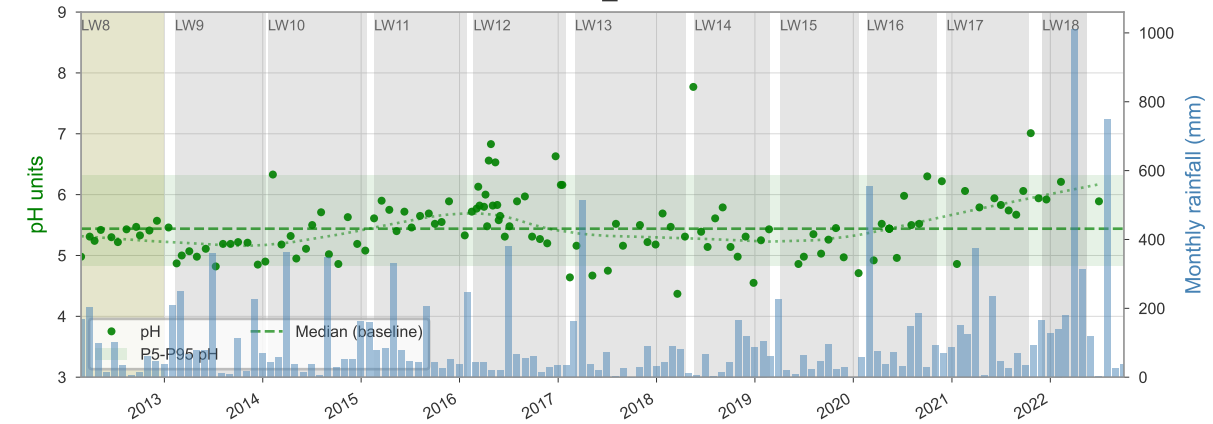
LA5_S1



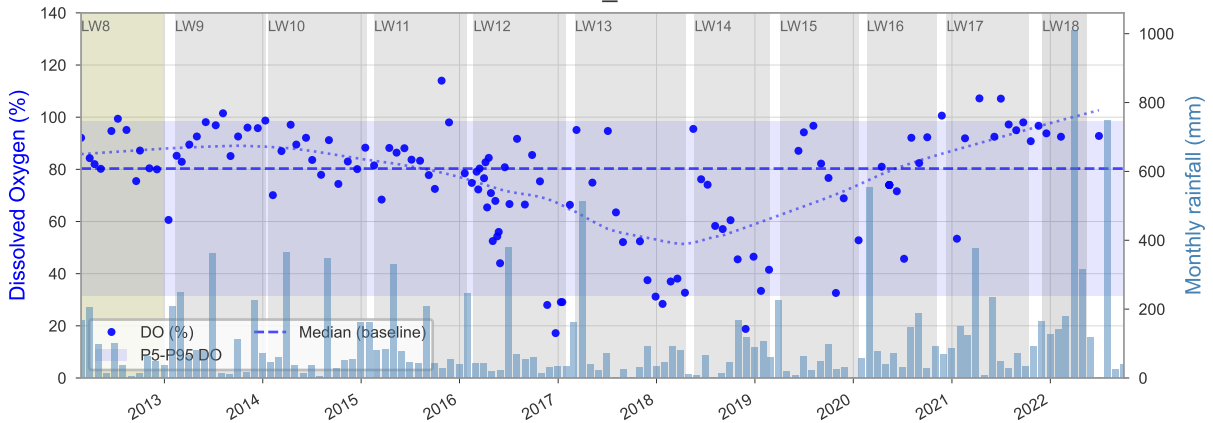
LA5_S1



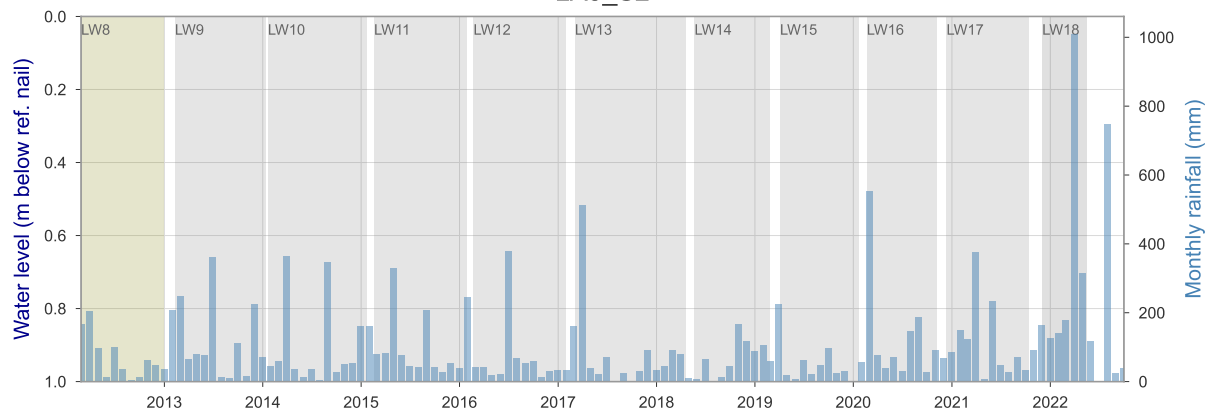
LA5_S1



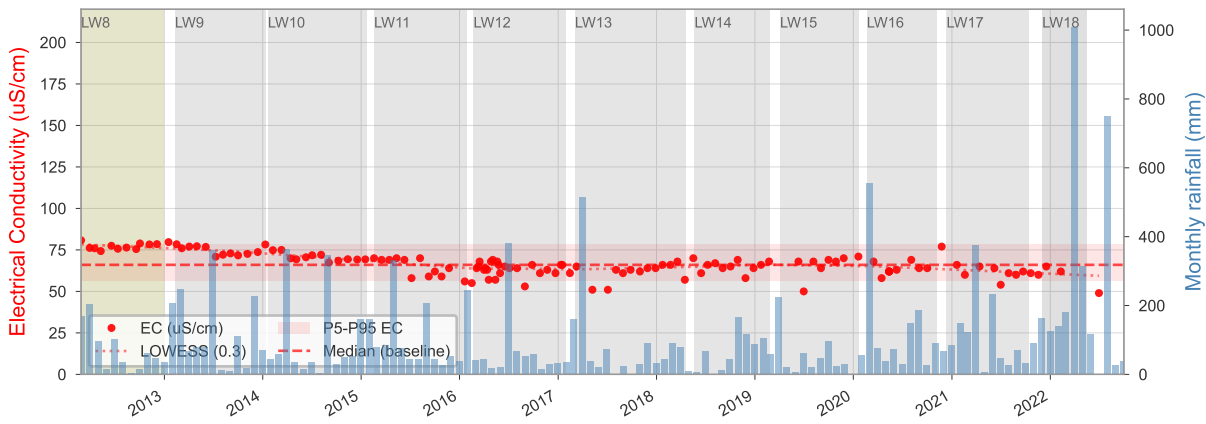
LA5_S1



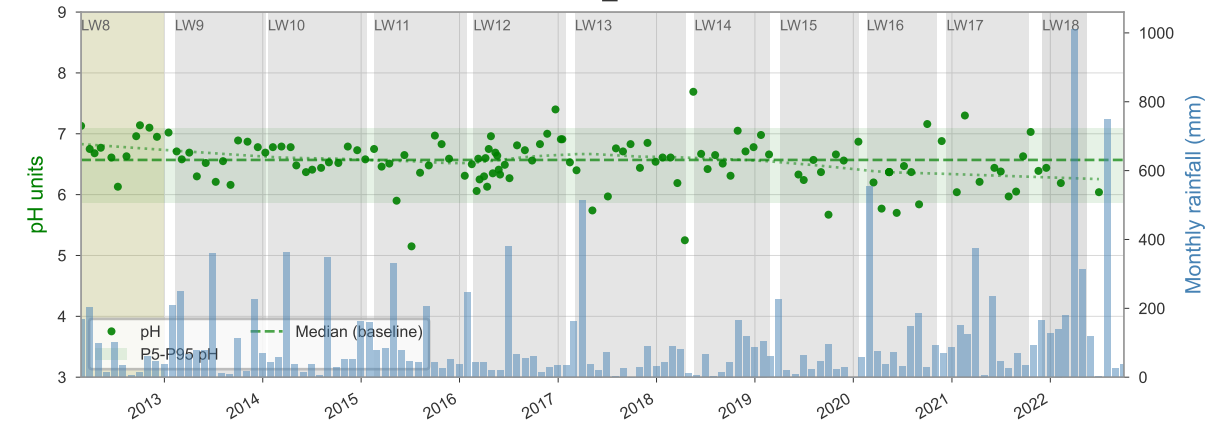
LA5_S2



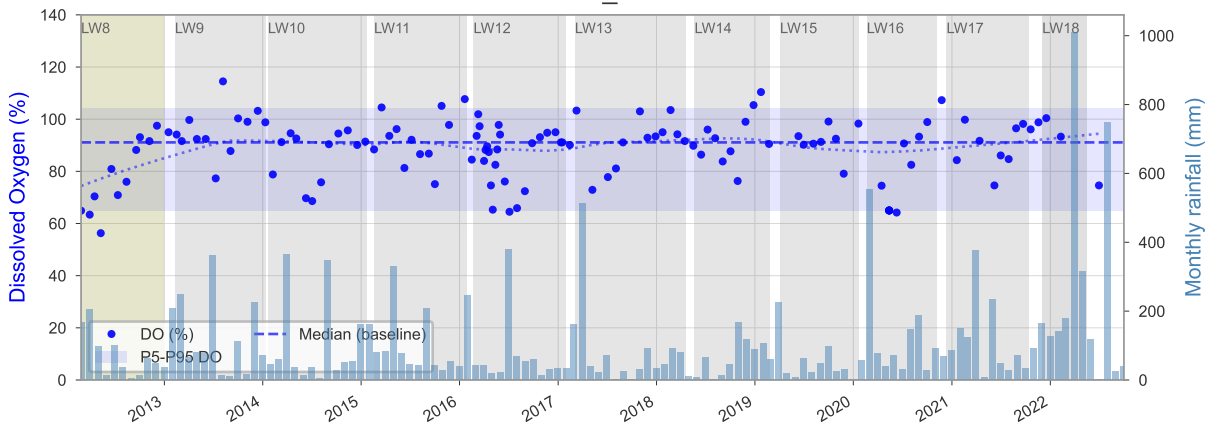
LA5_S2



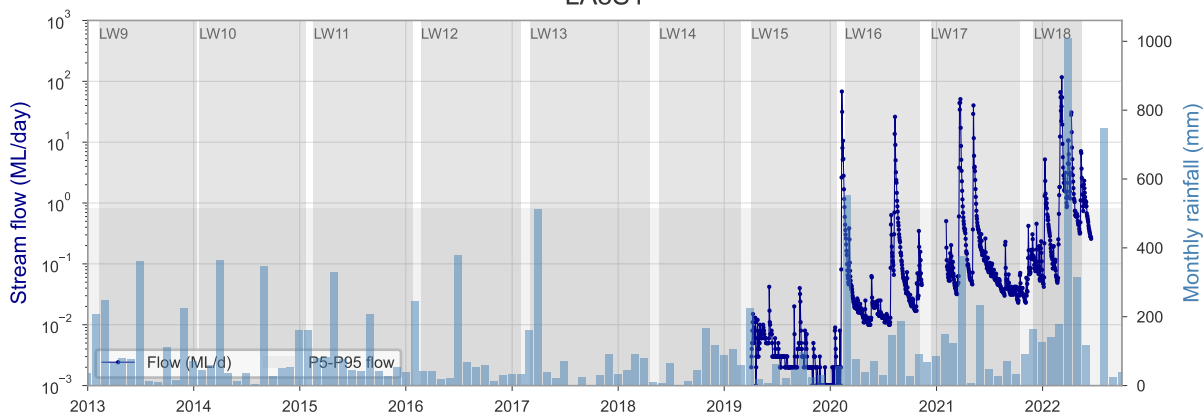
LA5_S2



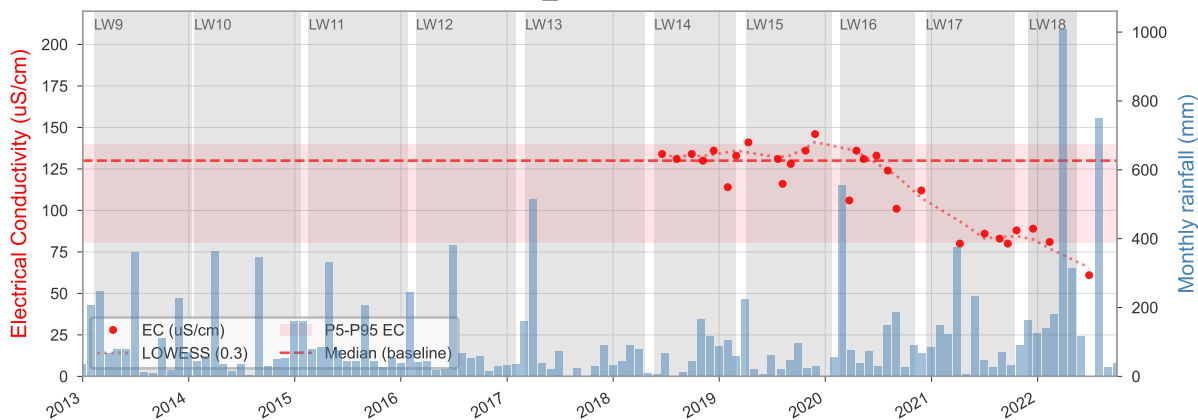
LA5_S2



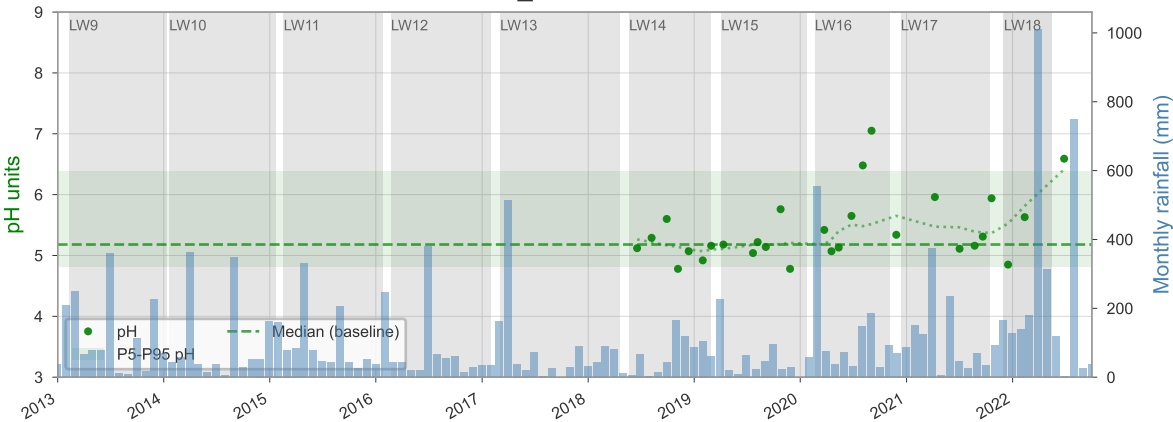
LA8S1



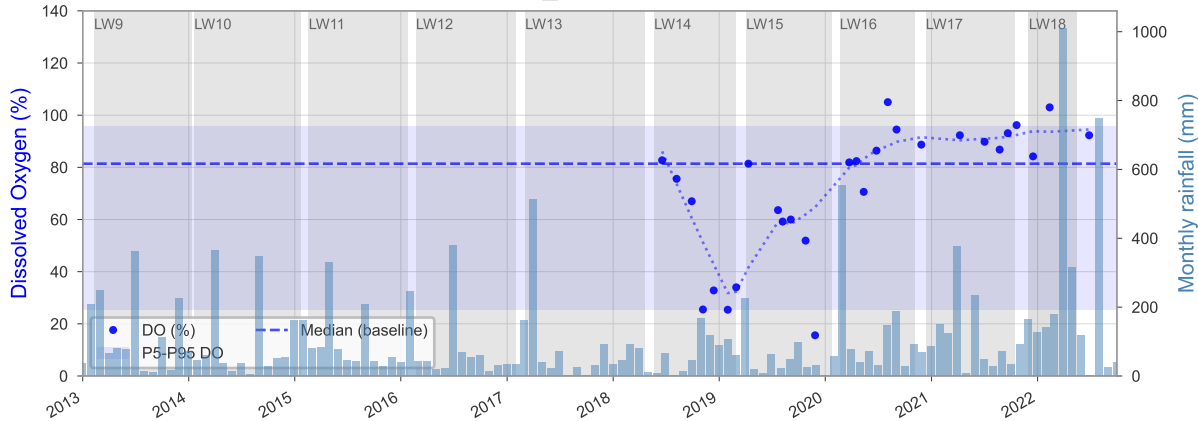
LA8_ROCKBAR1



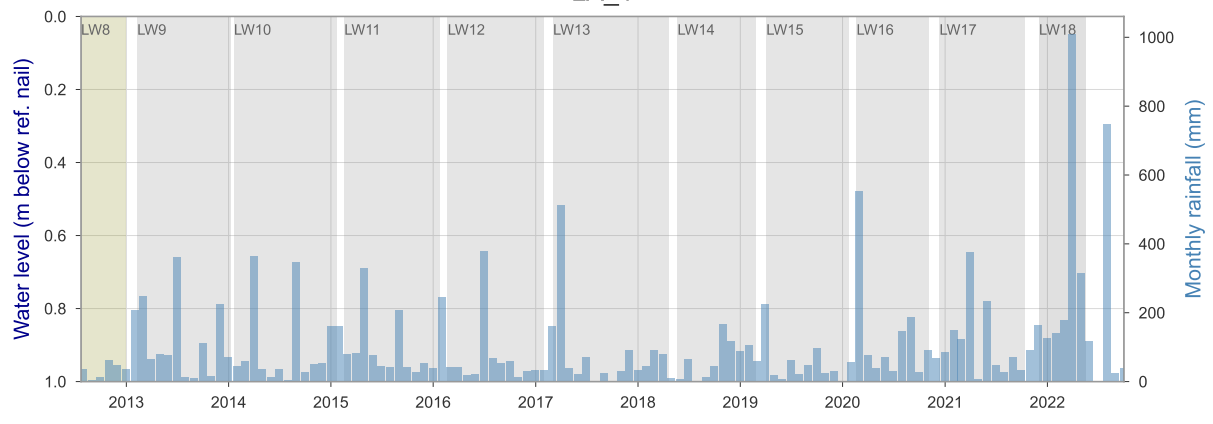
LA8_ROCKBAR1



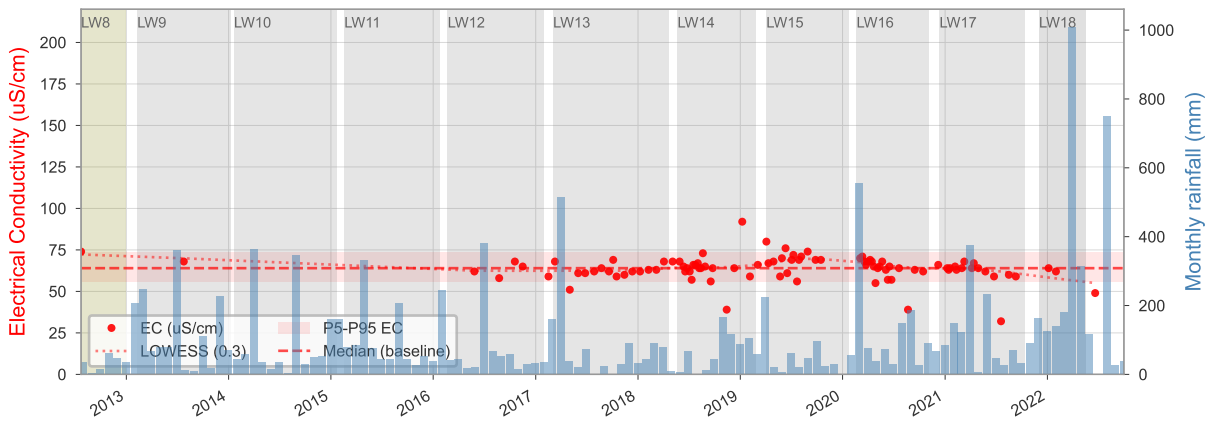
LA8_ROCKBAR1



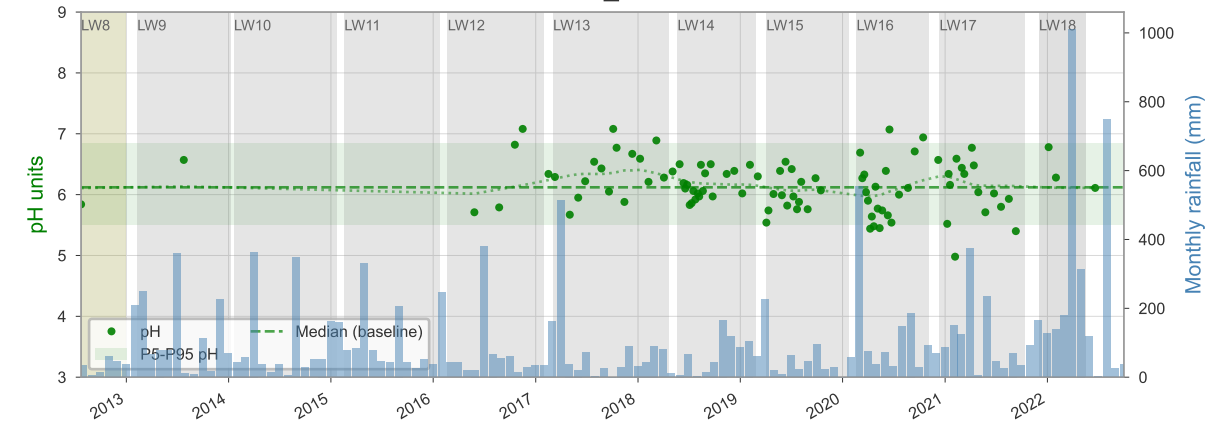
LA_1



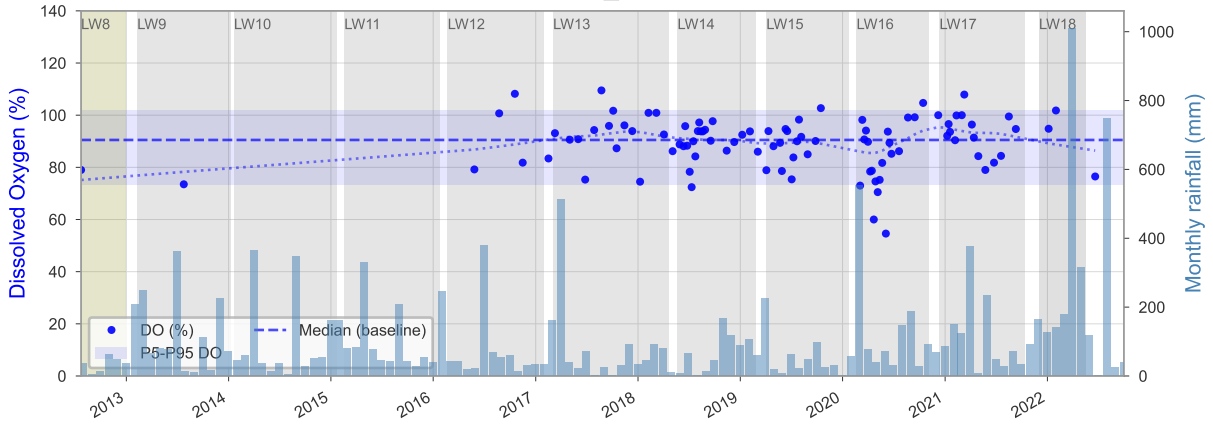
LA_1



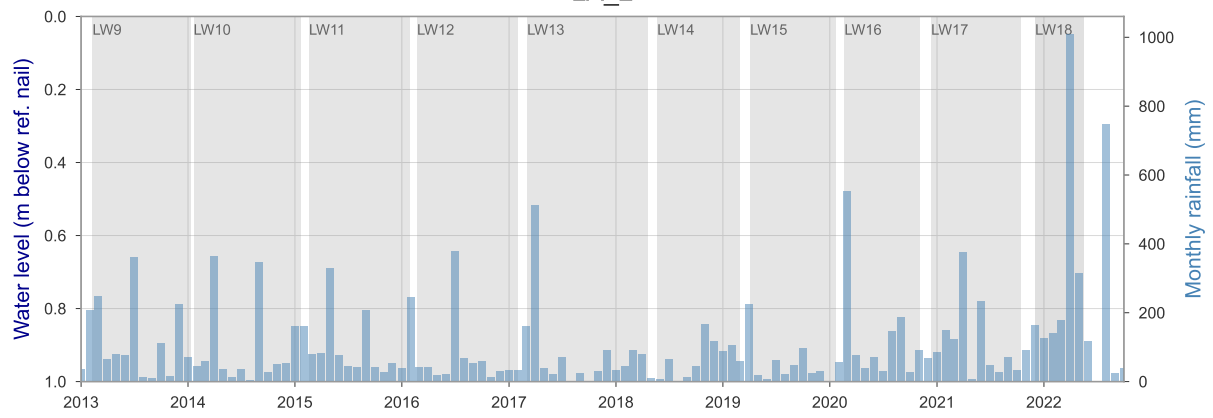
LA_1



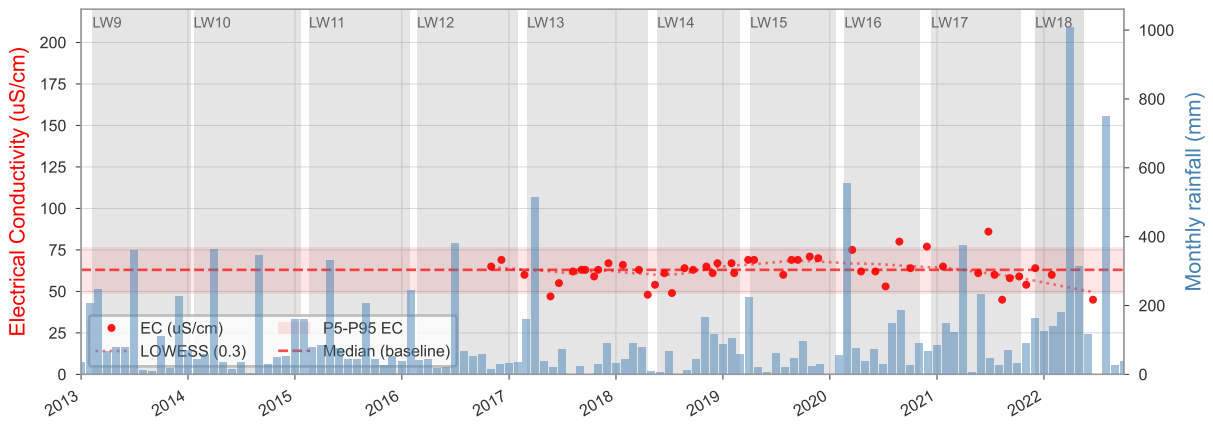
LA_1



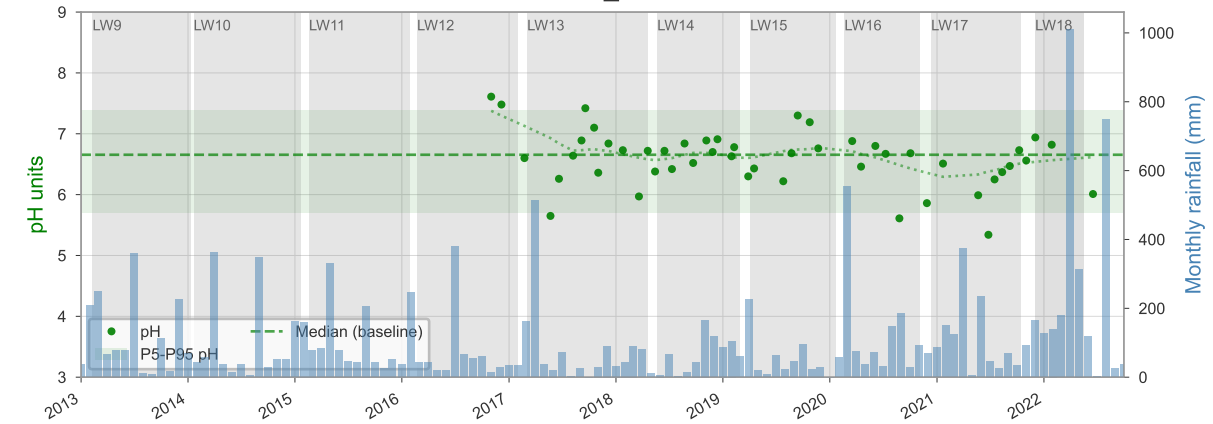
LA_2



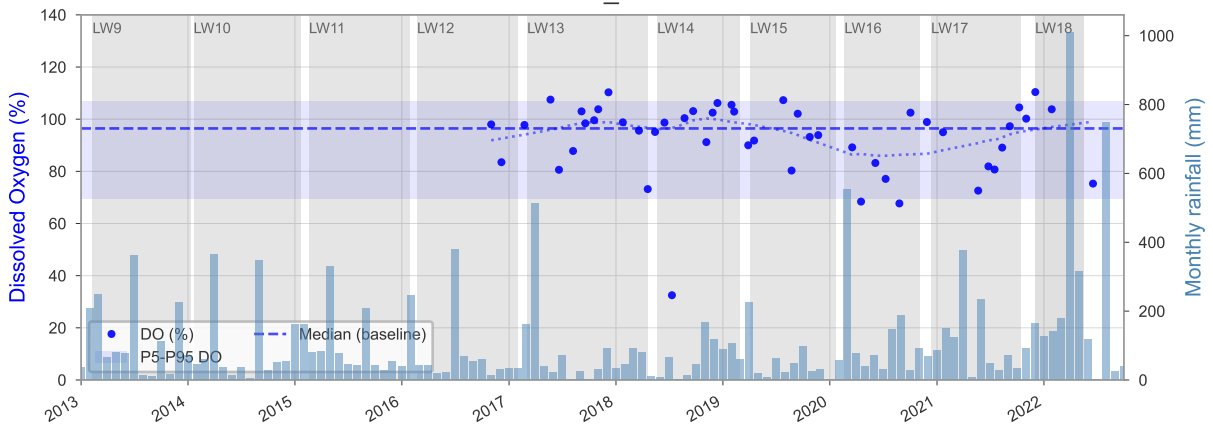
LA_2



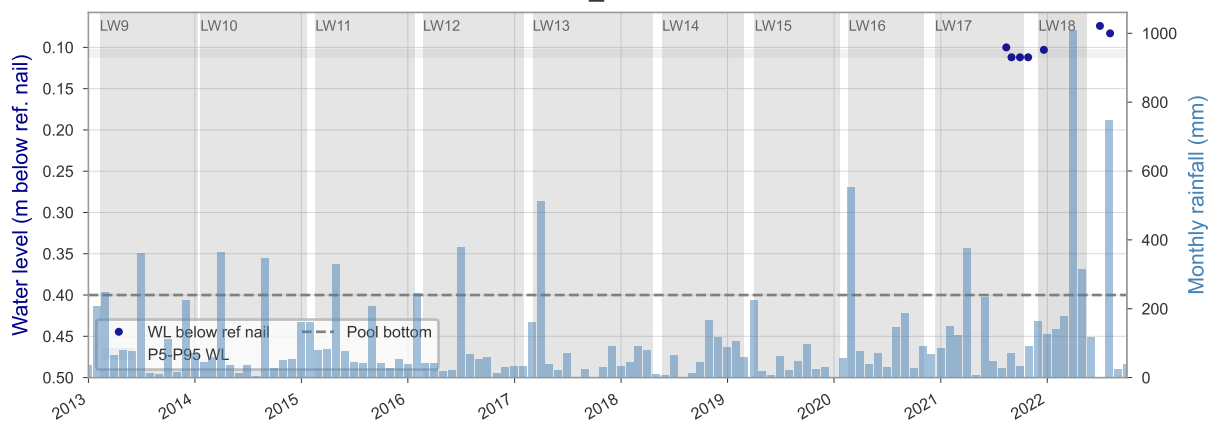
LA_2



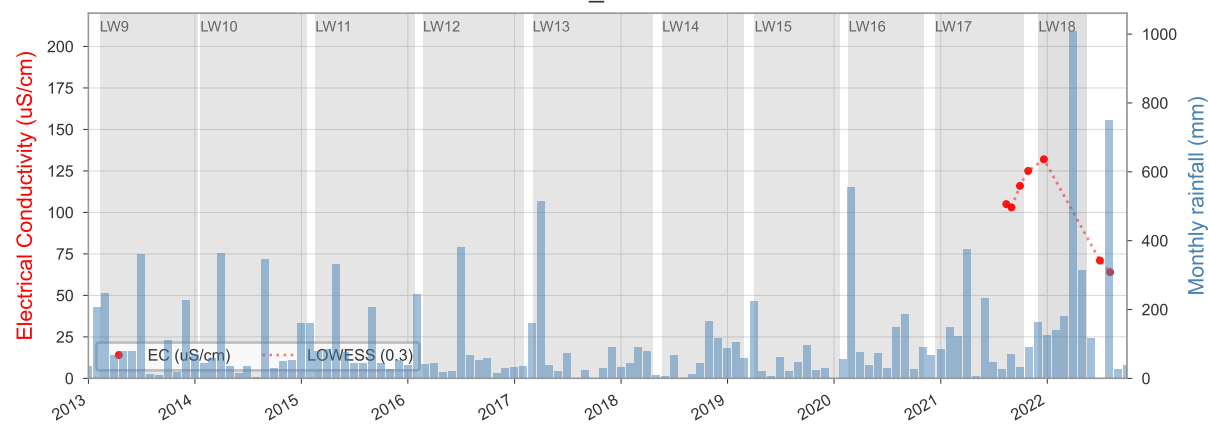
LA_2



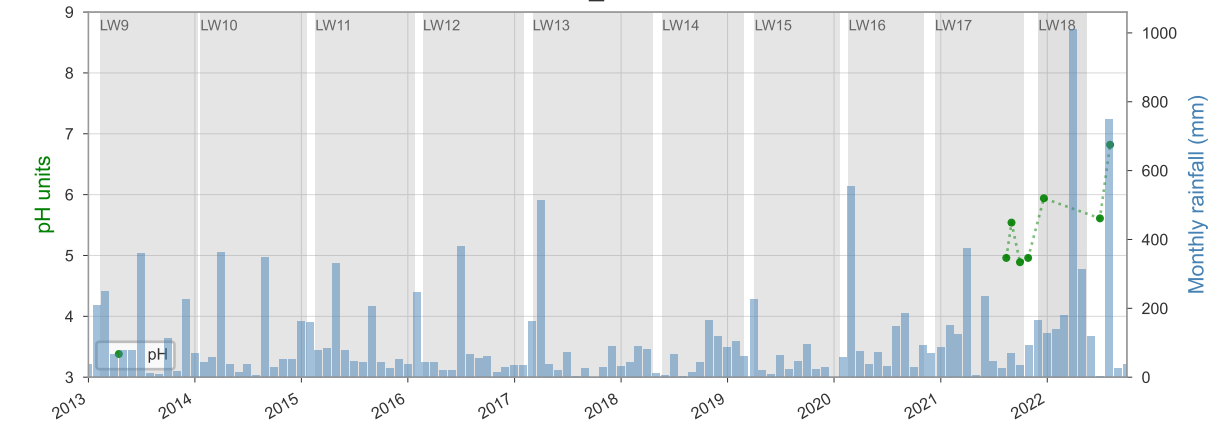
LC5A_POOL6



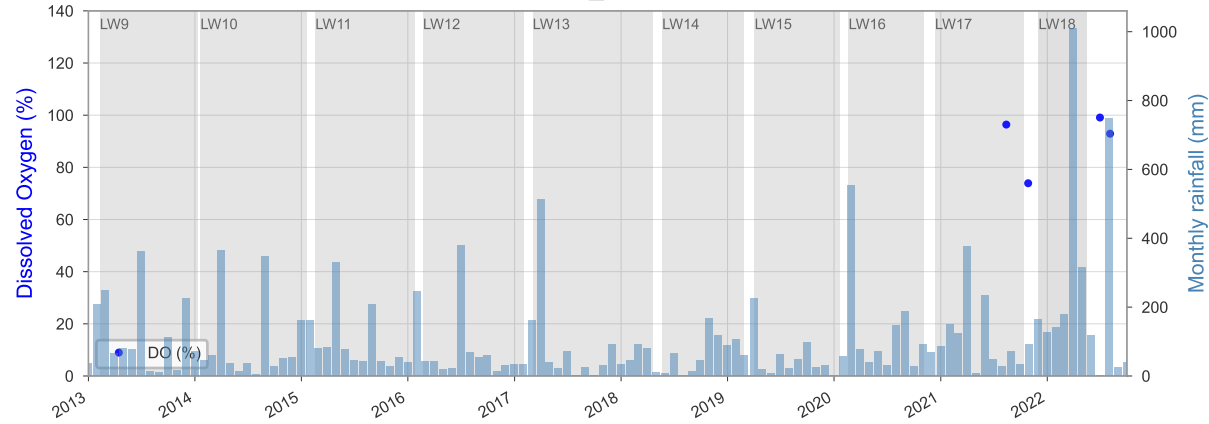
LC5A_POOL6



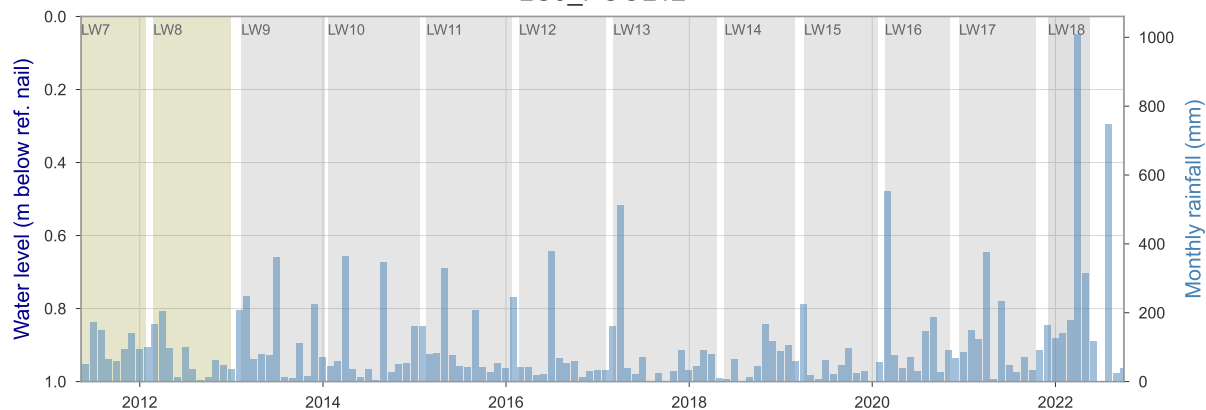
LC5A_POOL6



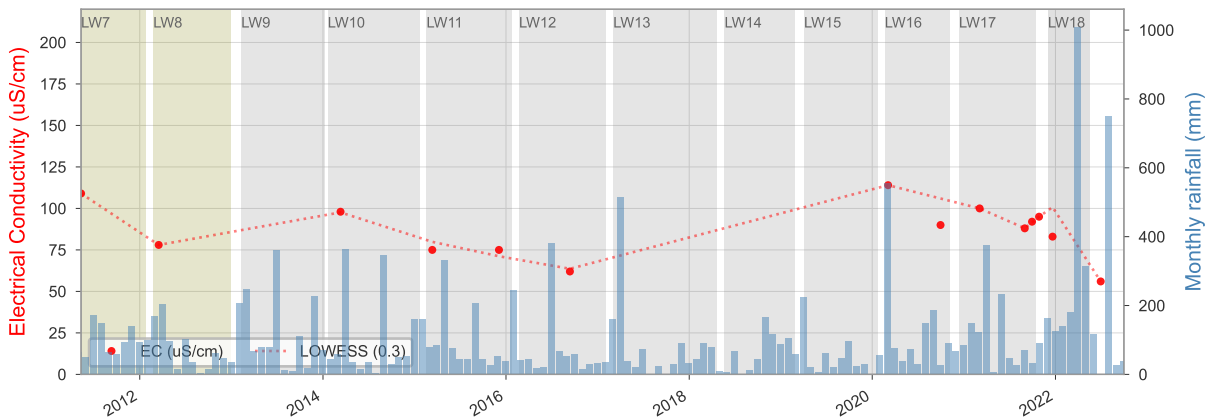
LC5A_POOL6



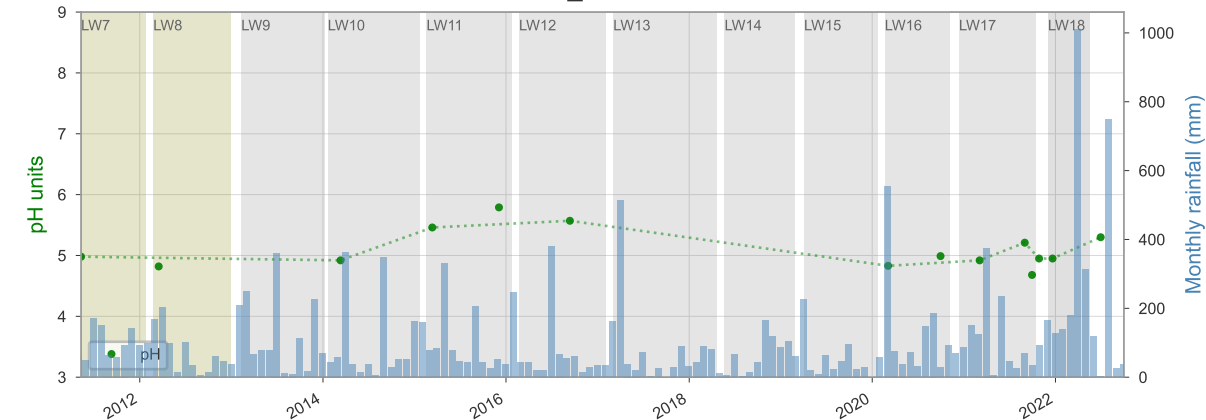
LC5_POOL12



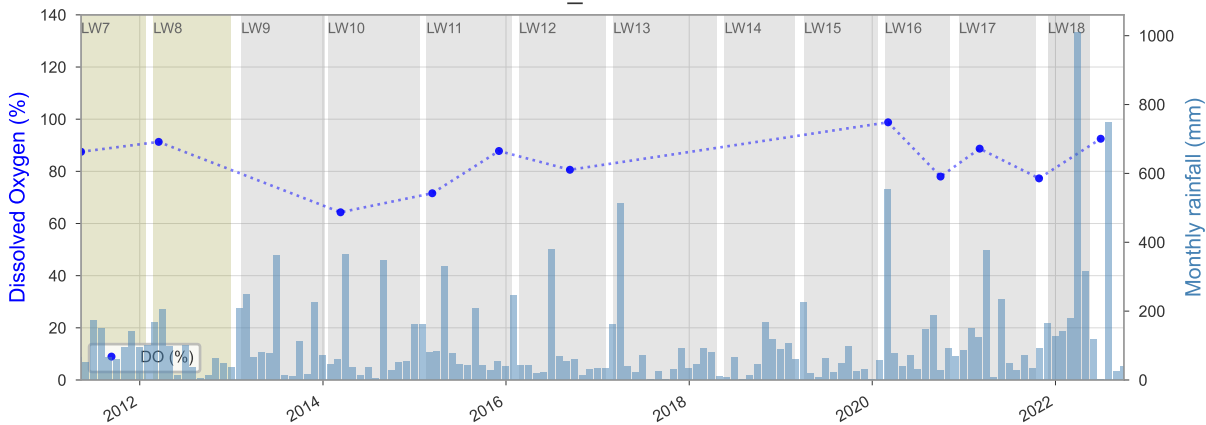
LC5_POOL12



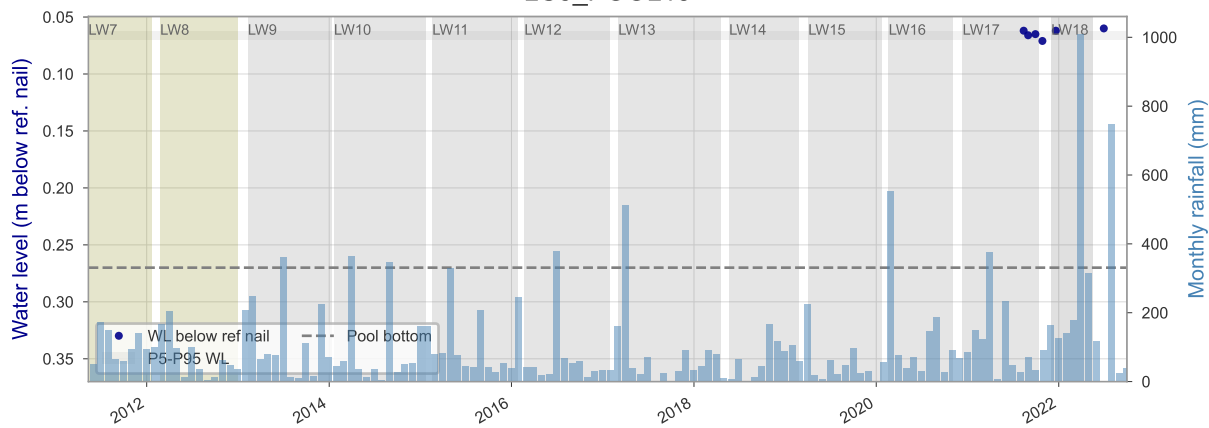
LC5_POOL12



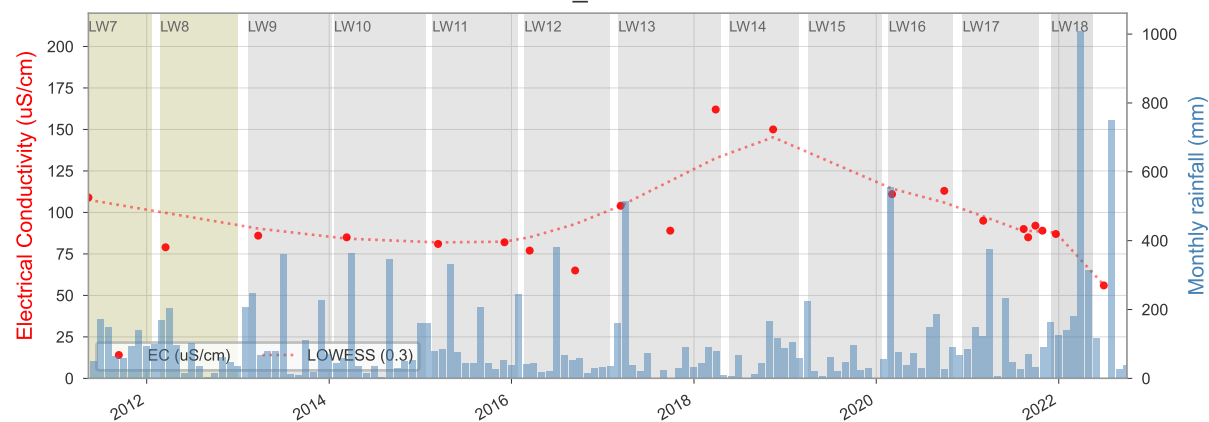
LC5_POOL12



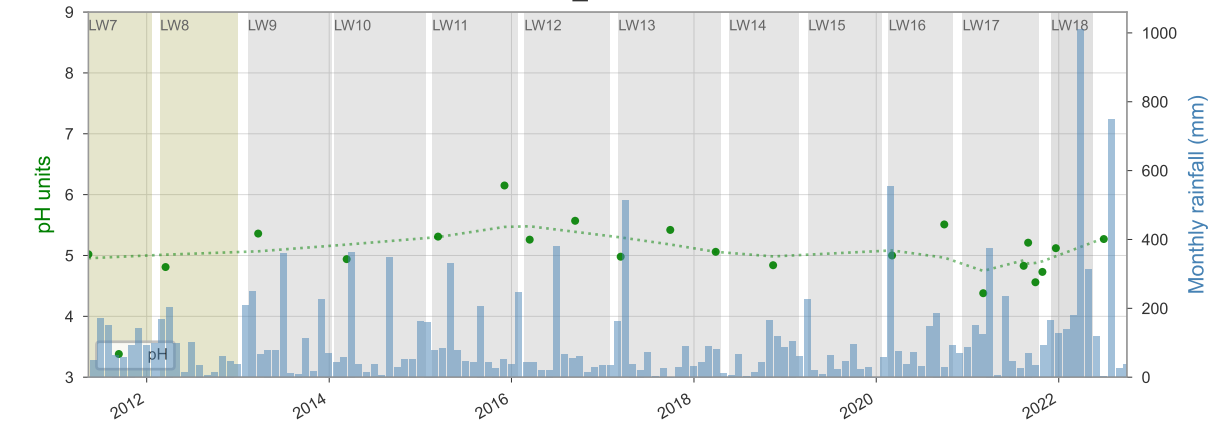
LC5_POOL13



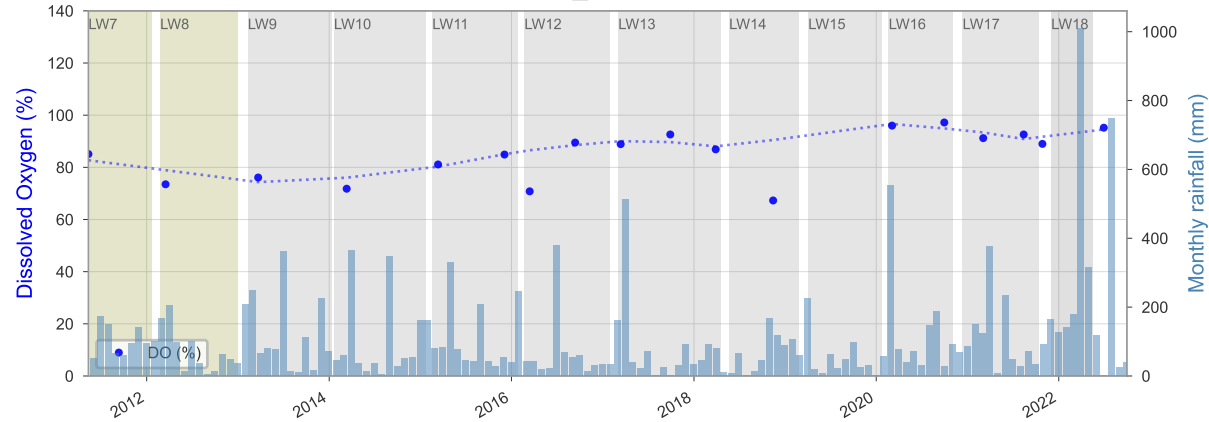
LC5_POOL13



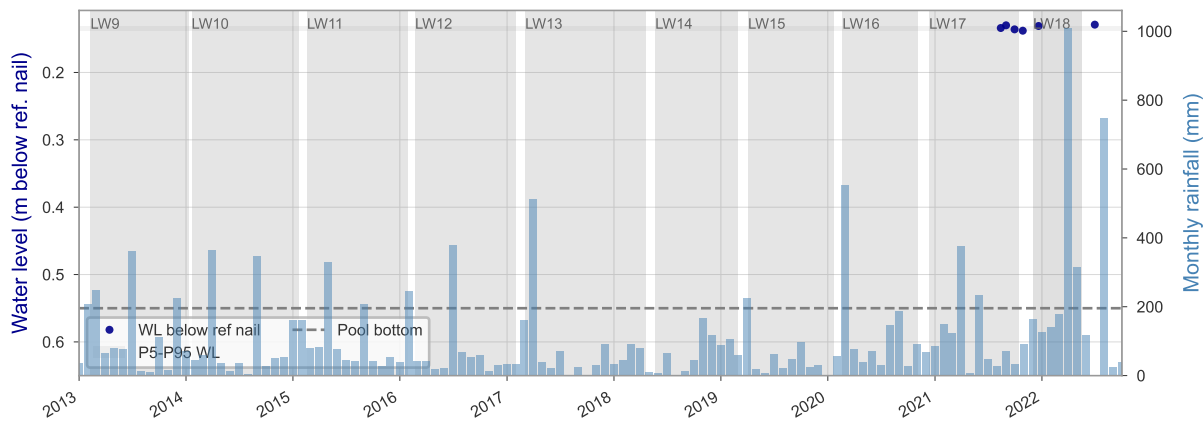
LC5_POOL13



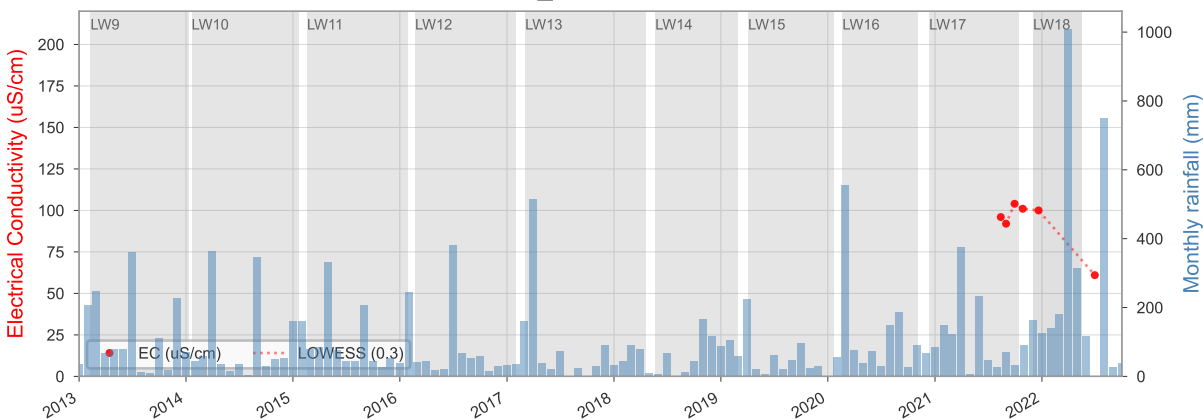
LC5_POOL13



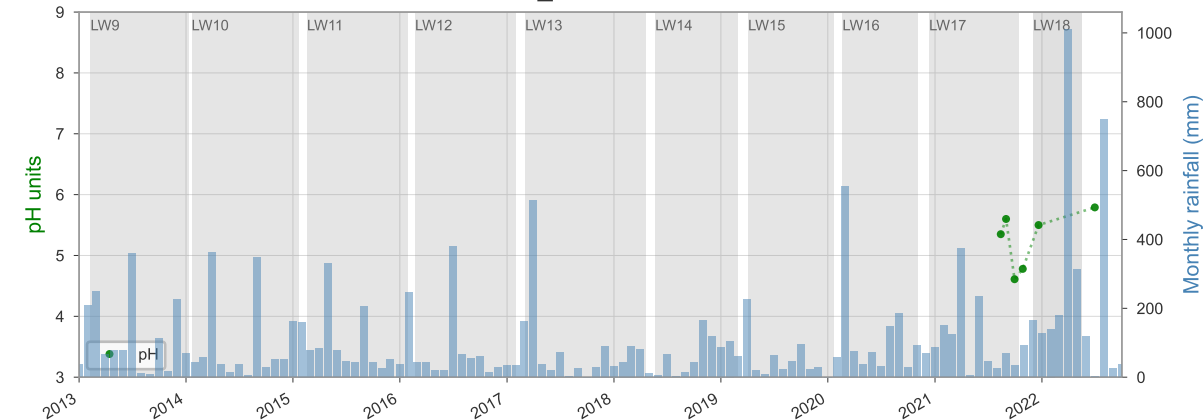
LC5_ROCKBAR11



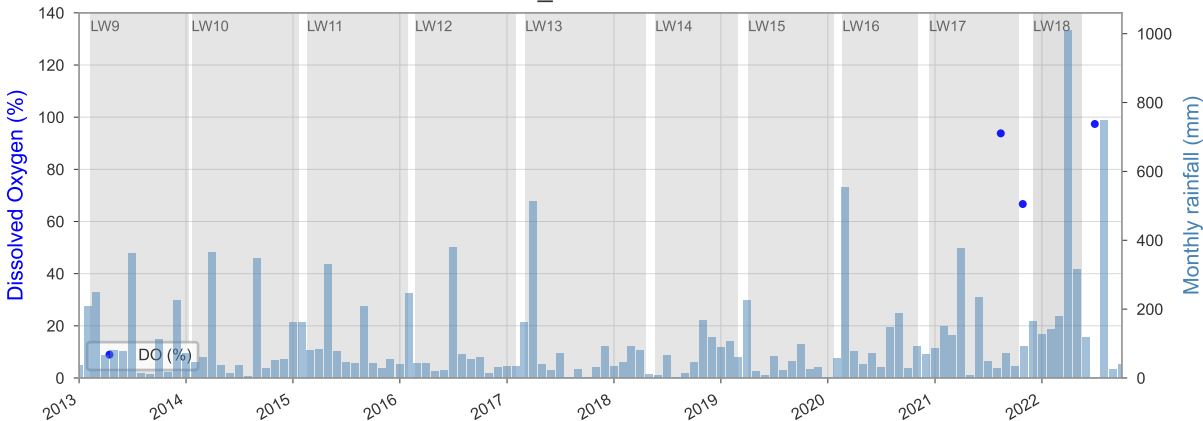
LC5_ROCKBAR11



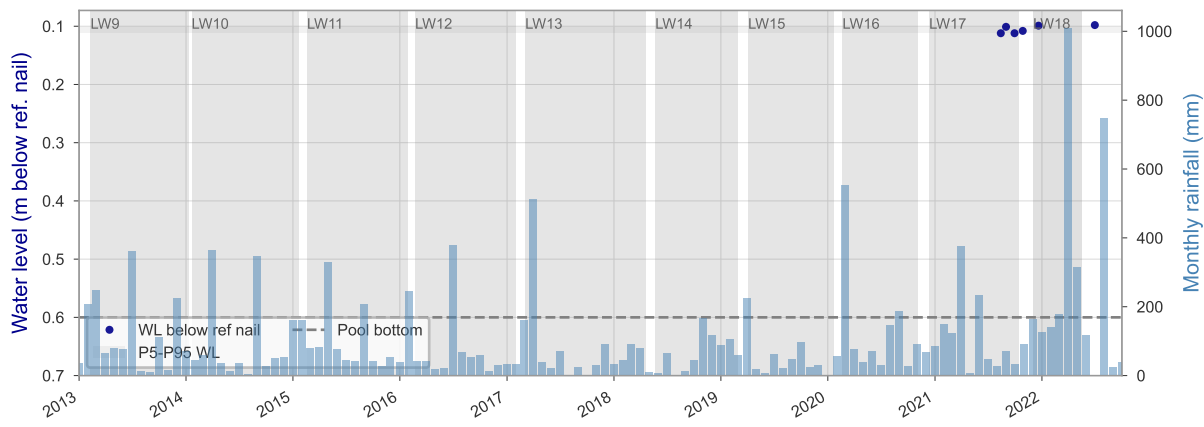
LC5_ROCKBAR11



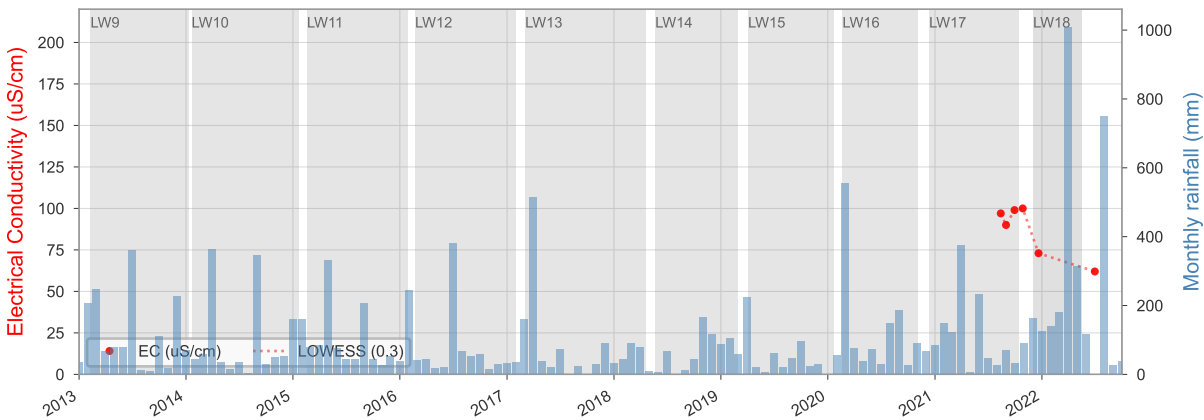
LC5_ROCKBAR11



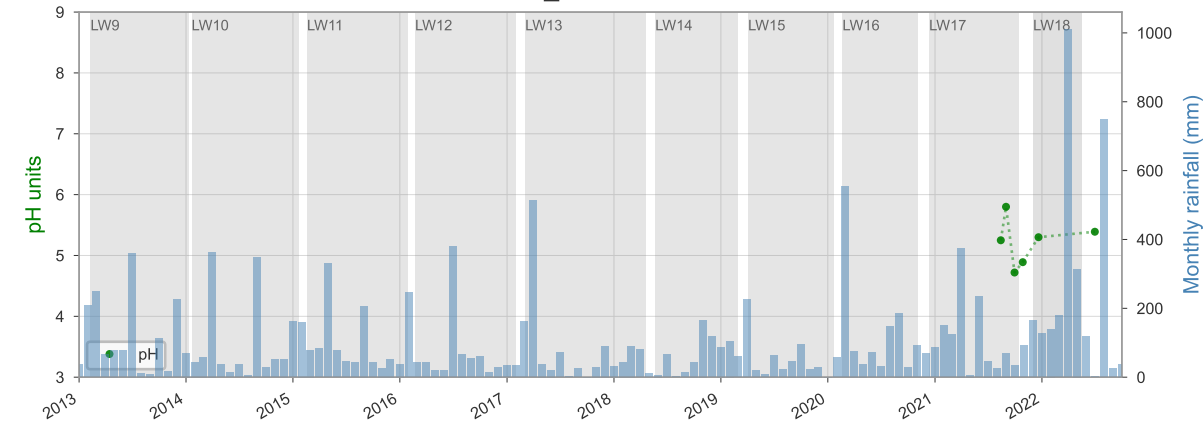
LC5_ROCKBAR7



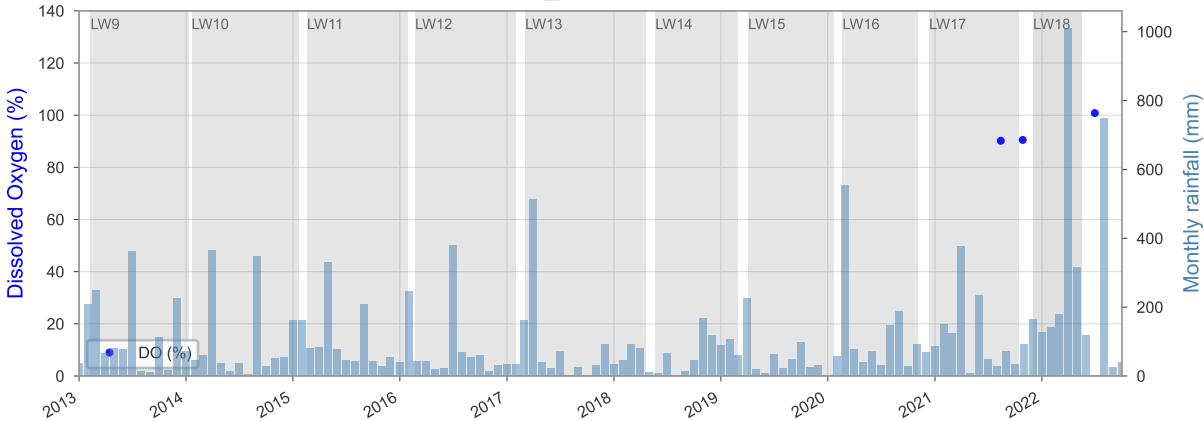
LC5_ROCKBAR7



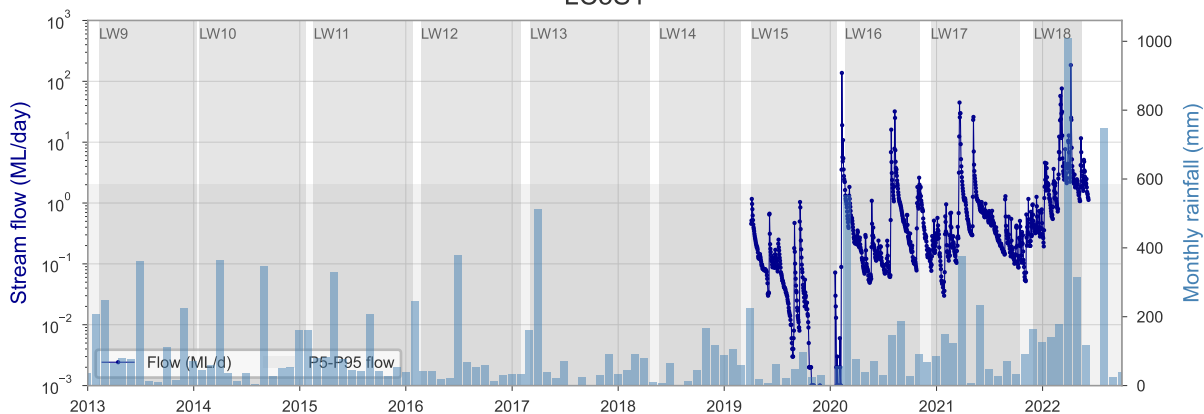
LC5_ROCKBAR7



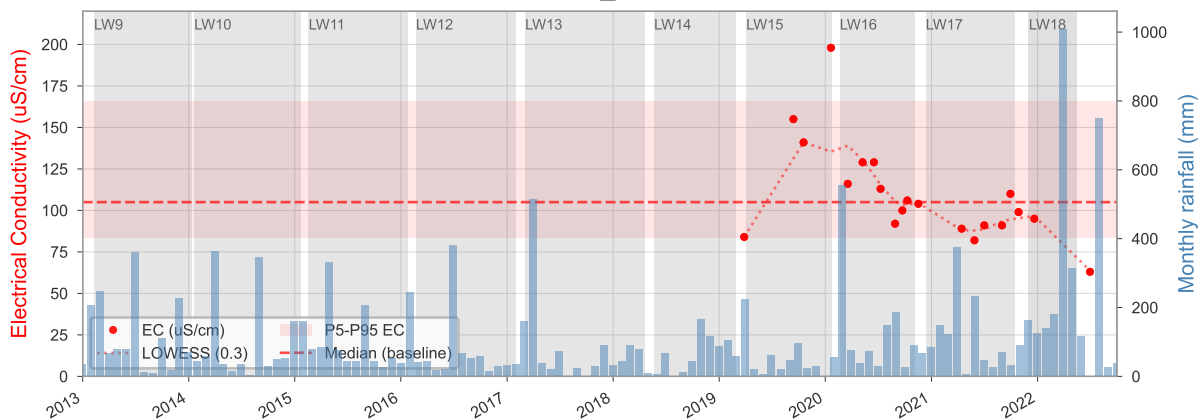
LC5_ROCKBAR7



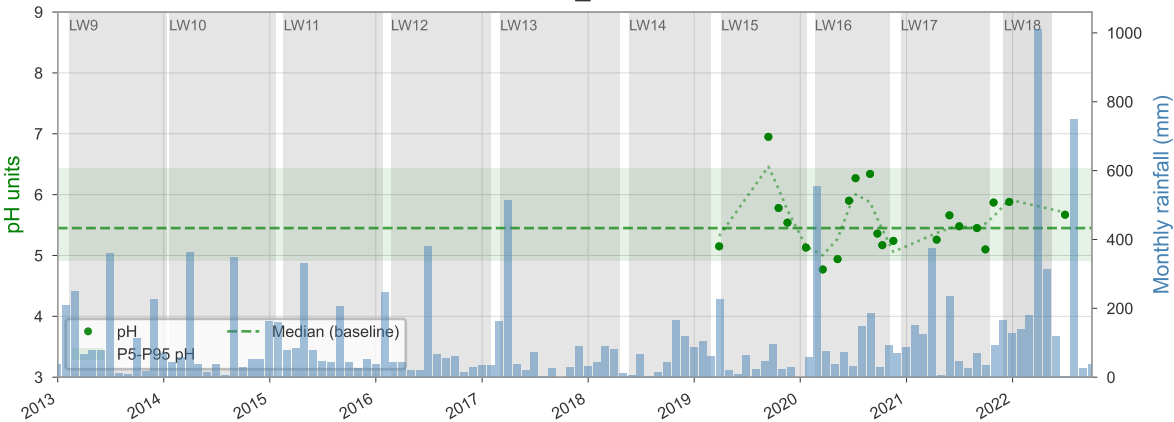
LC5S1



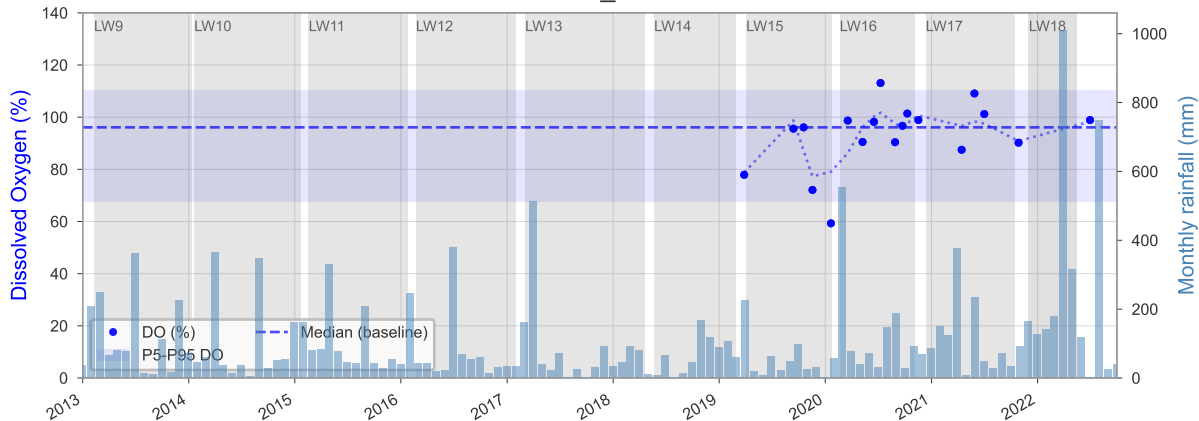
LC5_S1



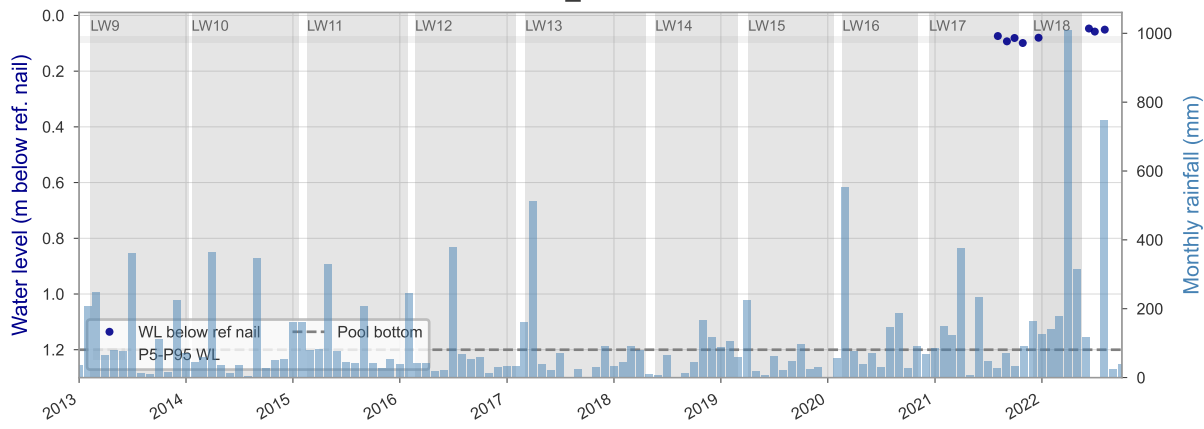
LC5_S1



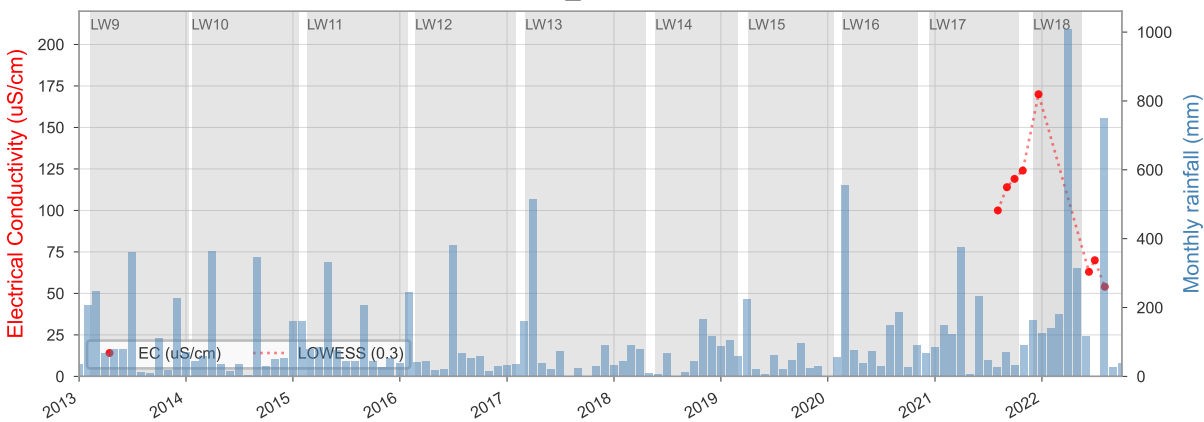
LC5_S1



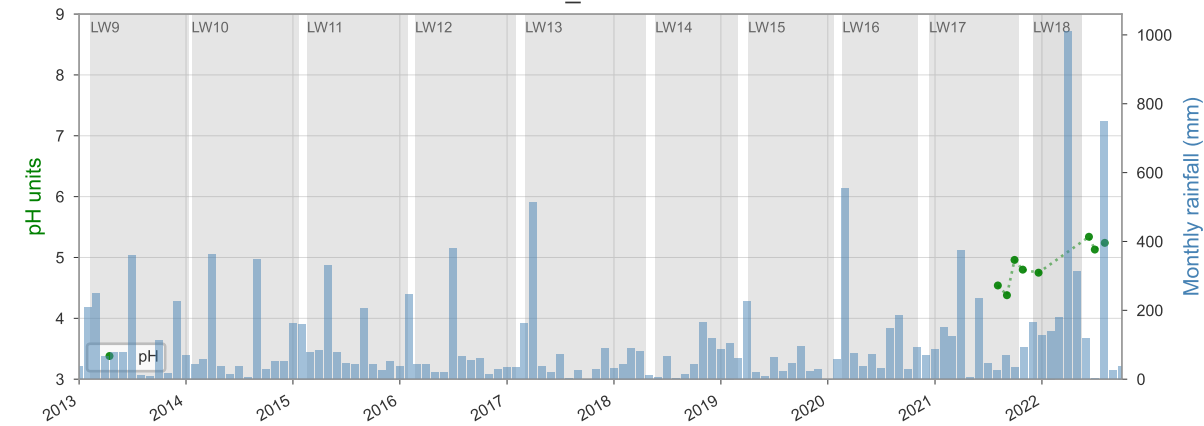
LC6_POOL16



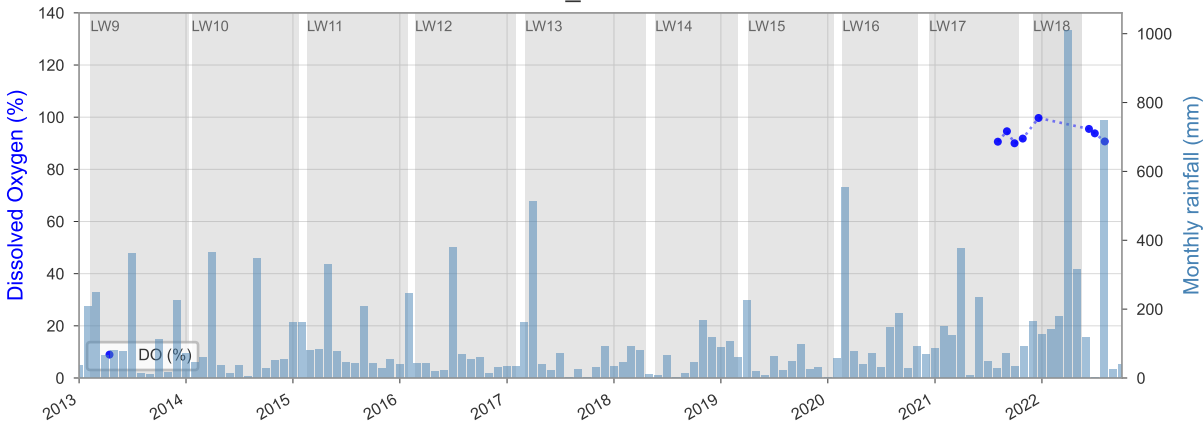
LC6_POOL16



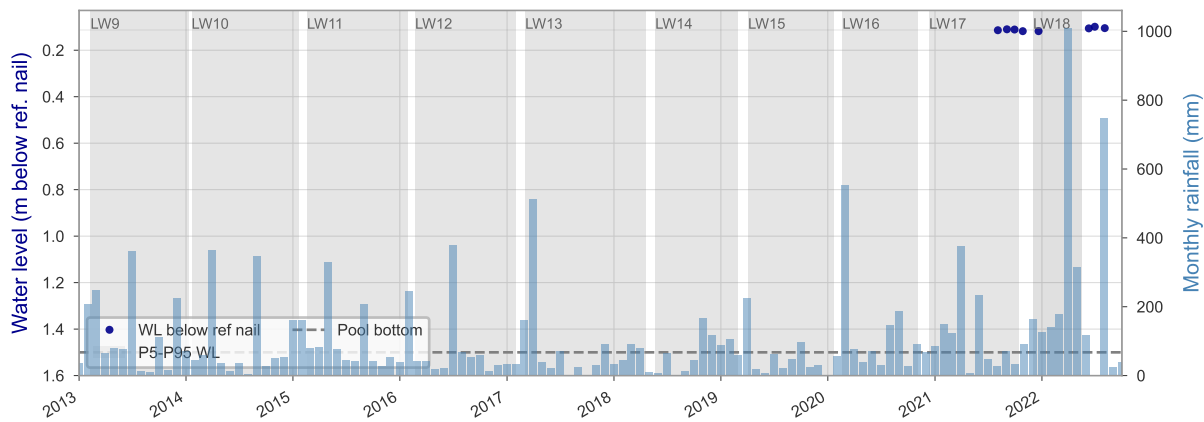
LC6_POOL16



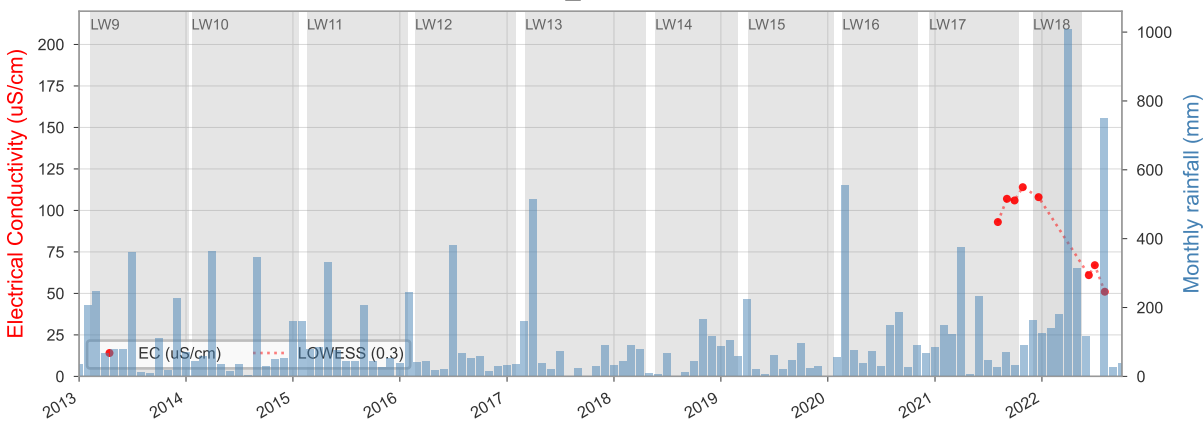
LC6_POOL16



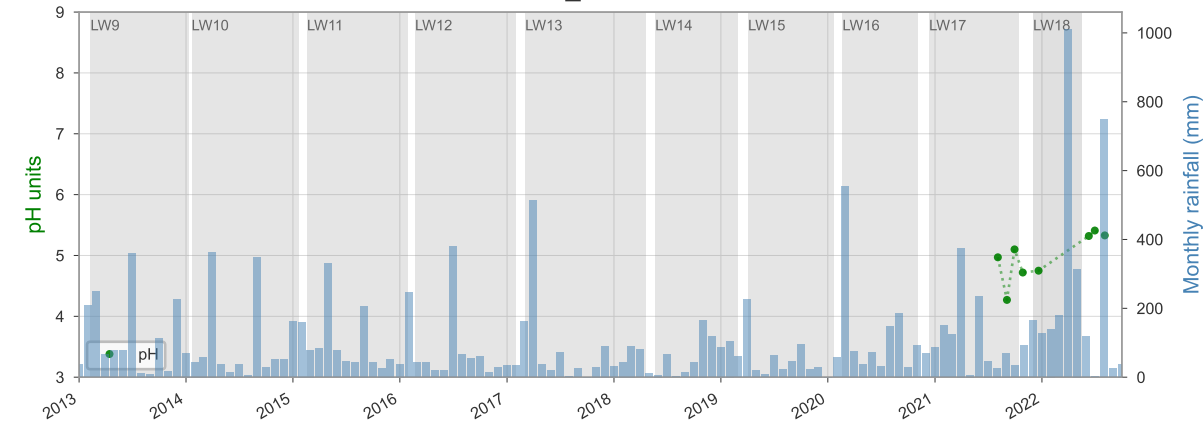
LC6_POOL36



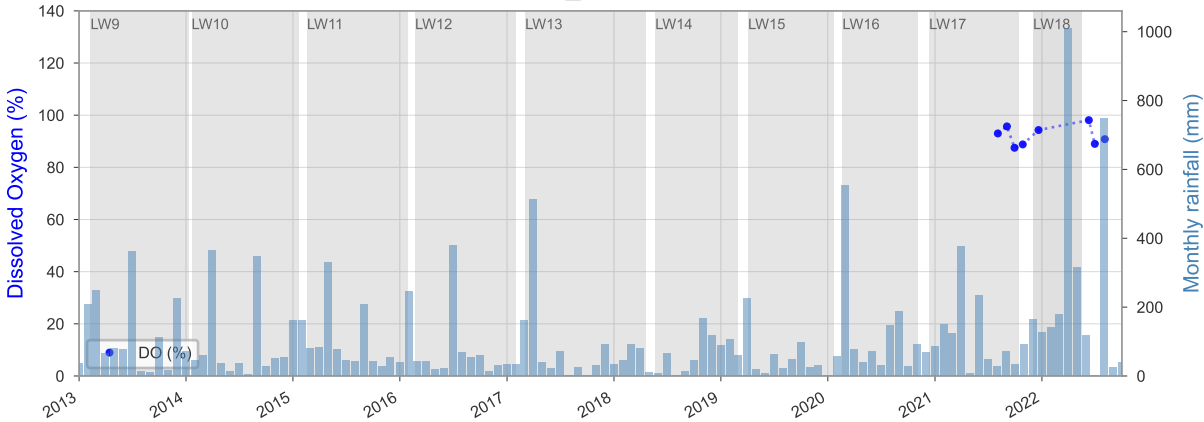
LC6_POOL36



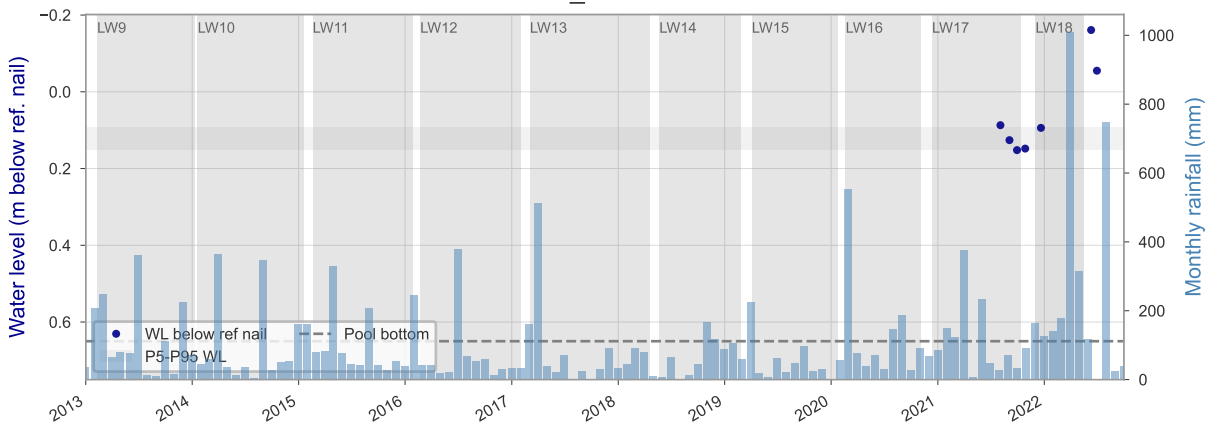
LC6_POOL36



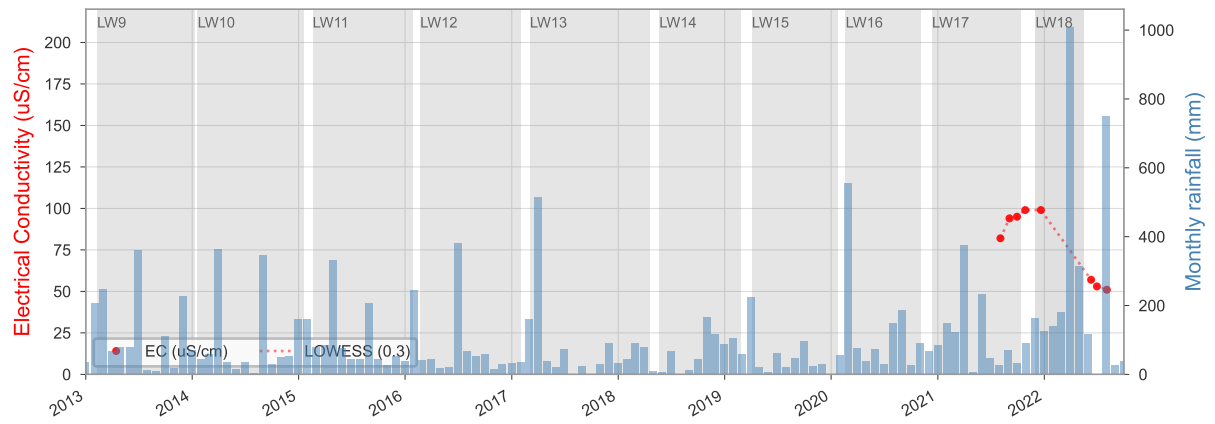
LC6_POOL36



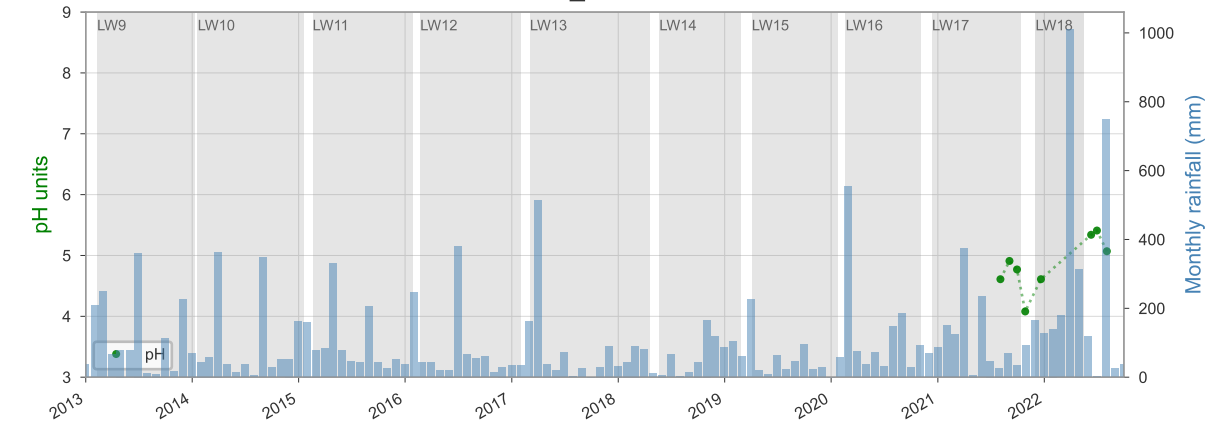
LC6_POOL51



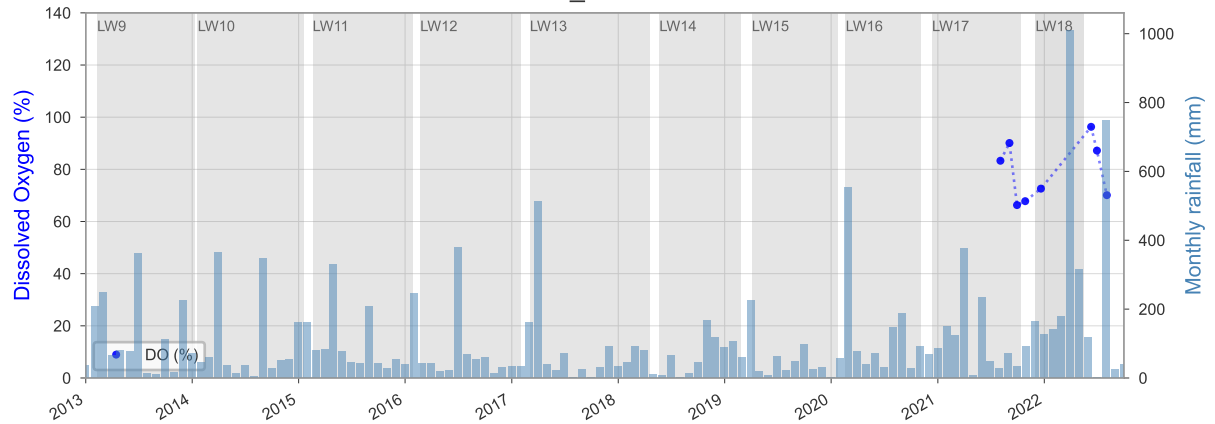
LC6_POOL51



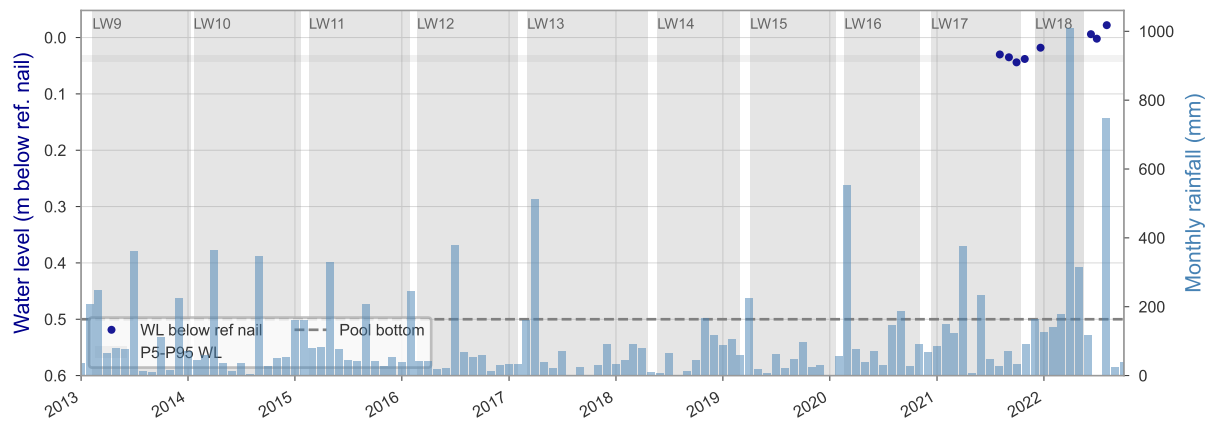
LC6_POOL51



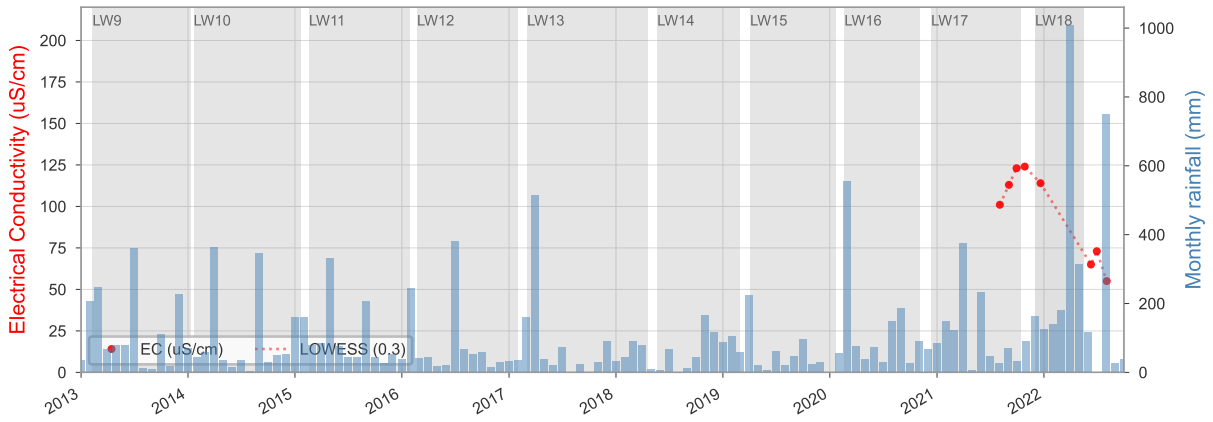
LC6_POOL51



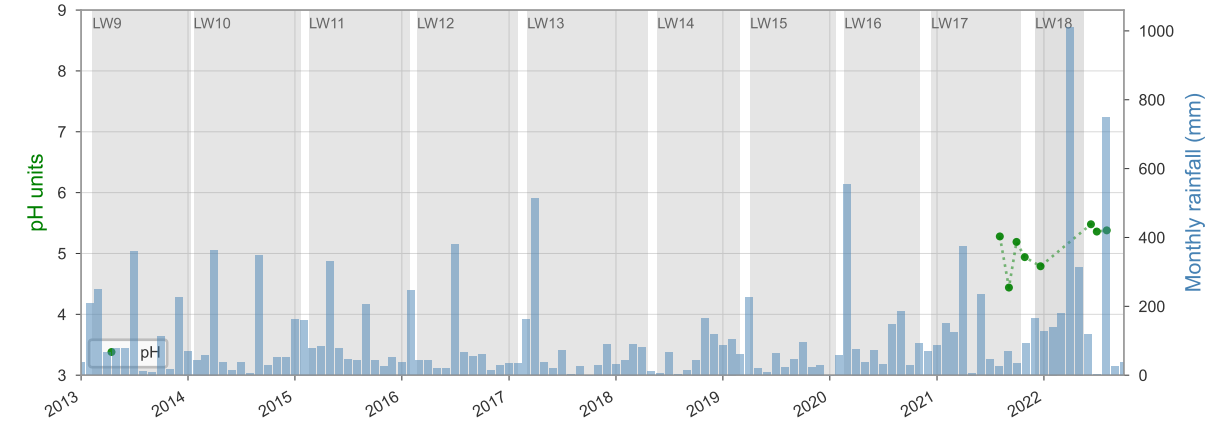
LC6_ROCKBAR1



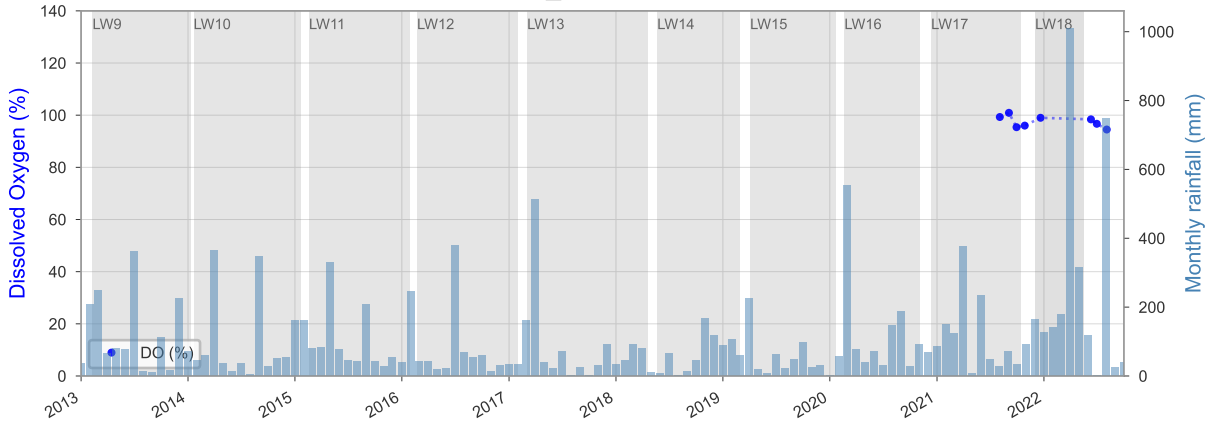
LC6_ROCKBAR1



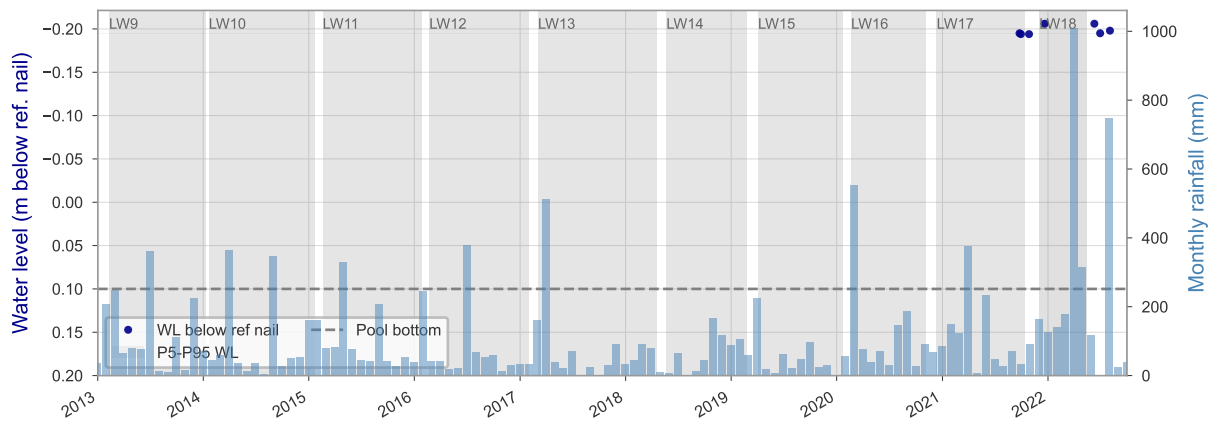
LC6_ROCKBAR1



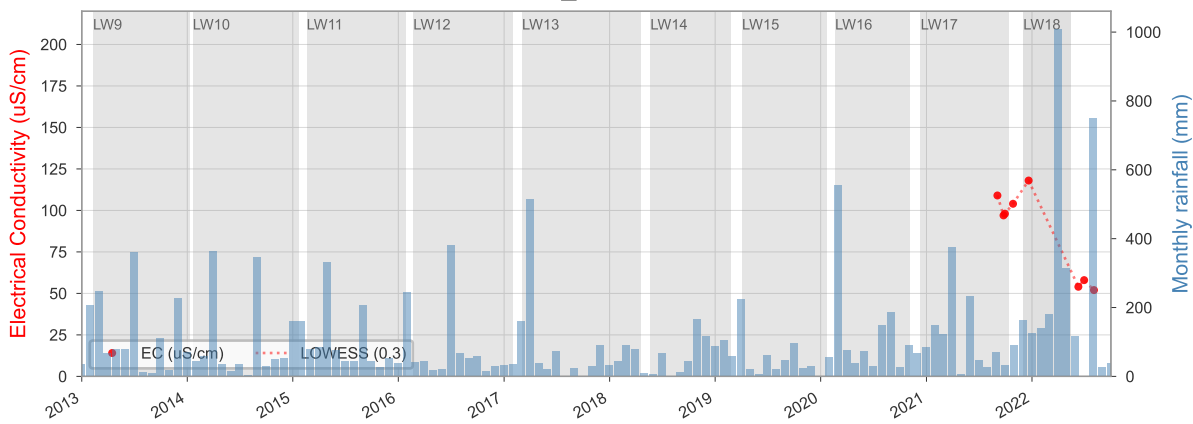
LC6_ROCKBAR1



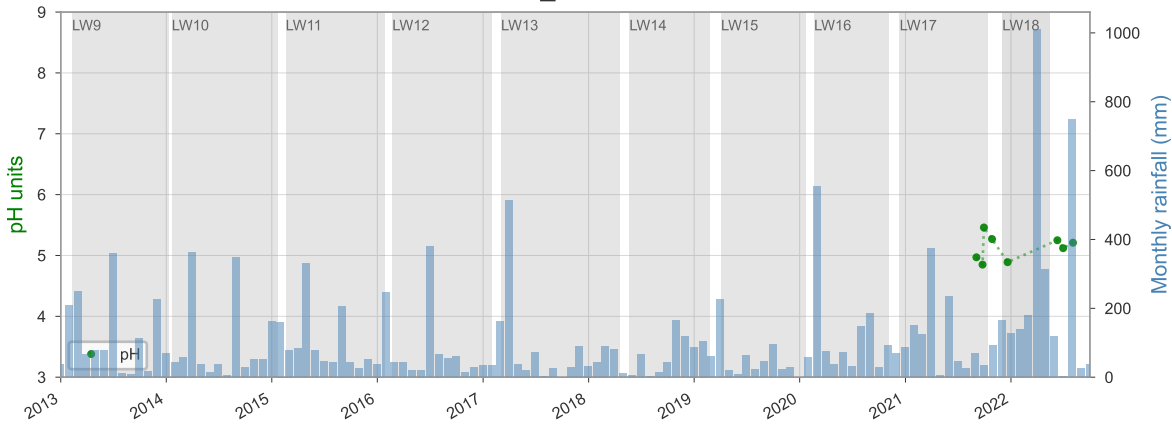
LC7_POOL17



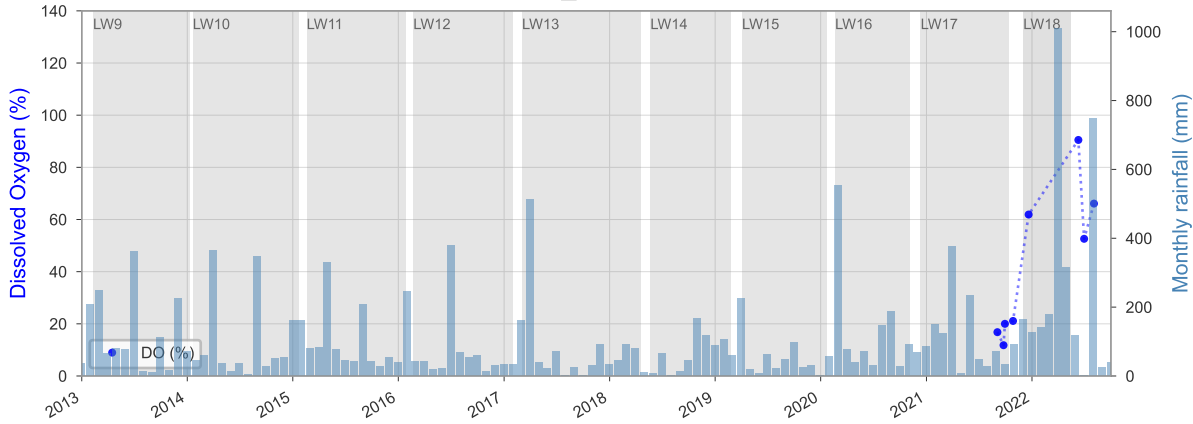
LC7_POOL17



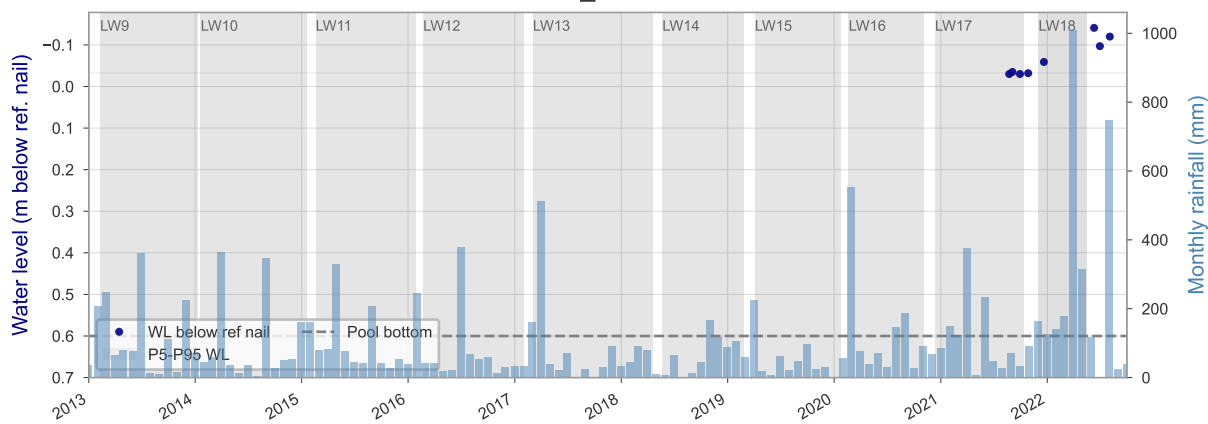
LC7_POOL17



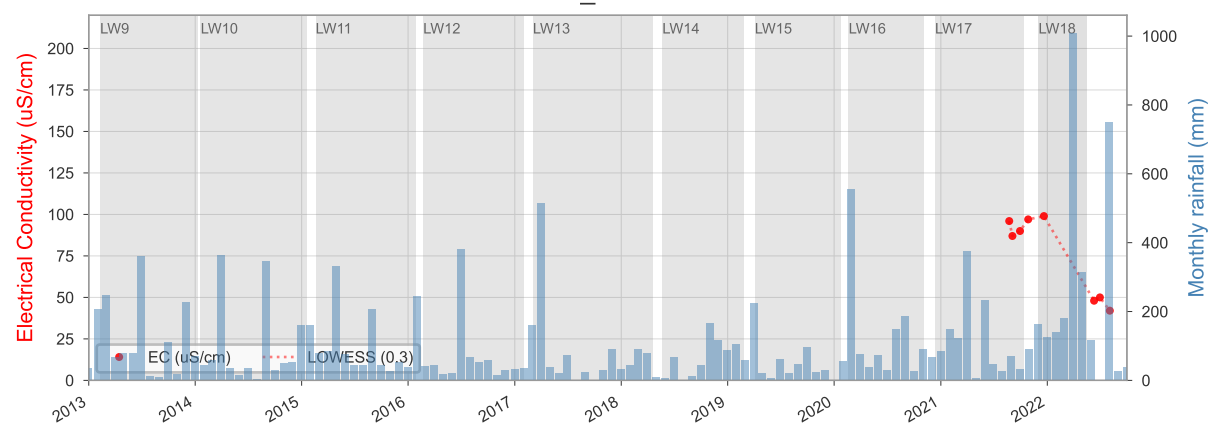
LC7_POOL17



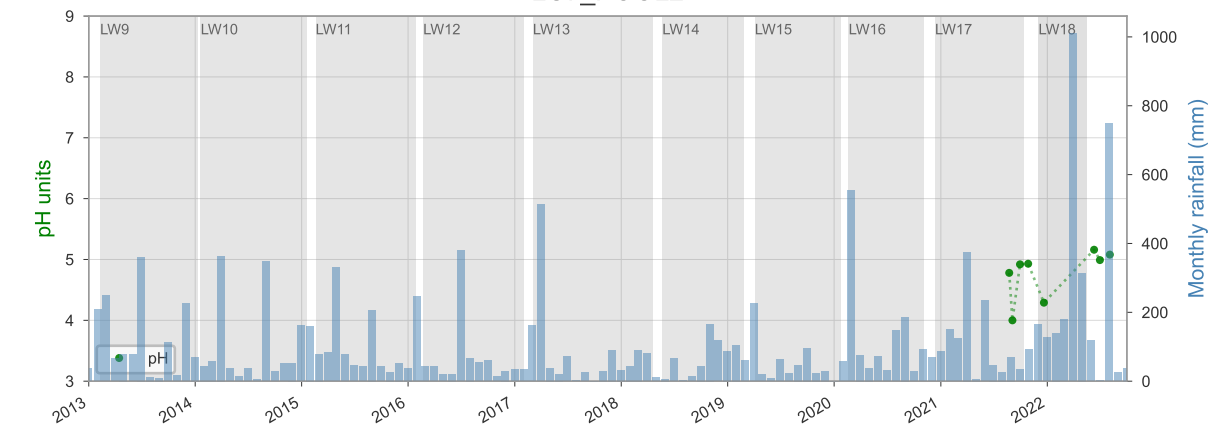
LC7_POOL2



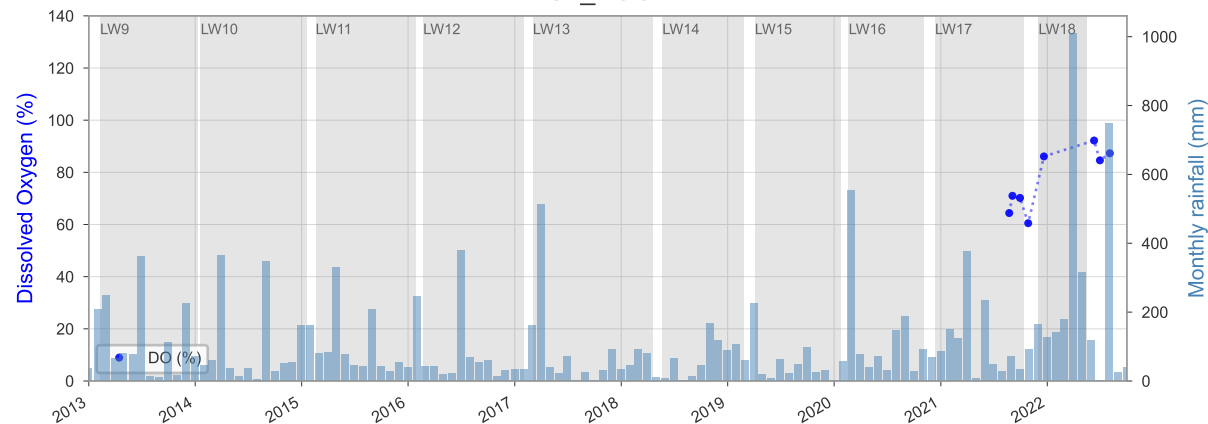
LC7_POOL2



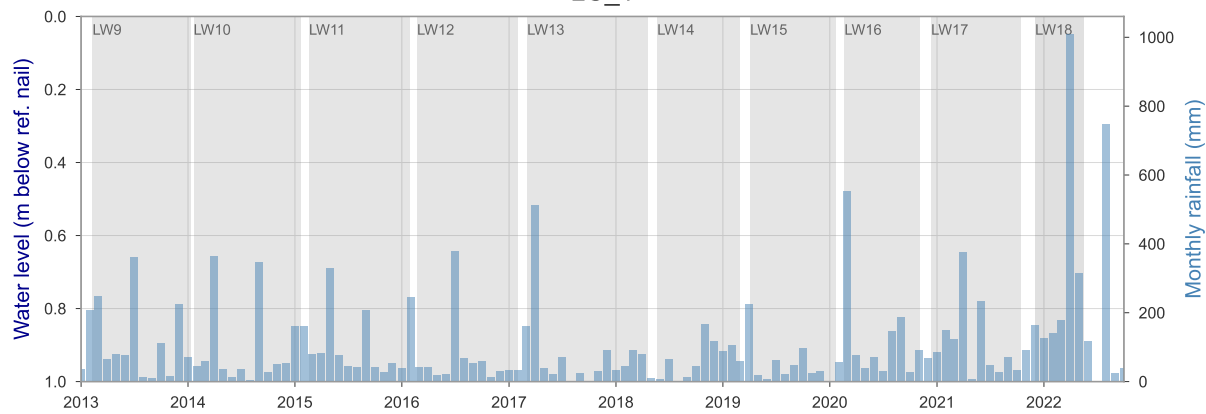
LC7_POOL2



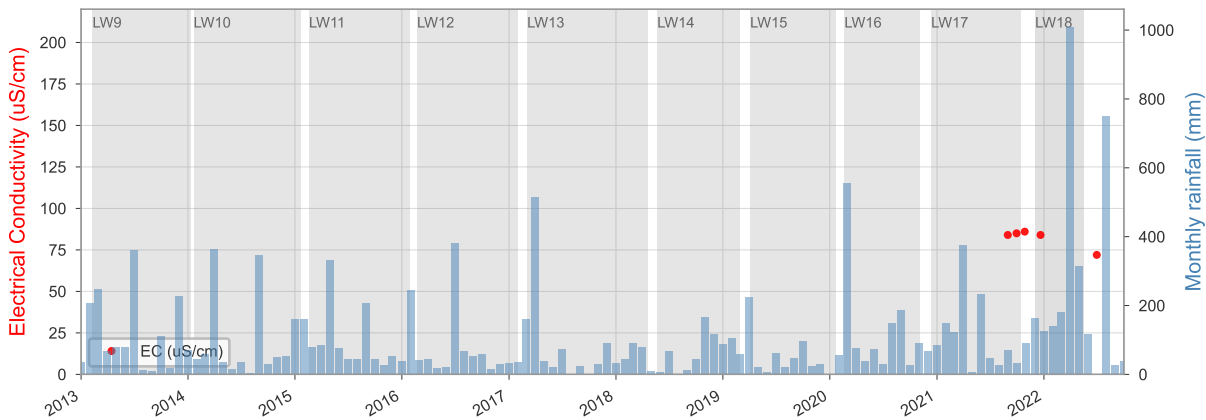
LC7_POOL2



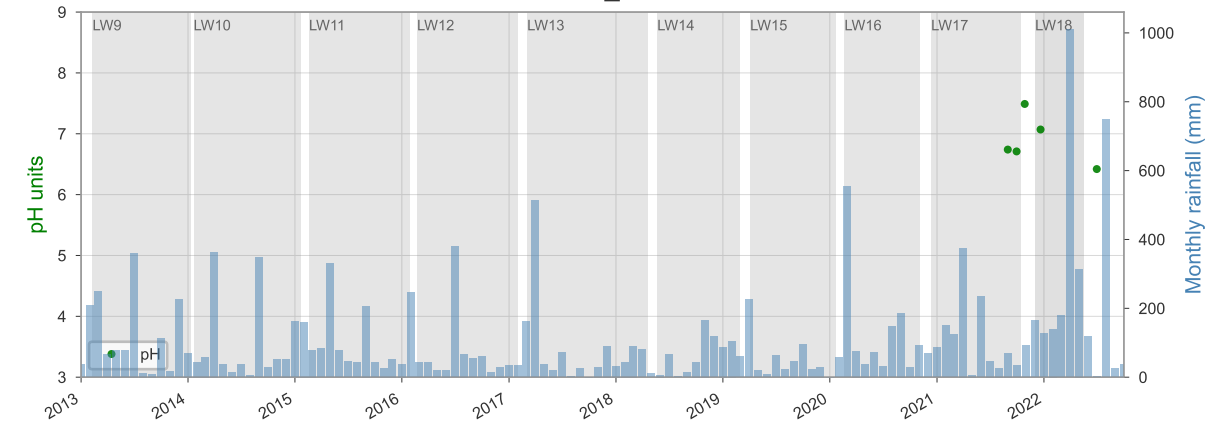
LC_1



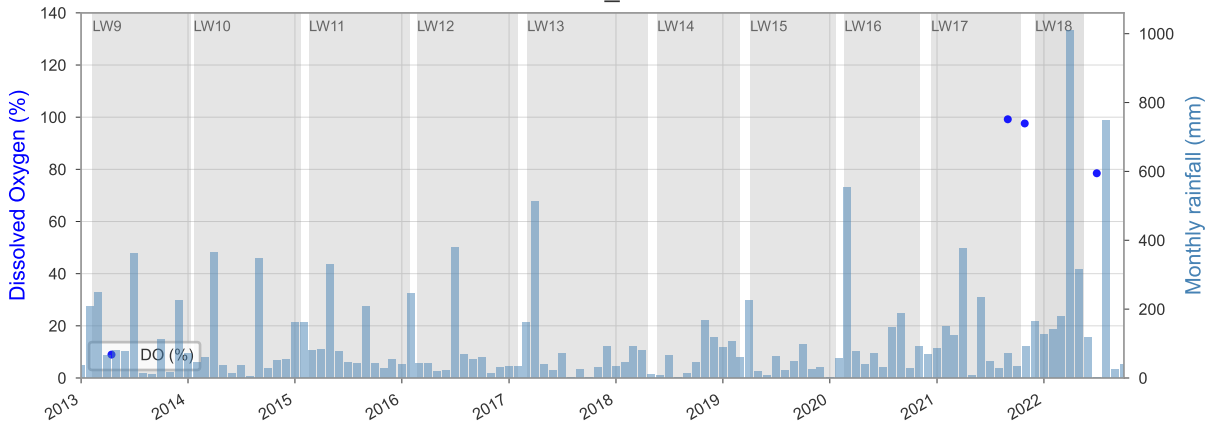
LC_1



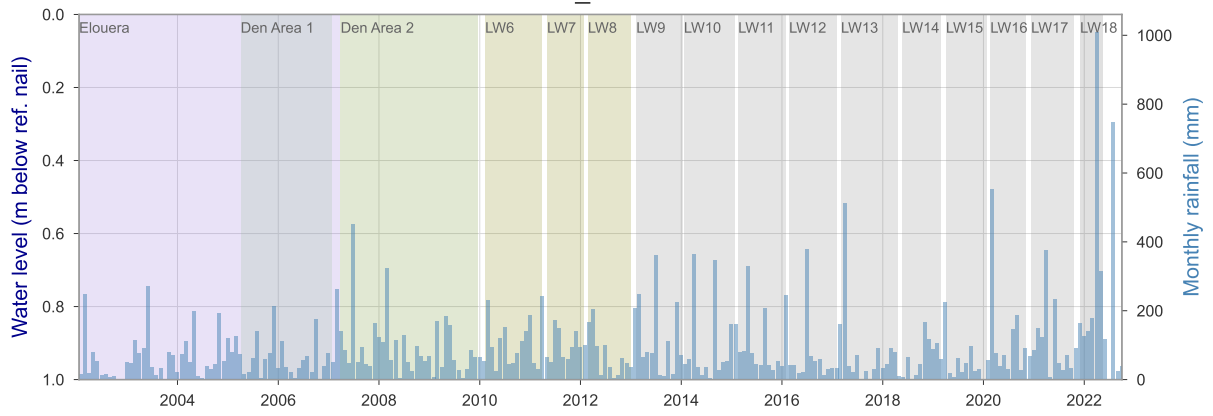
LC_1



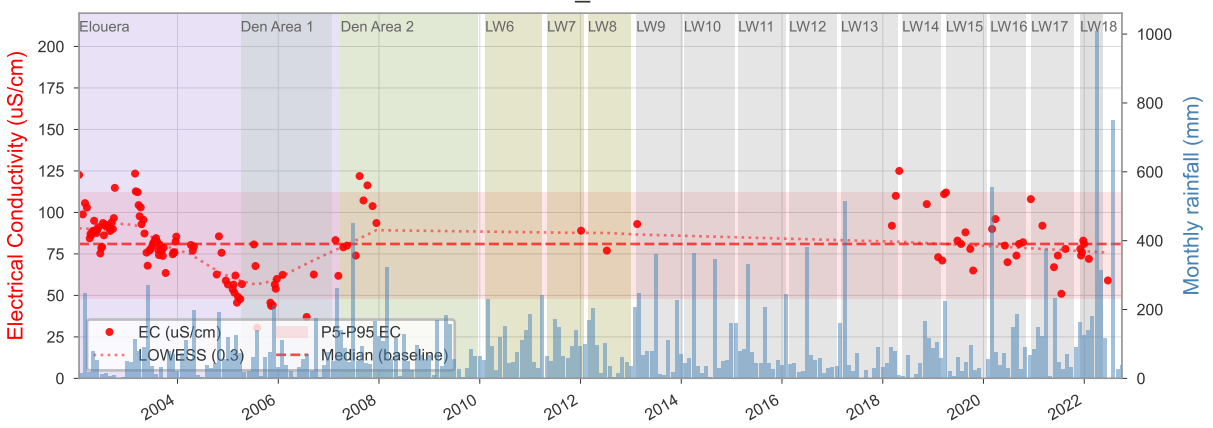
LC_1



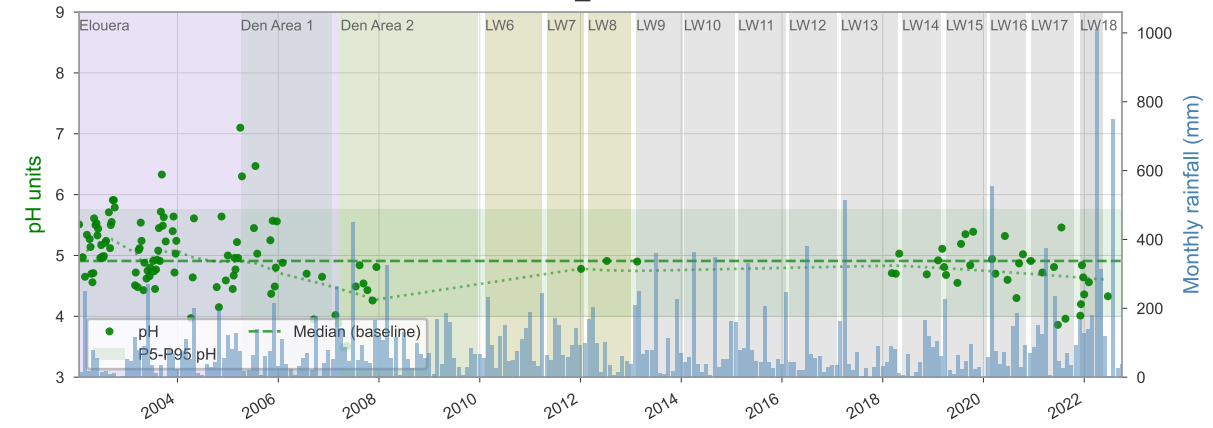
ND1_POOL2



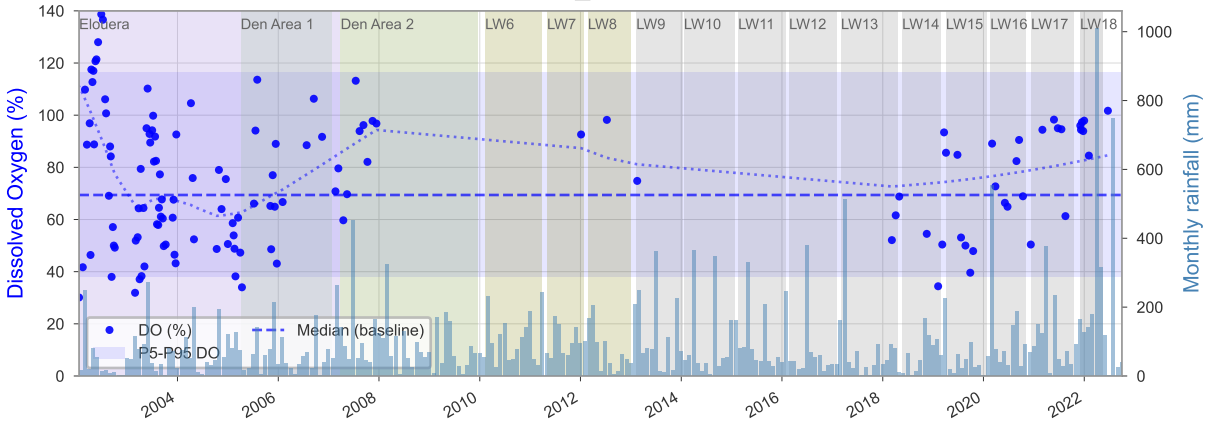
ND1_POOL2



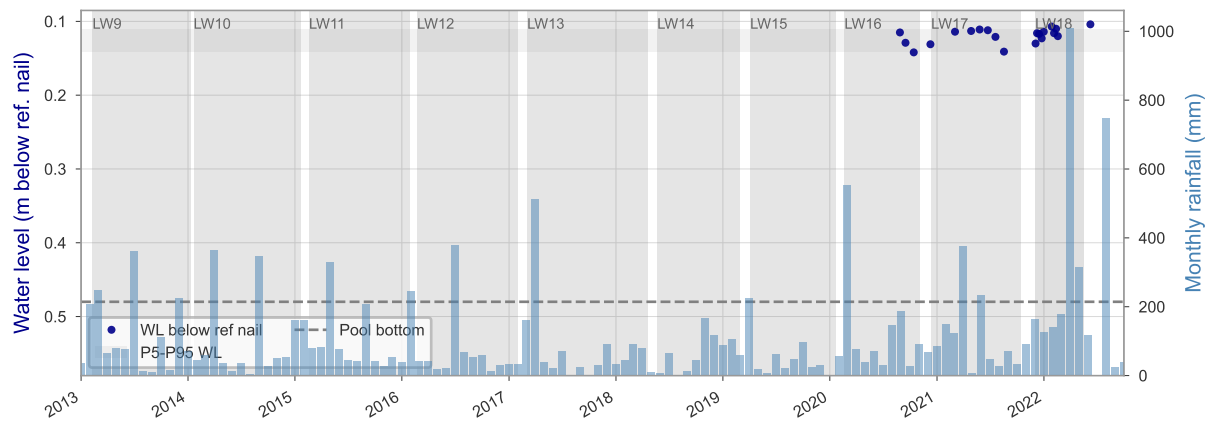
ND1_POOL2



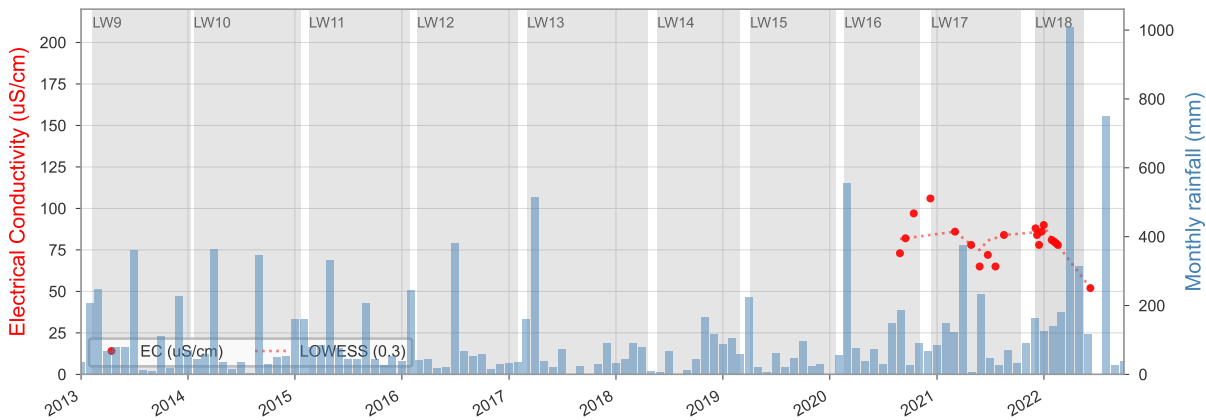
ND1_POOL2



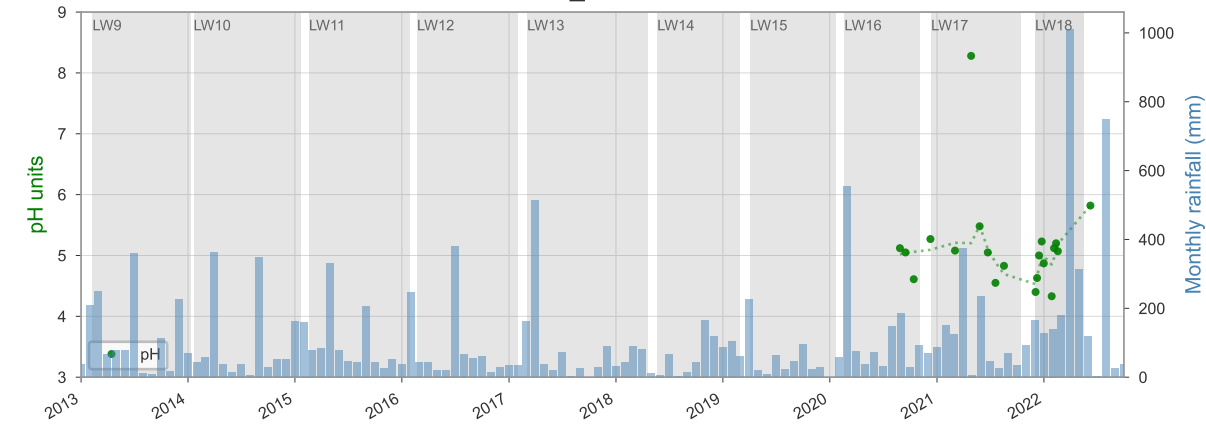
ND1_POOL23



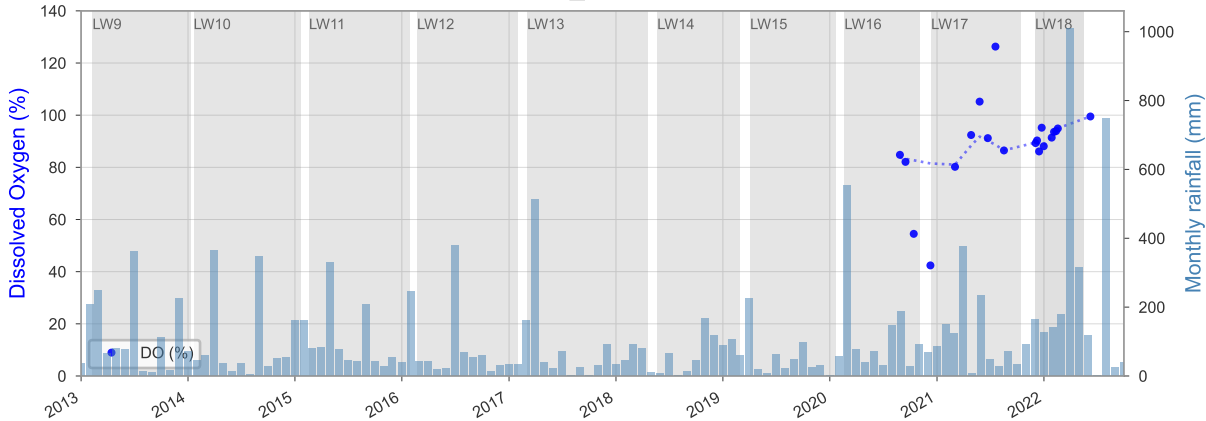
ND1_POOL23



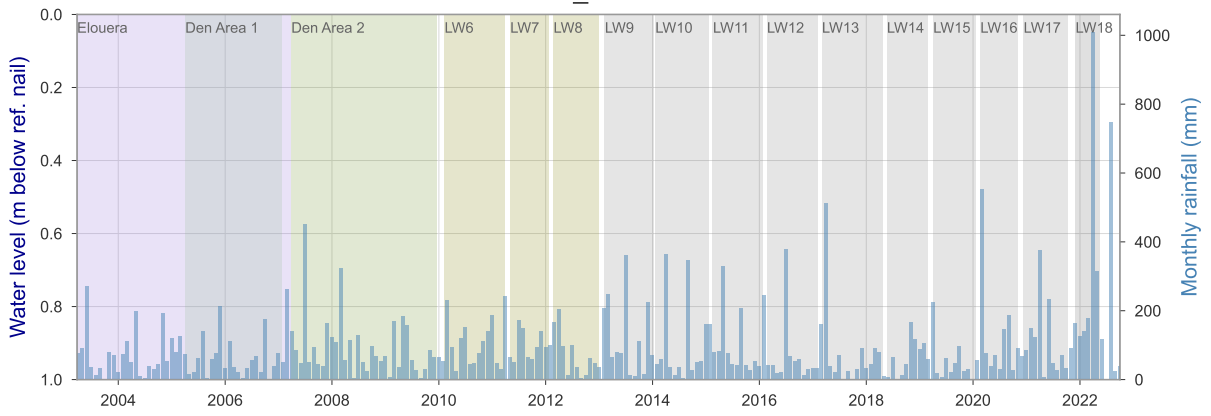
ND1_POOL23



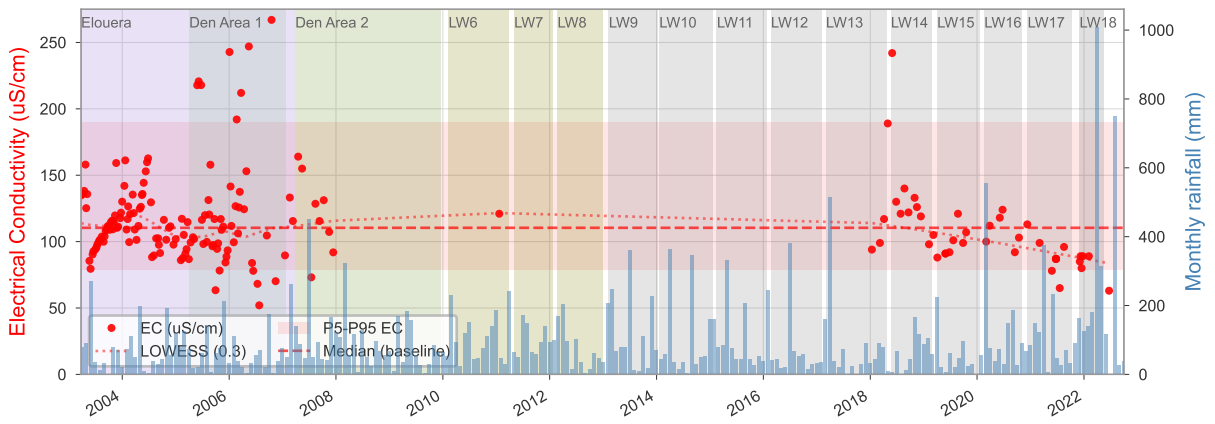
ND1_POOL23



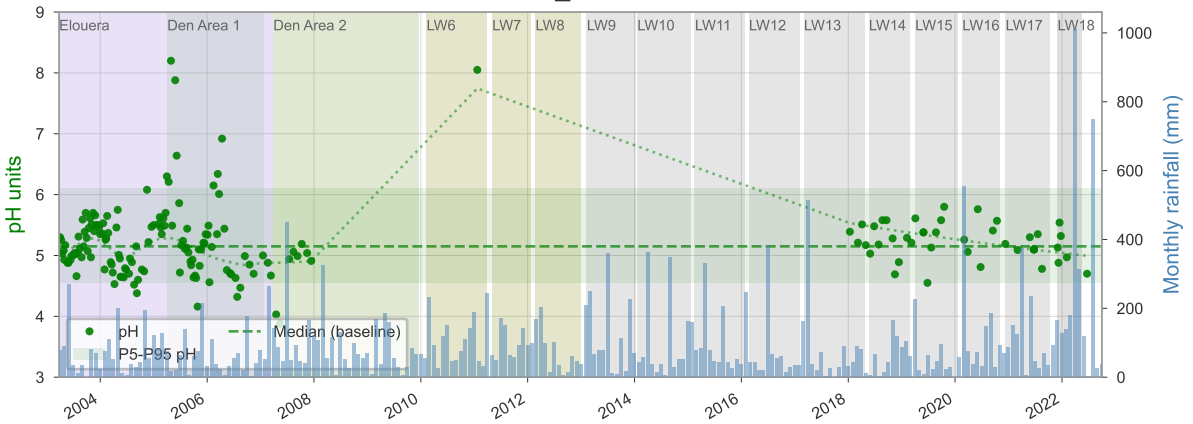
ND2_POOL3



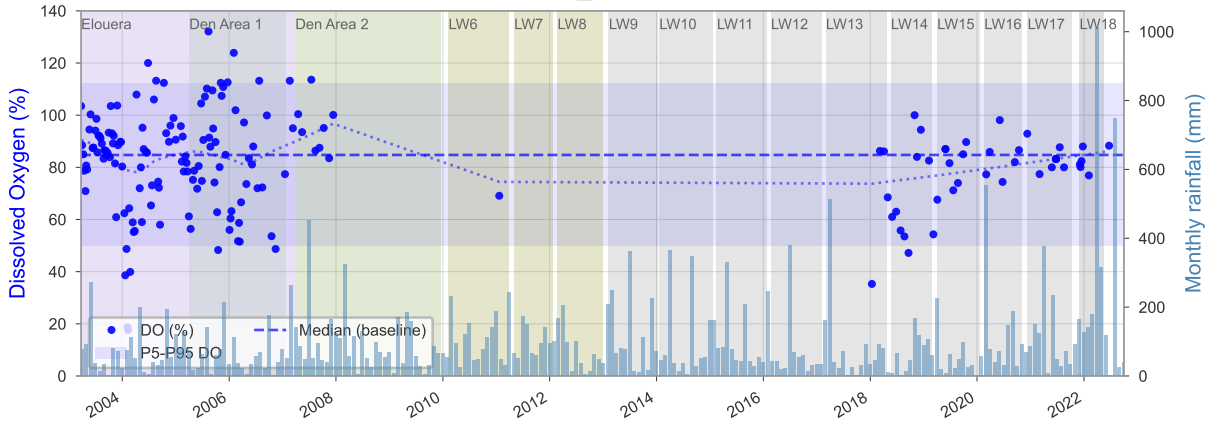
ND2_POOL3



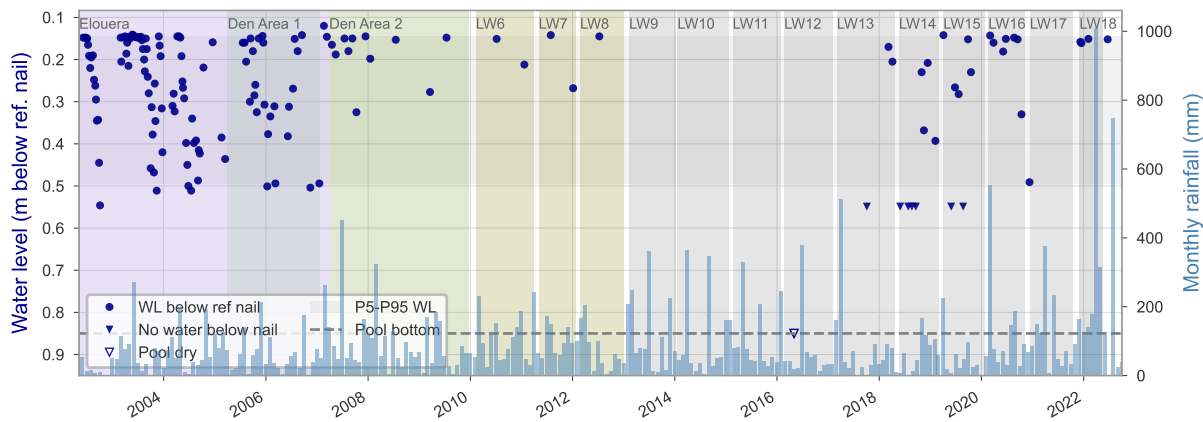
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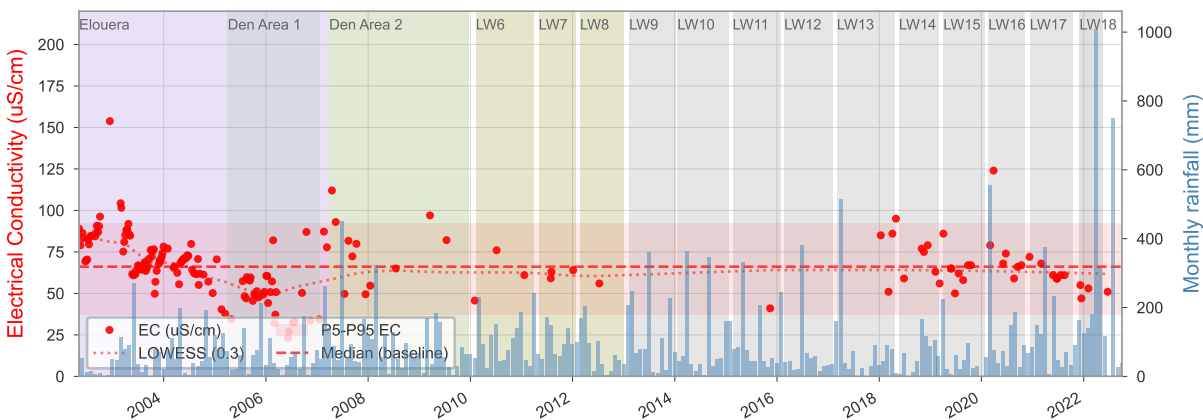
ND2_POOL3



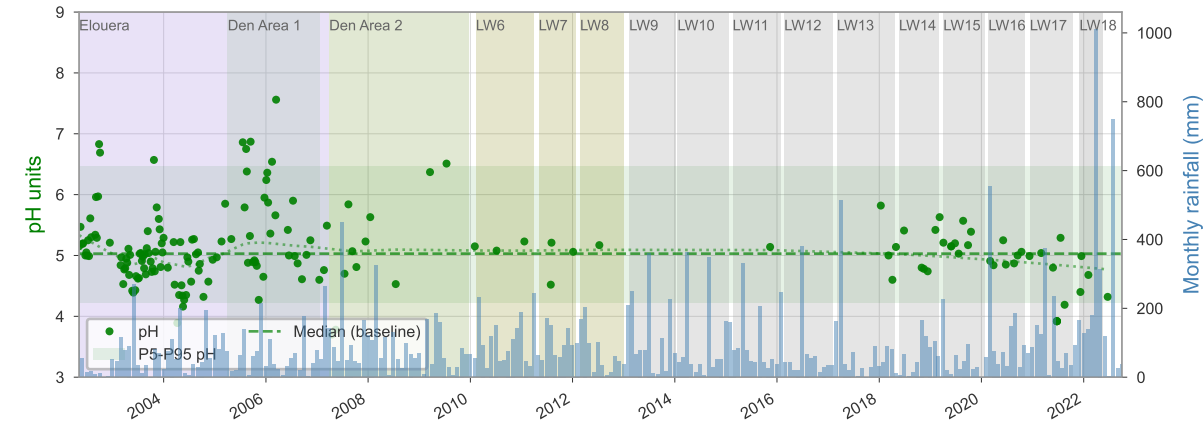
NDC1



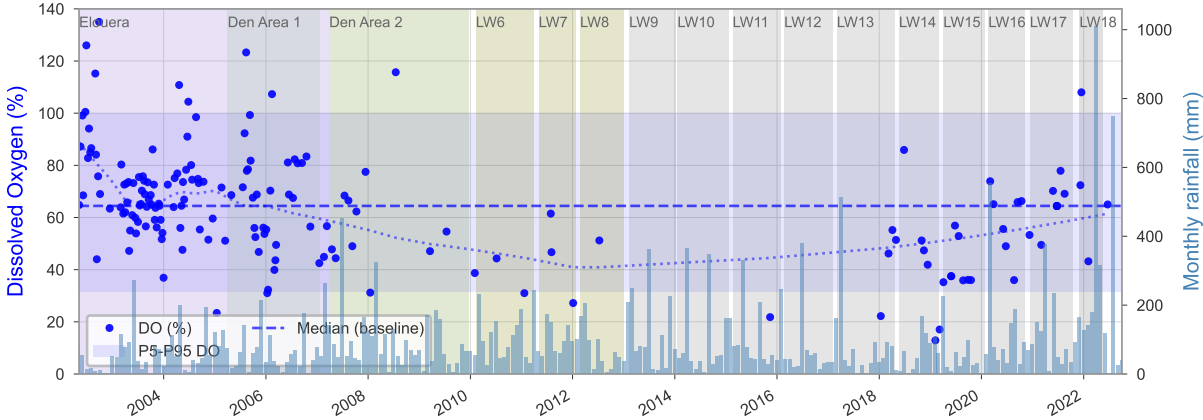
NDC1



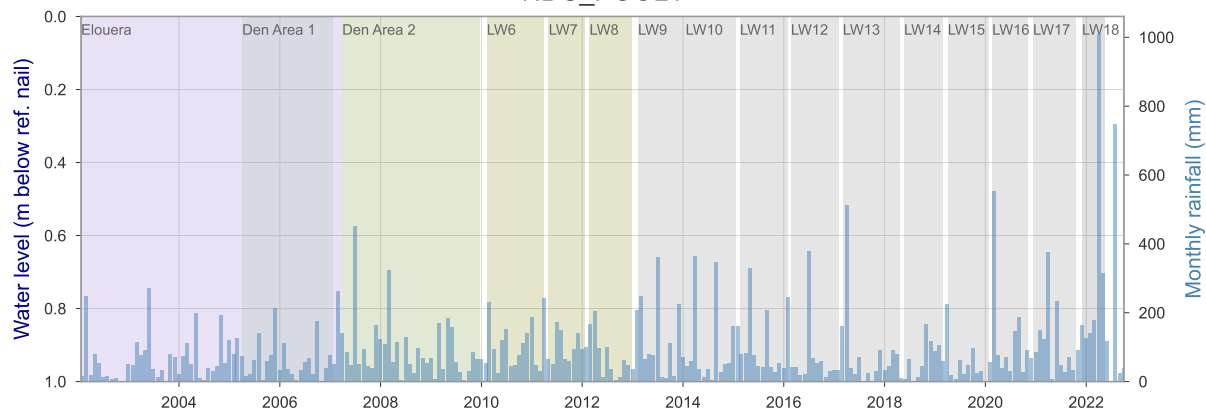
NDC1



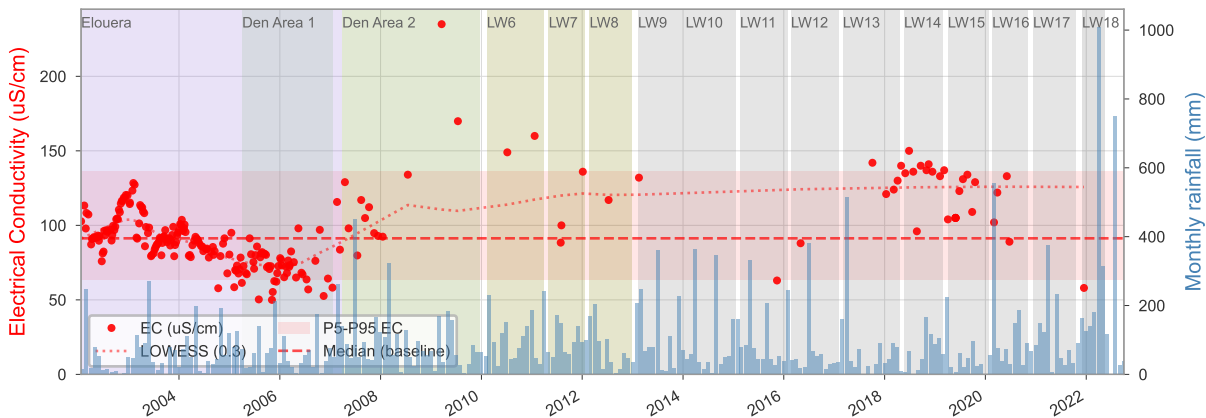
NDC1



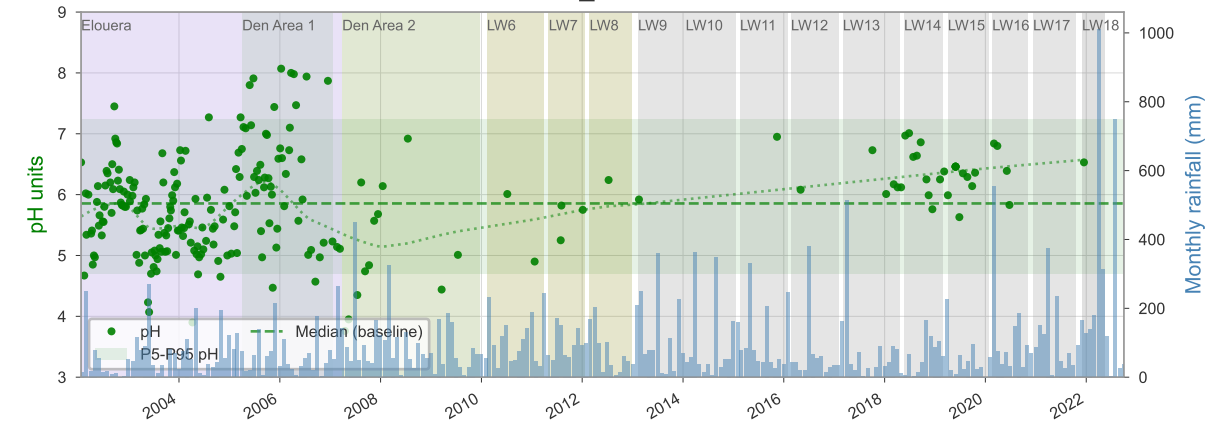
NDC_POOL1



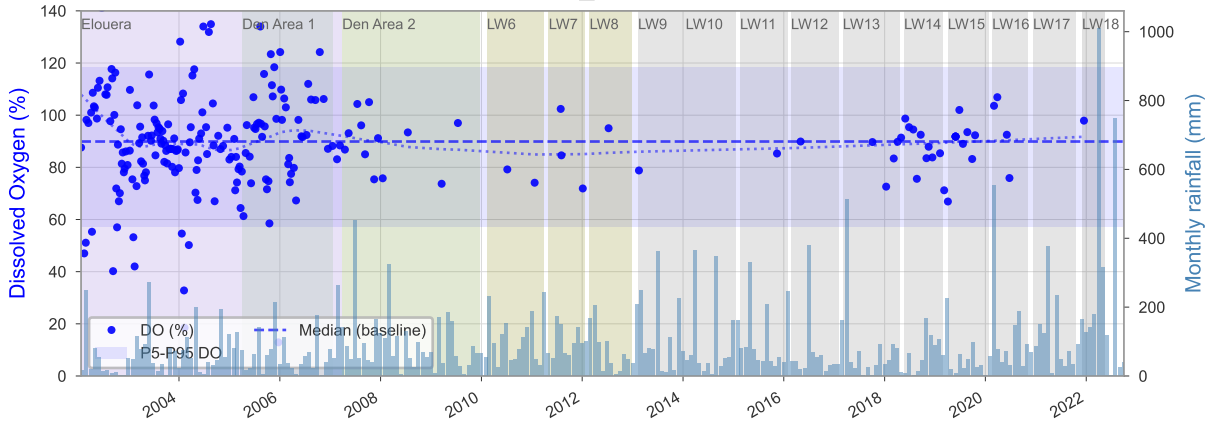
NDC_POOL1



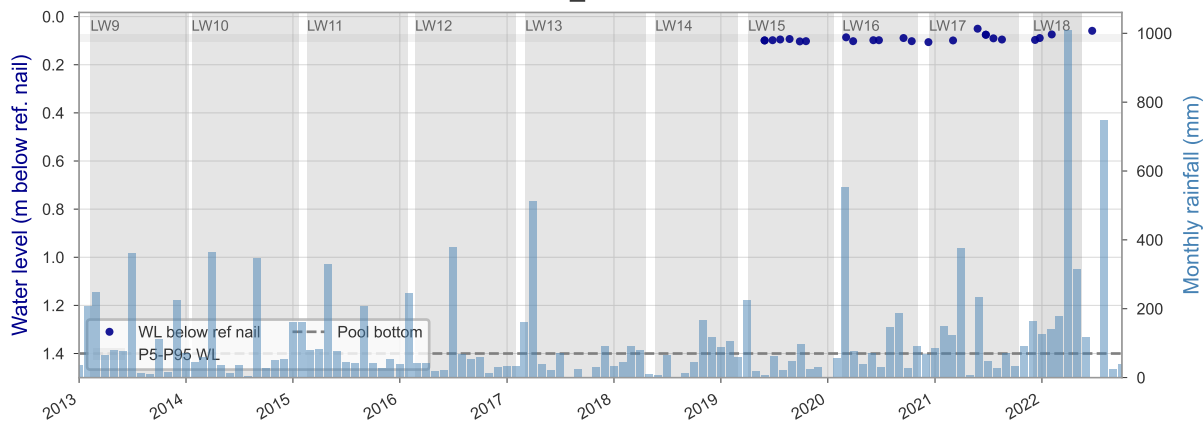
NDC_POOL1



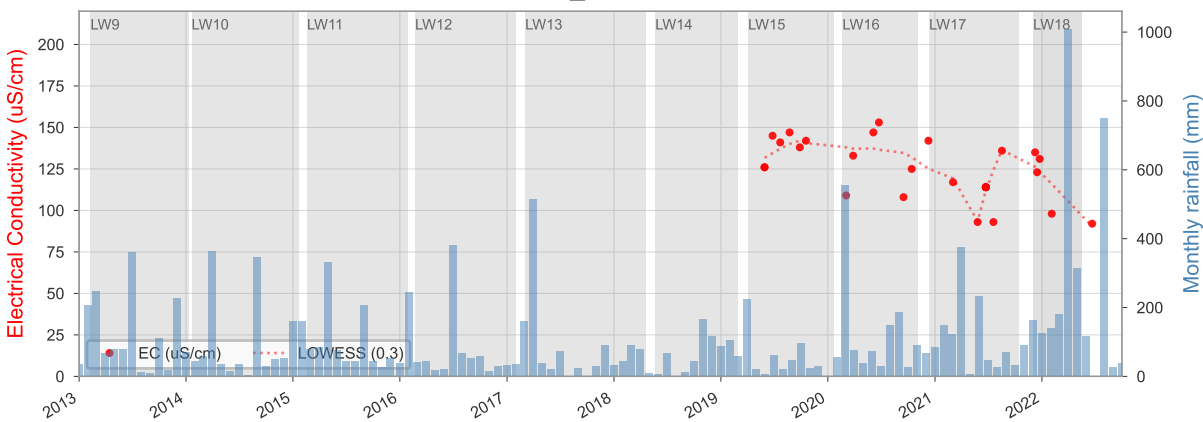
NDC_POOL1



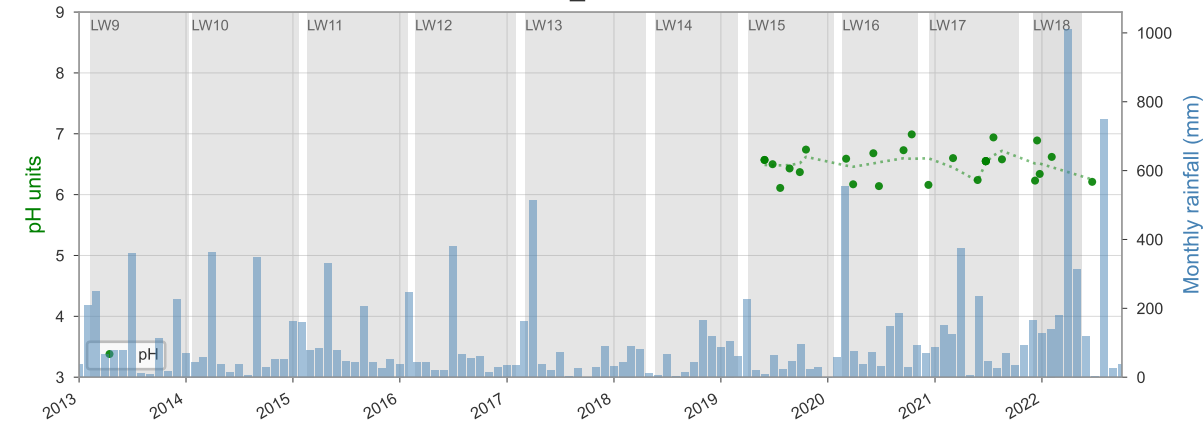
NDC_POOL15



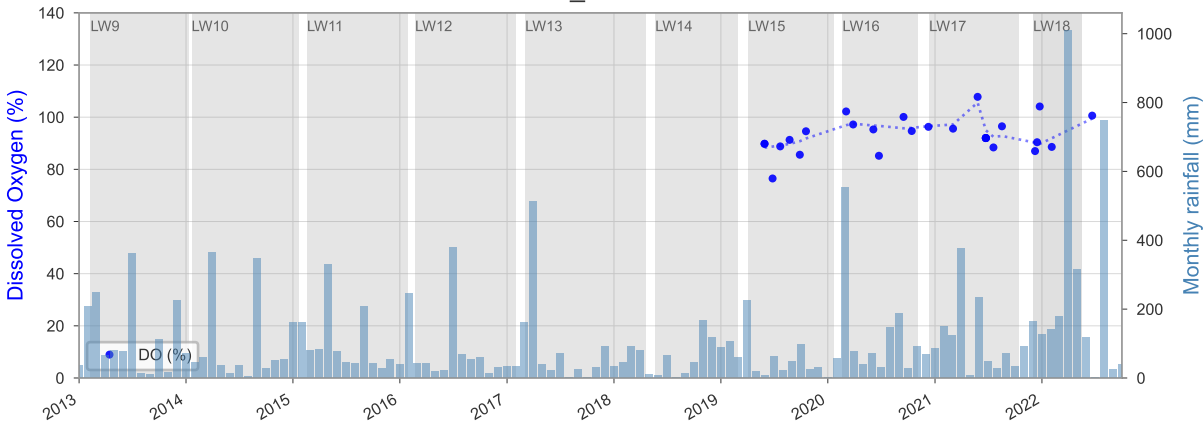
NDC_POOL15



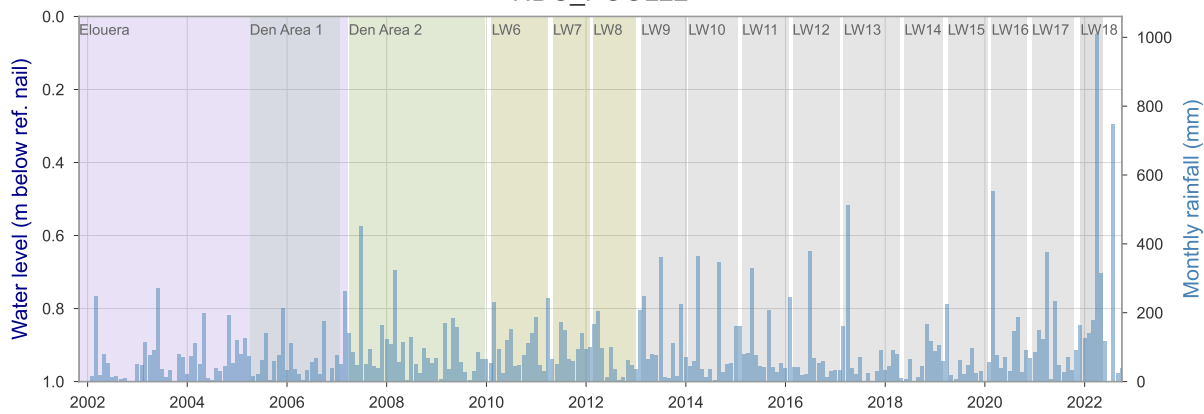
NDC_POOL15



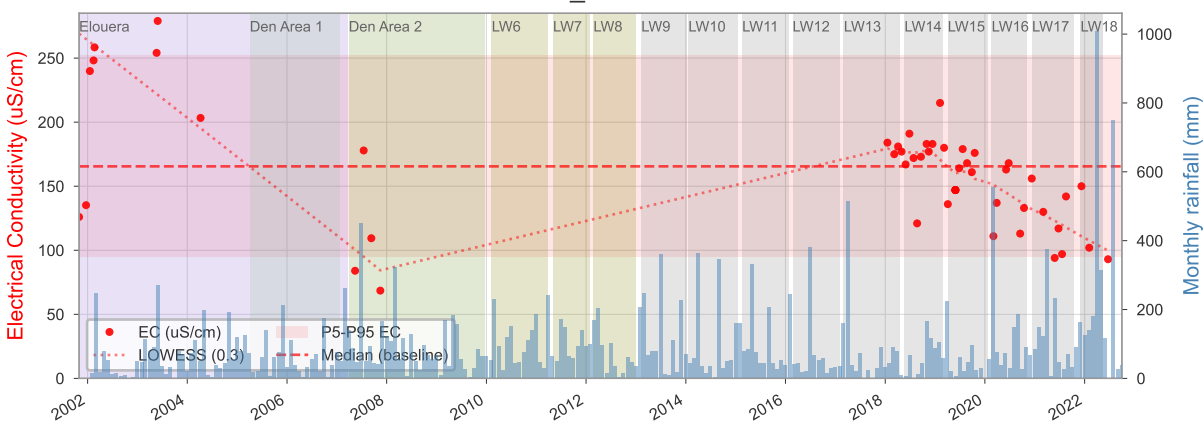
NDC_POOL15



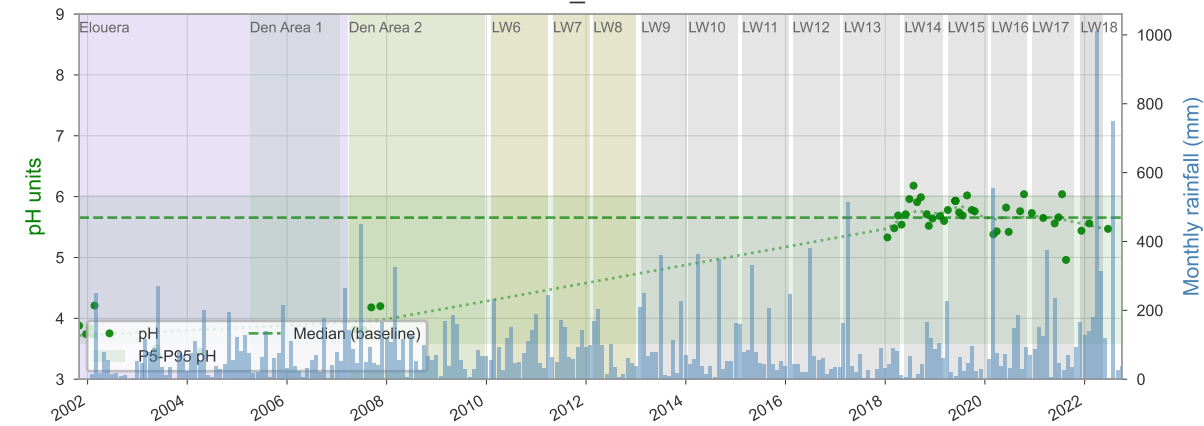
NDC_POOL22



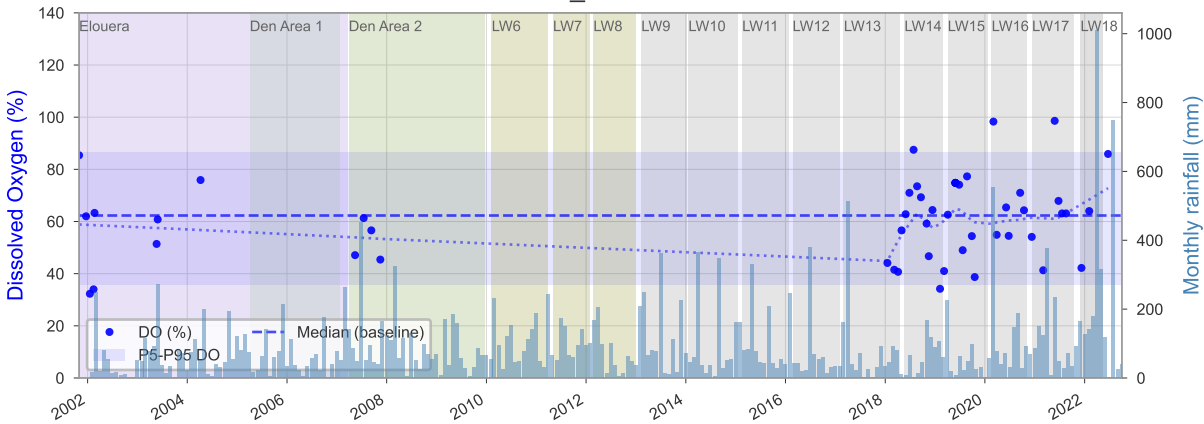
NDC_POOL22



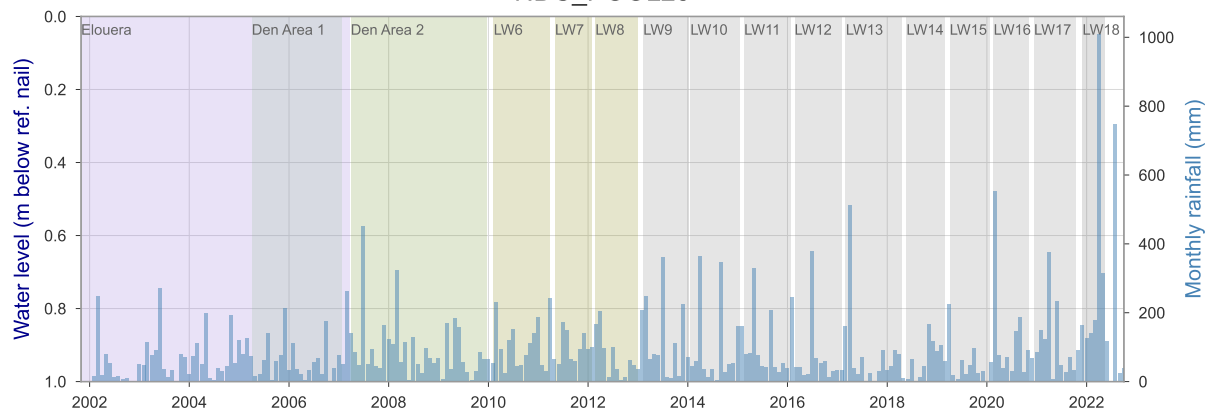
NDC_POOL22



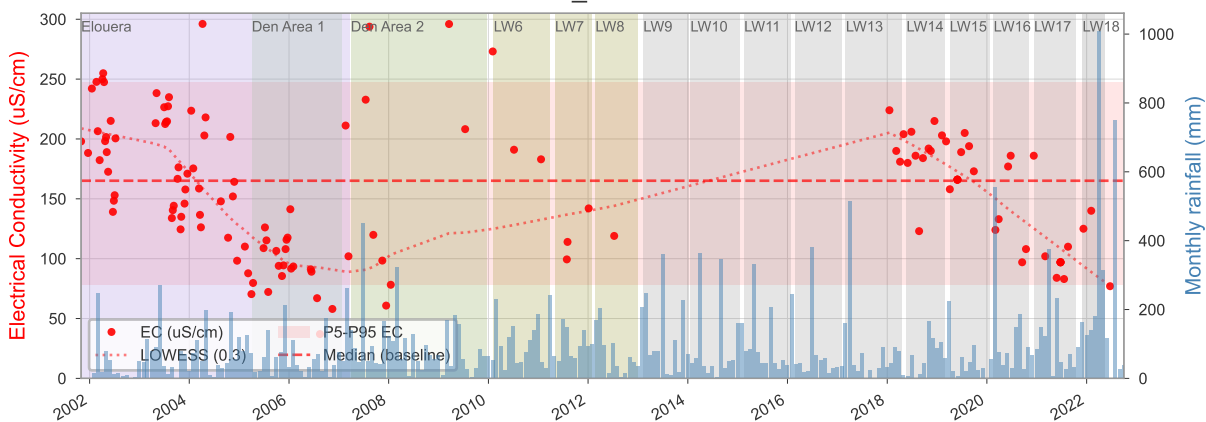
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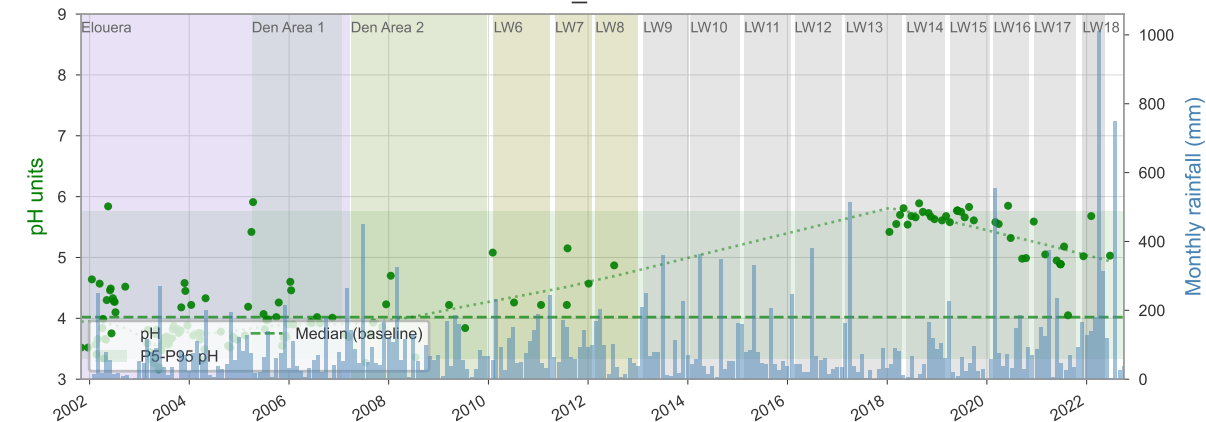
NDC_POOL25



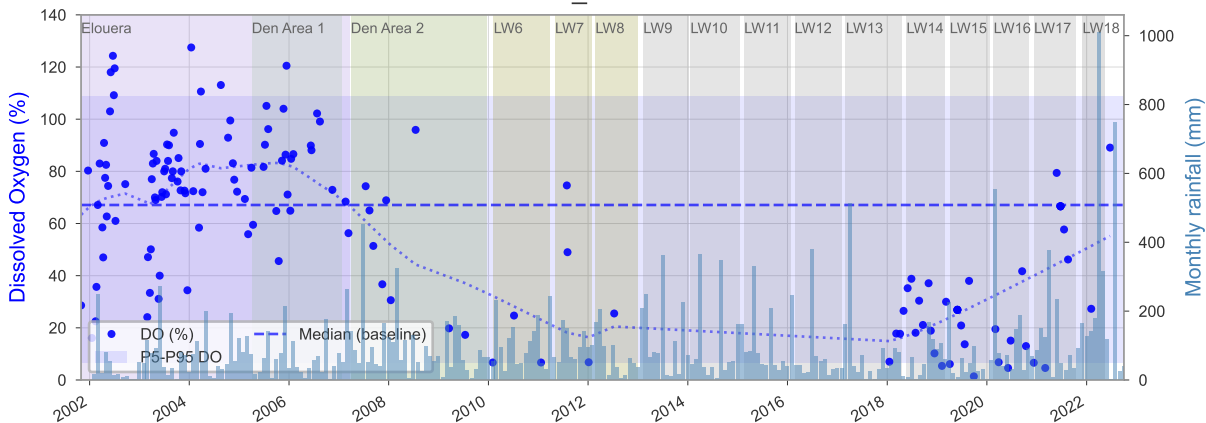
NDC_POOL25



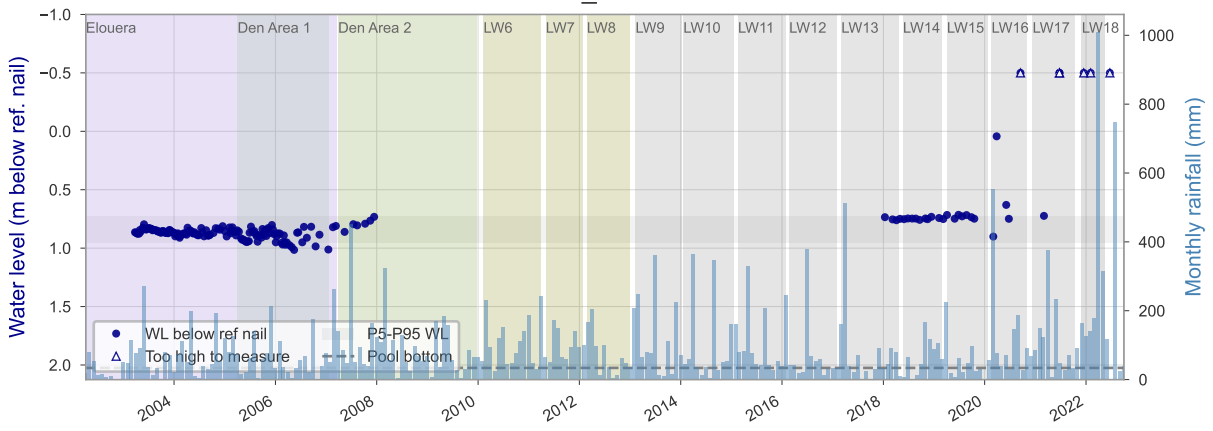
NDC_POOL25



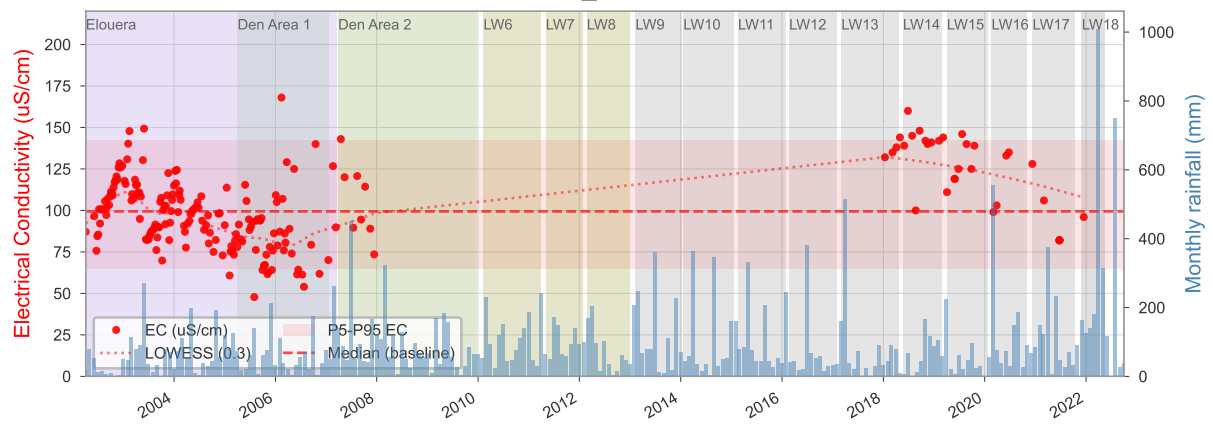
NDC_POOL25



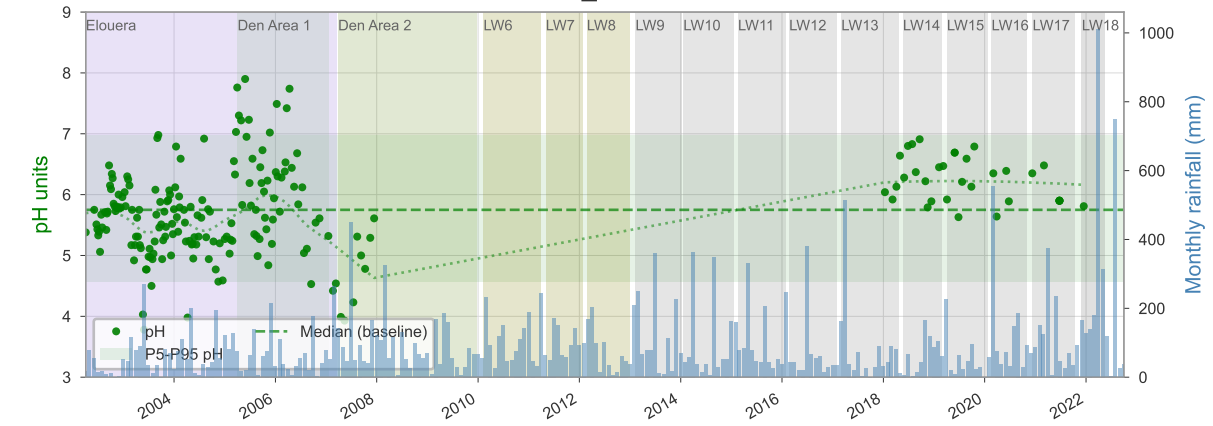
NDC_POOL6



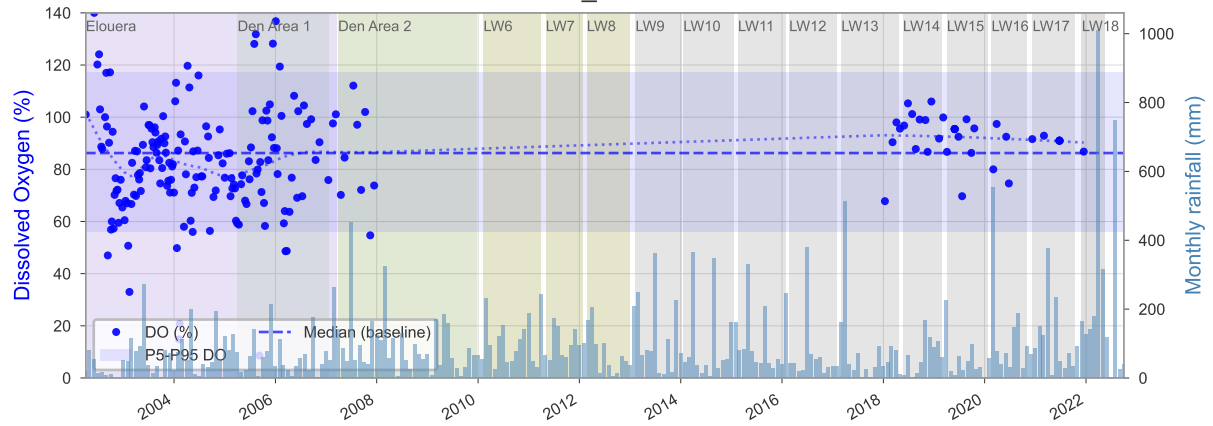
NDC_POOL6



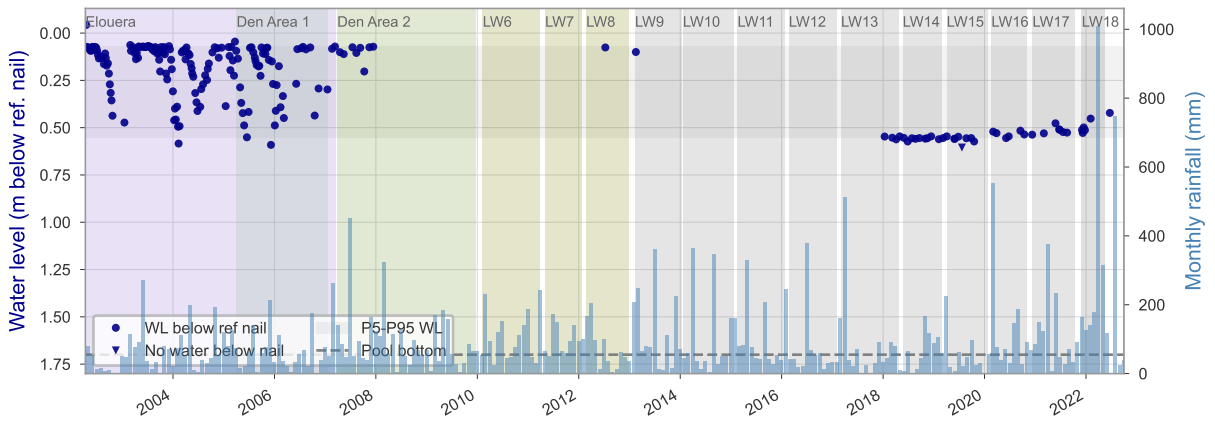
NDC_POOL6



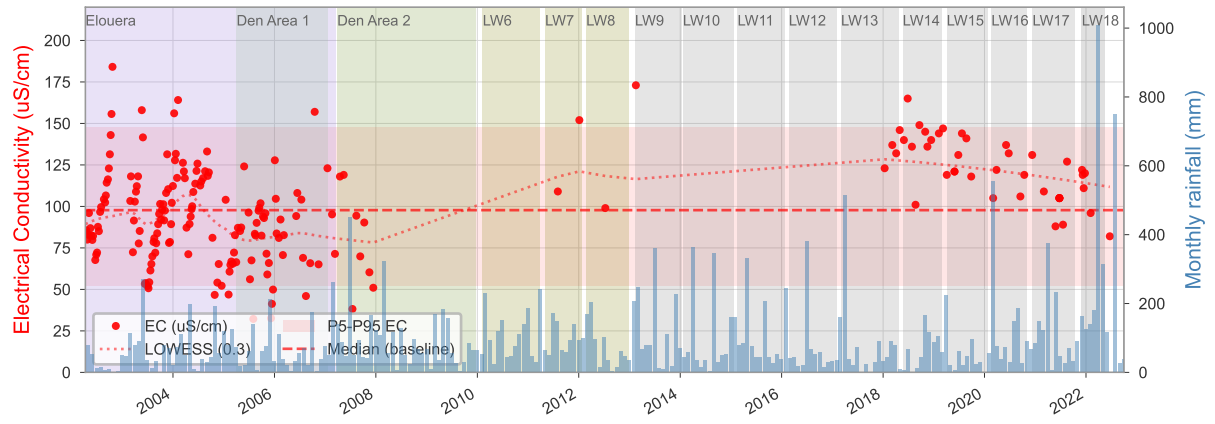
NDC_POOL6



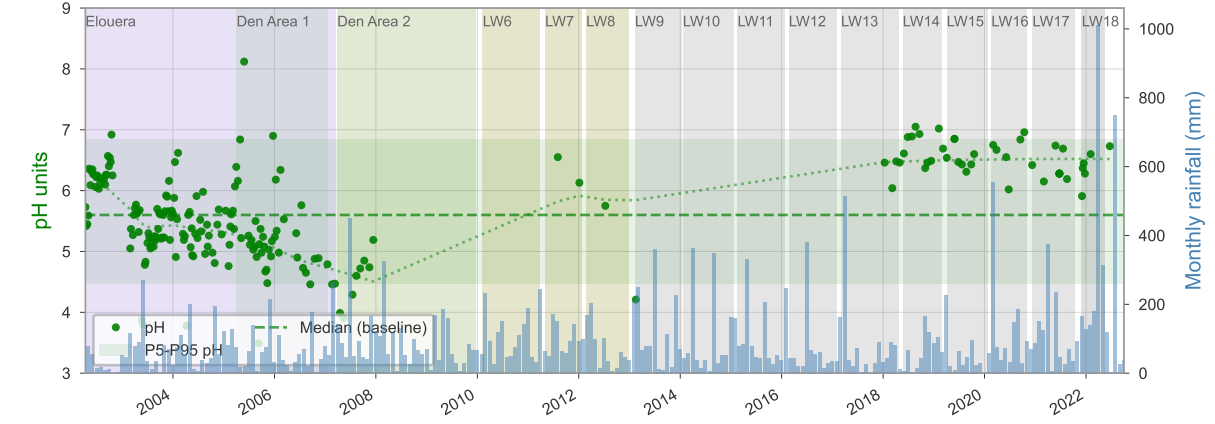
NDC_POOL7



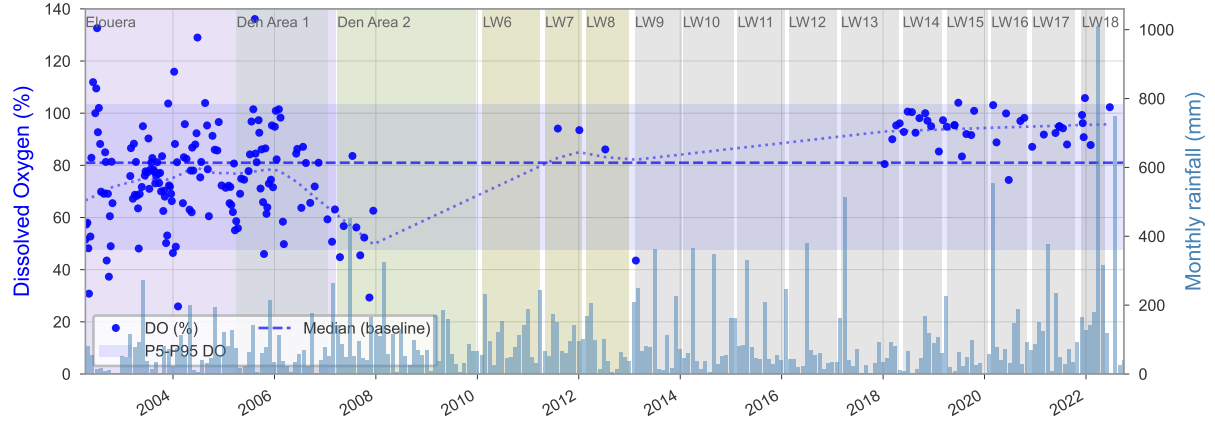
NDC_POOL7



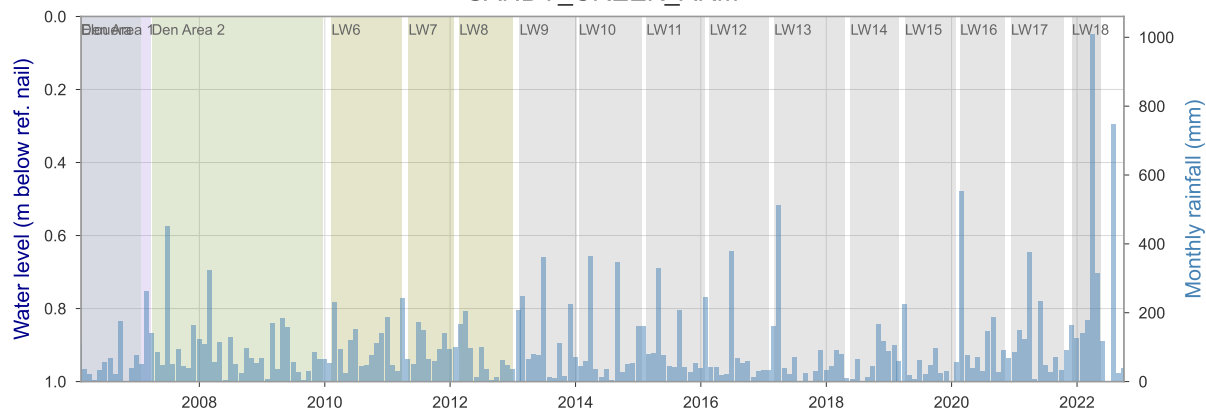
NDC_POOL7



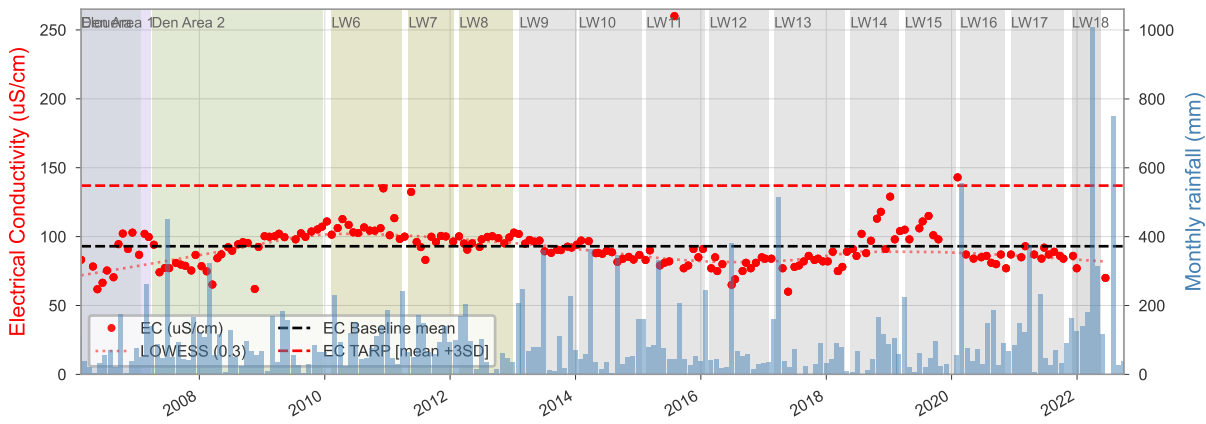
NDC_POOL7



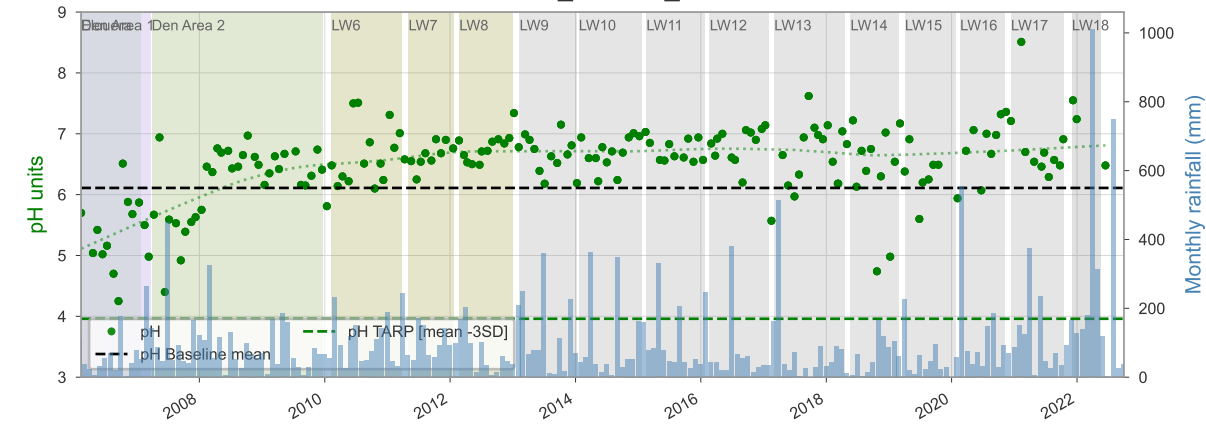
SANDY_CREEK_ARM



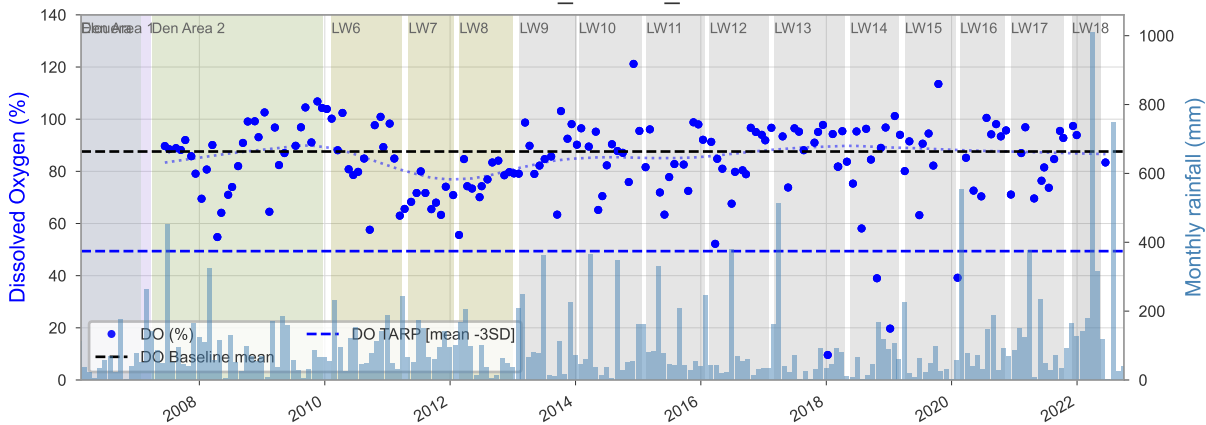
SANDY_CREEK_ARM



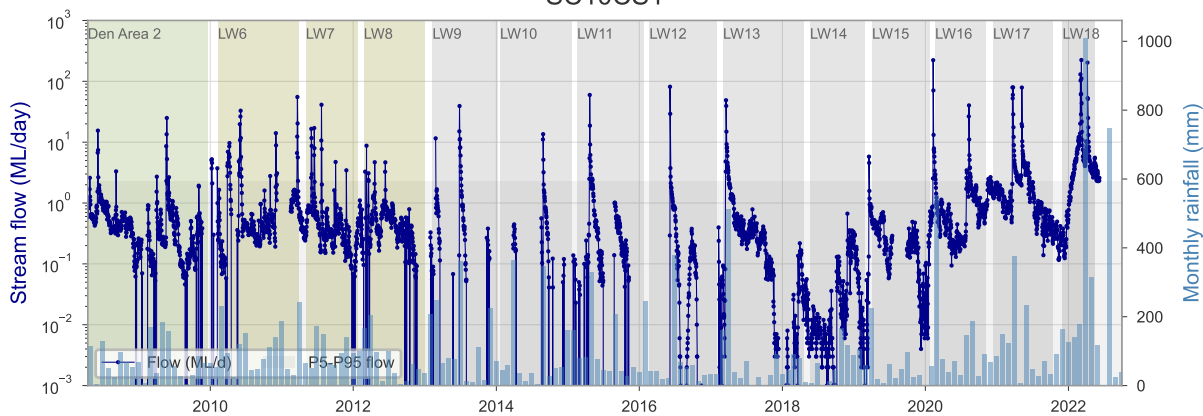
SANDY_CREEK_ARM



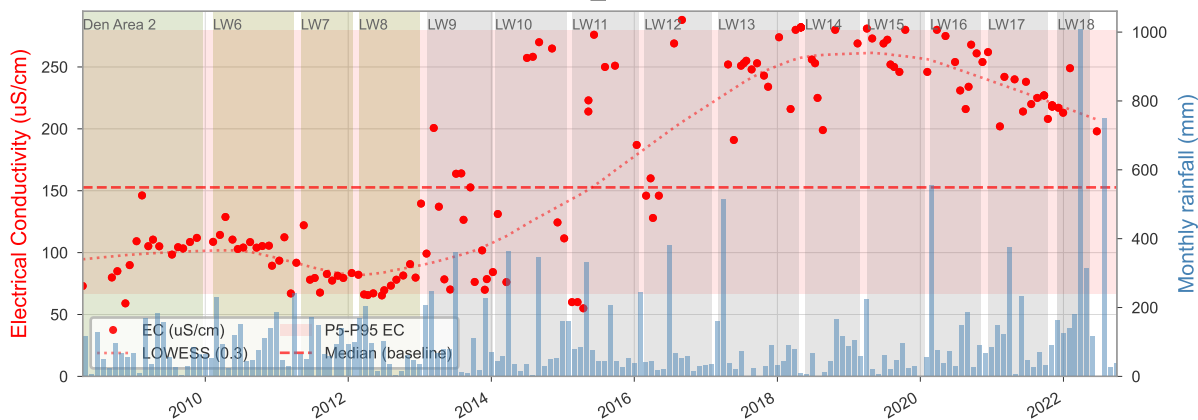
SANDY_CREEK_ARM



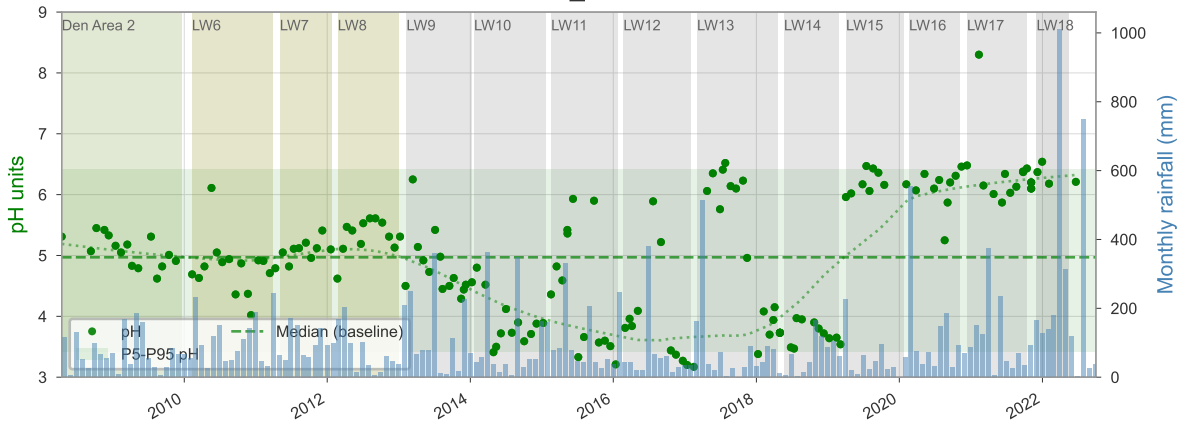
SC10CS1



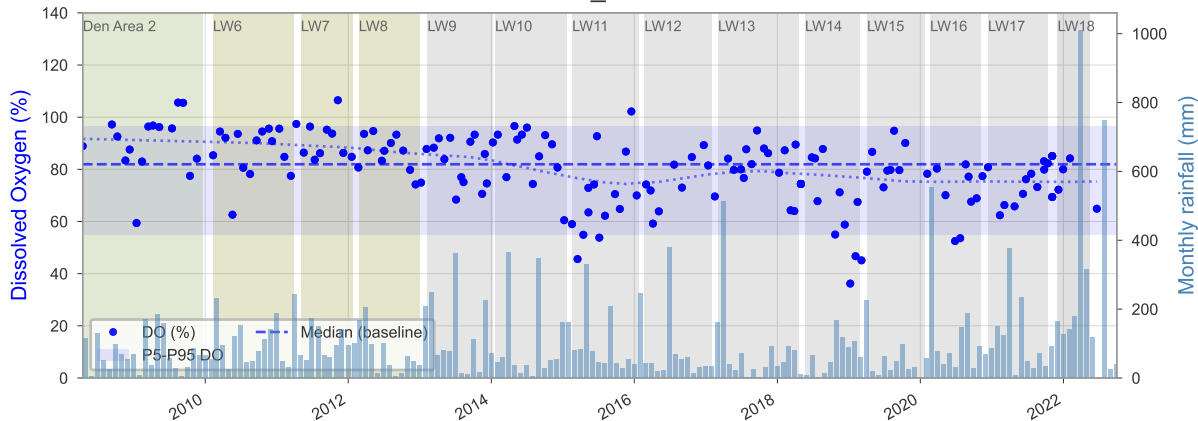
SC10C_POOL1



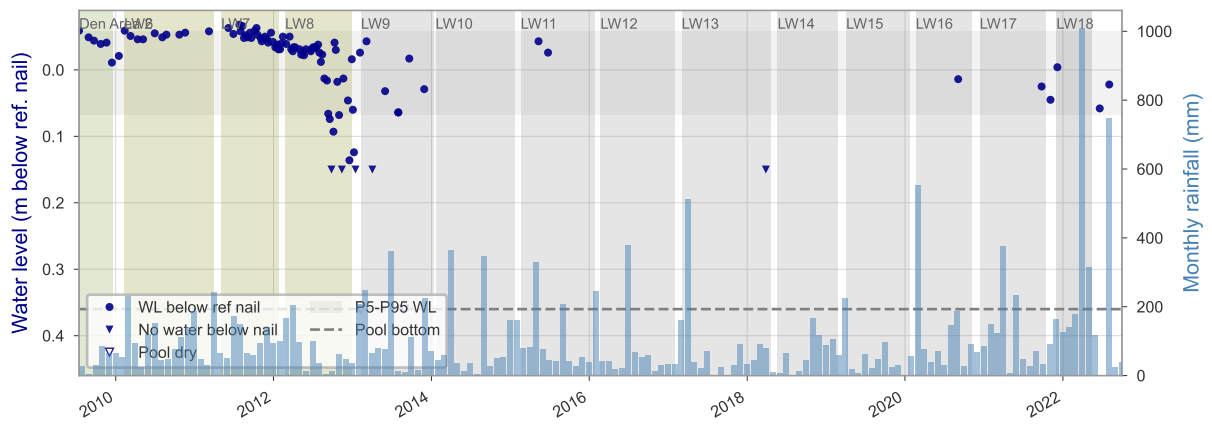
SC10C_POOL1



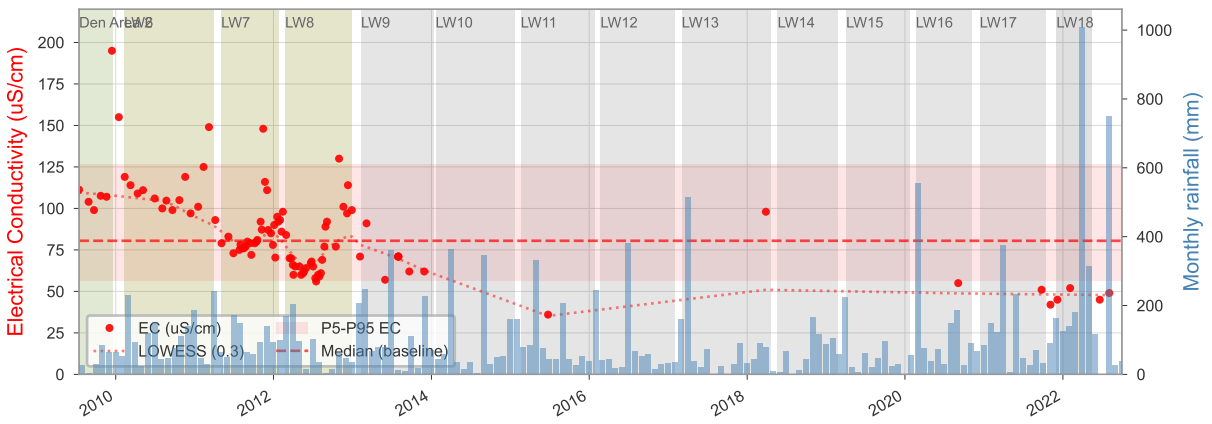
SC10C_POOL1



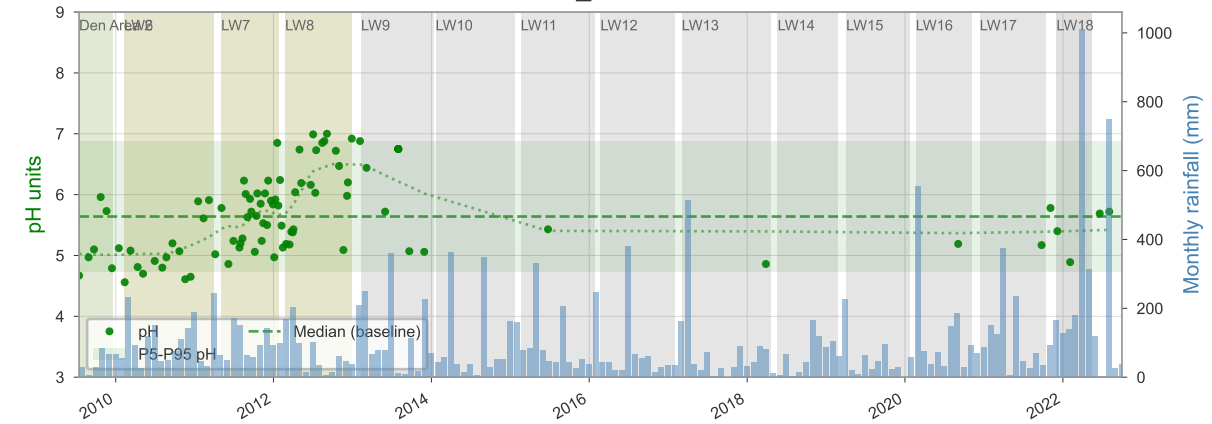
SC10C_POOL11A



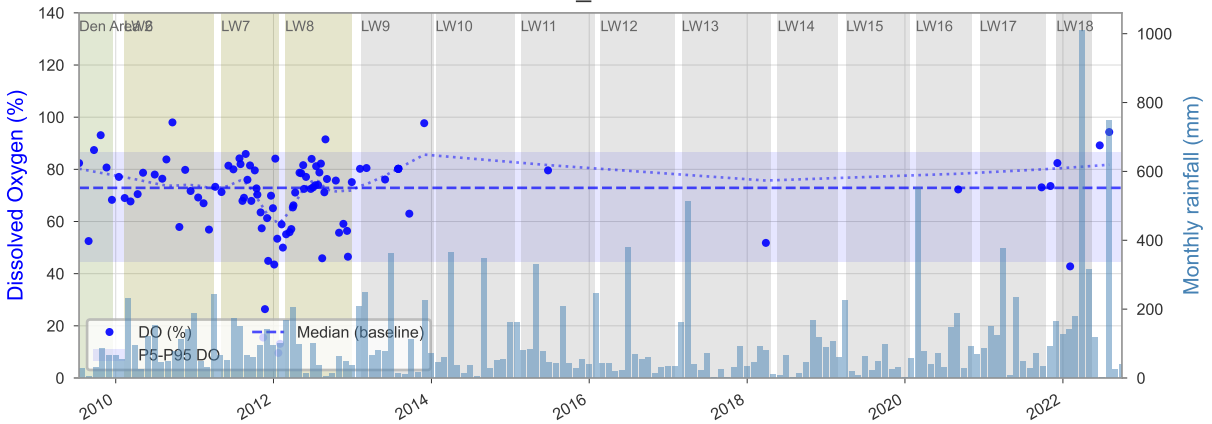
SC10C_POOL11A



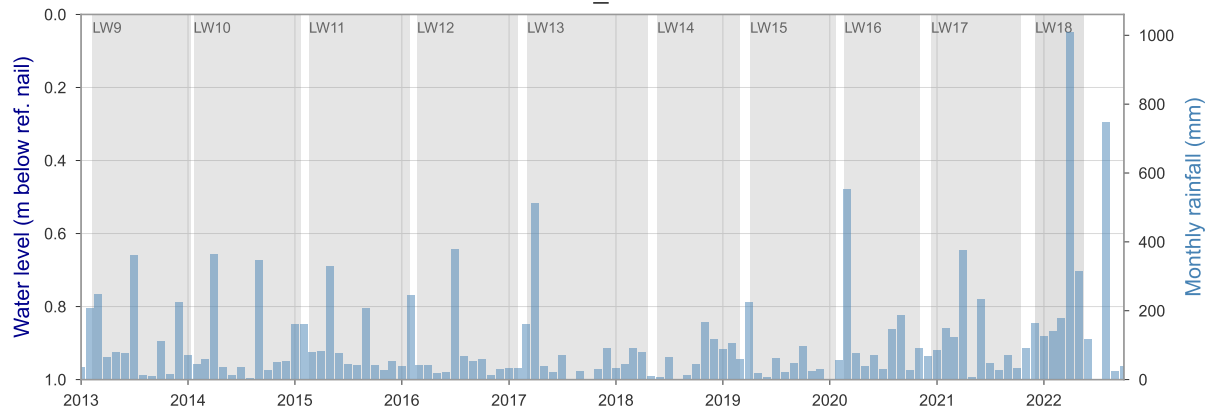
SC10C_POOL11A



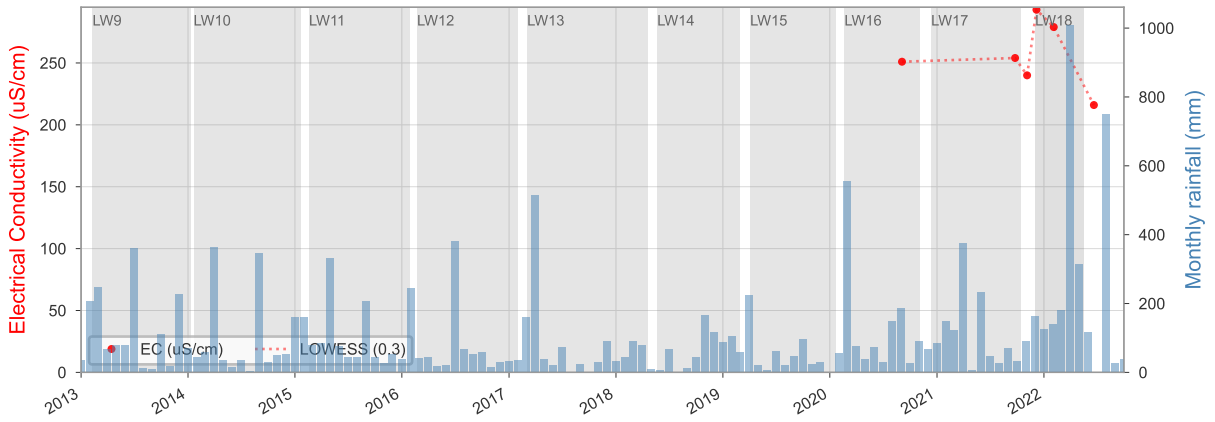
SC10C_POOL11A



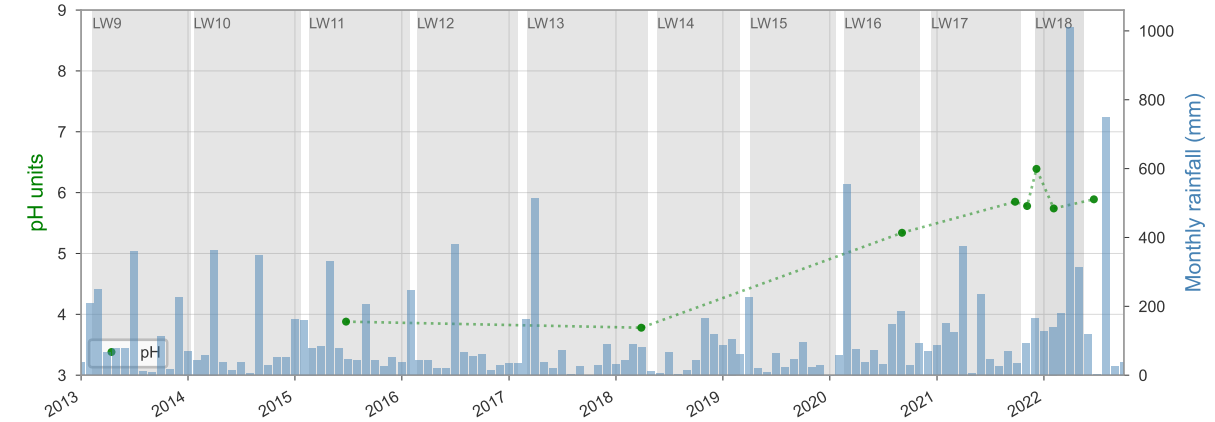
SC10C_POOL3



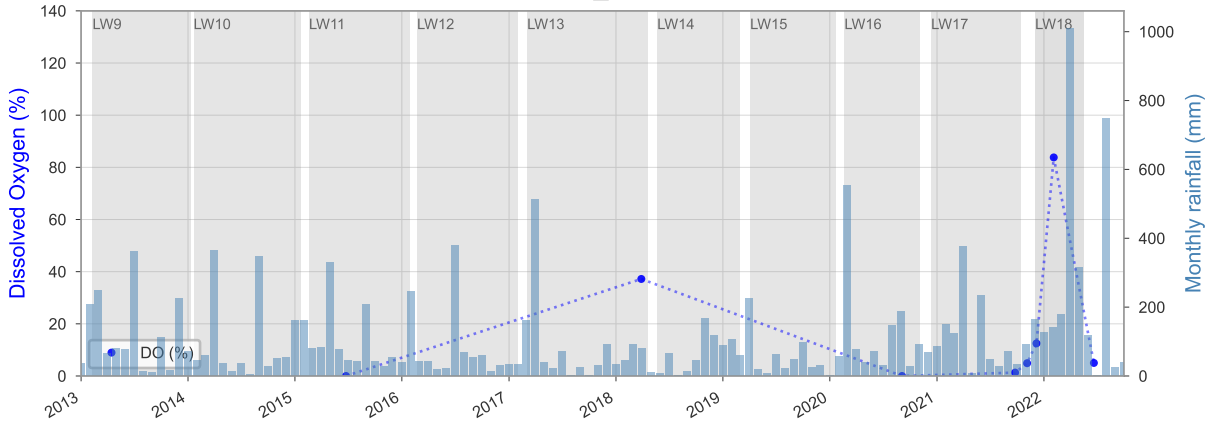
SC10C_POOL3



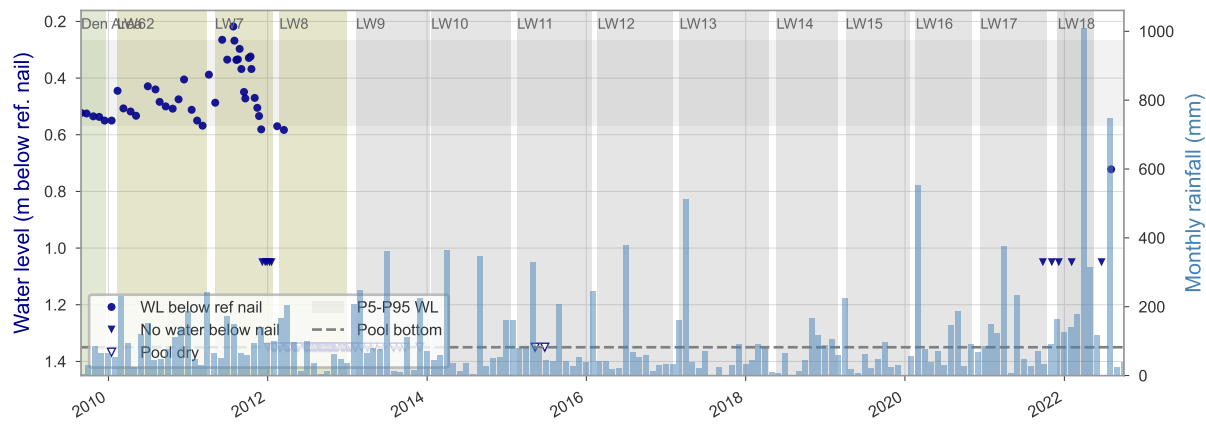
SC10C_POOL3



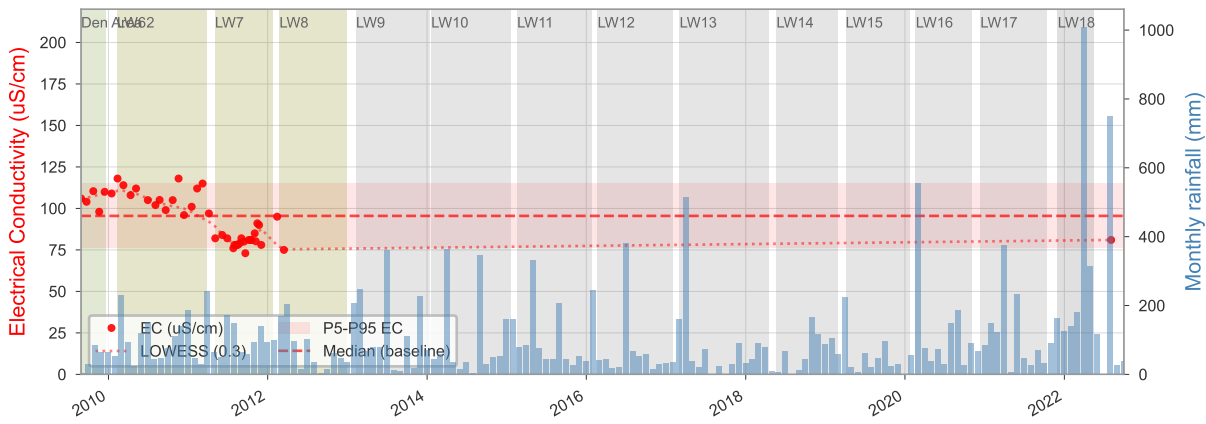
SC10C_POOL3



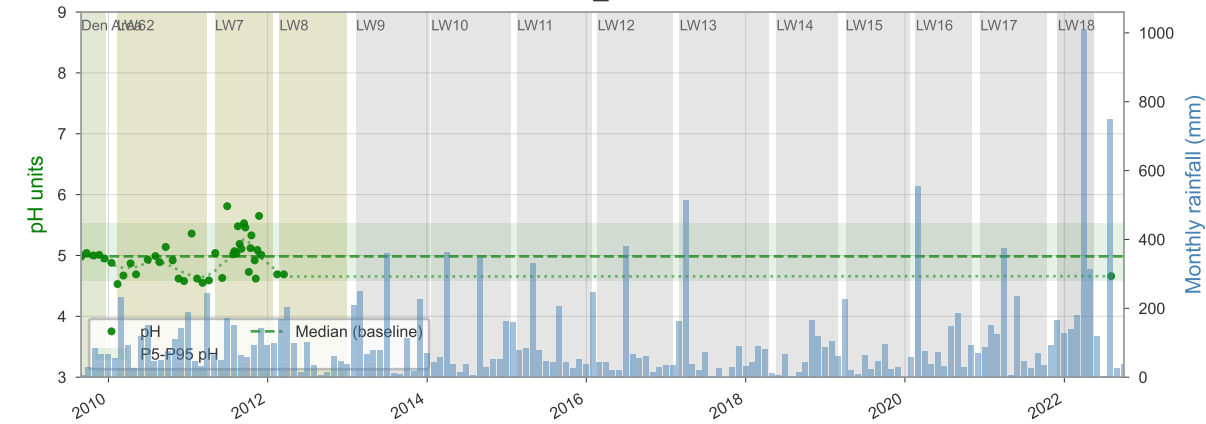
SC10C_POOL8



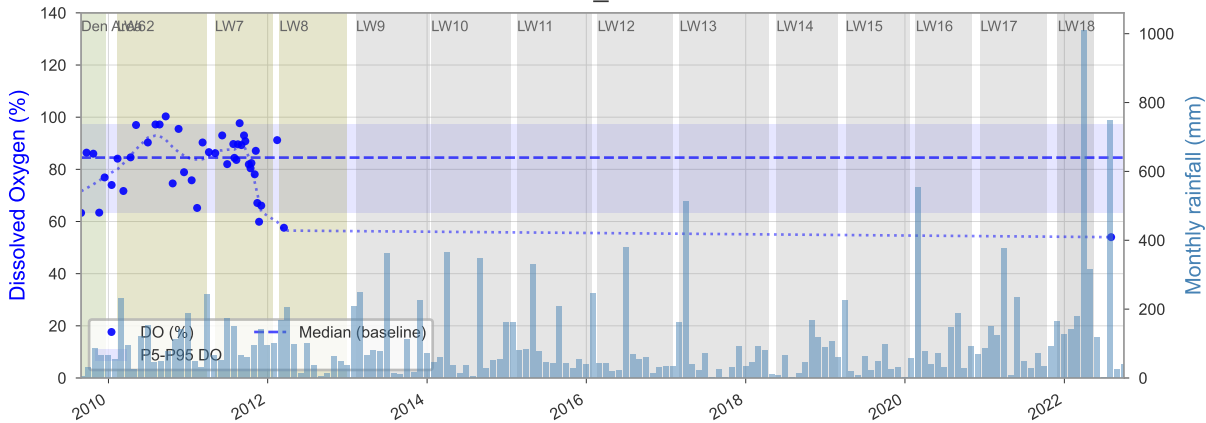
SC10C_POOL8



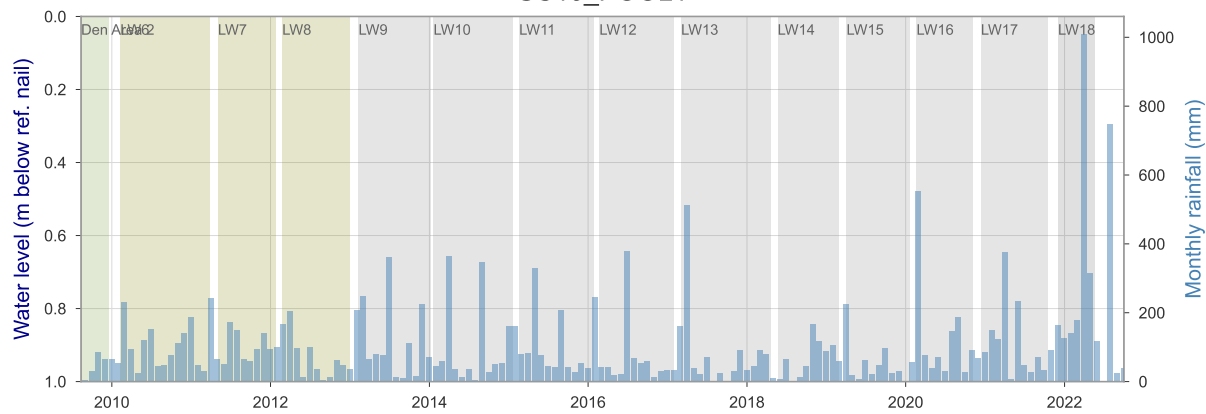
SC10C_POOL8



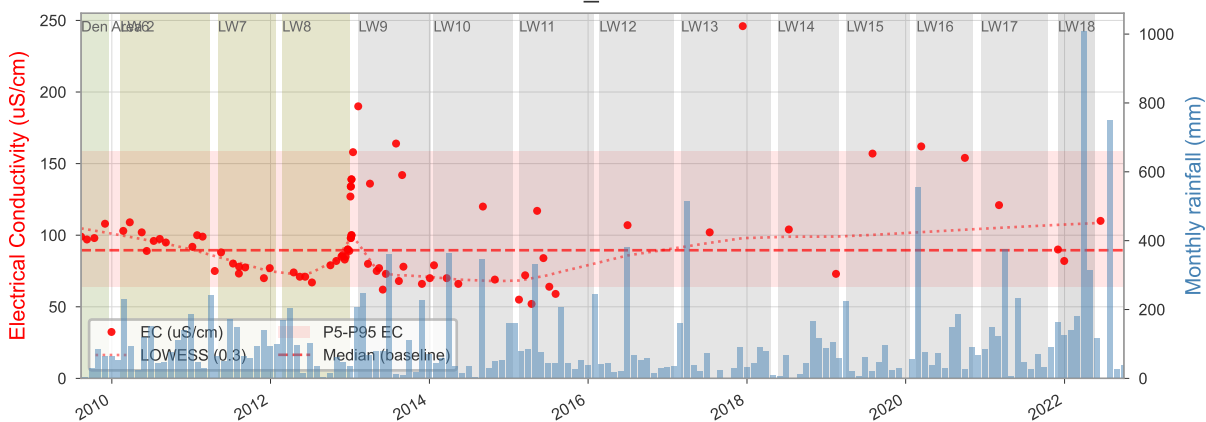
SC10C_POOL8



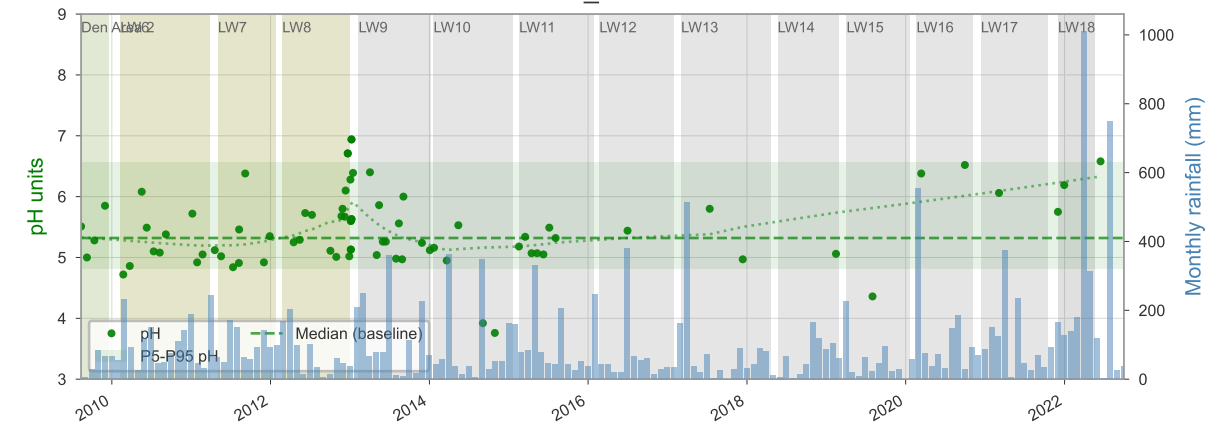
SC10_POOL1



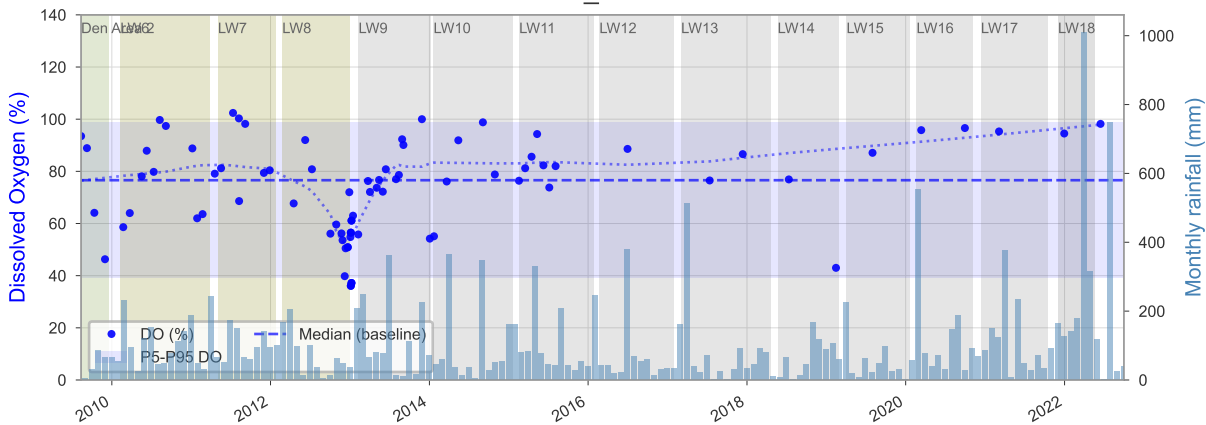
SC10_POOL1



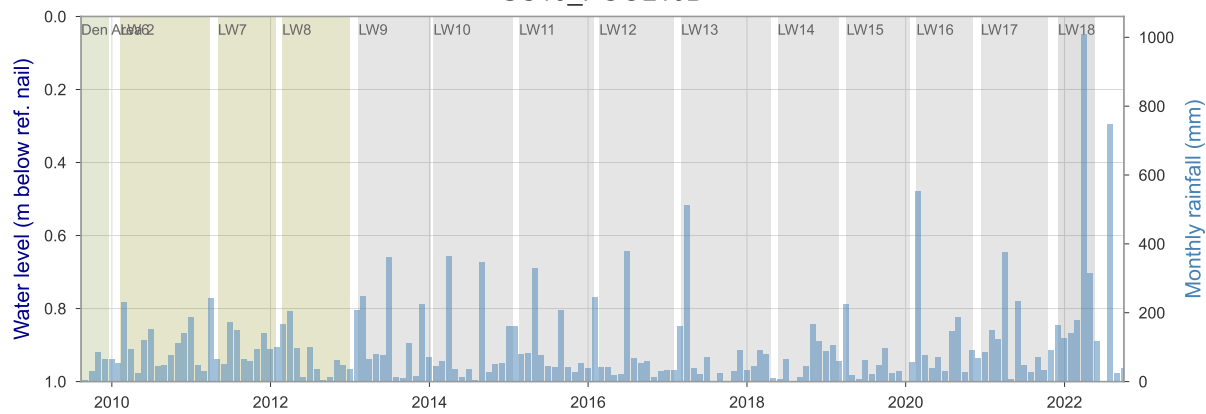
SC10_POOL1



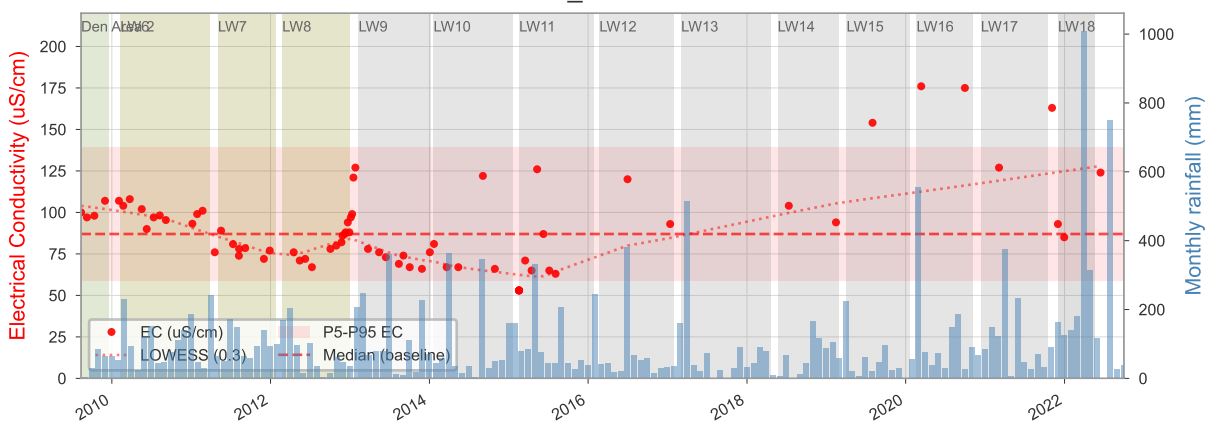
SC10_POOL1



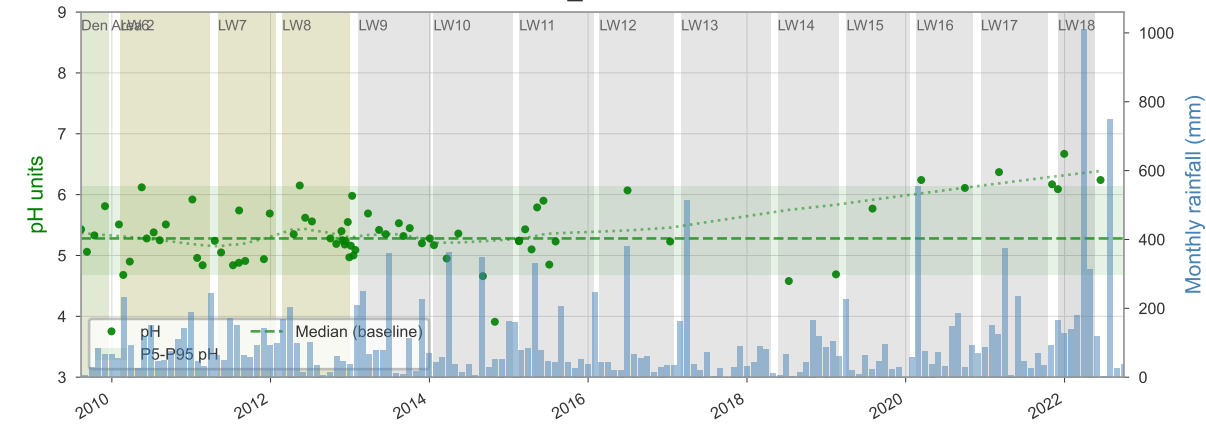
SC10_POOL10B



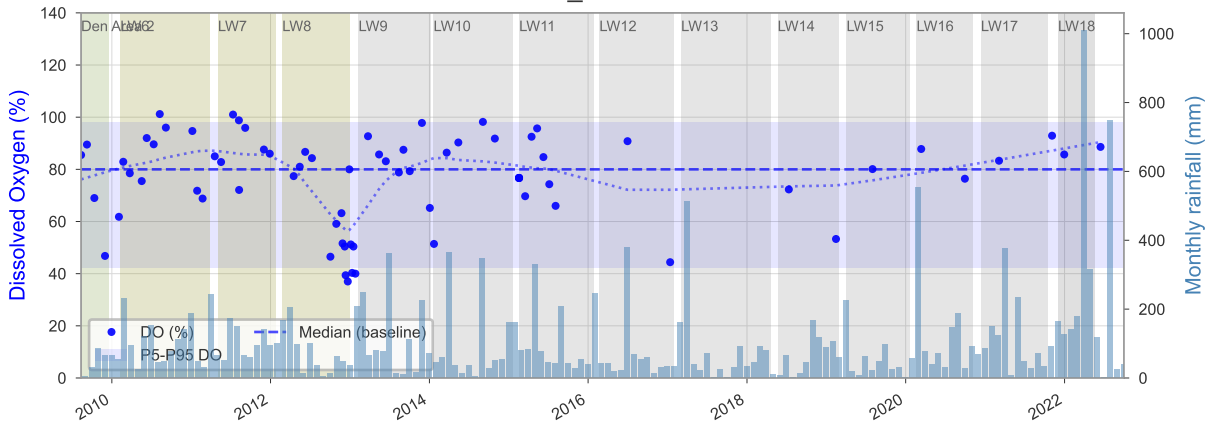
SC10_POOL10B



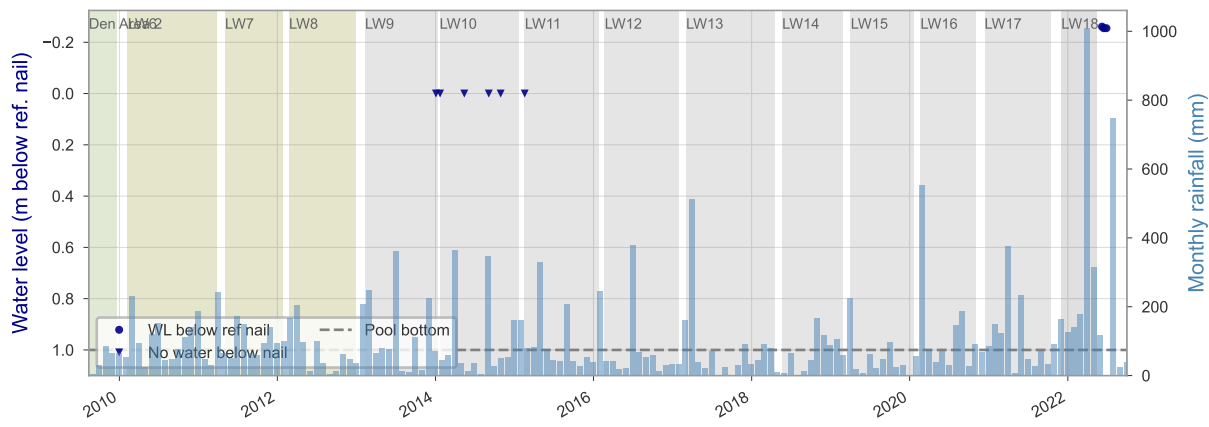
SC10_POOL10B



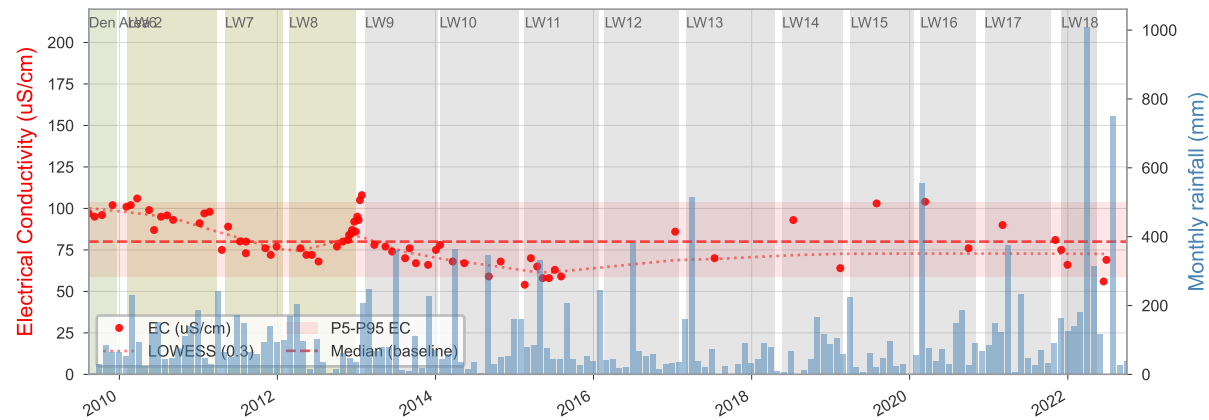
SC10_POOL10B



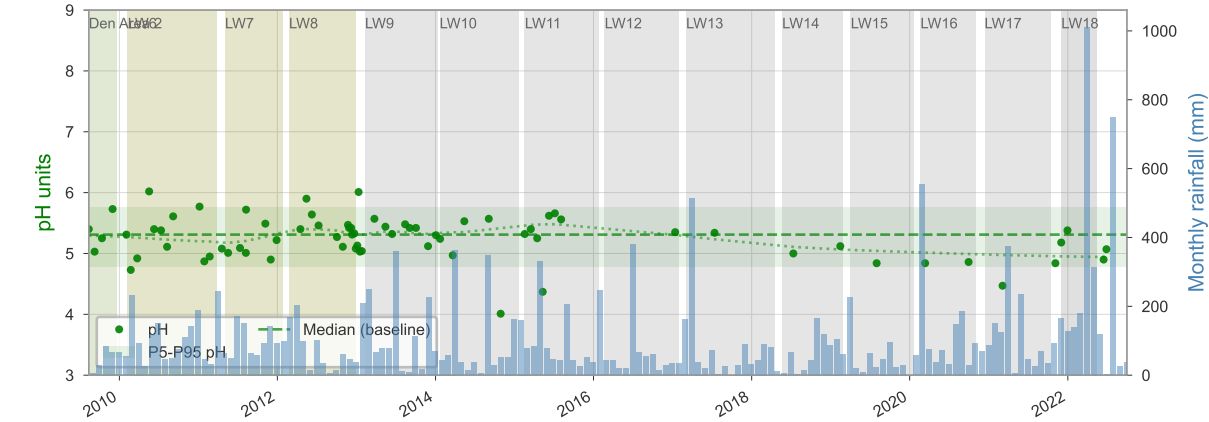
SC10_POOL11



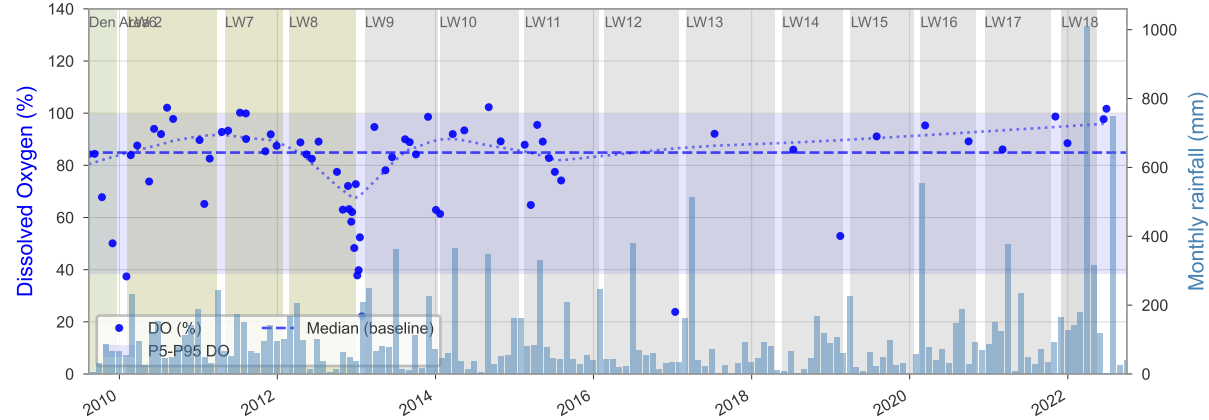
SC10_POOL11



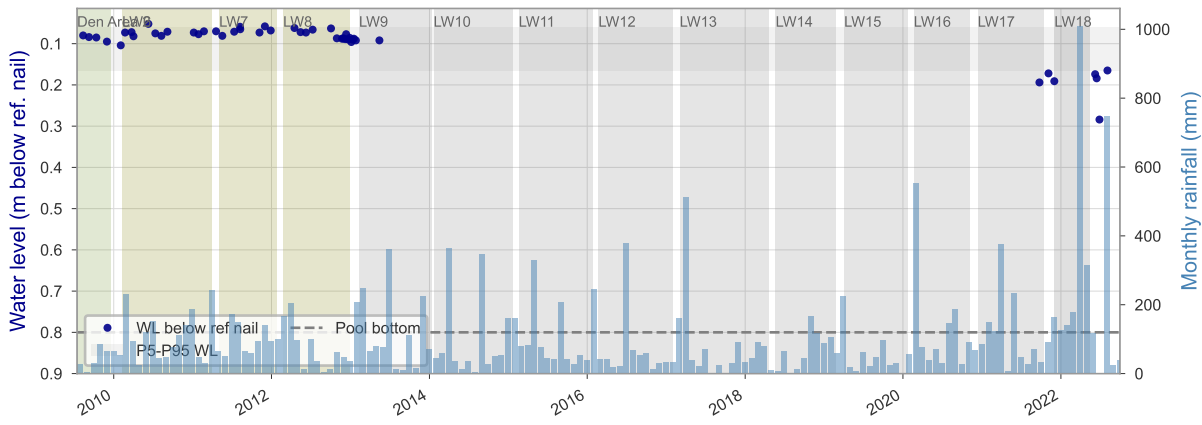
SC10_POOL11



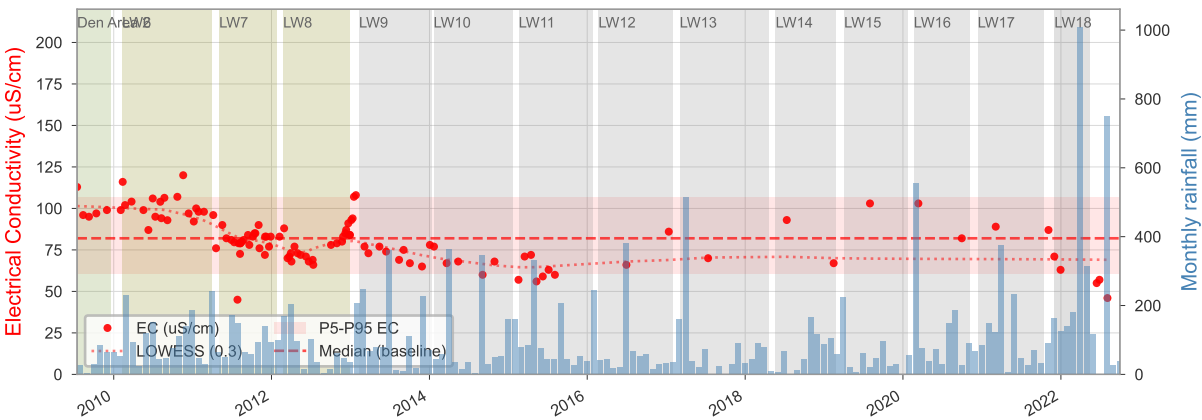
SC10_POOL11



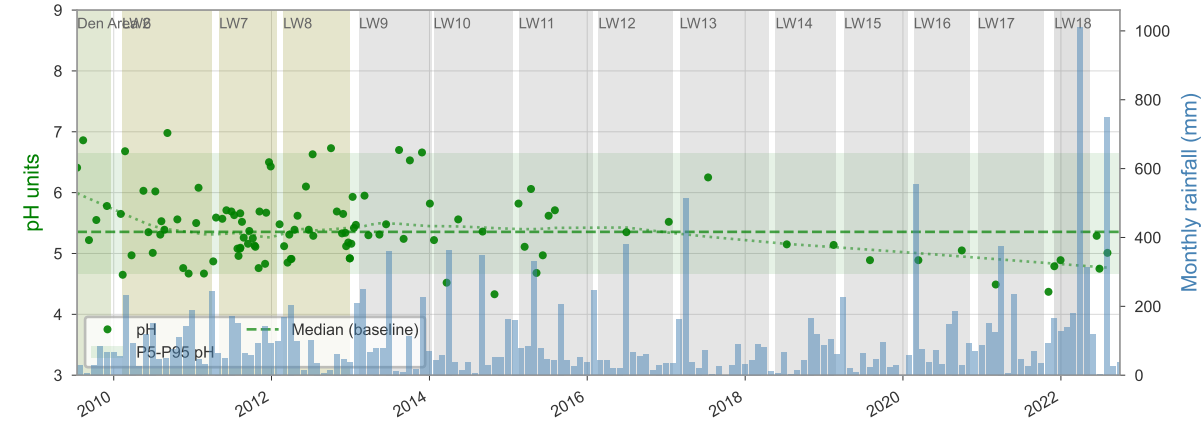
SC10_POOL14



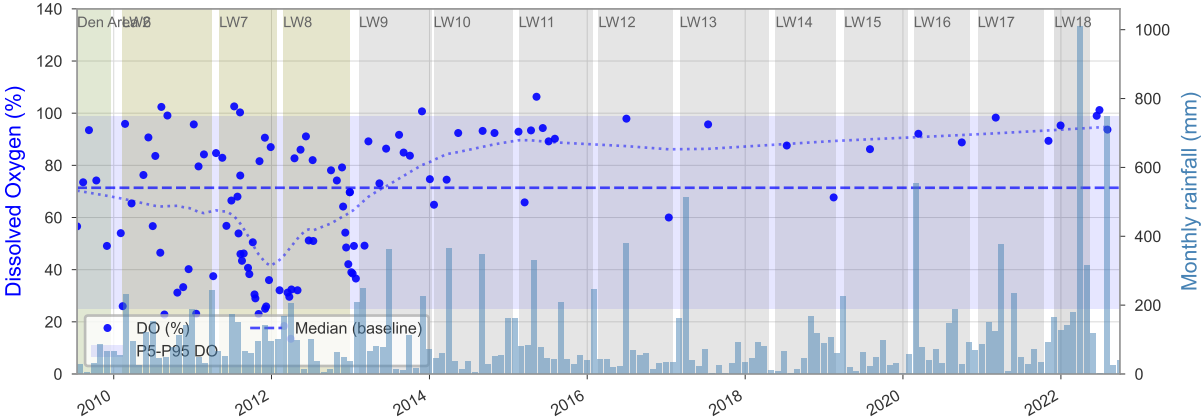
SC10_POOL14



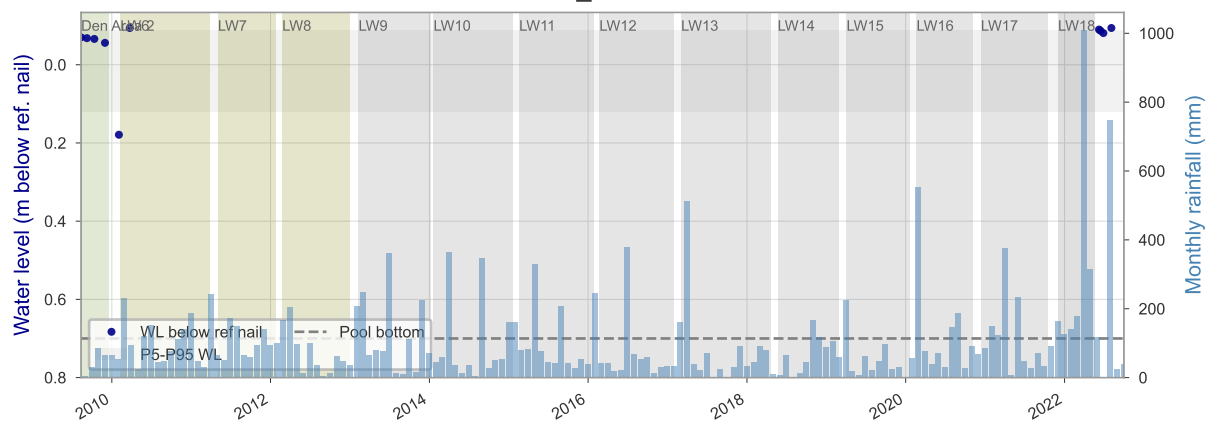
SC10_POOL14



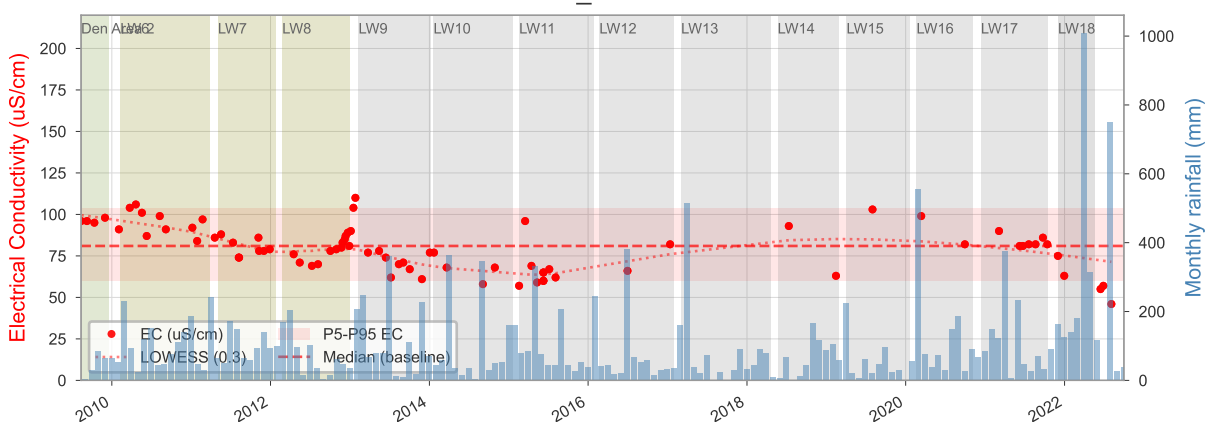
SC10_POOL14



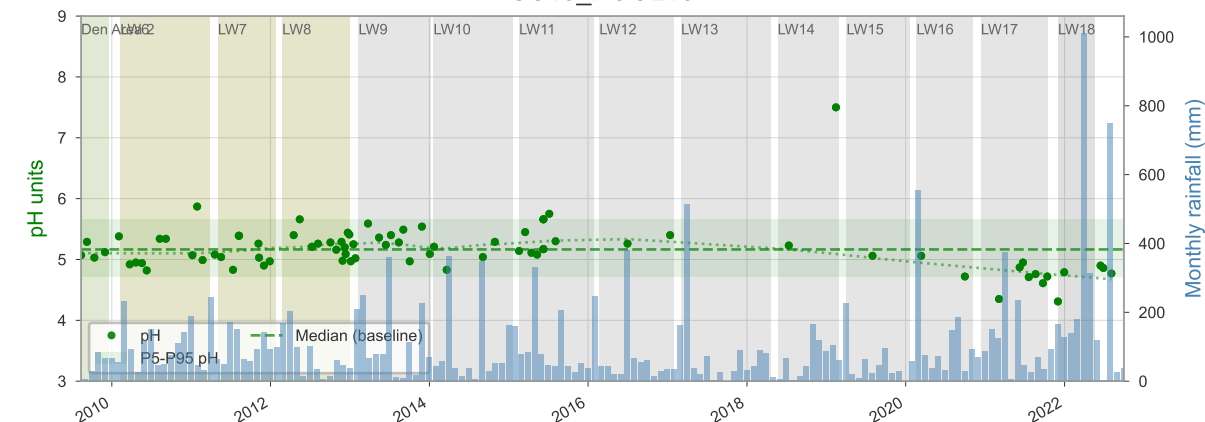
SC10_POOL15



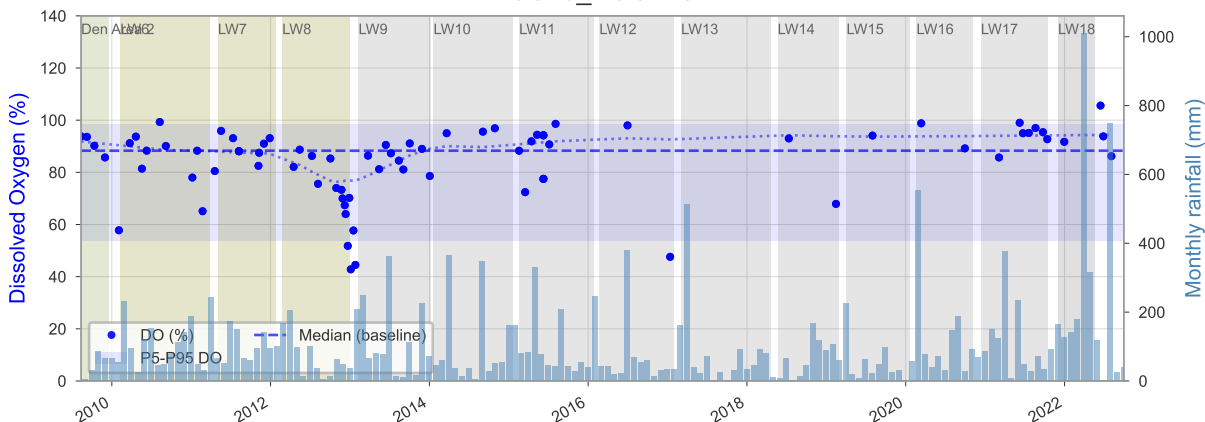
SC10_POOL15



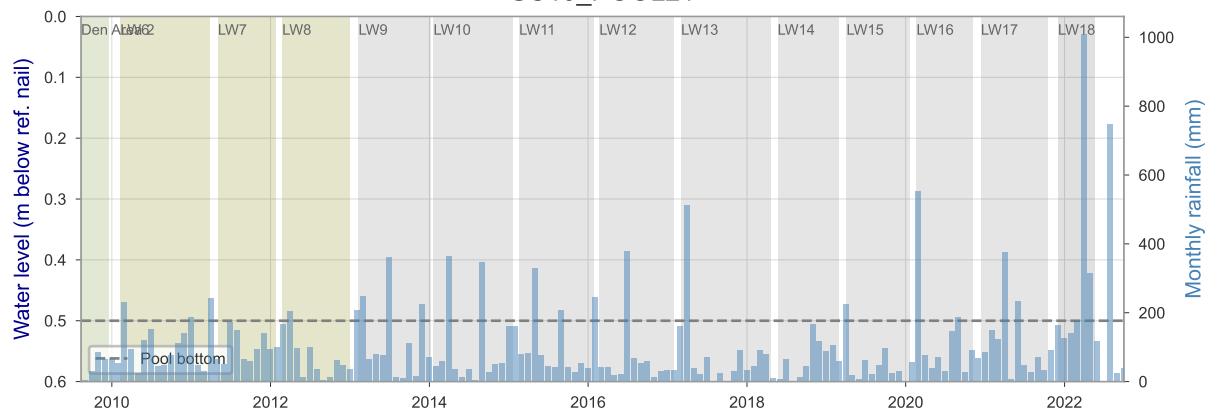
SC10_POOL15



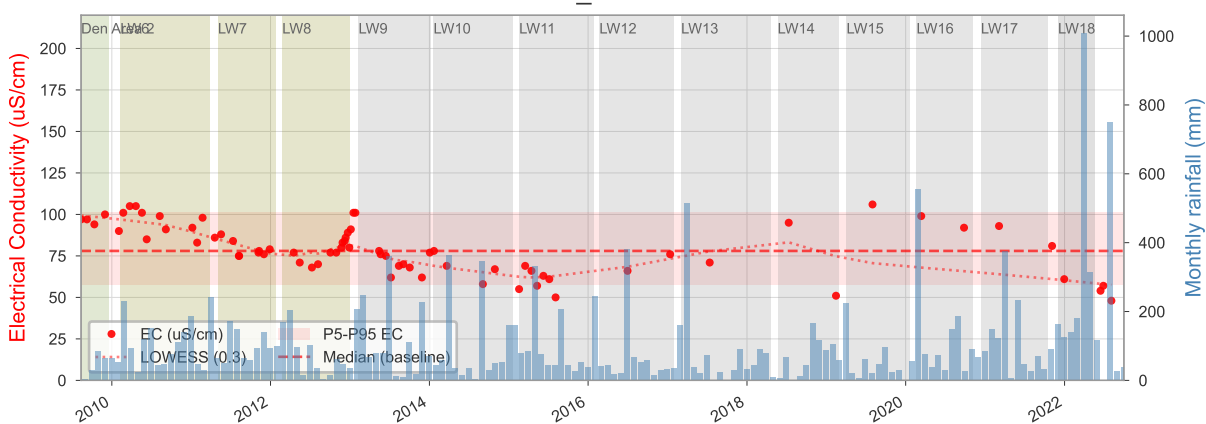
SC10_POOL15



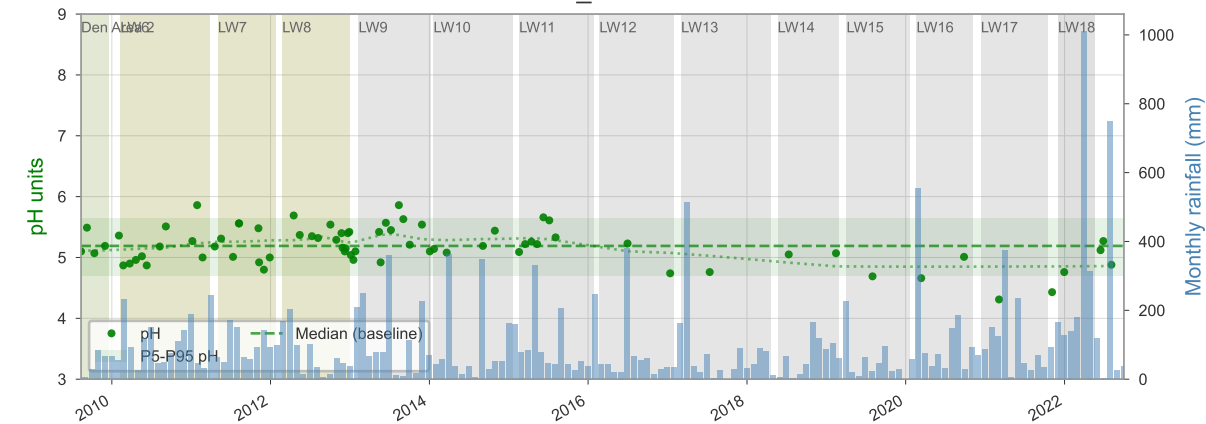
SC10_POOL21



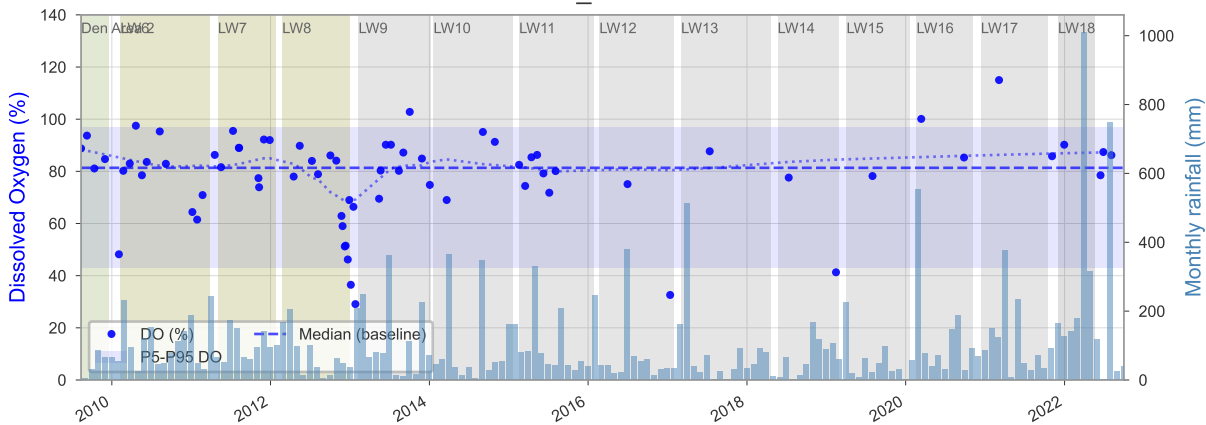
SC10_POOL21



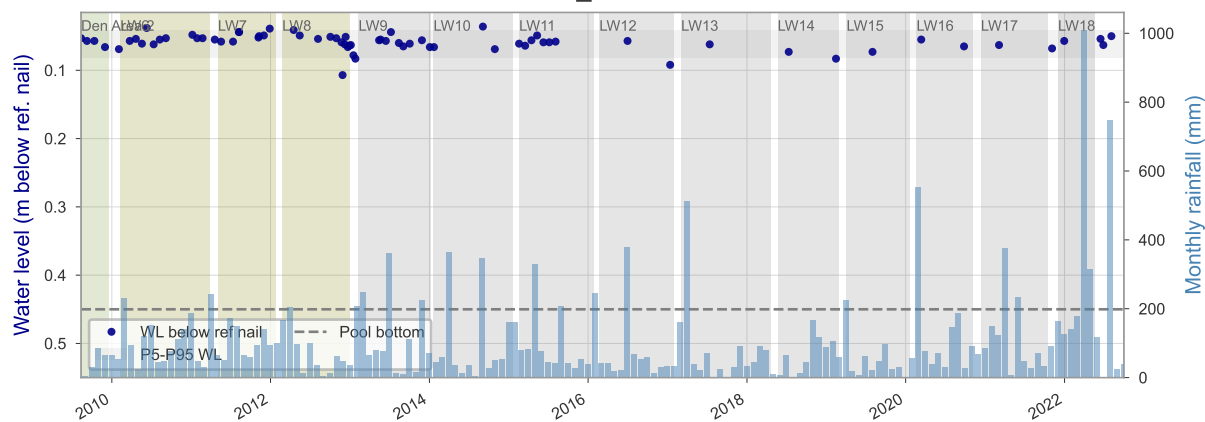
SC10_POOL21



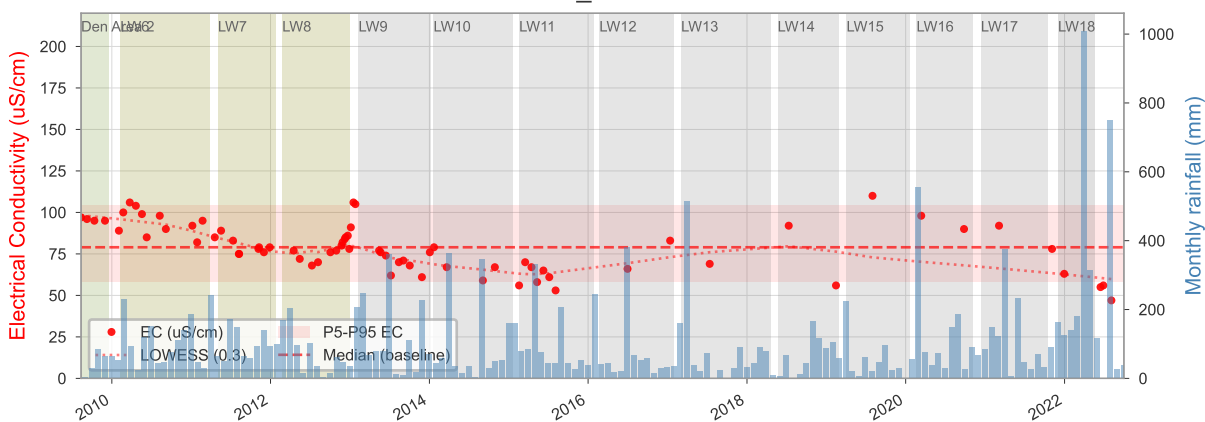
SC10_POOL21



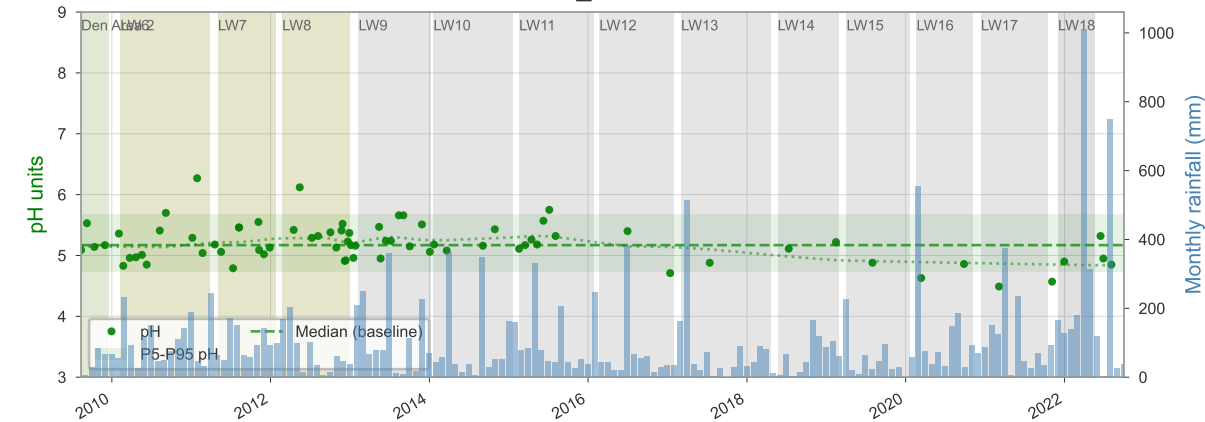
SC10_POOL23



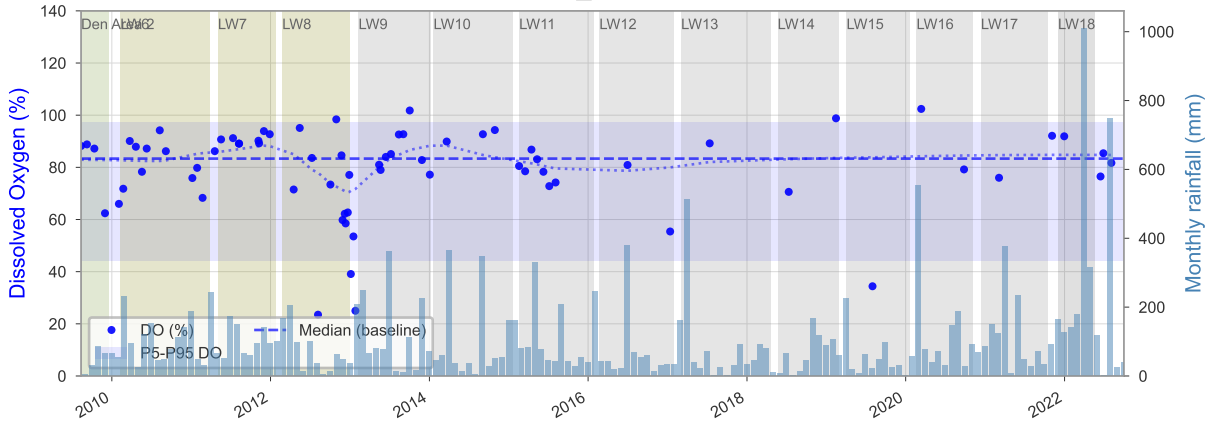
SC10_POOL23



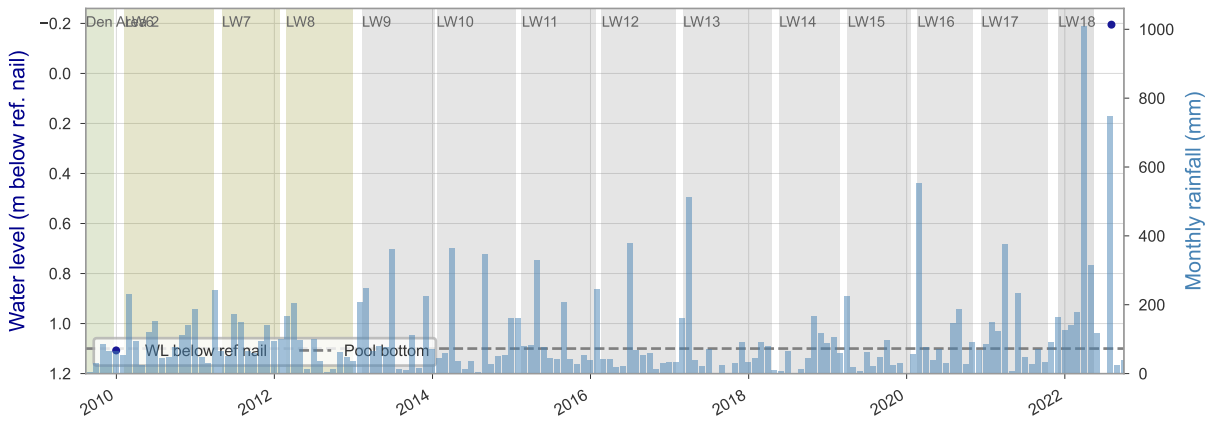
SC10_POOL23



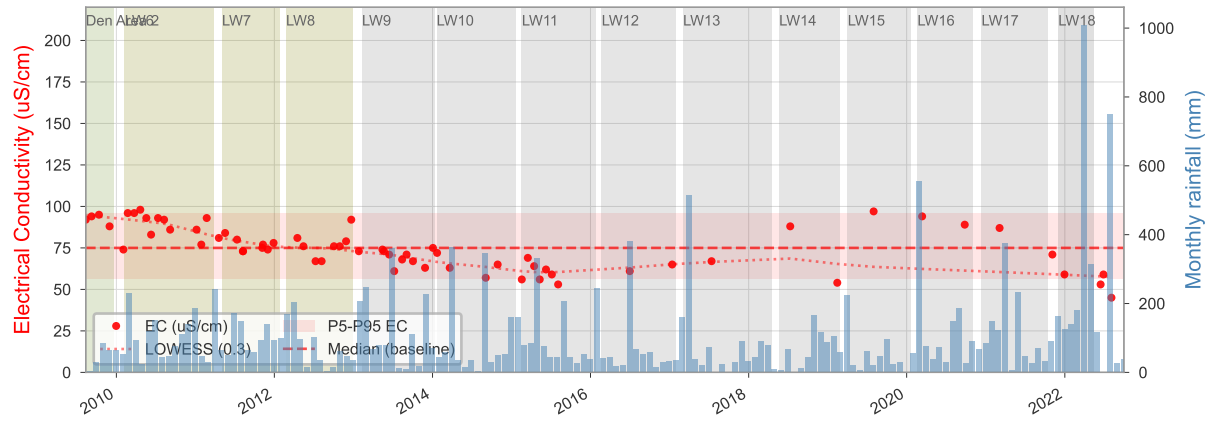
SC10_POOL23



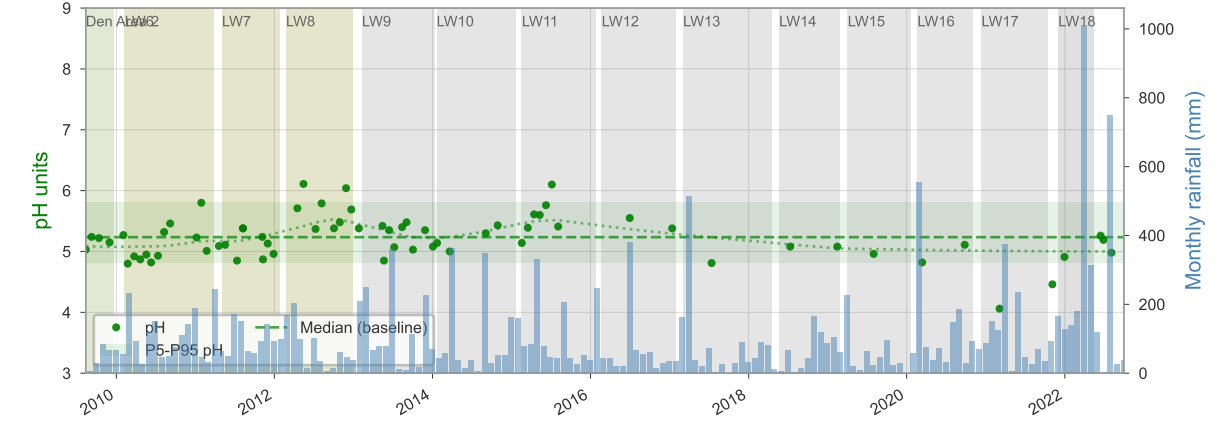
SC10_POOL26A



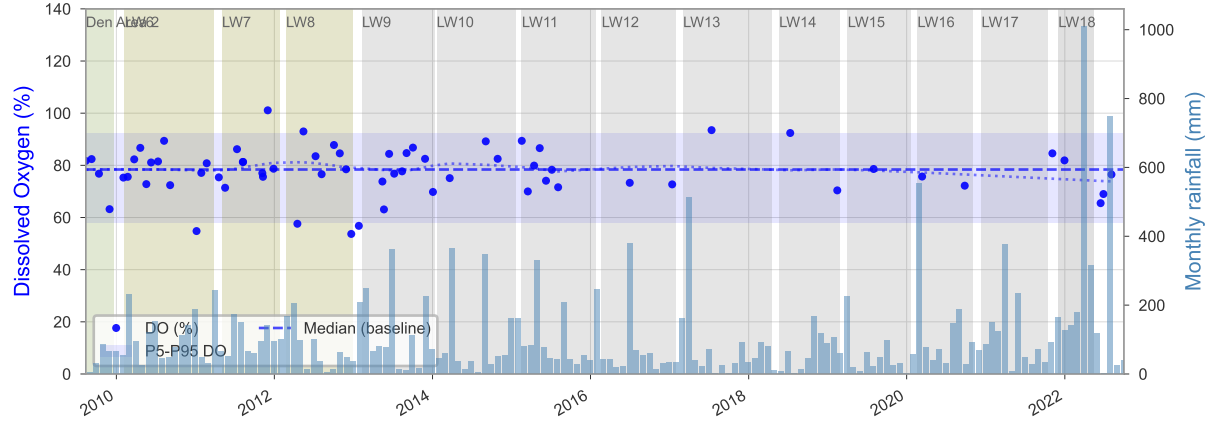
SC10_POOL26A



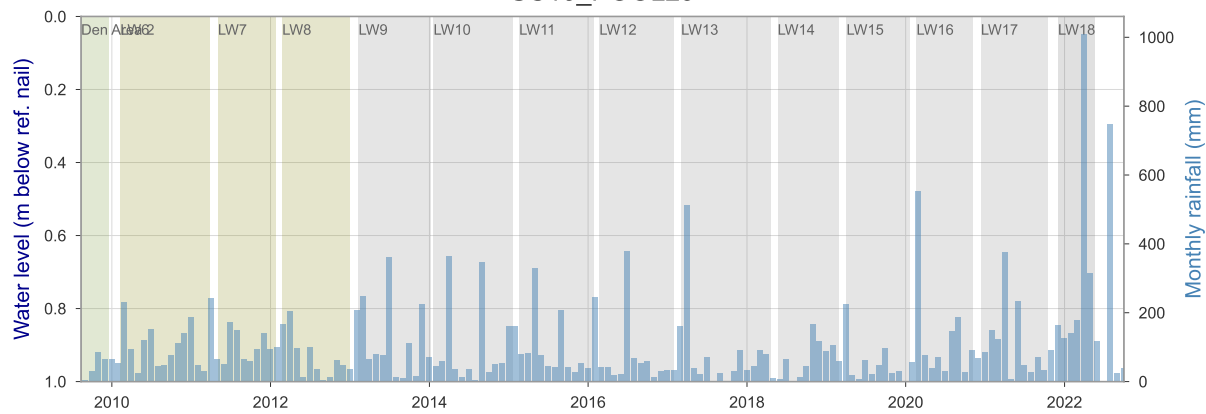
SC10_POOL26A



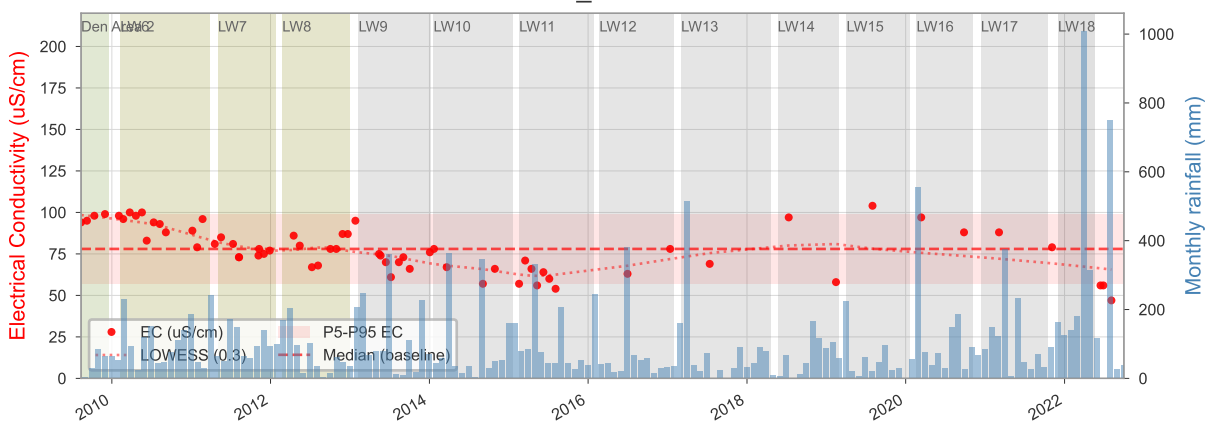
SC10_POOL26A



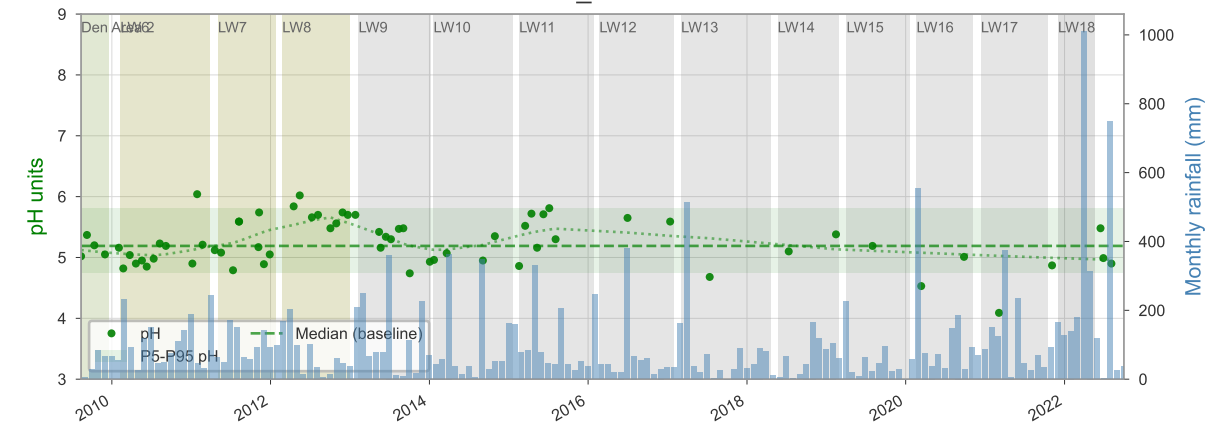
SC10_POOL29



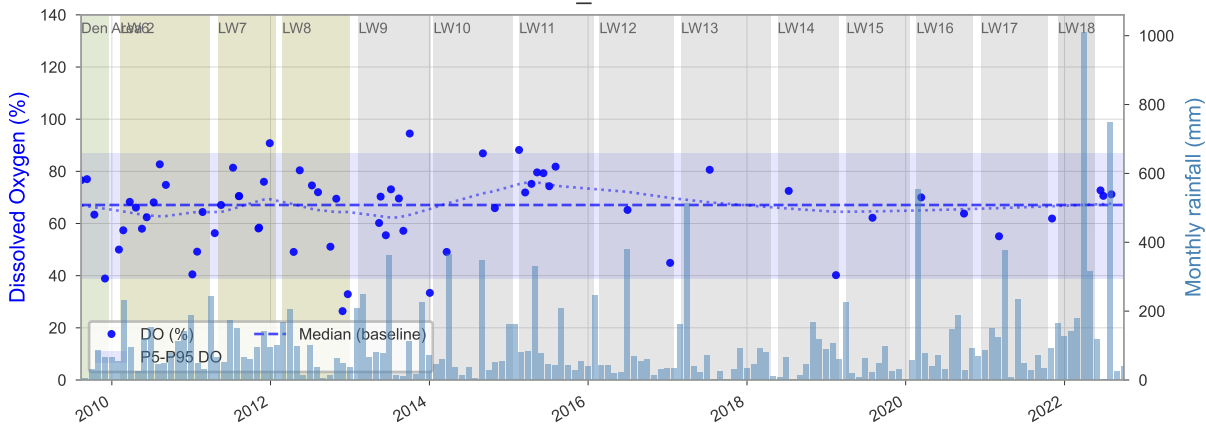
SC10_POOL29



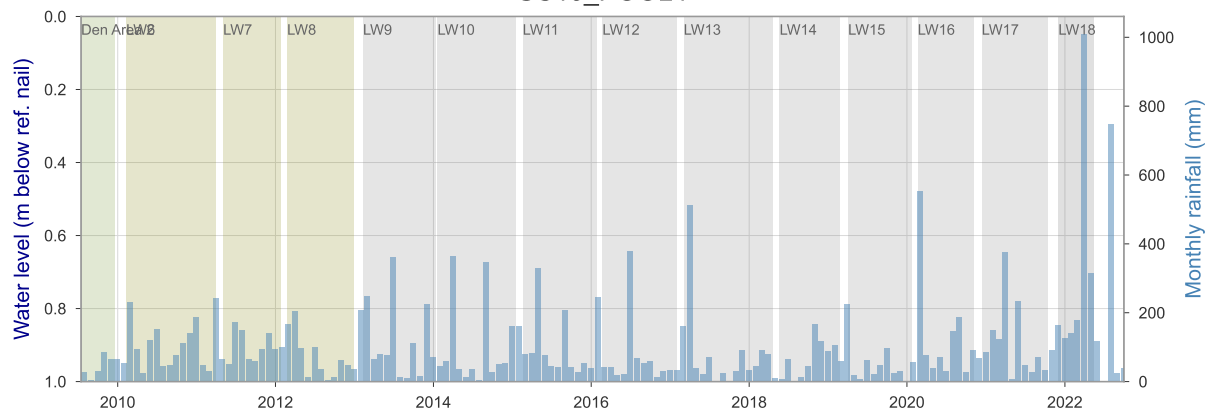
SC10_POOL29



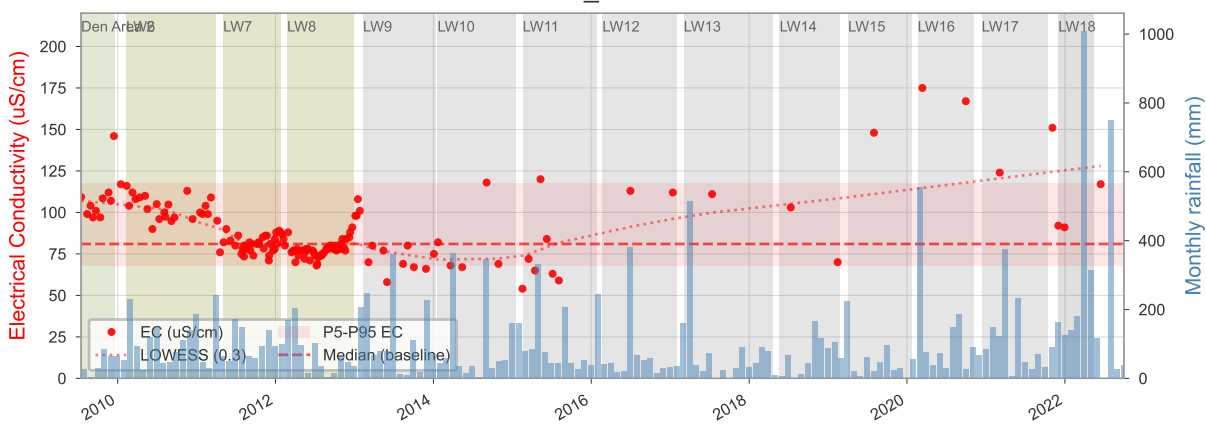
SC10_POOL29



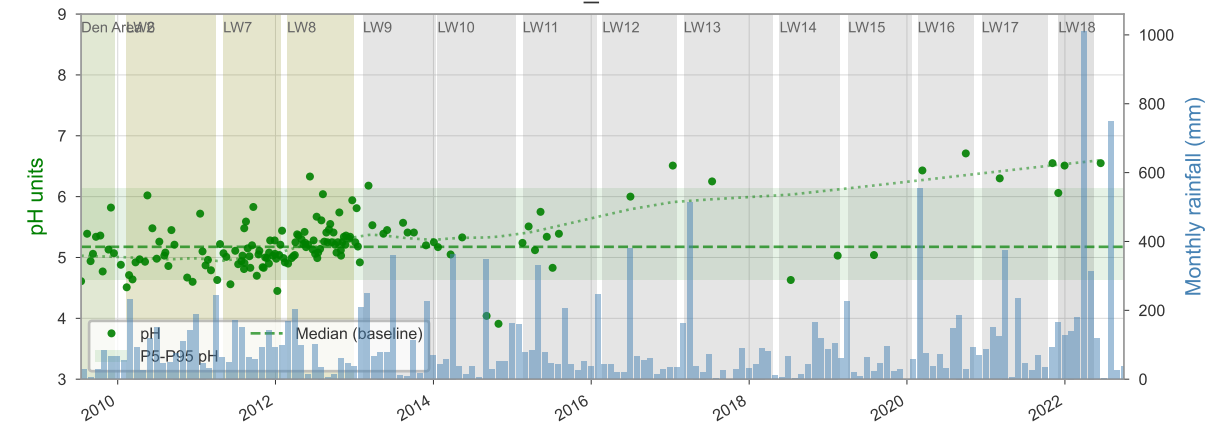
SC10_POOL4



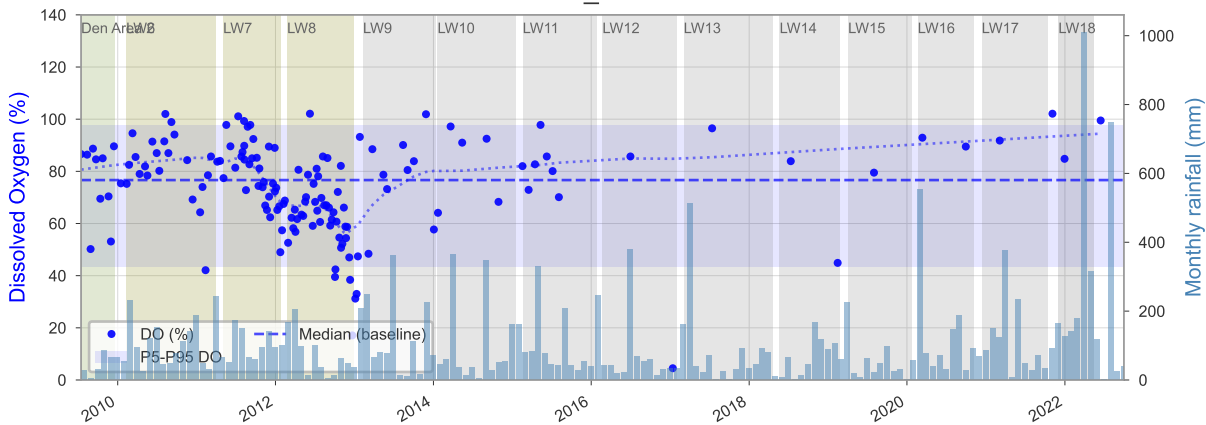
SC10_POOL4



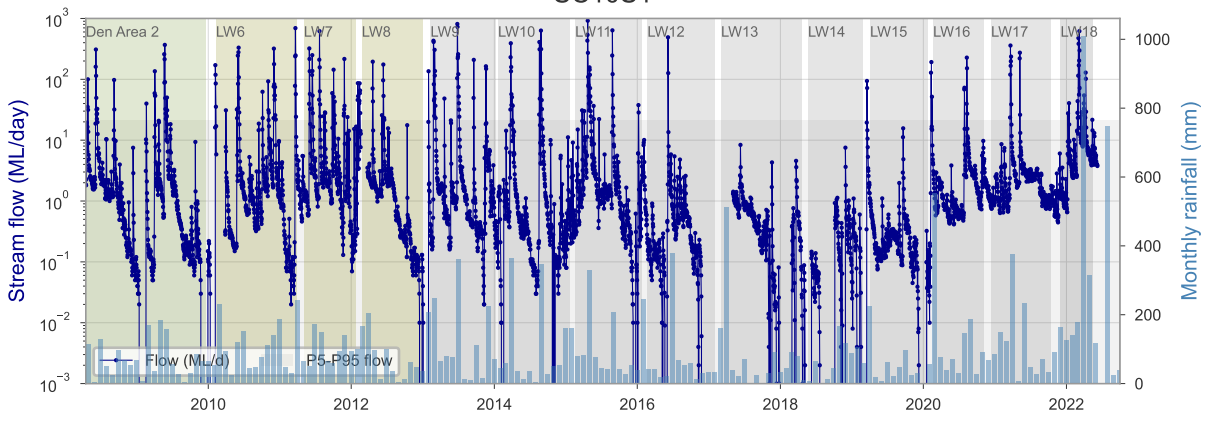
SC10_POOL4



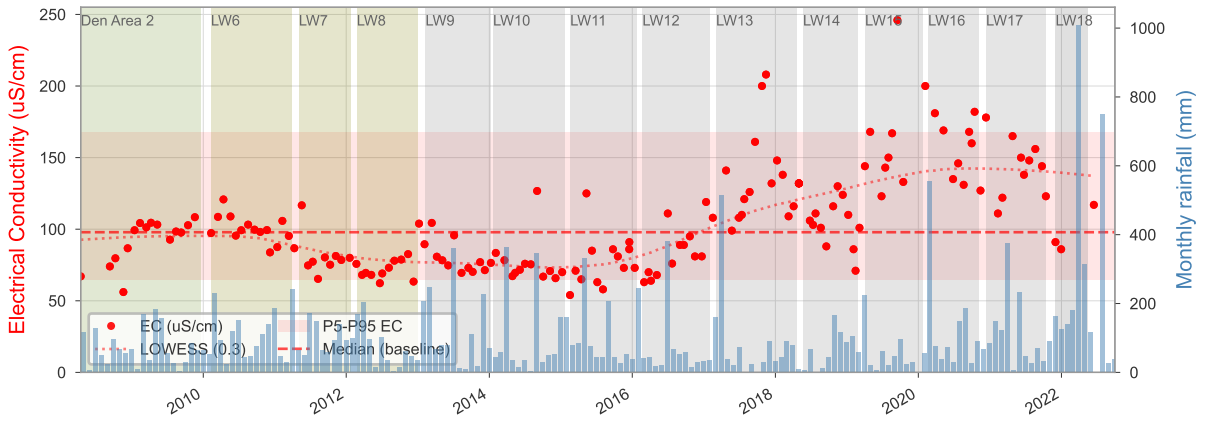
SC10_POOL4



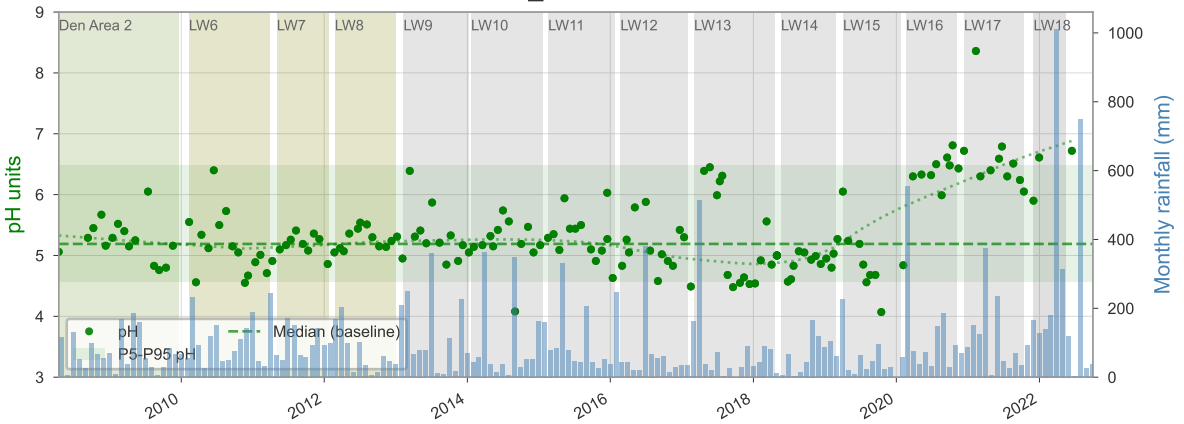
SC10S1



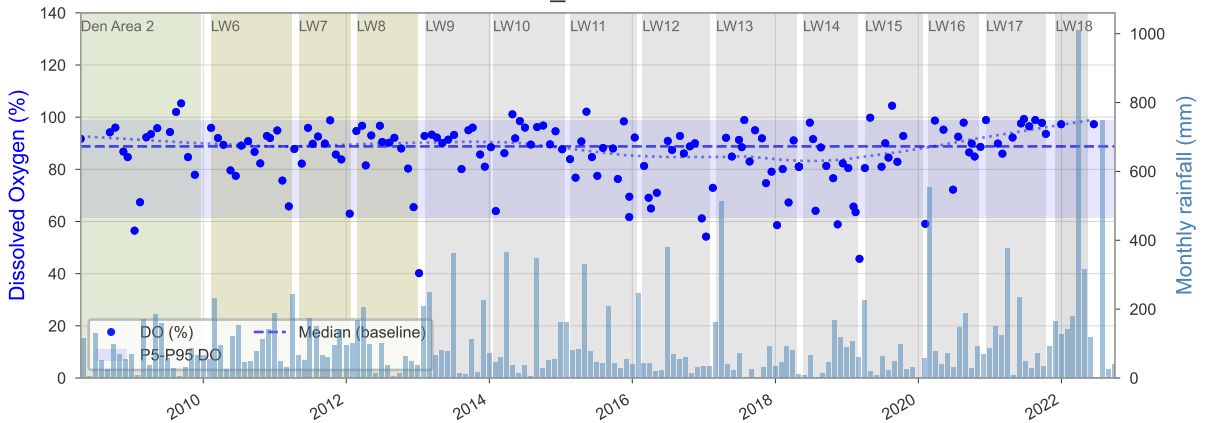
SC10_ROCKBAR3



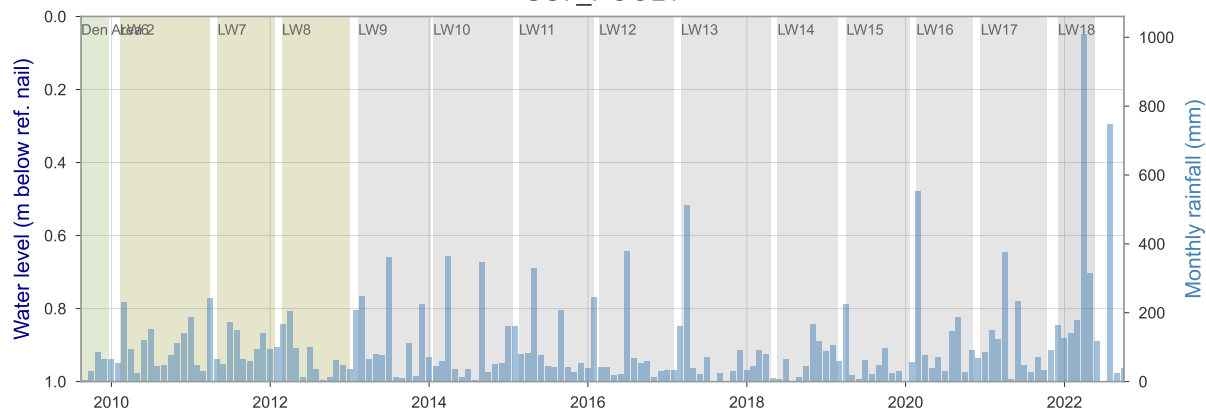
SC10_ROCKBAR3



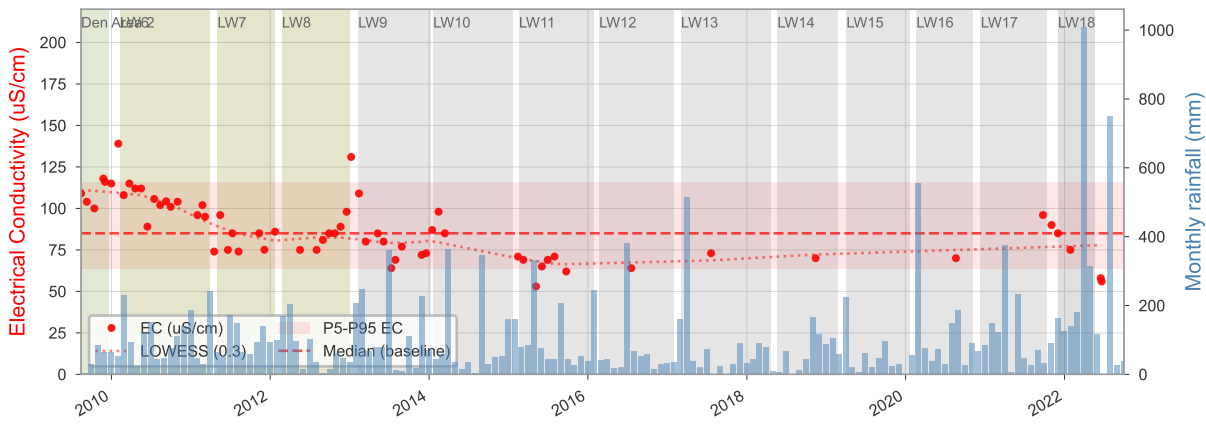
SC10_ROCKBAR3



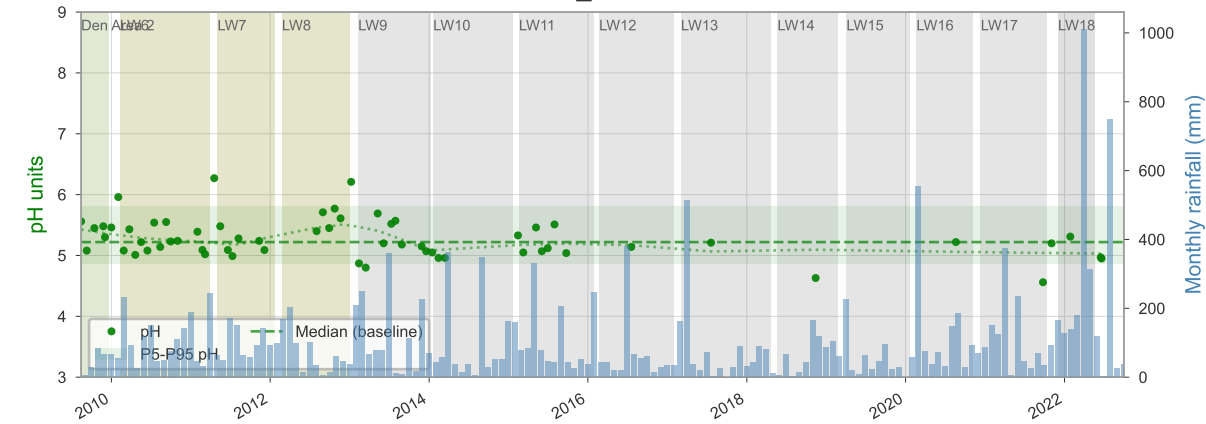
SC7_POOL1



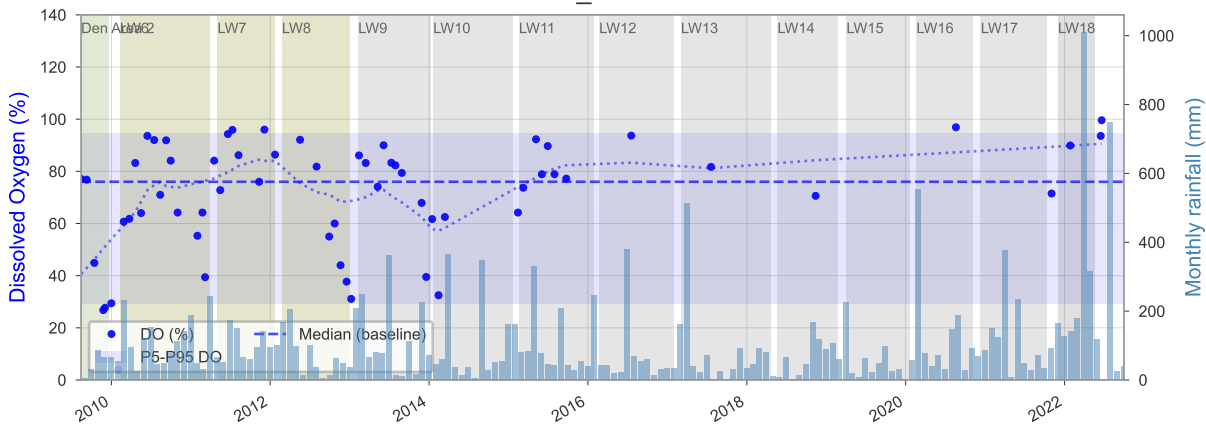
SC7_POOL1



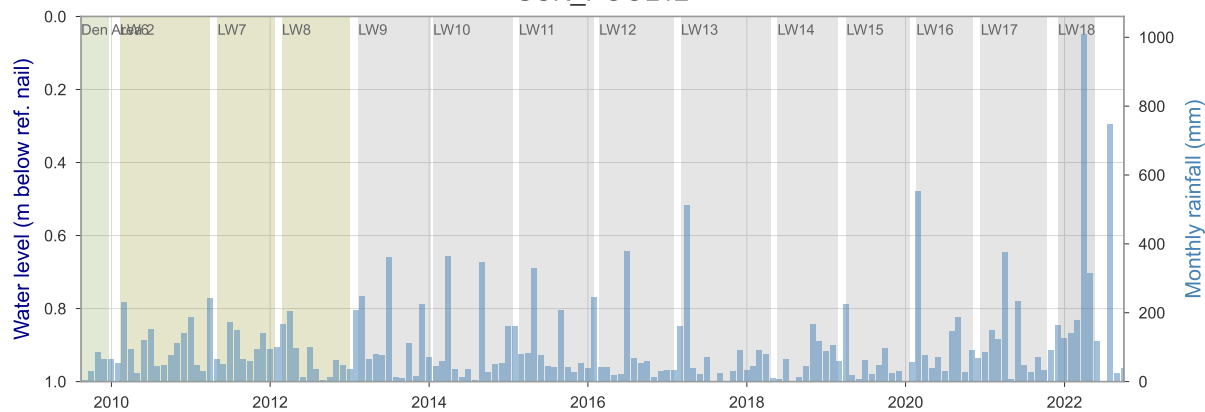
SC7_POOL1



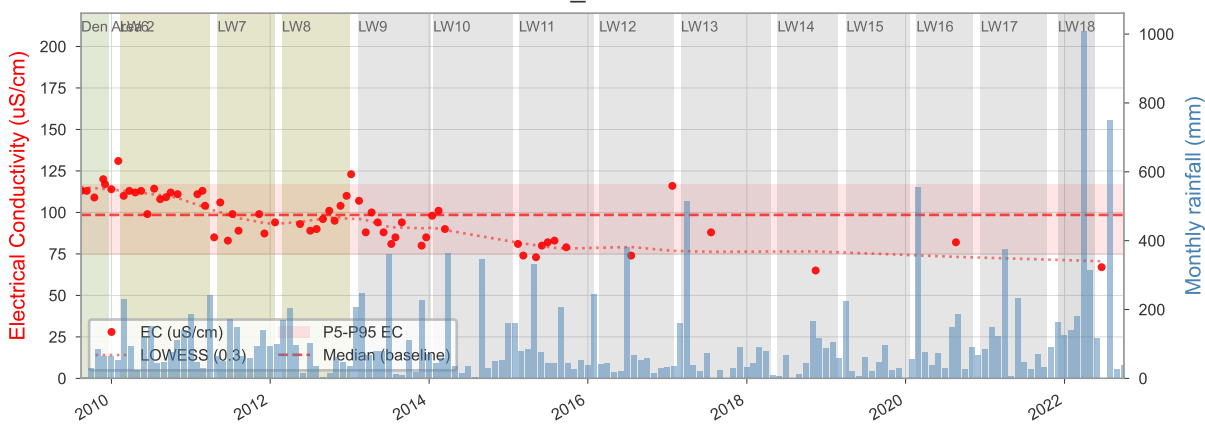
SC7_POOL1



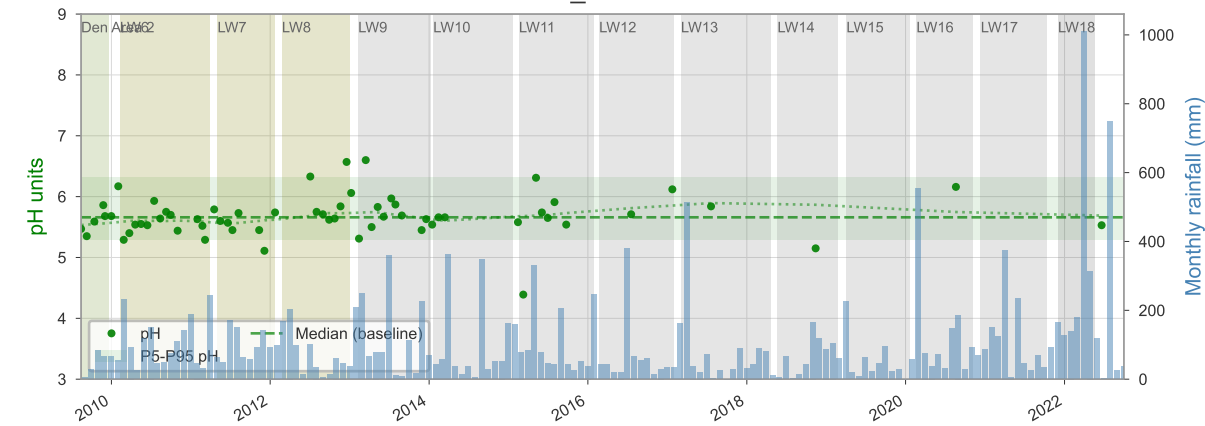
SCK_POOL12



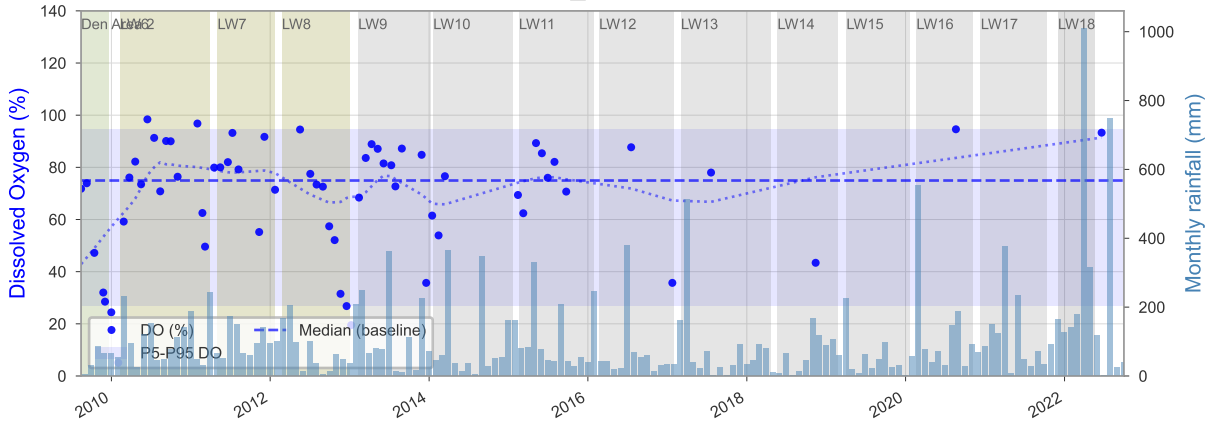
SCK_POOL12



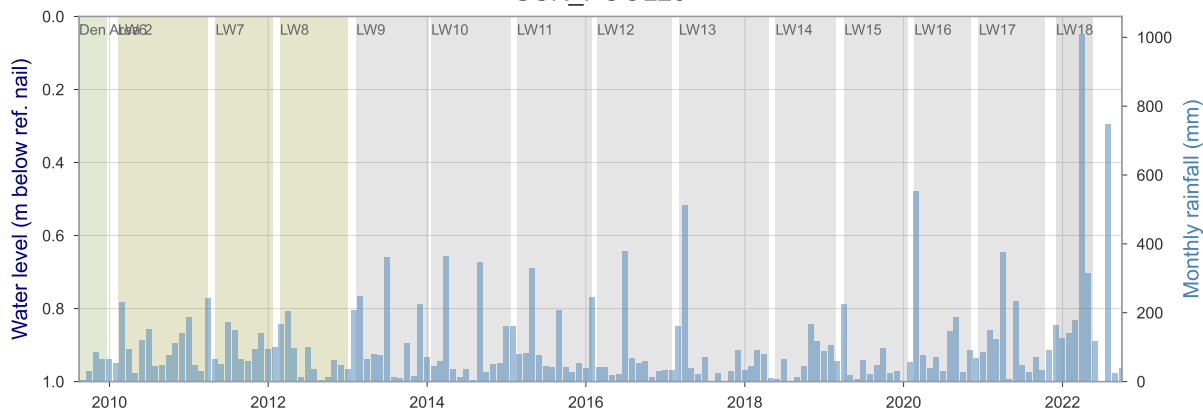
SCK_POOL12



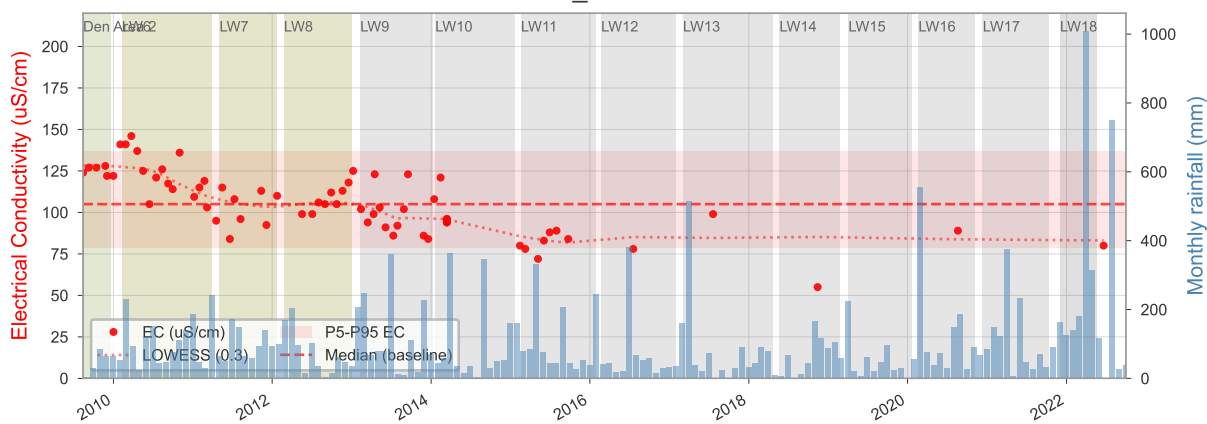
SCK_POOL12



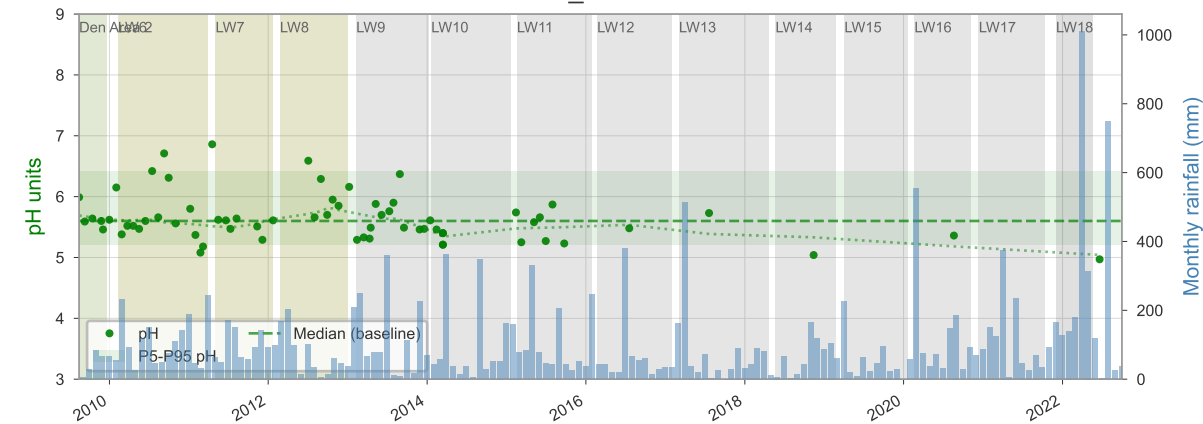
SCK_POOL23



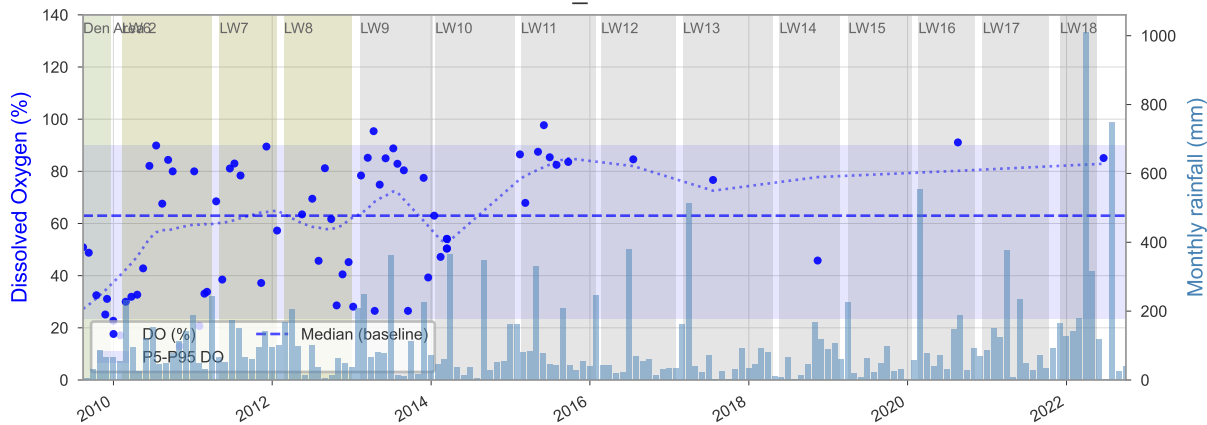
SCK_POOL23



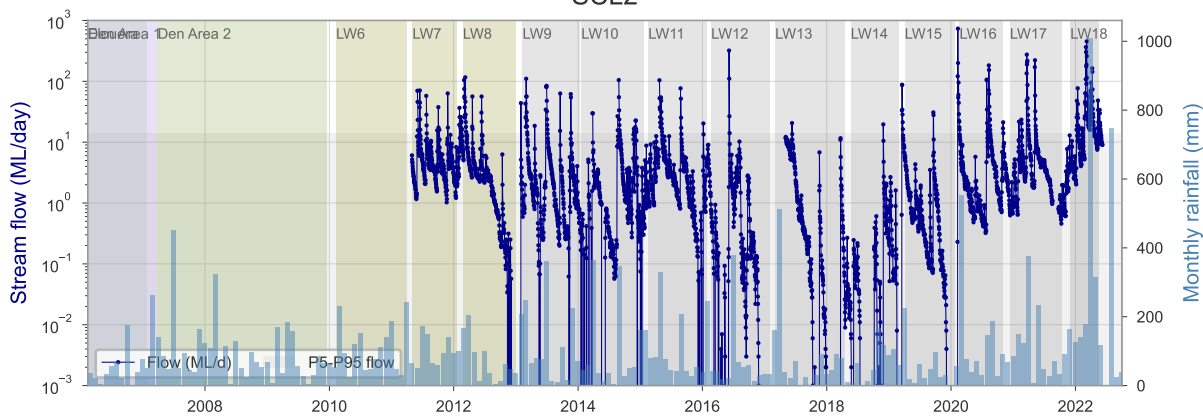
SCK_POOL23



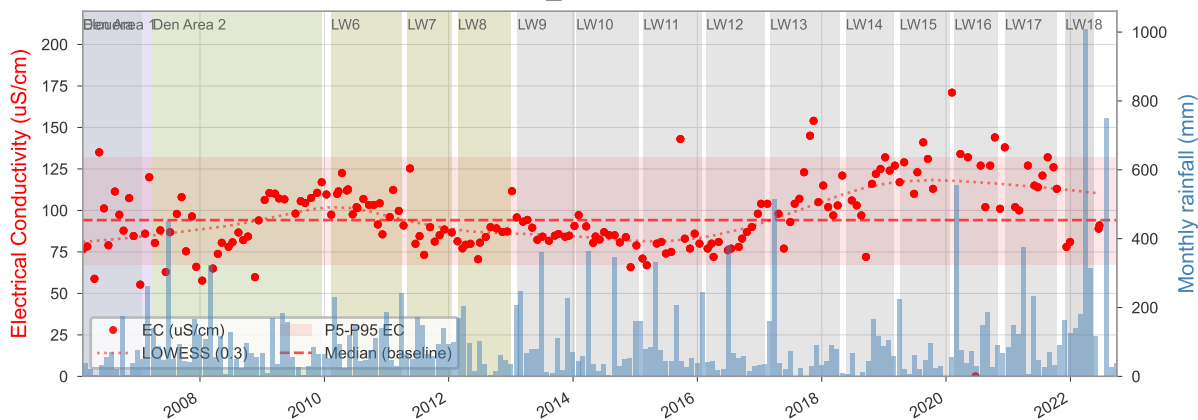
SCK_POOL23



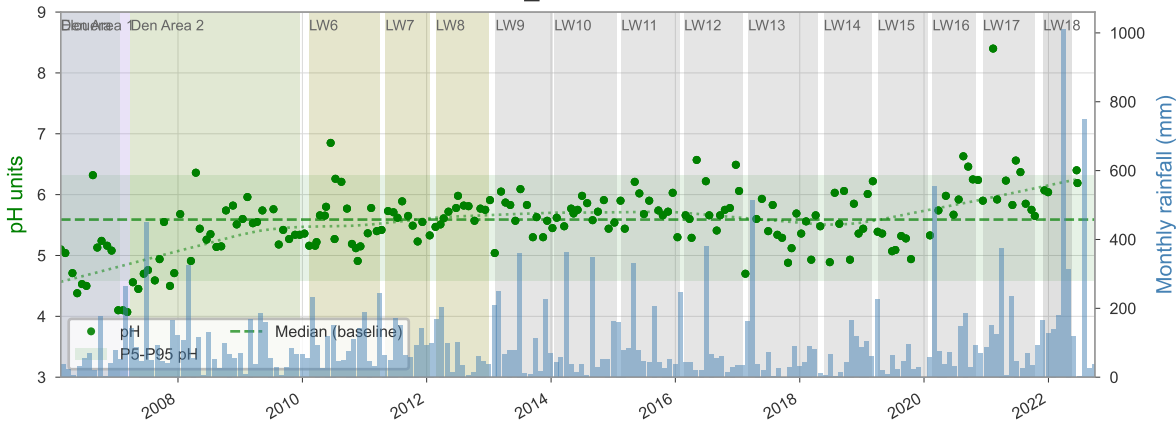
SCL2



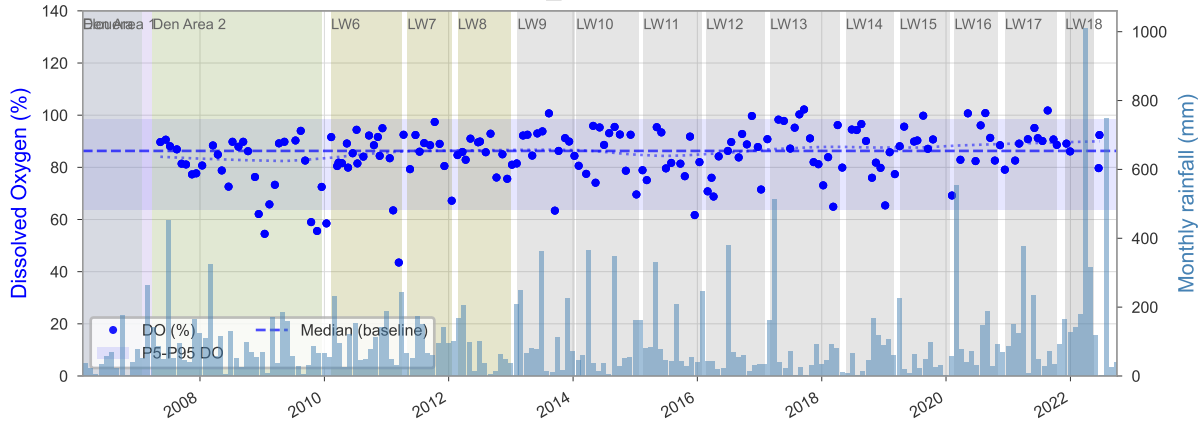
SCK_ROCKBAR5



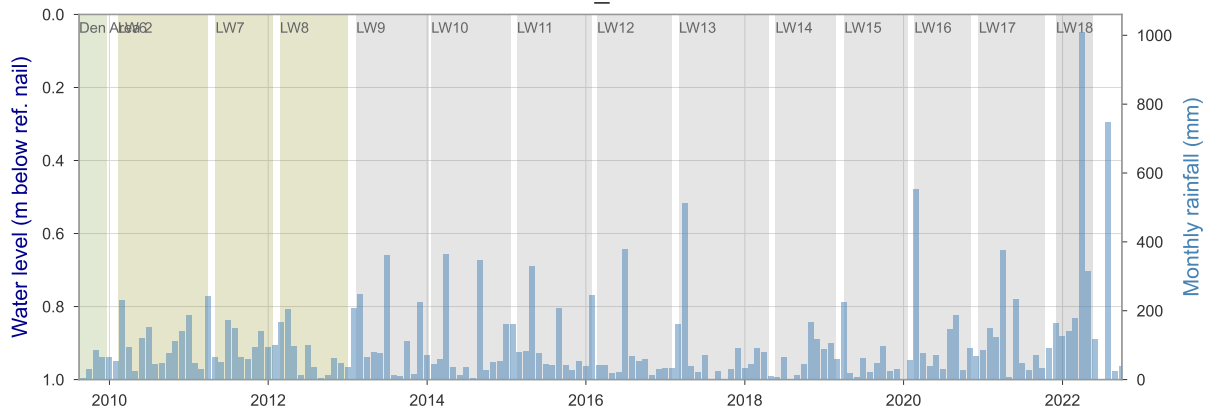
SCK_ROCKBAR5



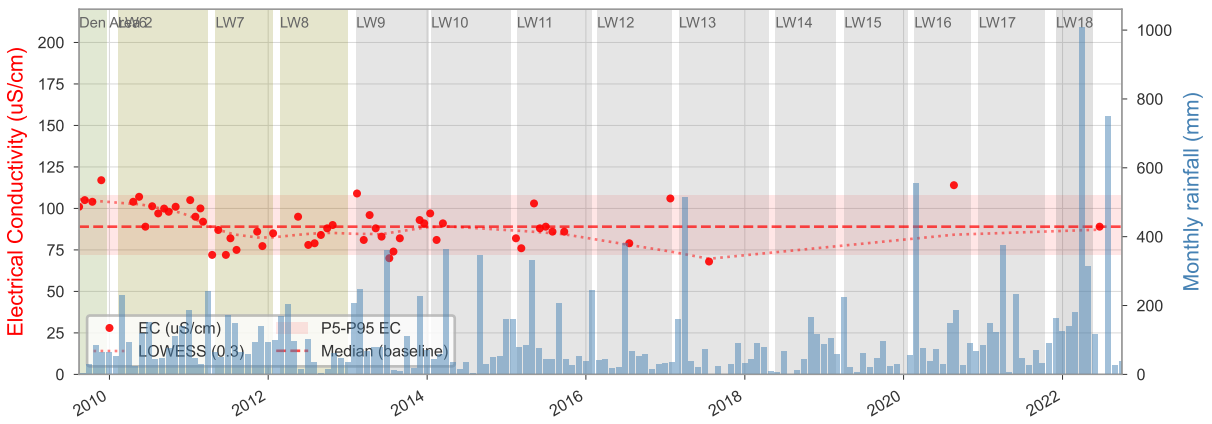
SCK_ROCKBAR5



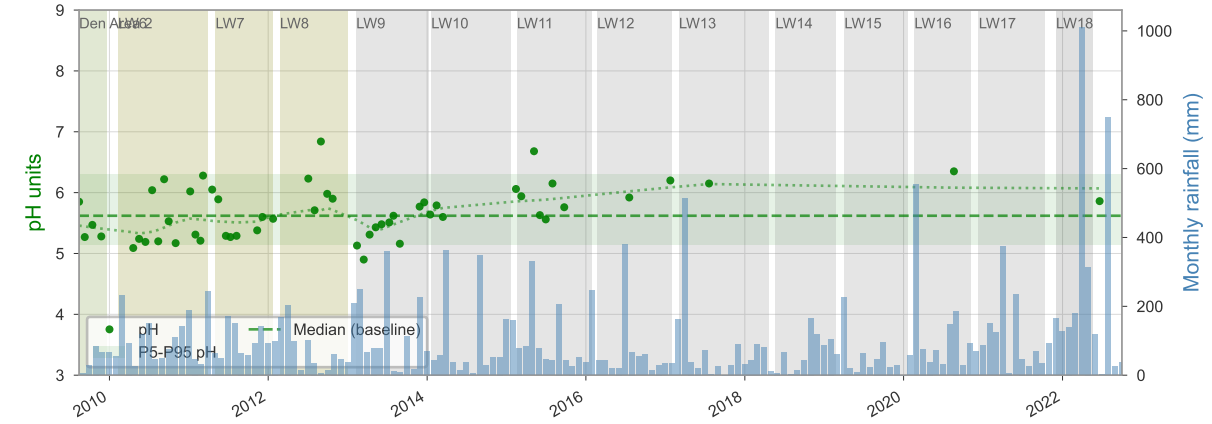
SCK_SC7



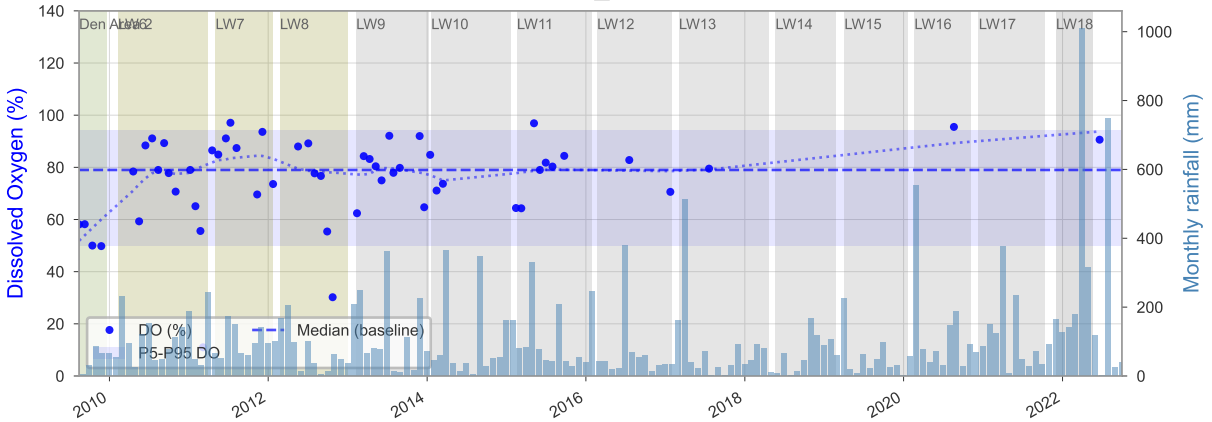
SCK_SC7



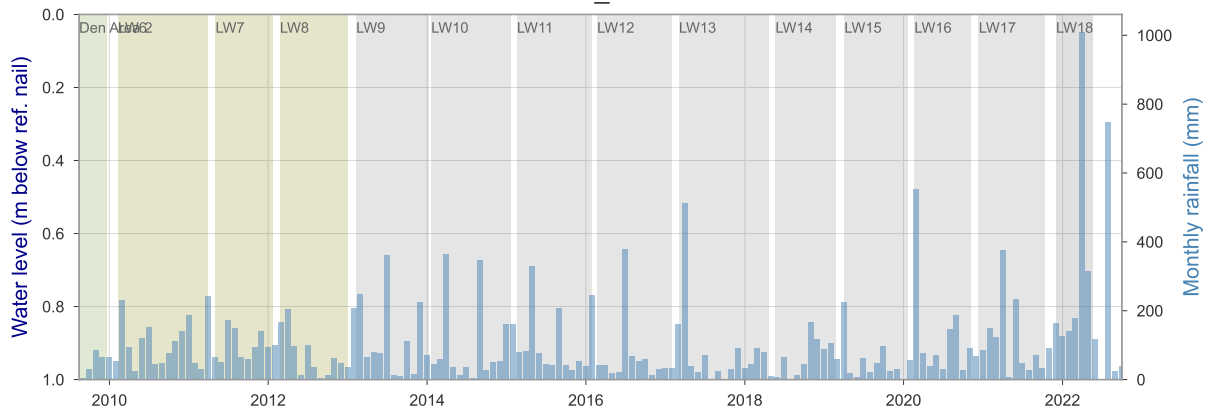
SCK_SC7



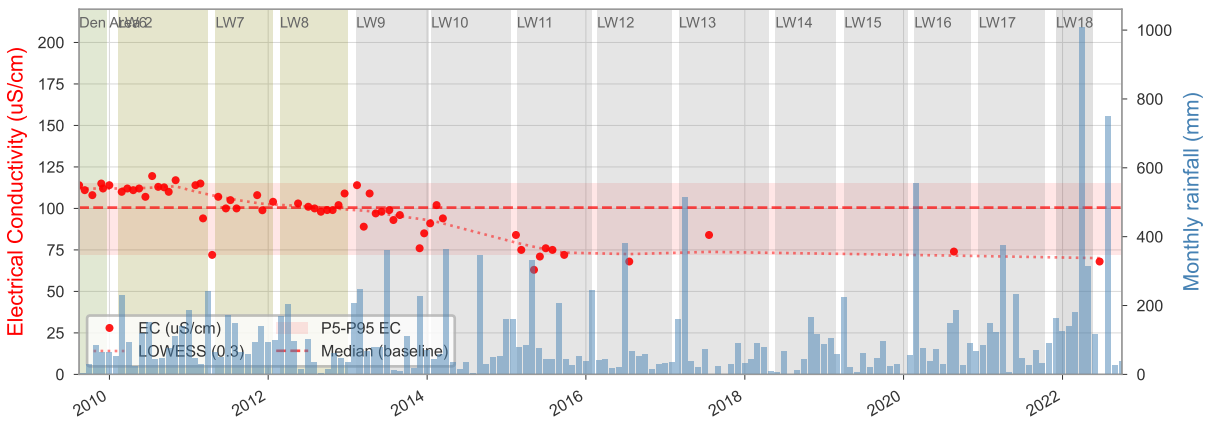
SCK_SC7



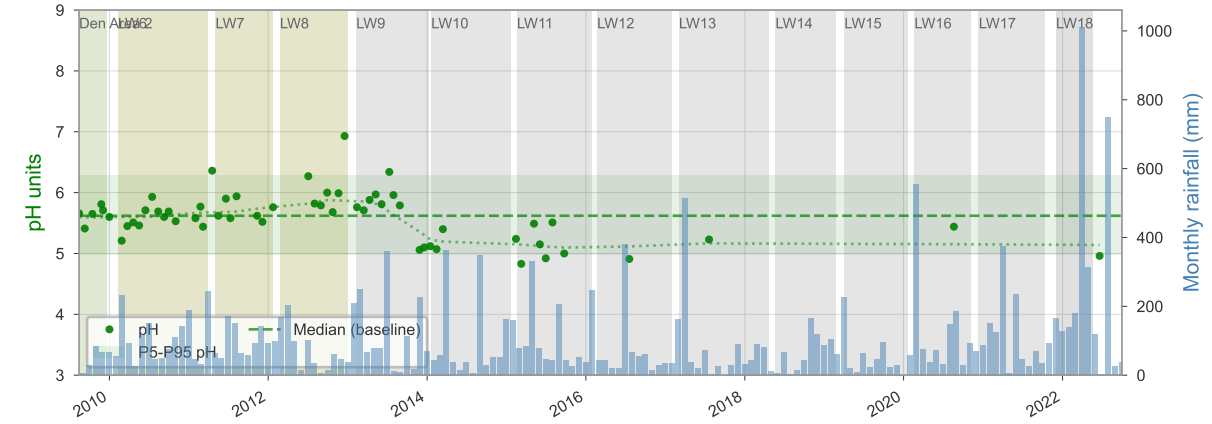
SCK_SC9



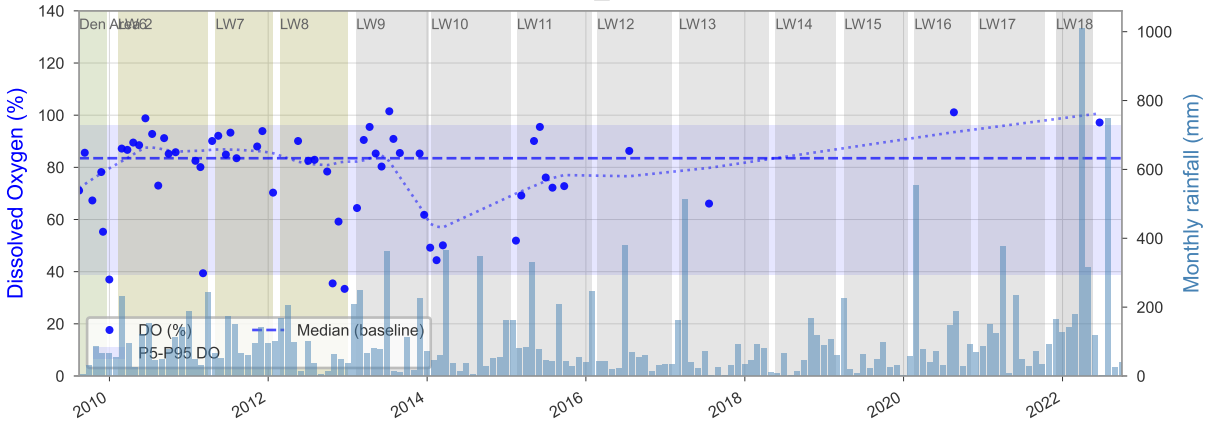
SCK_SC9



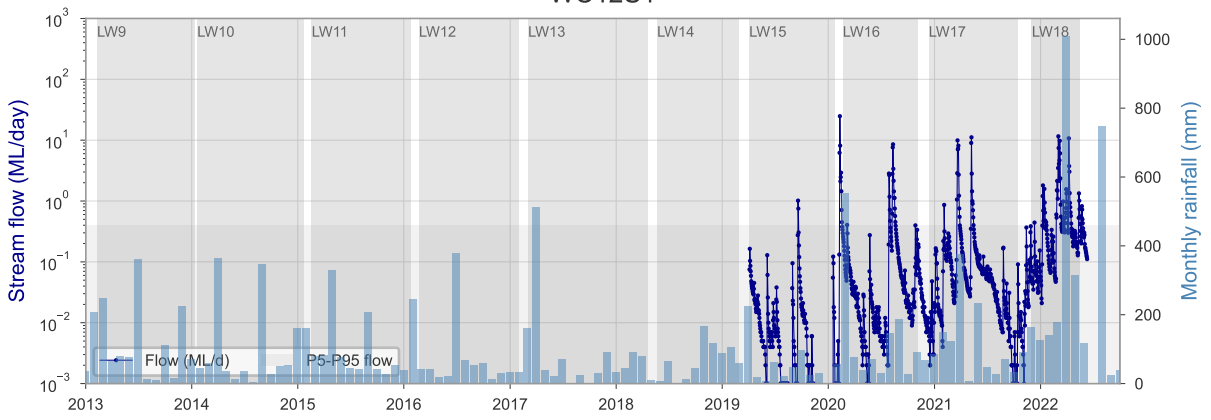
SCK_SC9



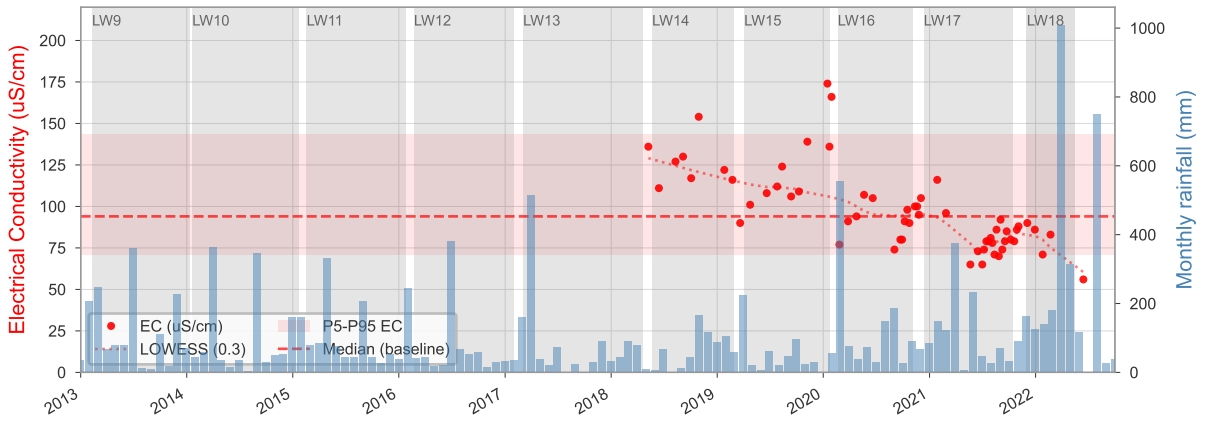
SCK_SC9



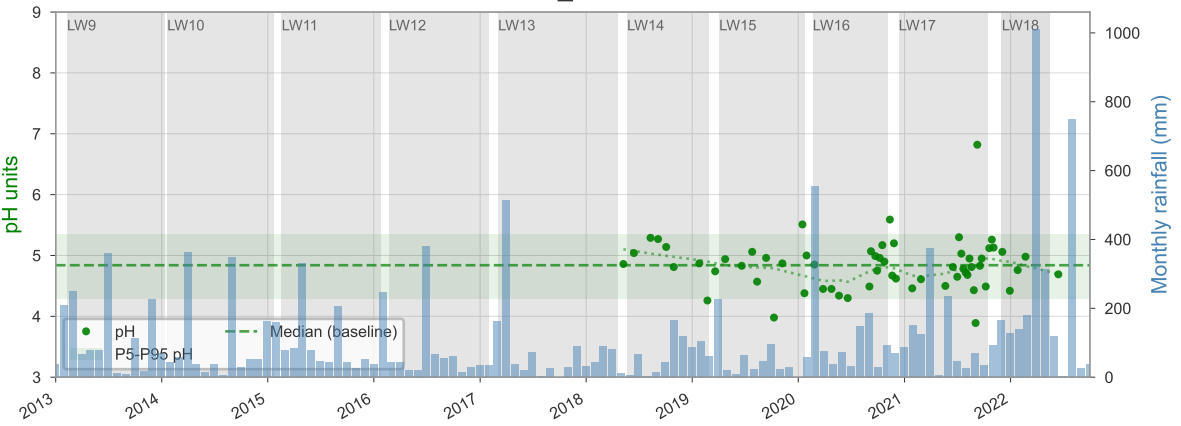
WC12S1



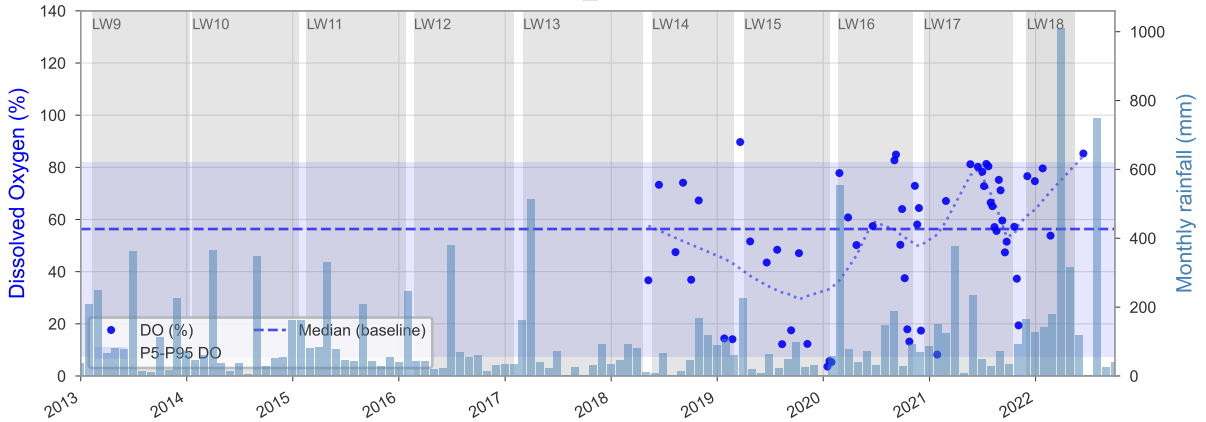
WC12_POOL1



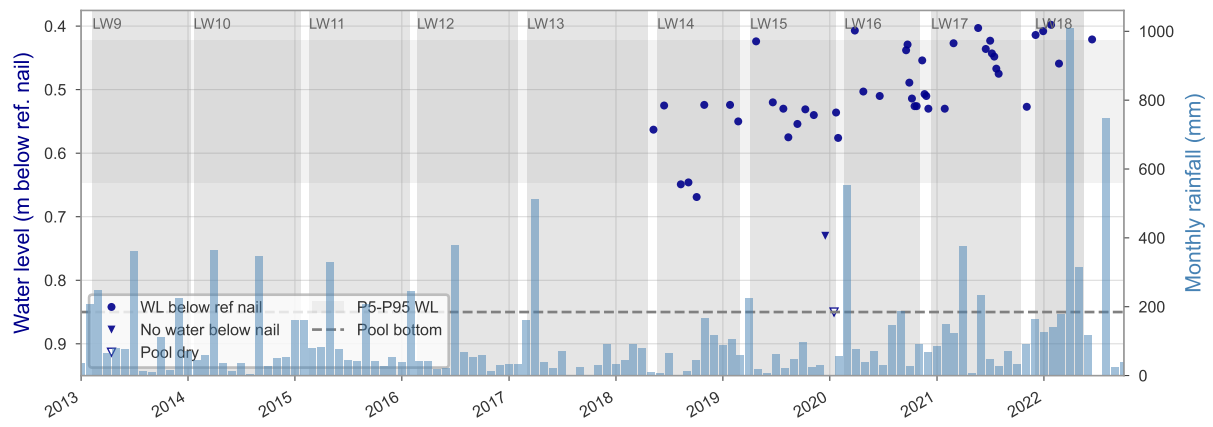
WC12_POOL1



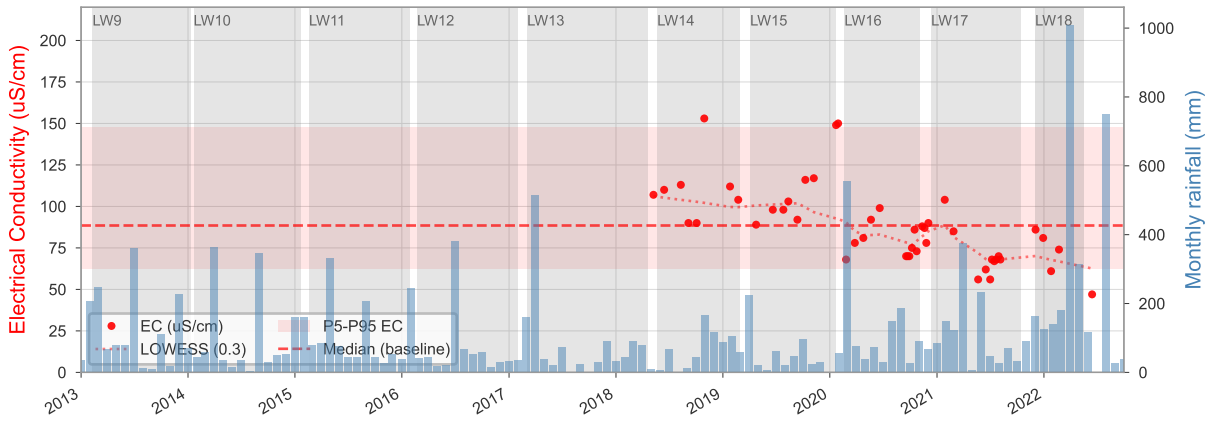
WC12_POOL1



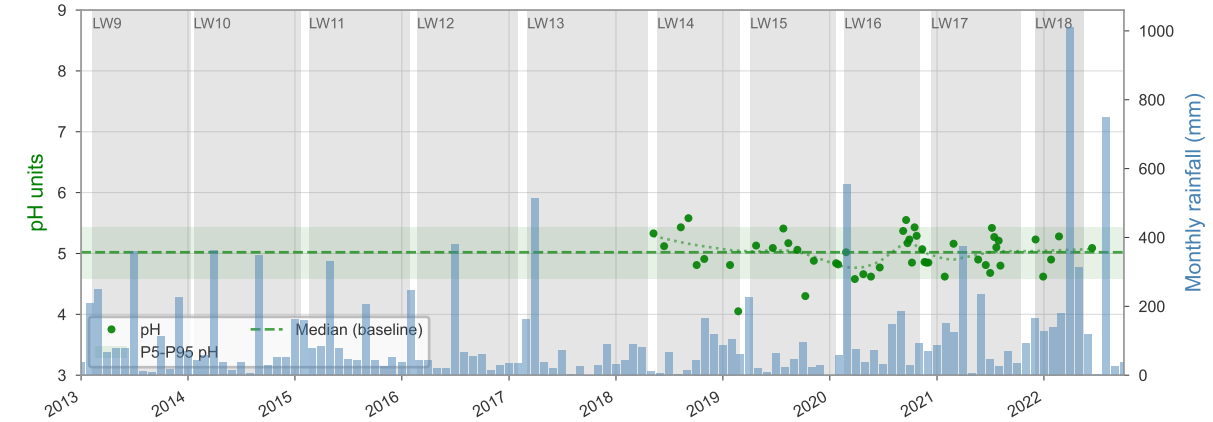
WC12_POOL12



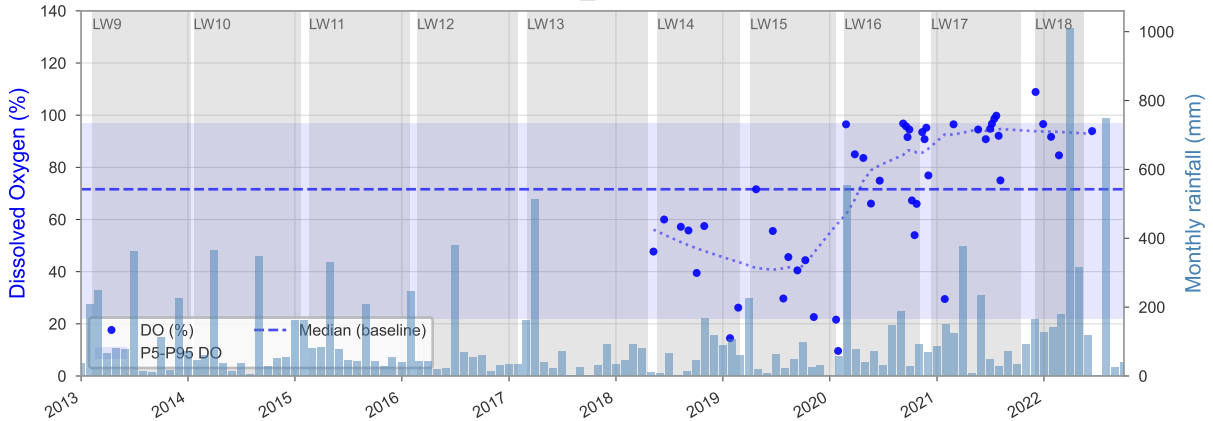
WC12_POOL12



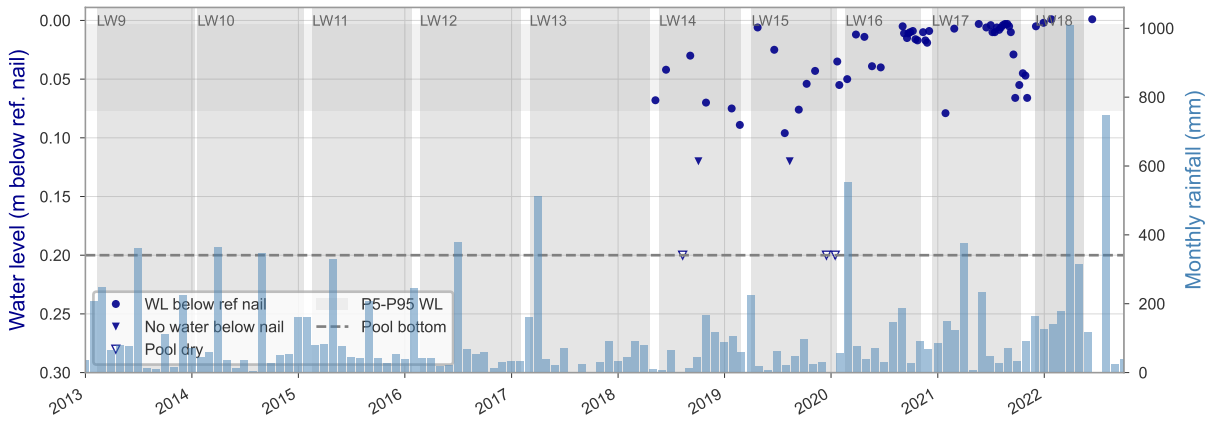
WC12_POOL12



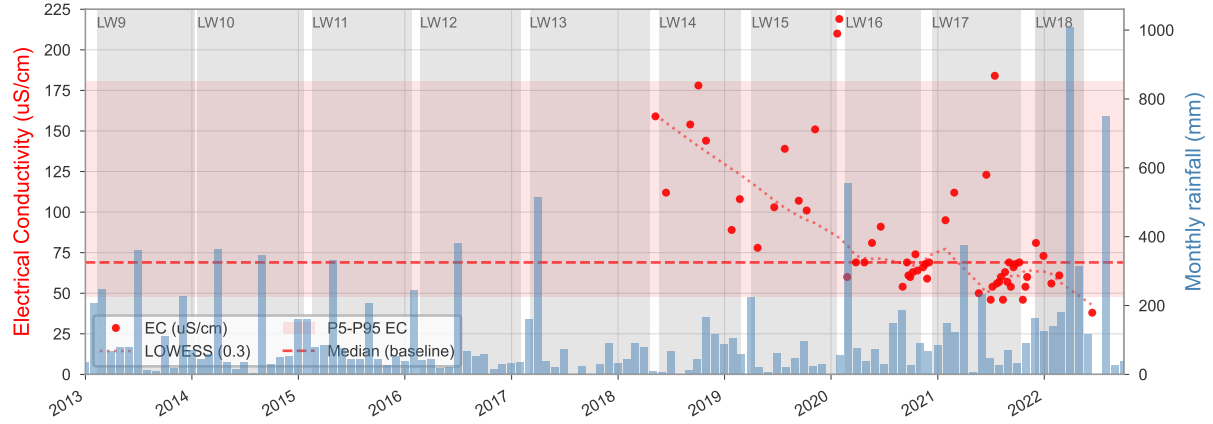
WC12_POOL12



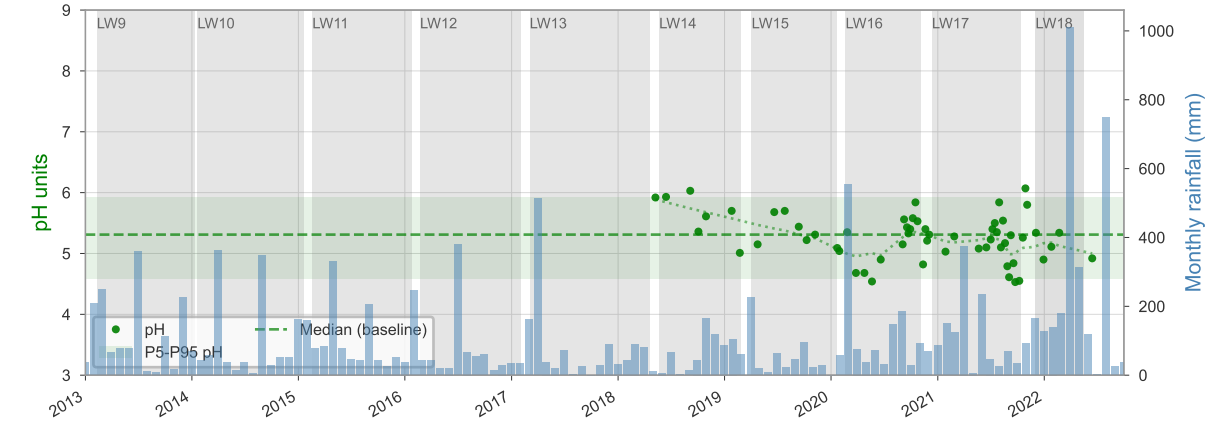
WC12_ROCKBAR18



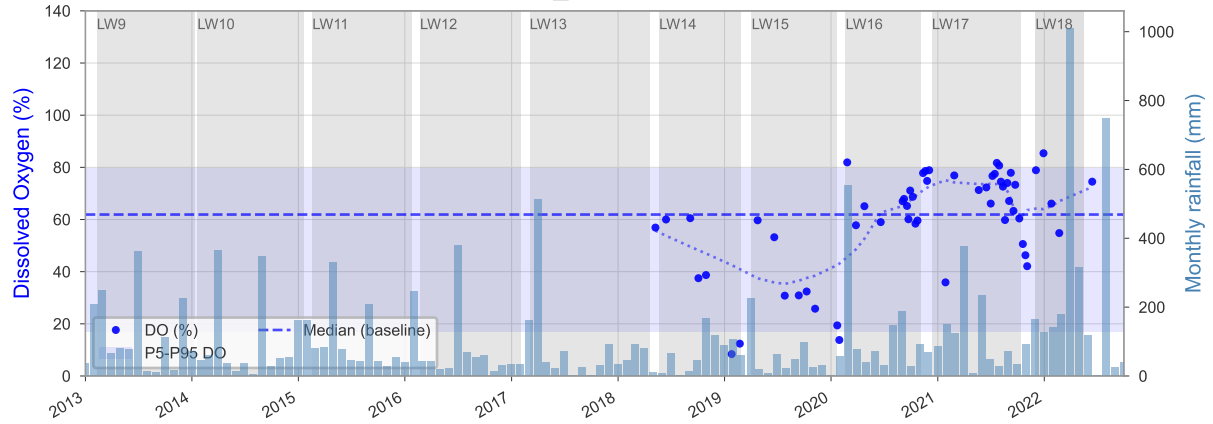
WC12_ROCKBAR18



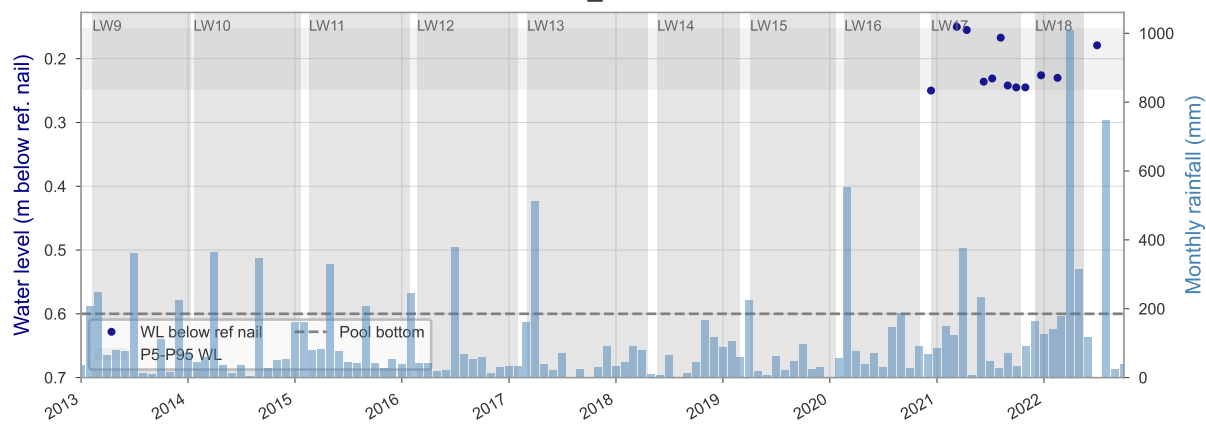
WC12_ROCKBAR18



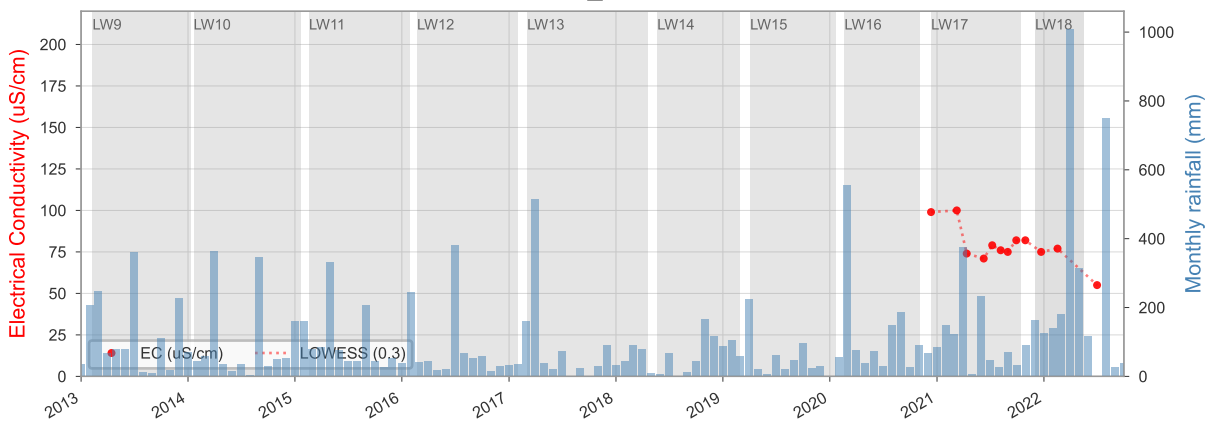
WC12_ROCKBAR18



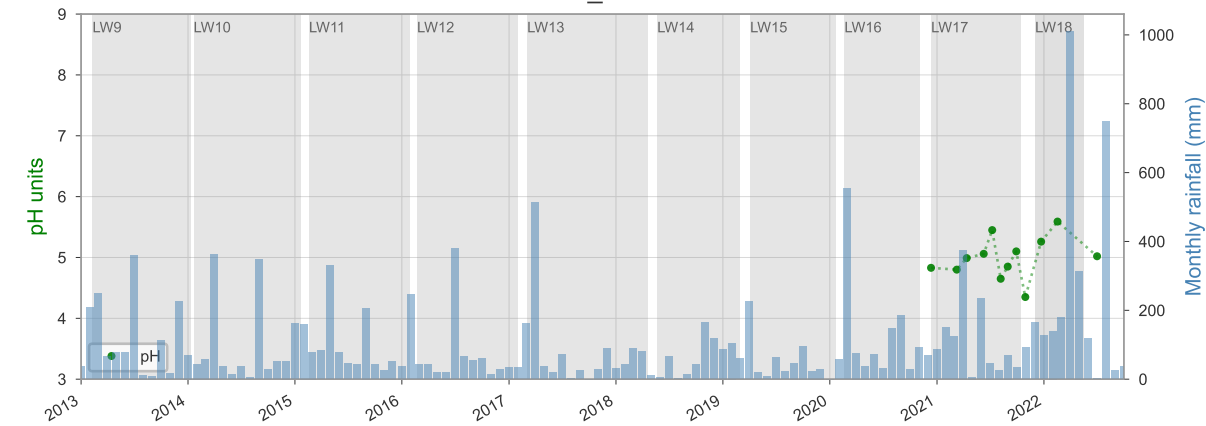
WC13_POOL1



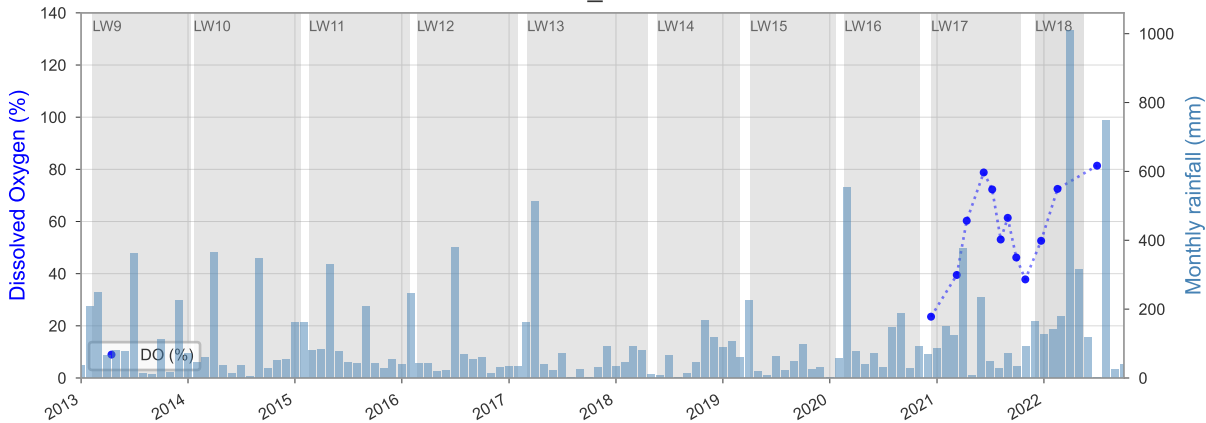
WC13_POOL1



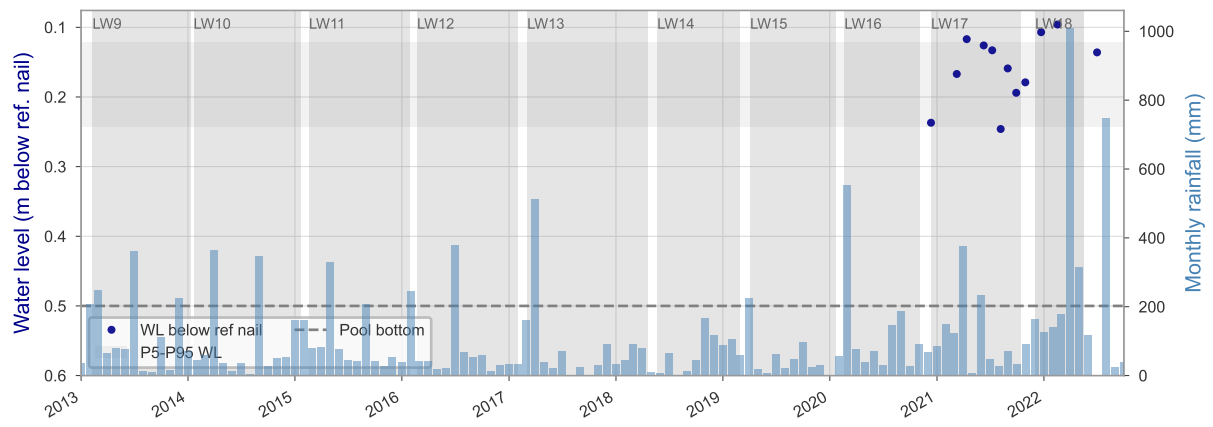
WC13_POOL1



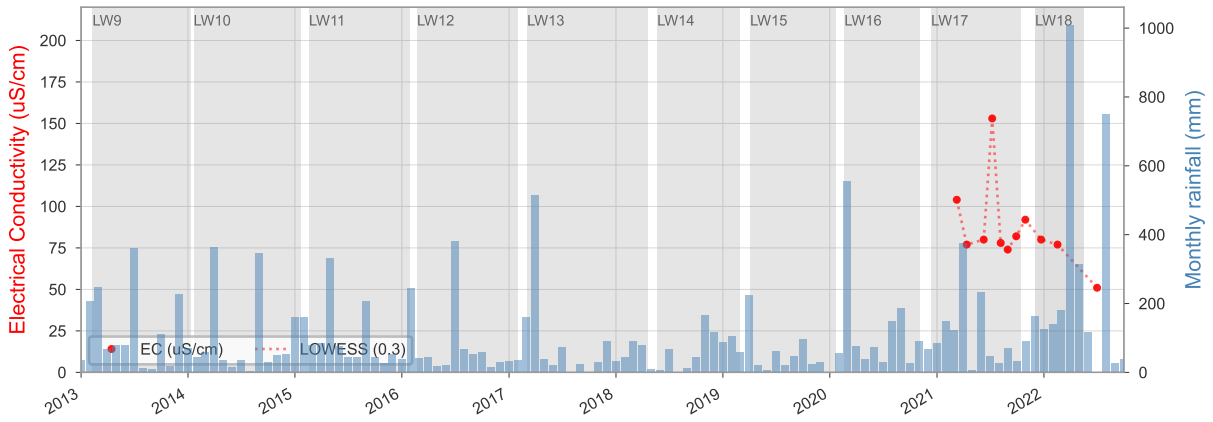
WC13_POOL1



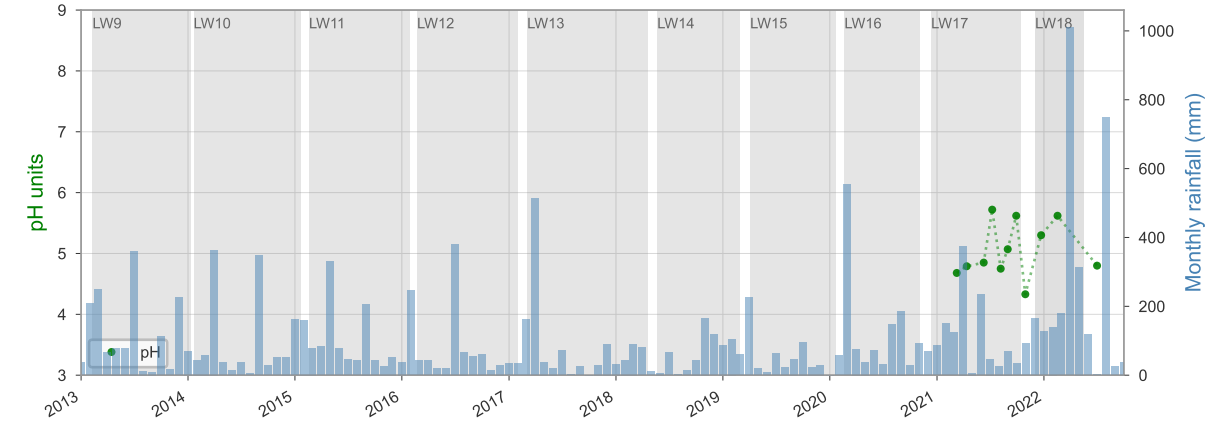
WC13_POOL3



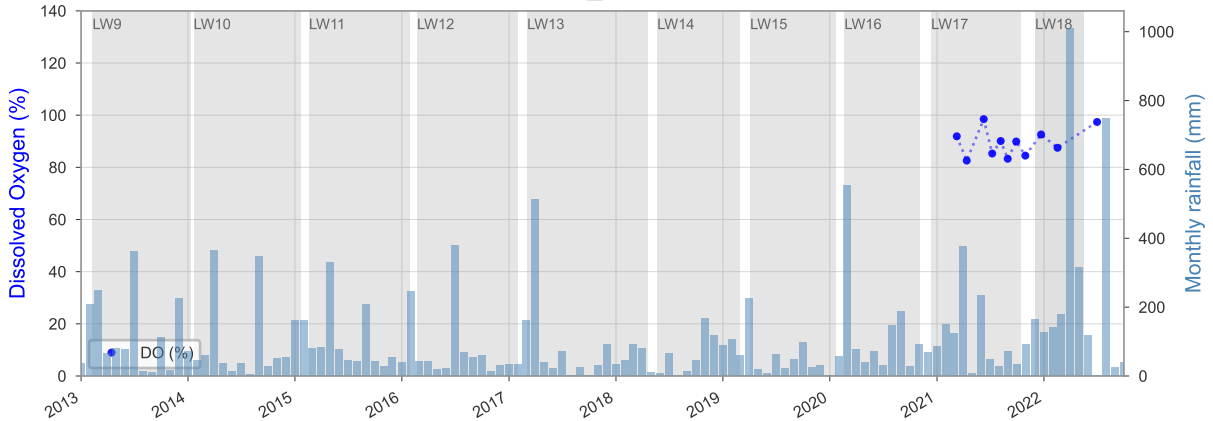
WC13_POOL3



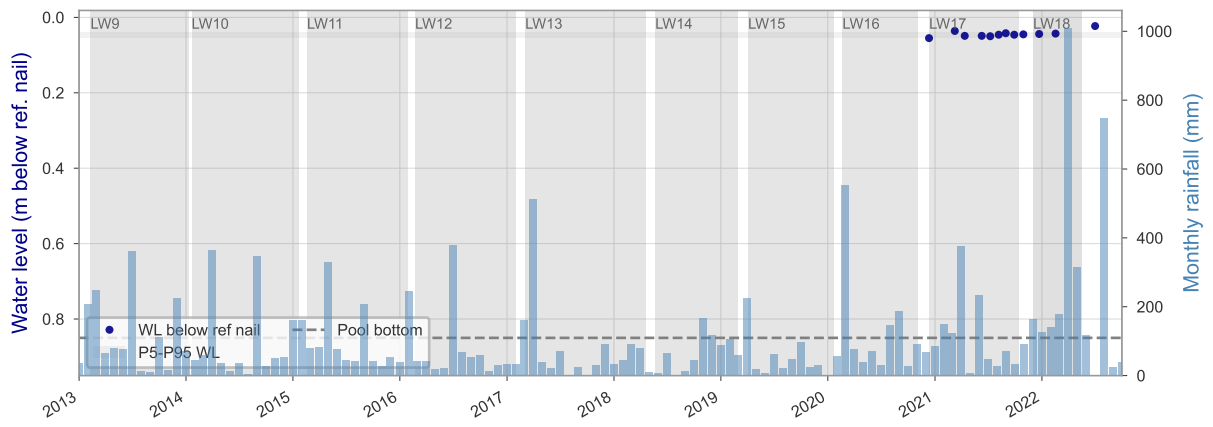
WC13_POOL3



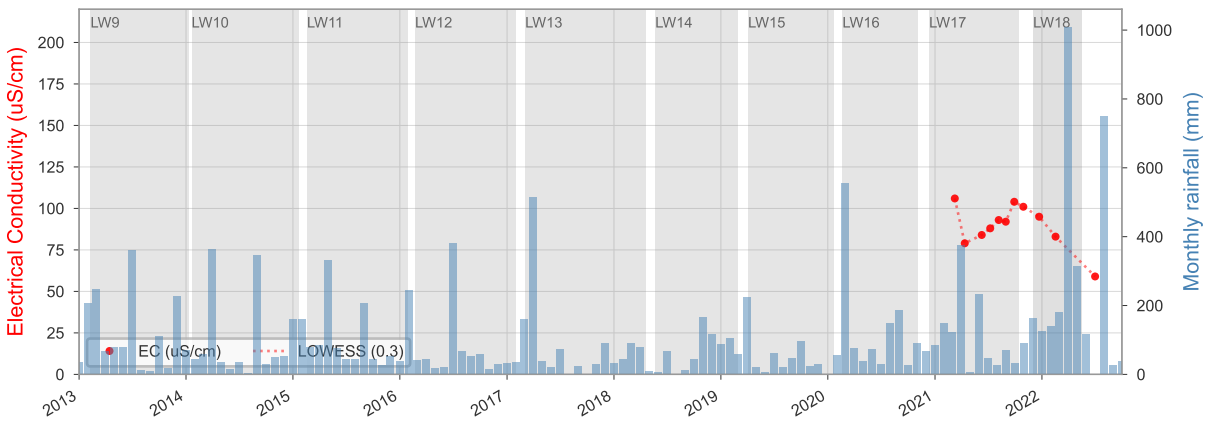
WC13_POOL3



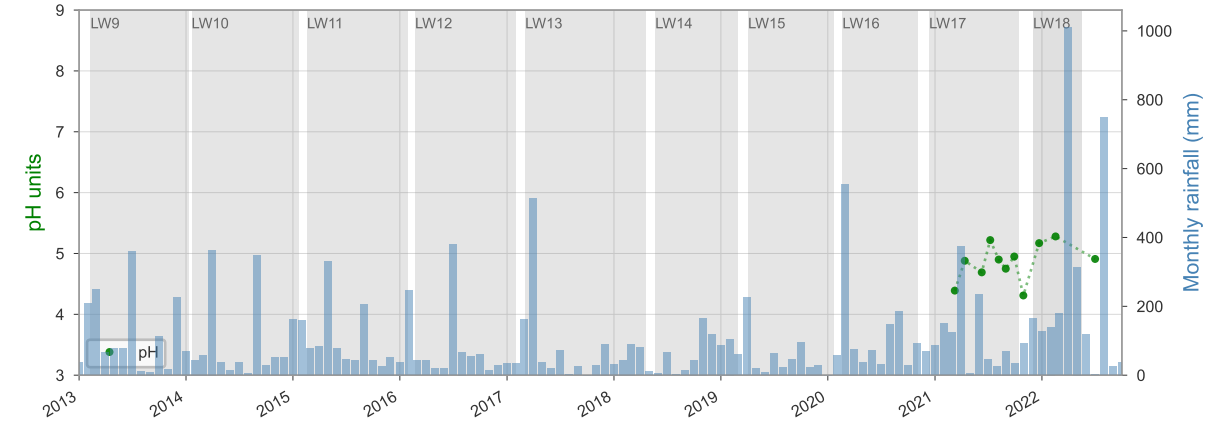
WC14_POOL16



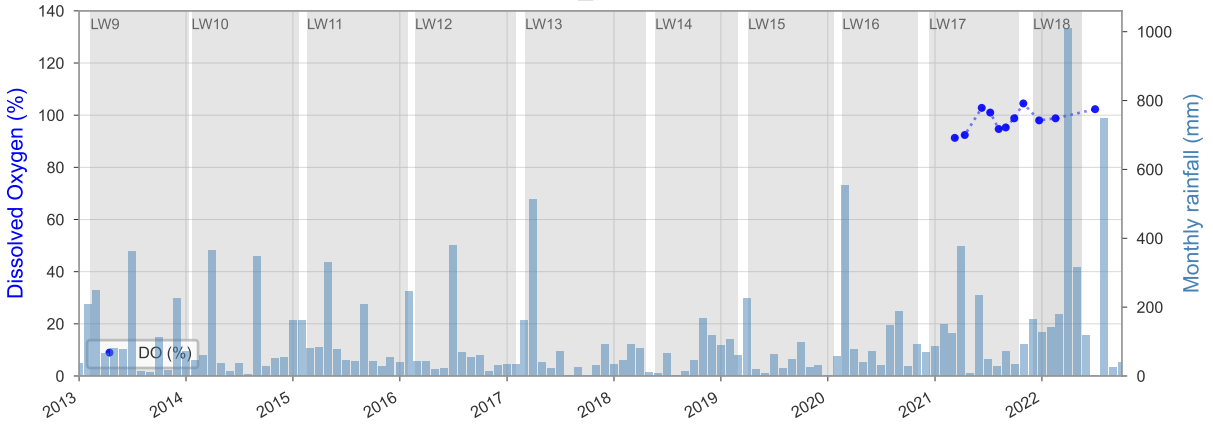
WC14_POOL16



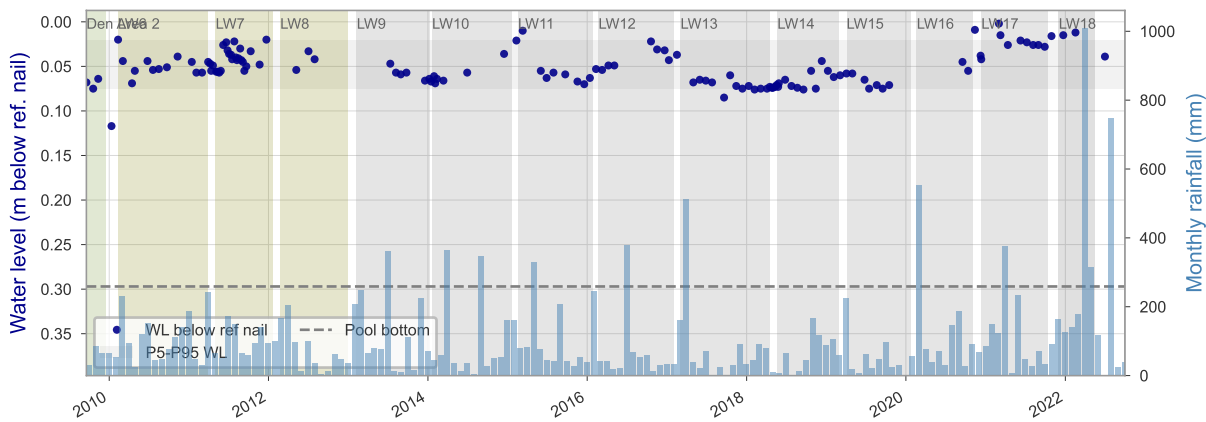
WC14_POOL16



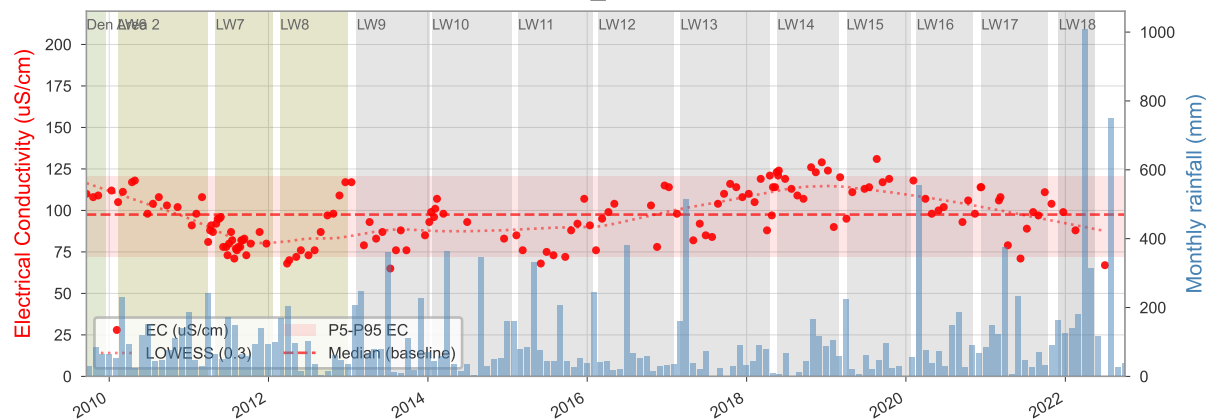
WC14_POOL16



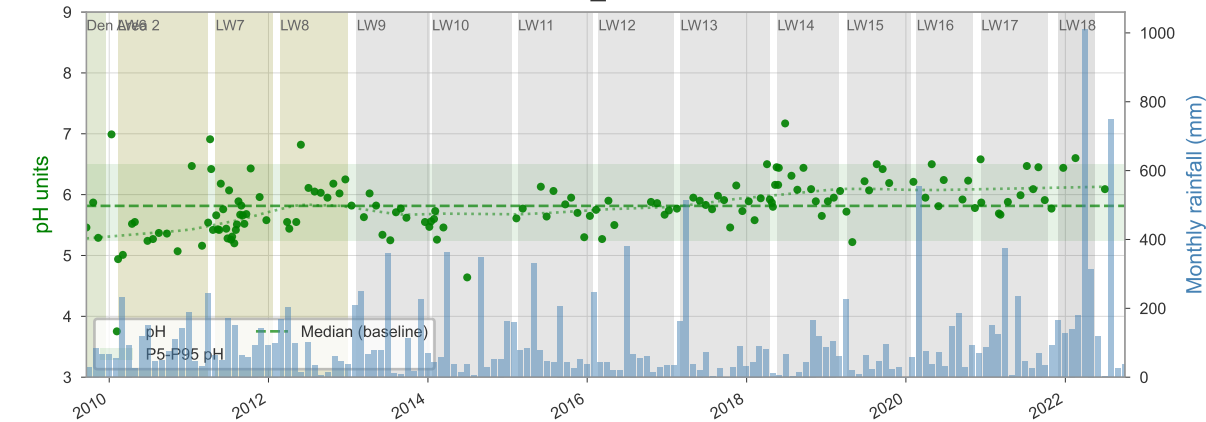
WC14_POOL3



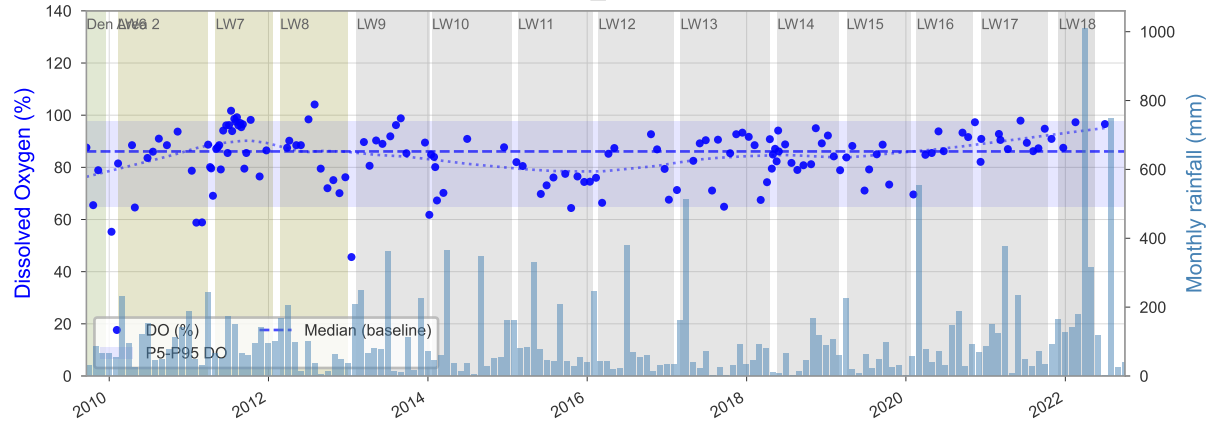
WC14_POOL3



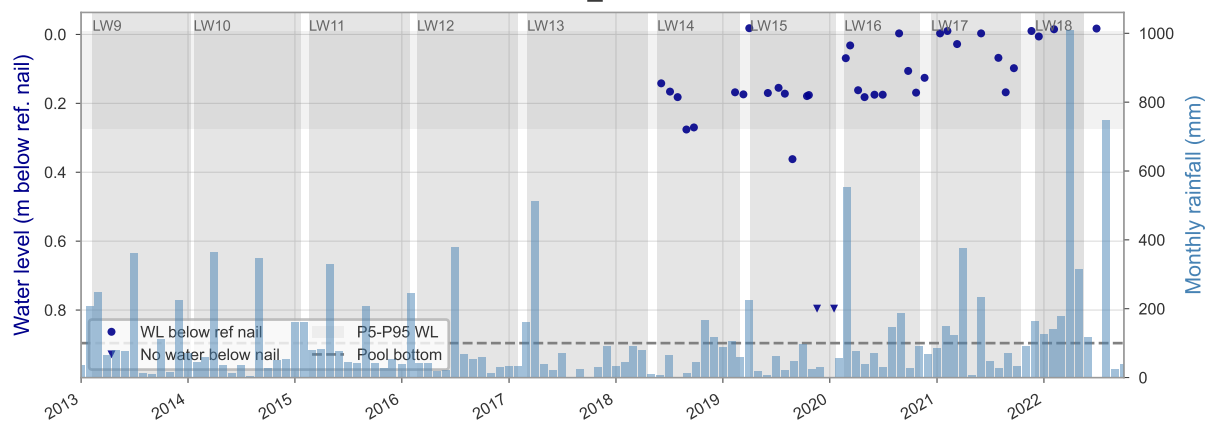
WC14_POOL3



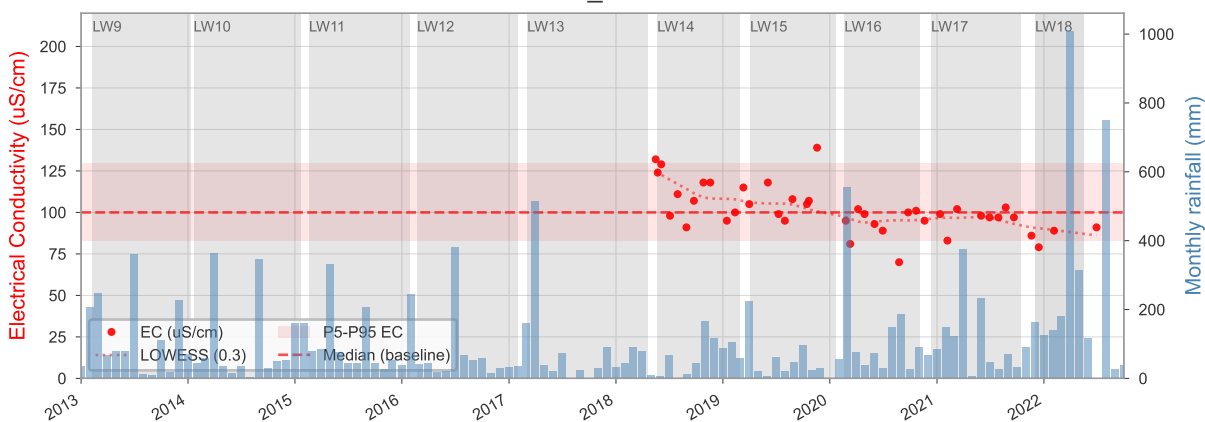
WC14_POOL3



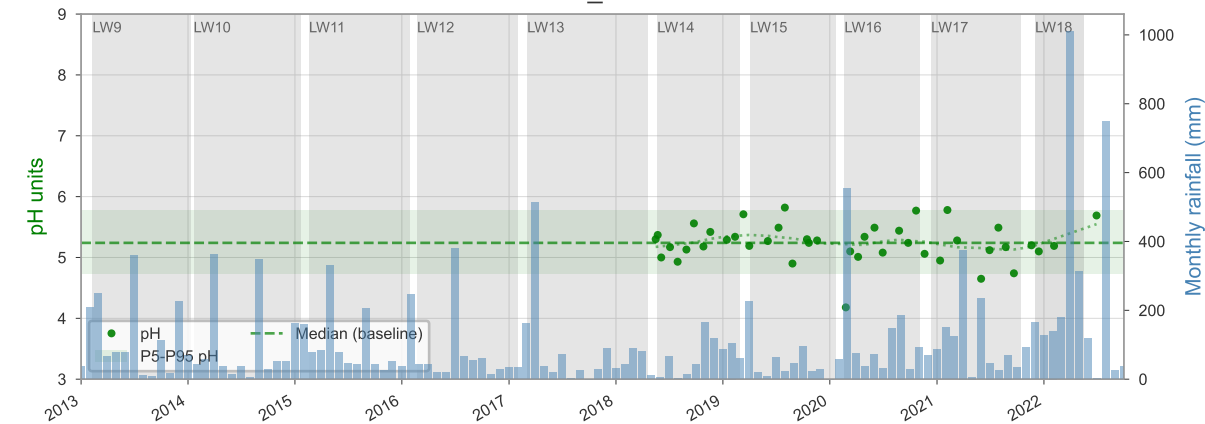
WC15_POOL2



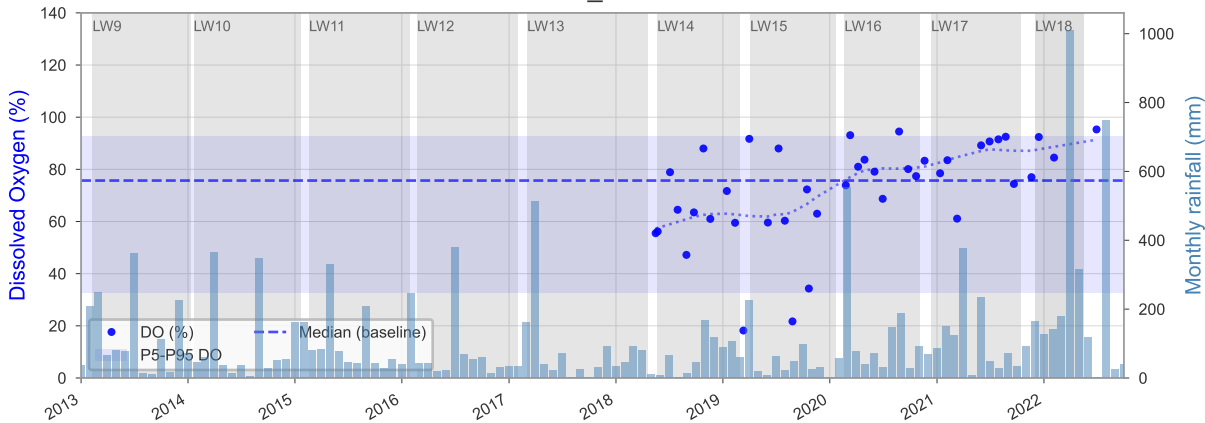
WC15_POOL2



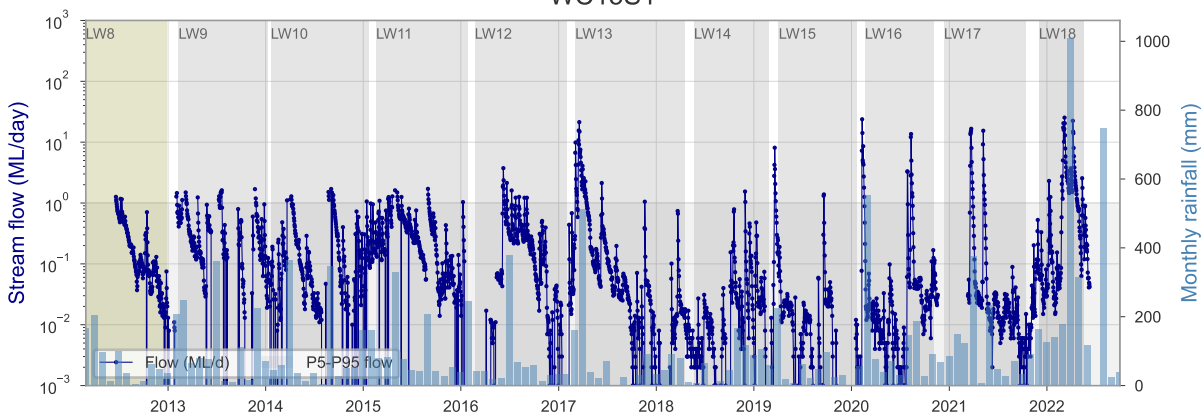
WC15_POOL2



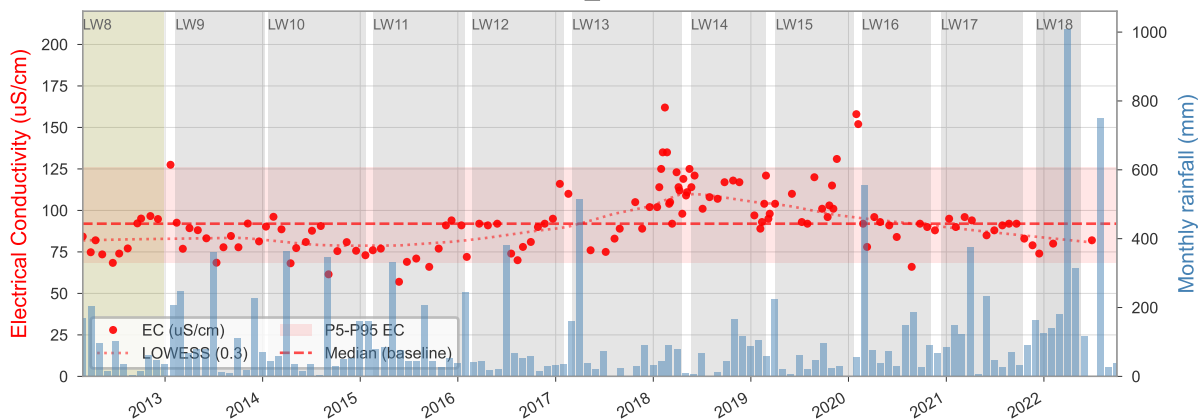
WC15_POOL2



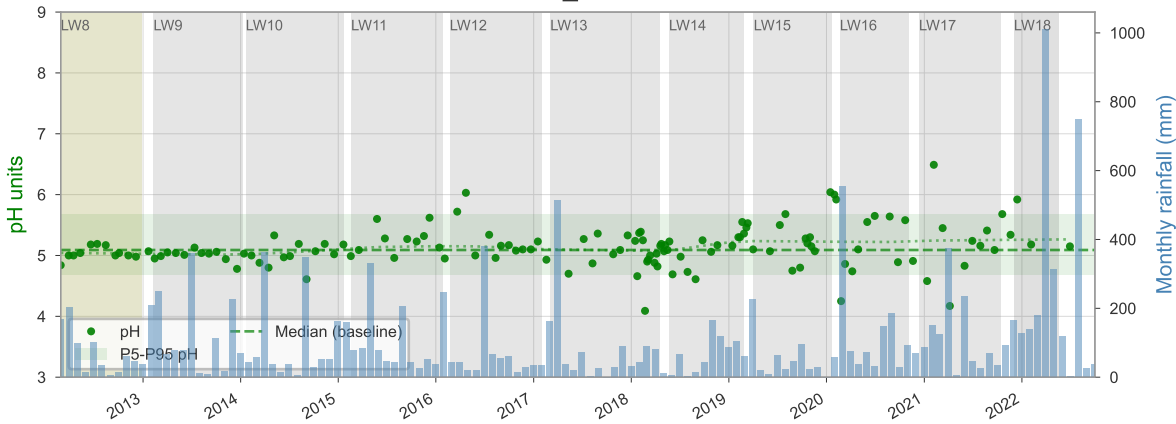
WC15S1



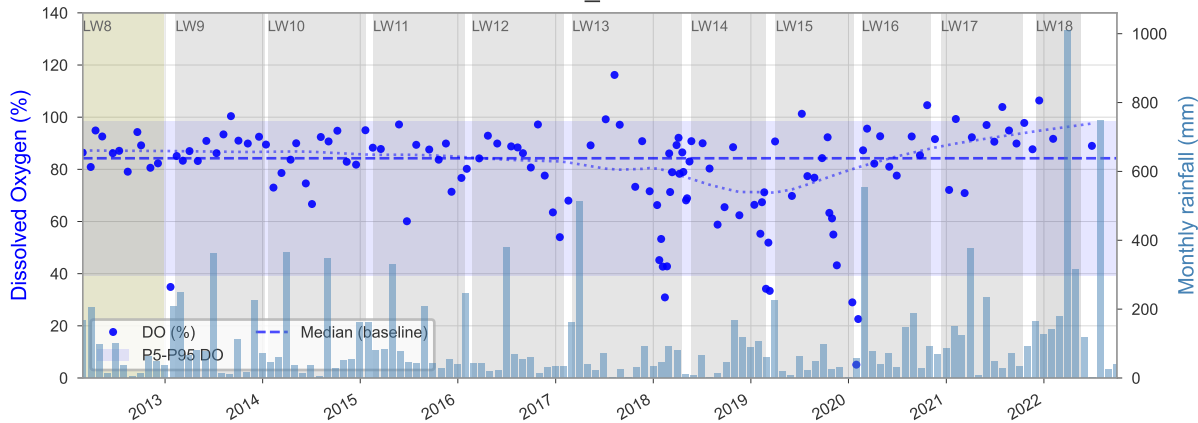
WC15_POOL9



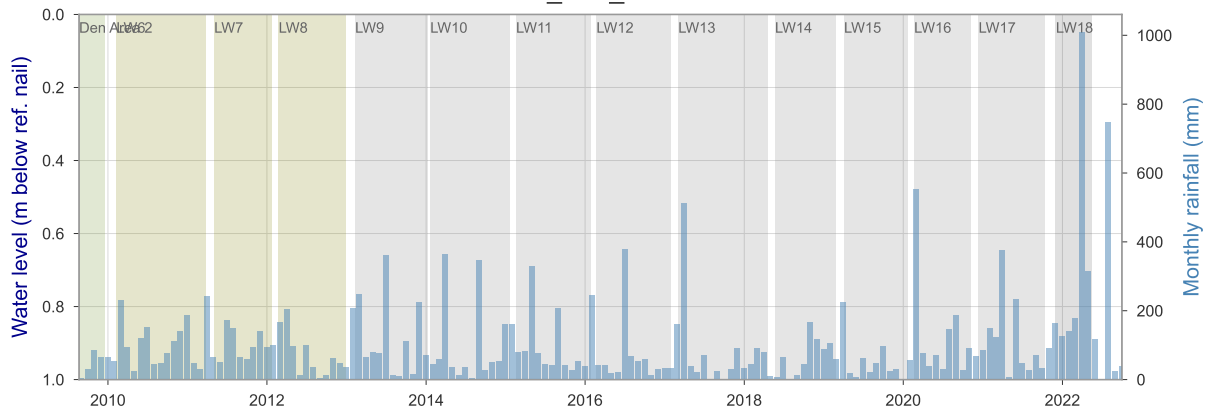
WC15_POOL9



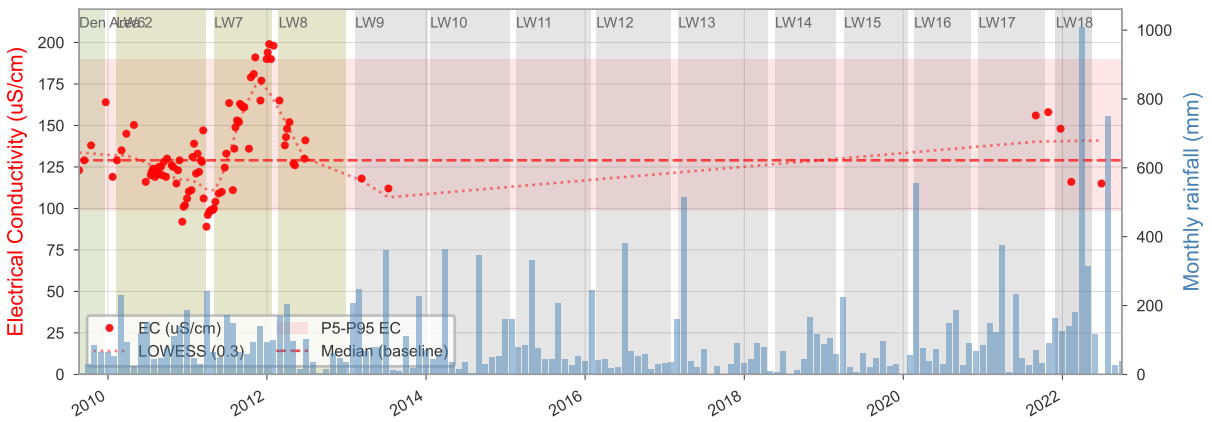
WC15_POOL9



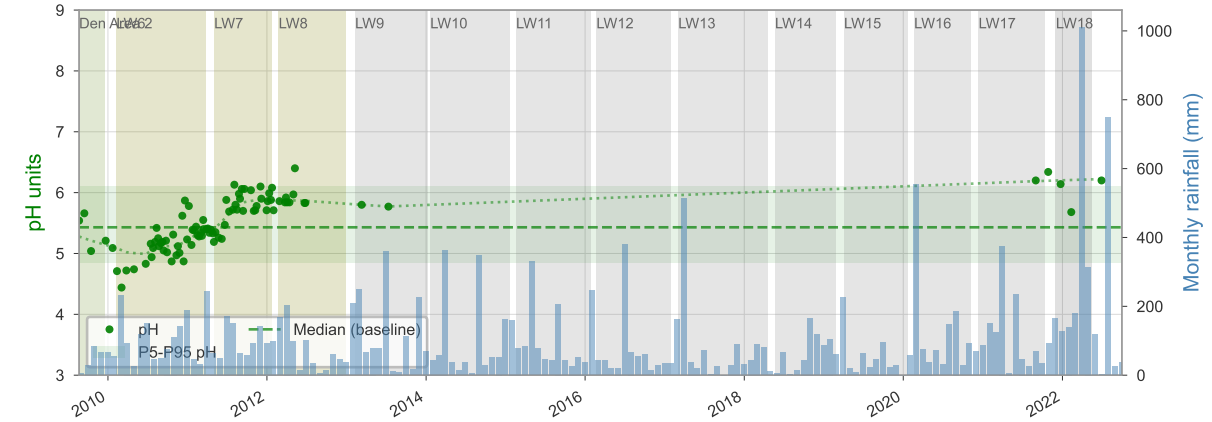
WC17_S12_POOL10



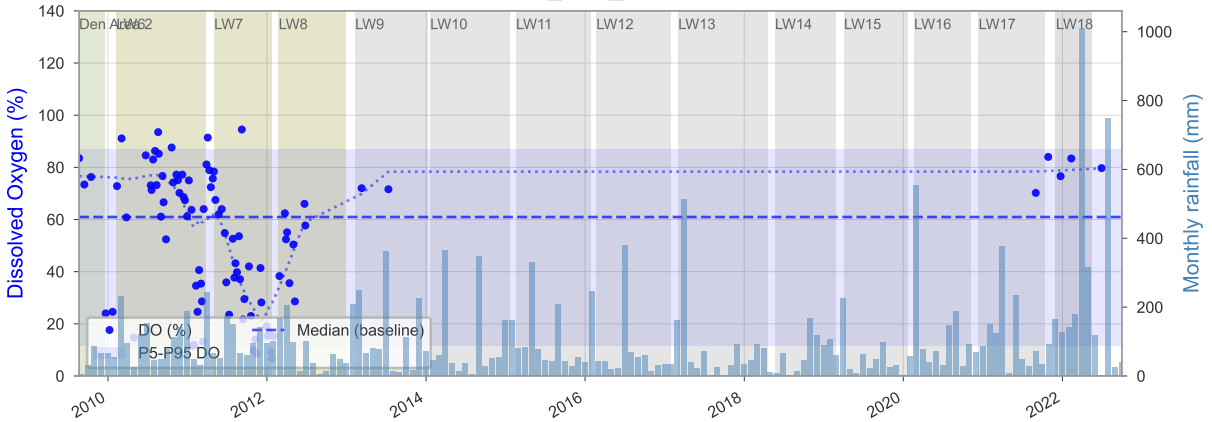
WC17_S12_POOL10



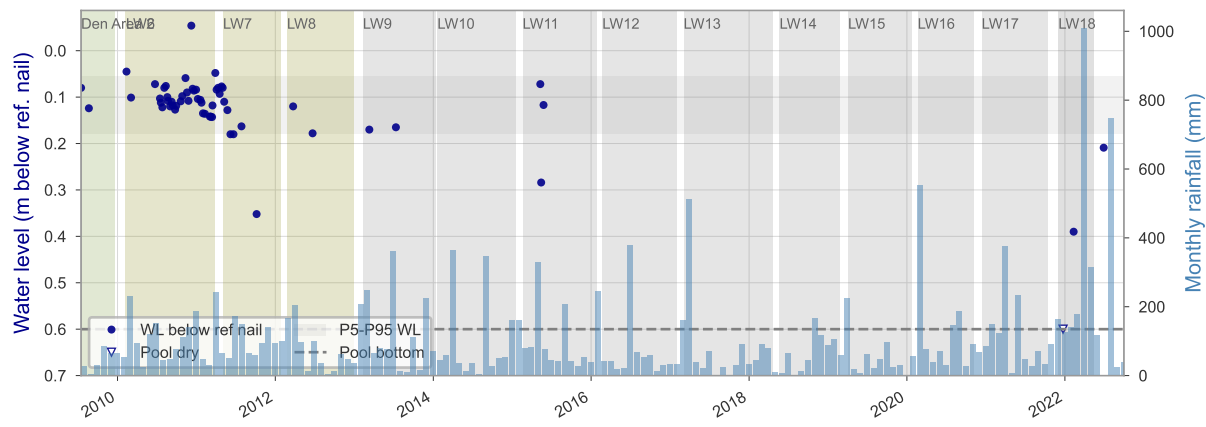
WC17_S12_POOL10



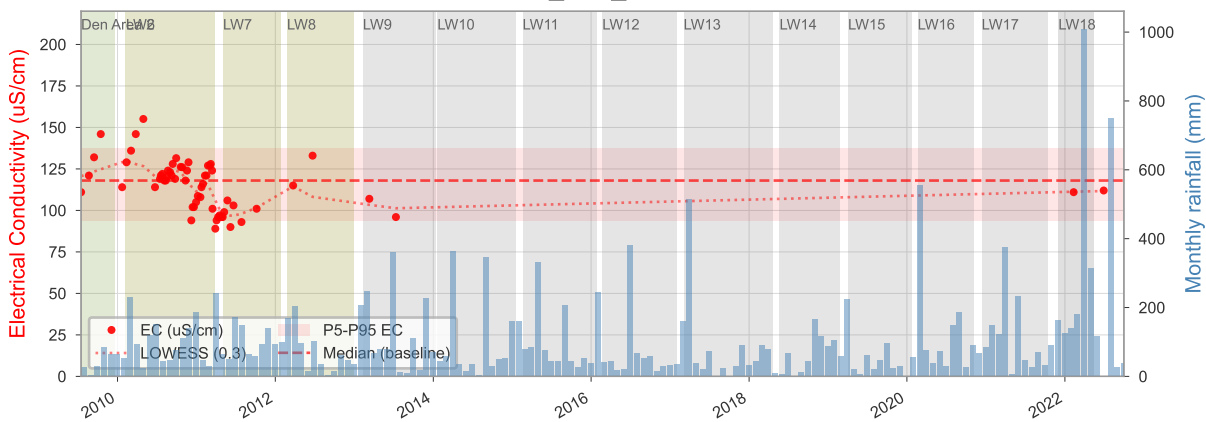
WC17_S12_POOL10



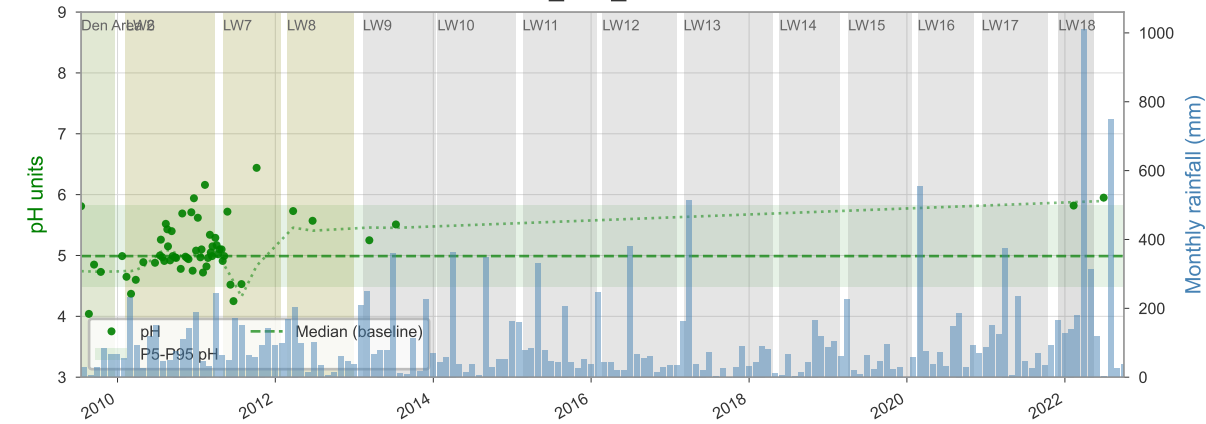
WC17_S12_POOL12



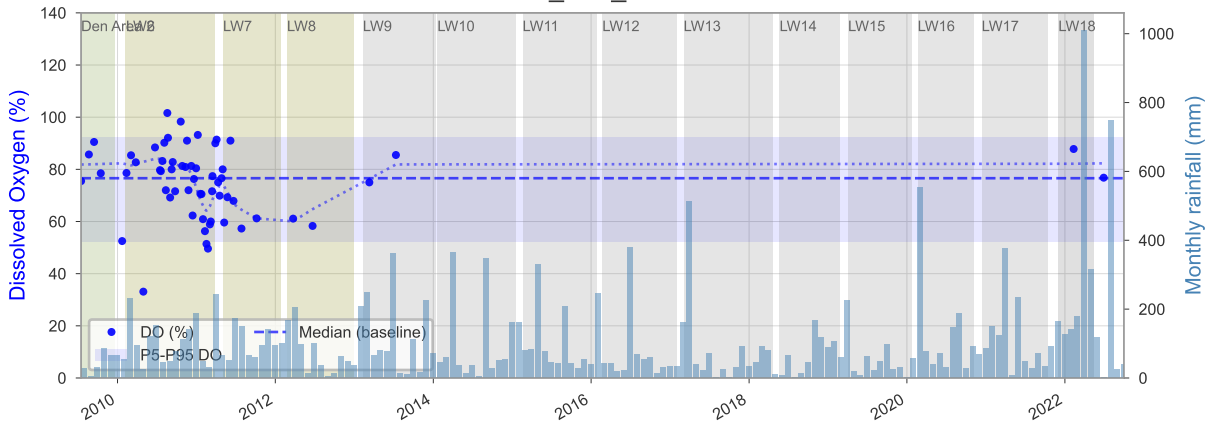
WC17_S12_POOL12



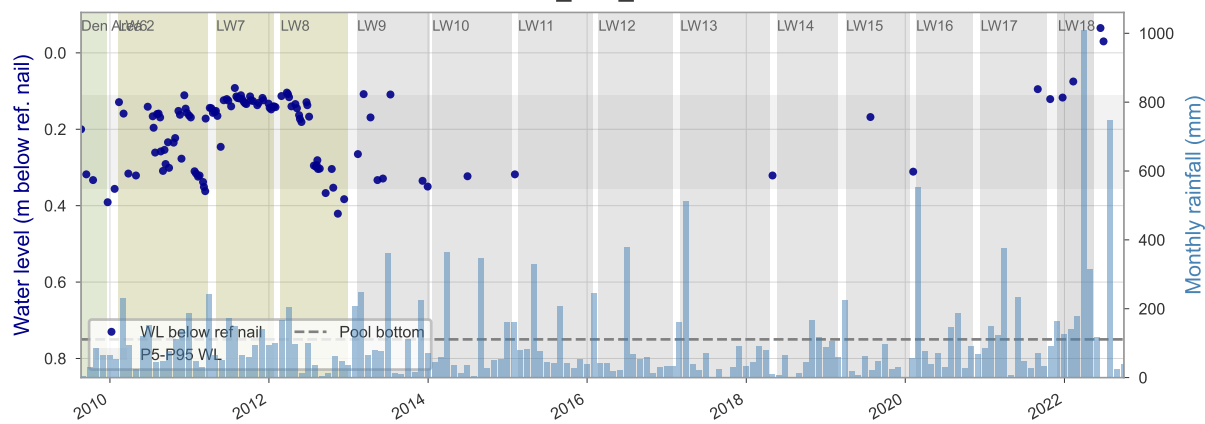
WC17_S12_POOL12



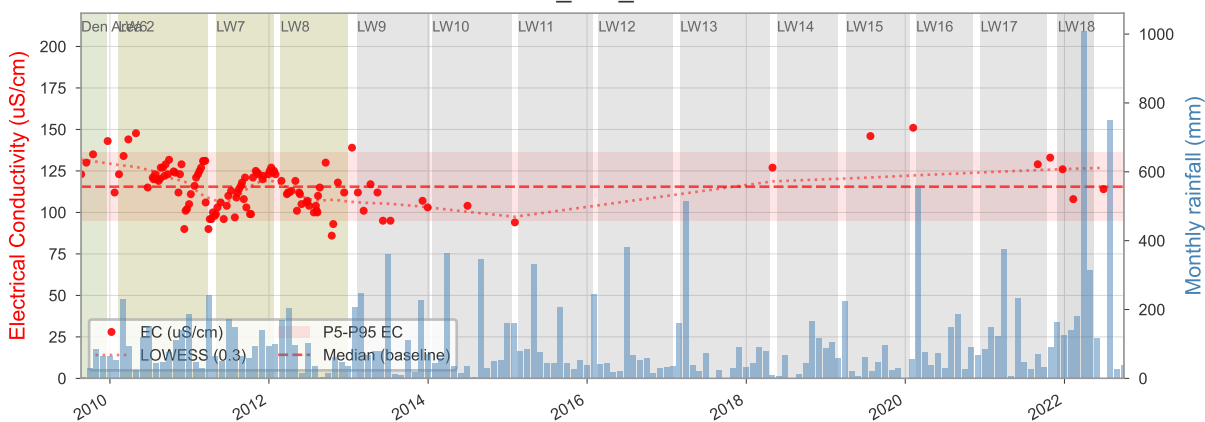
WC17_S12_POOL12



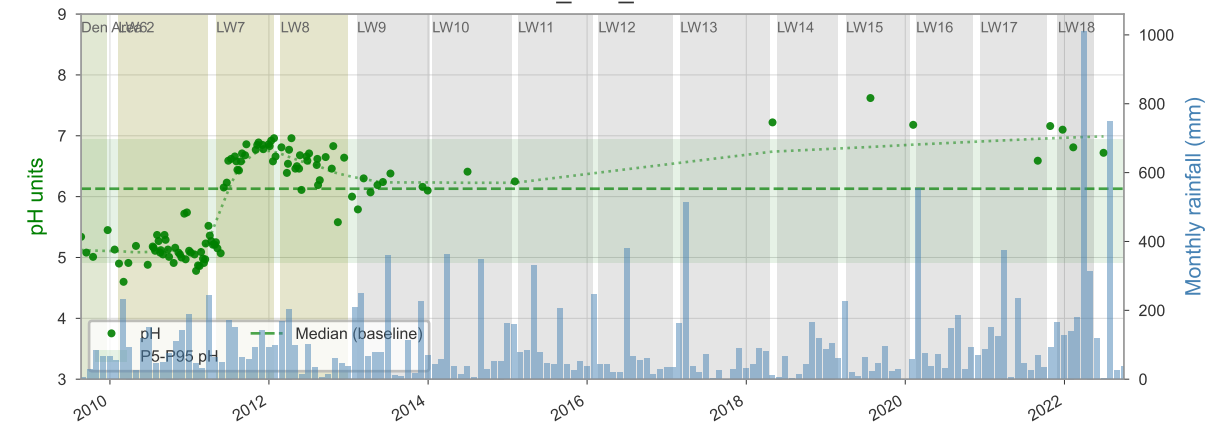
WC17_S12_POOL4



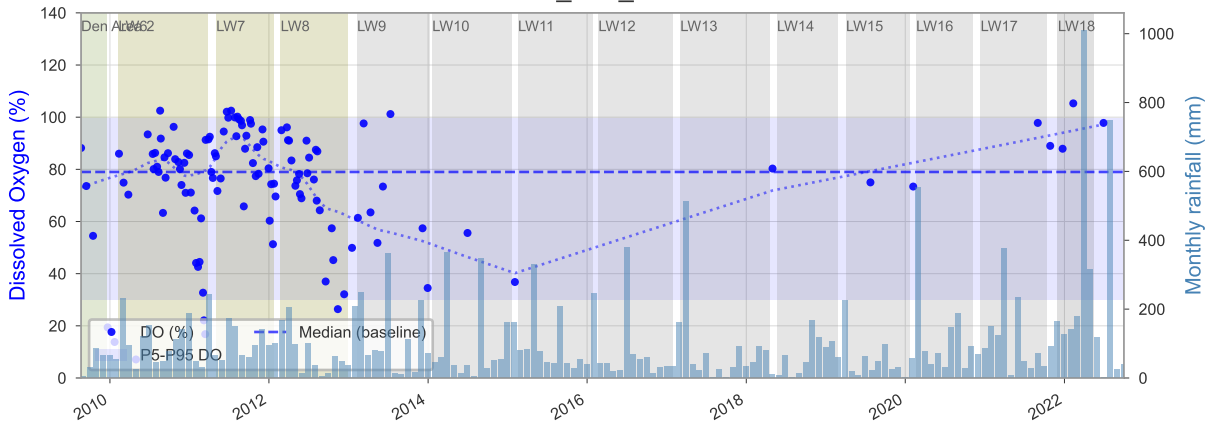
WC17_S12_POOL4



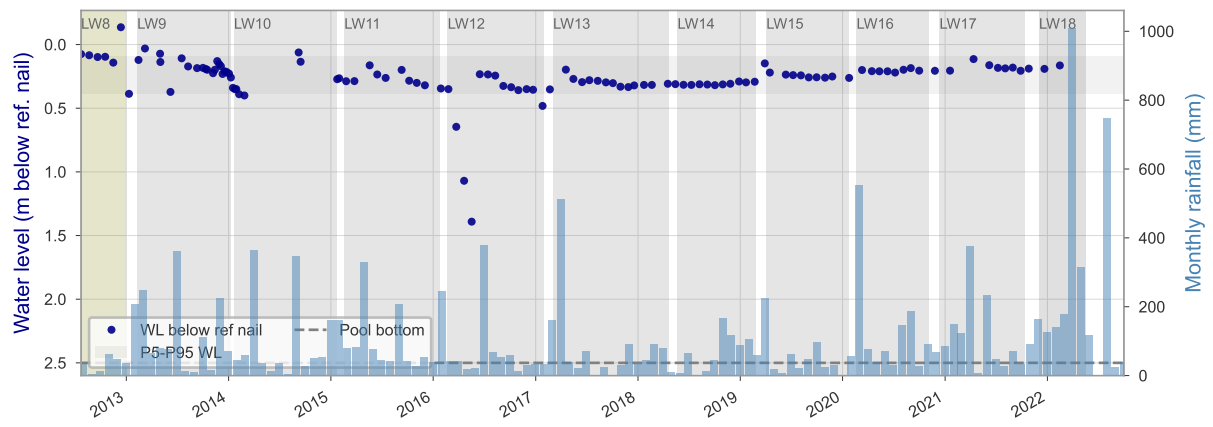
WC17_S12_POOL4



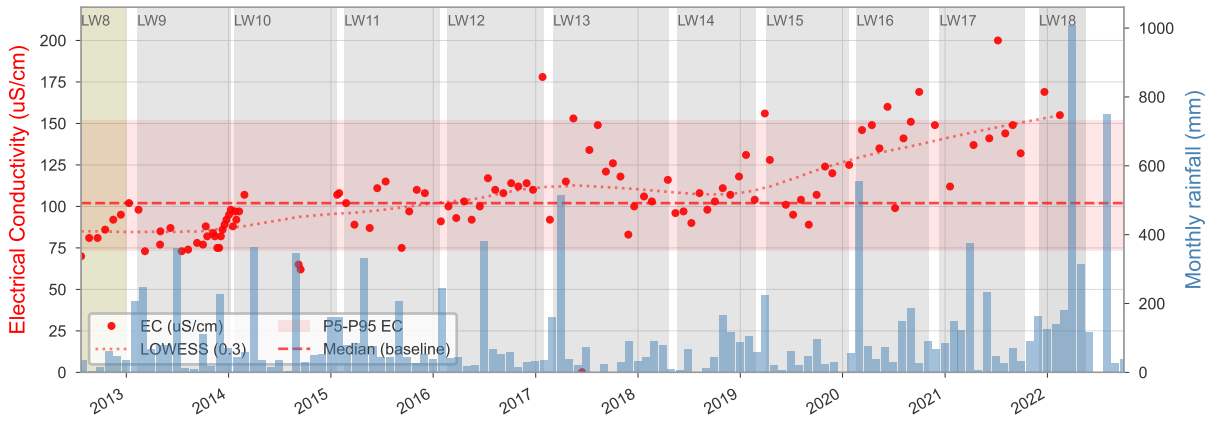
WC17_S12_POOL4



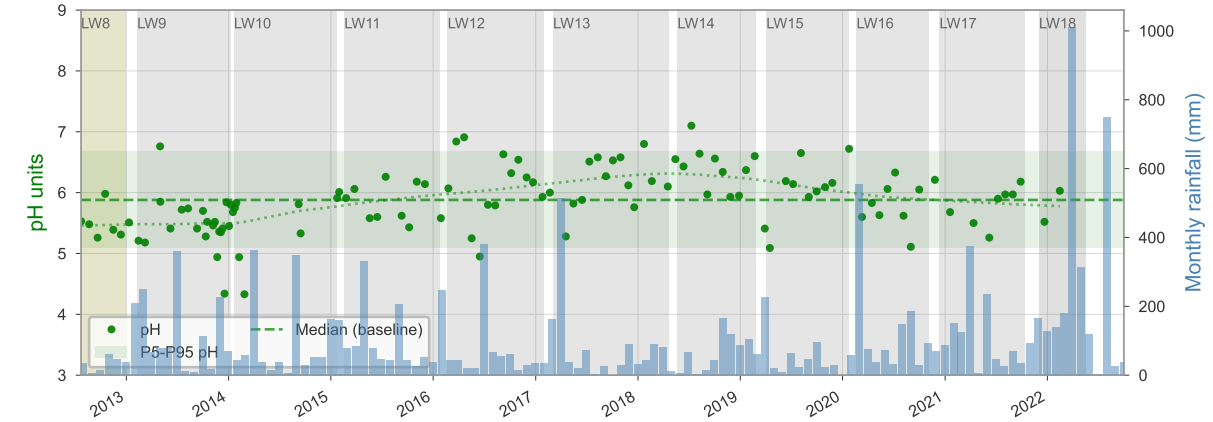
WC21_POOL10



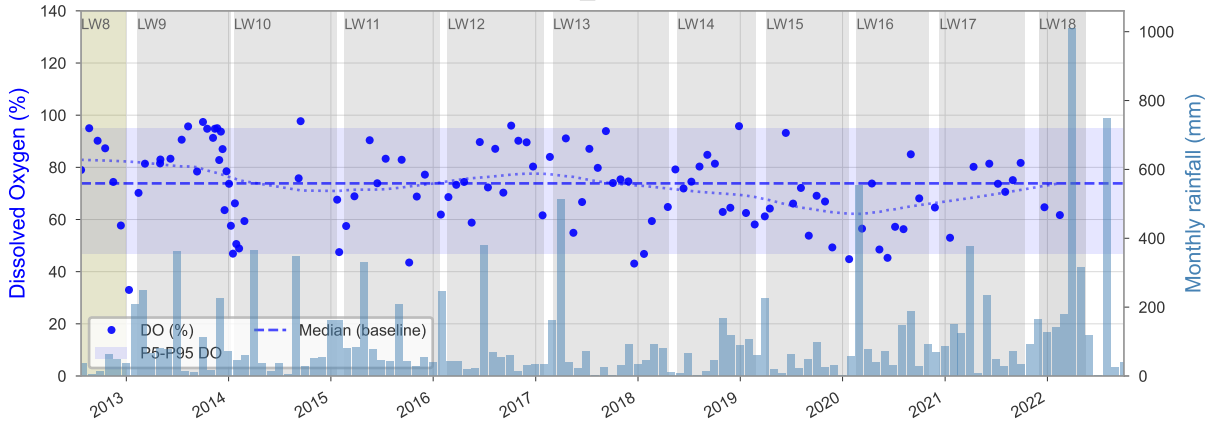
WC21_POOL10



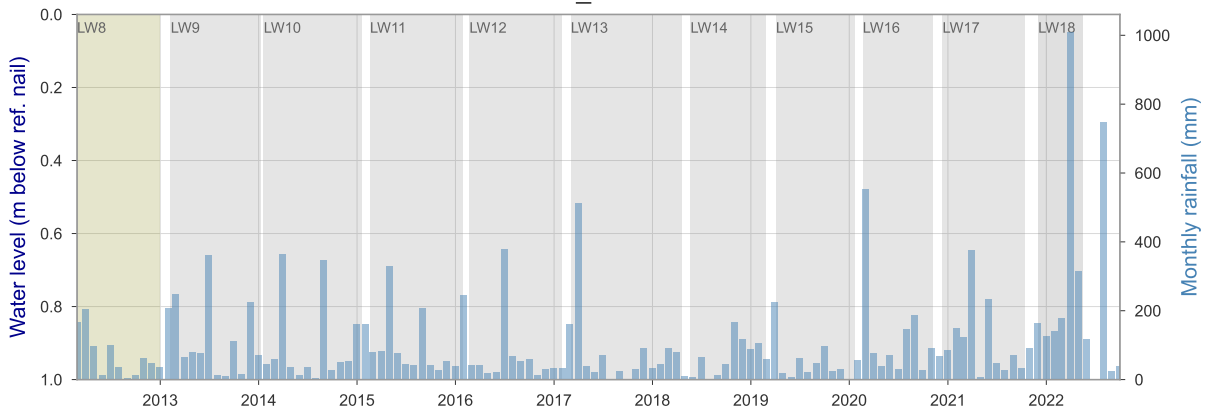
WC21_POOL10



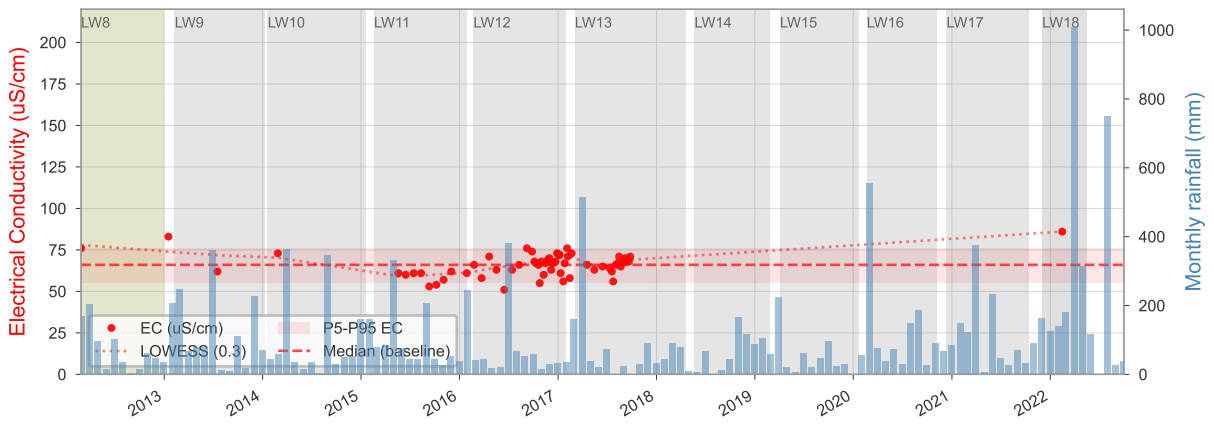
WC21_POOL10



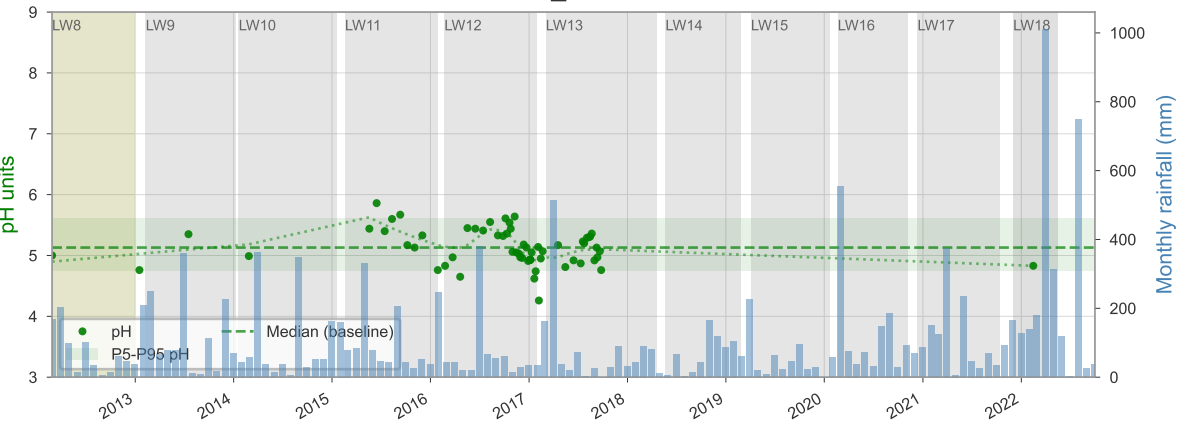
WC21_POOL48



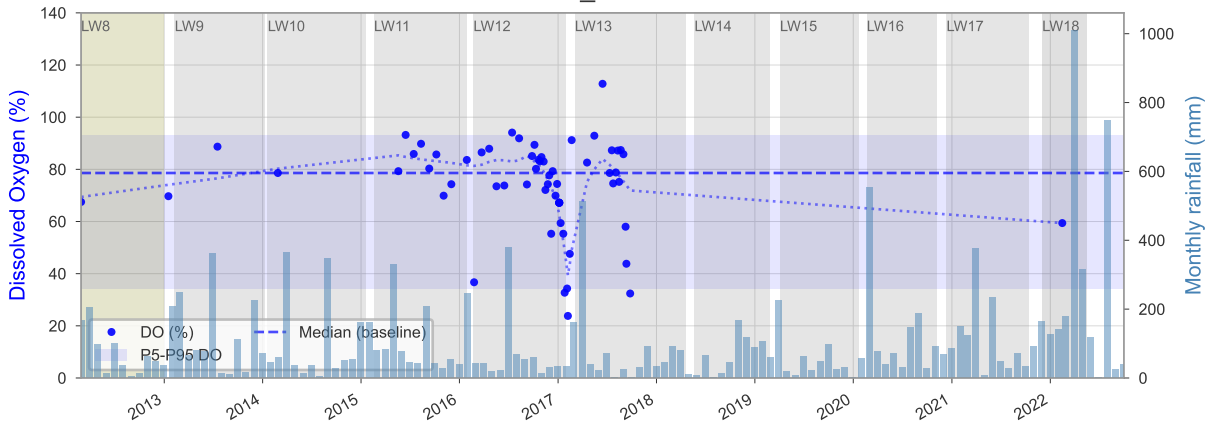
WC21_POOL48



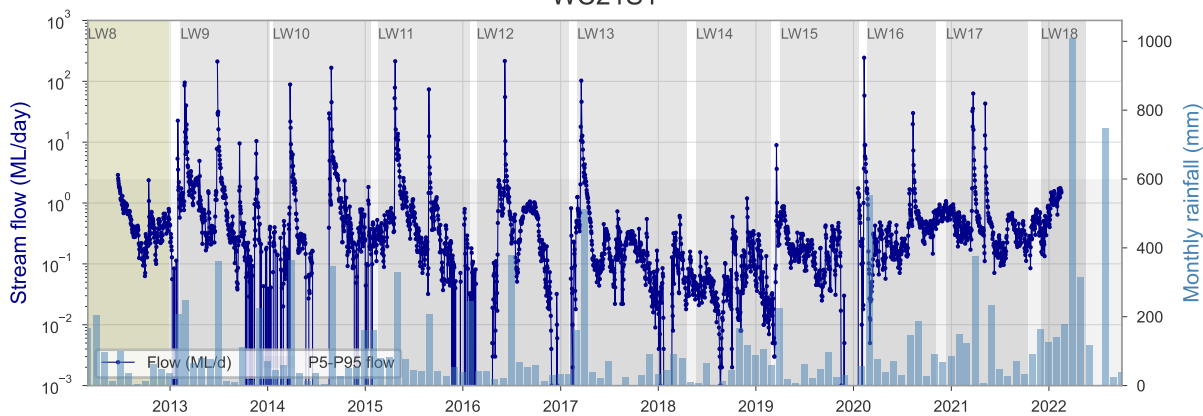
WC21_POOL48



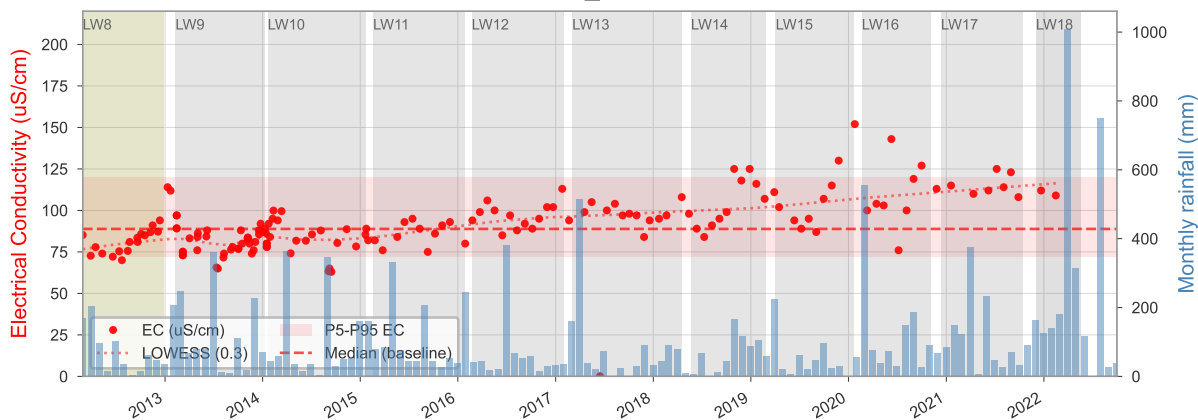
WC21_POOL48



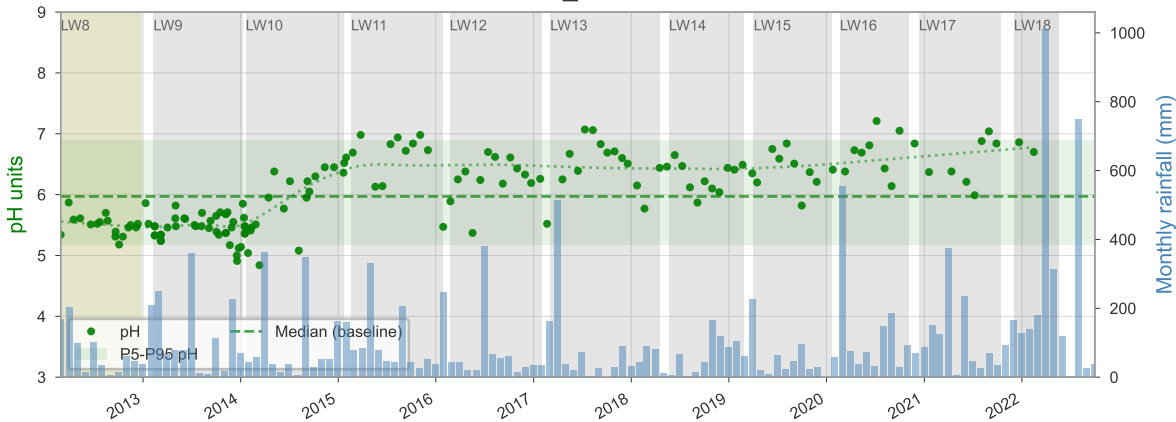
WC21S1



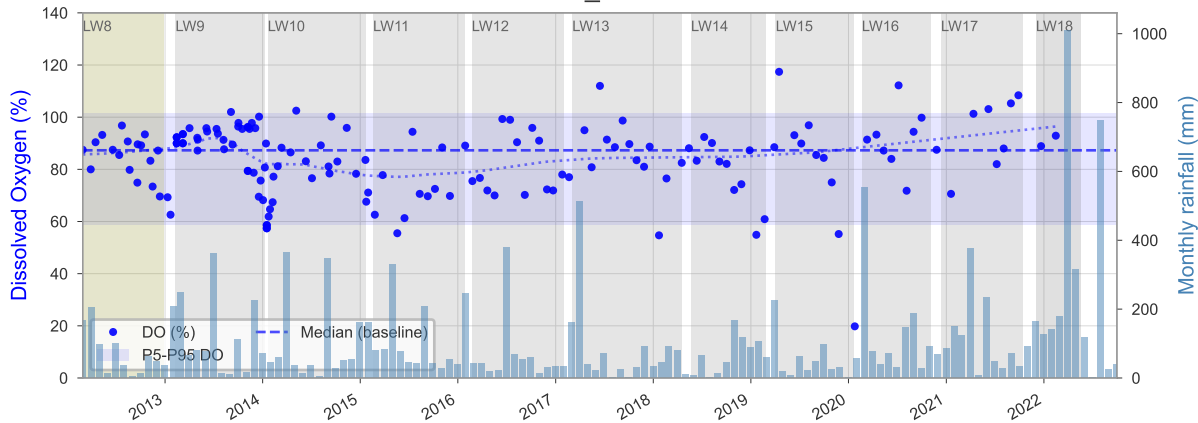
WC21_POOL5



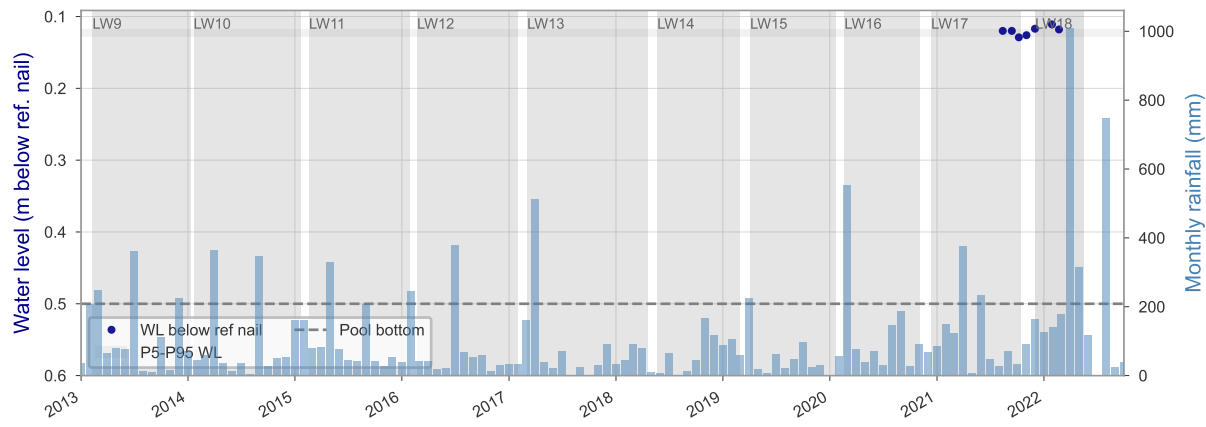
WC21_POOL5



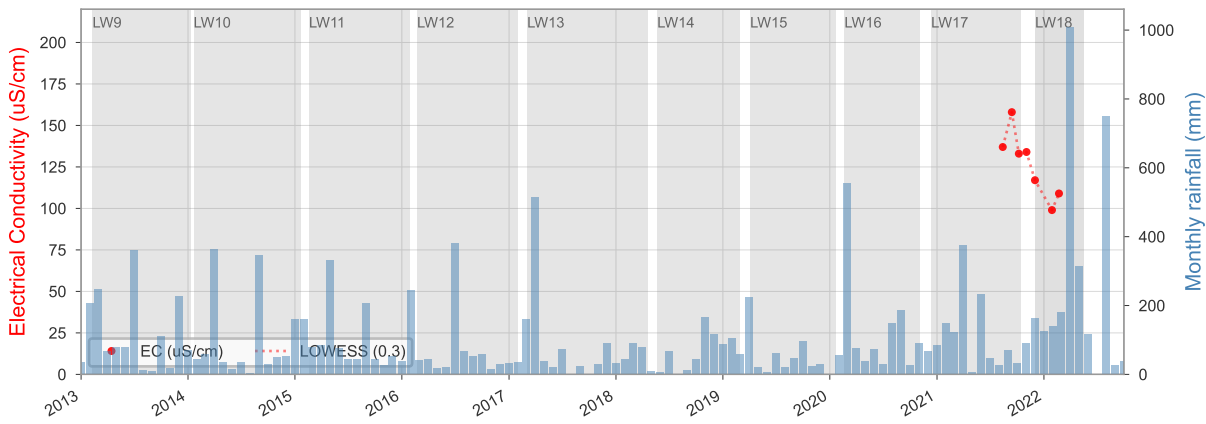
WC21_POOL5



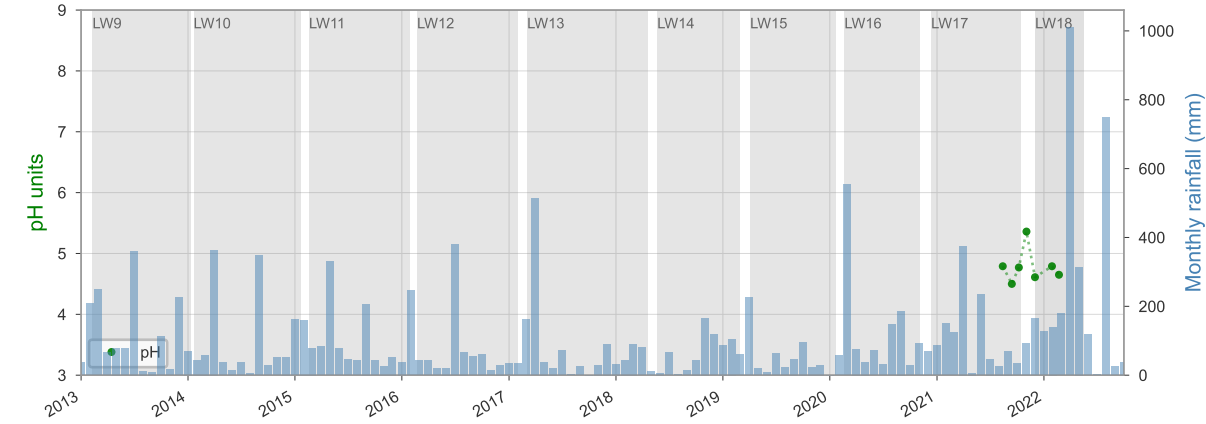
WC24A_POOL1



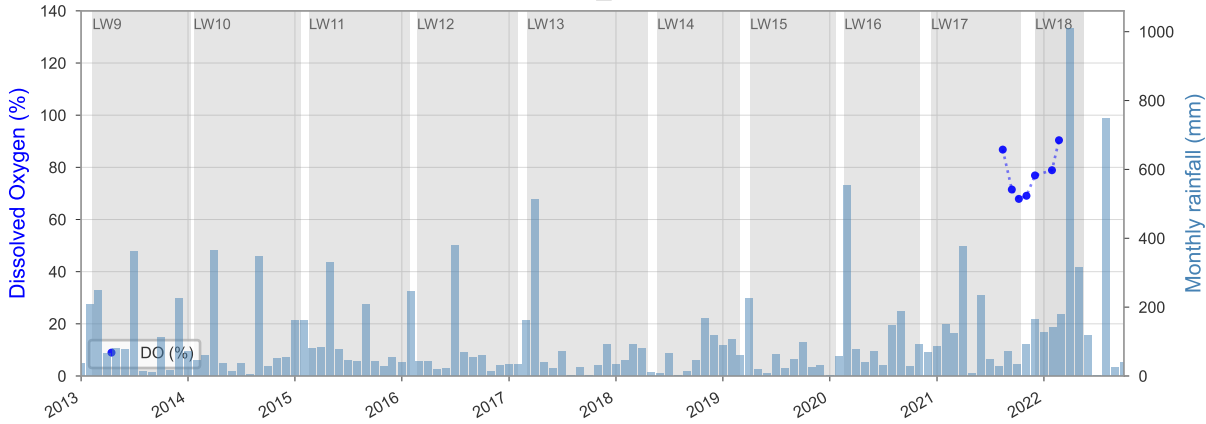
WC24A_POOL1



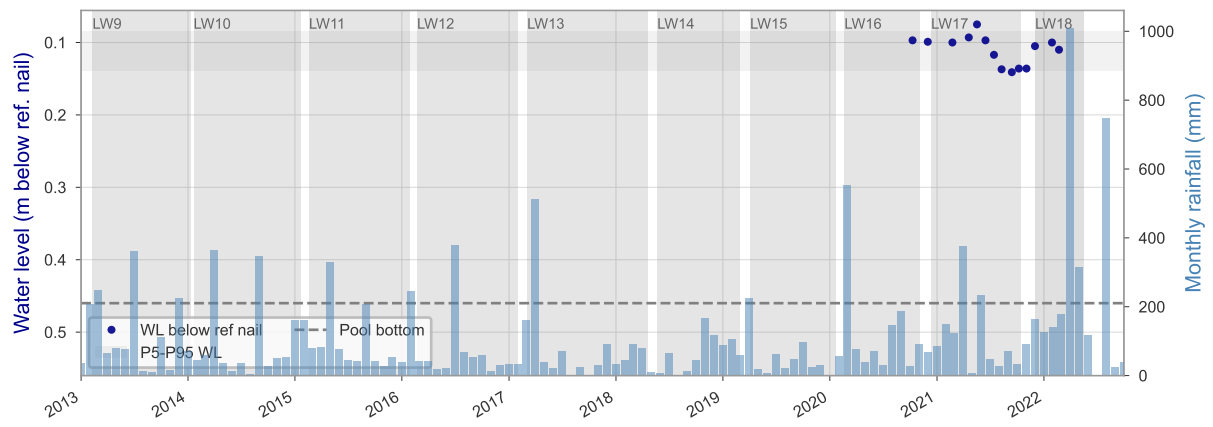
WC24A_POOL1



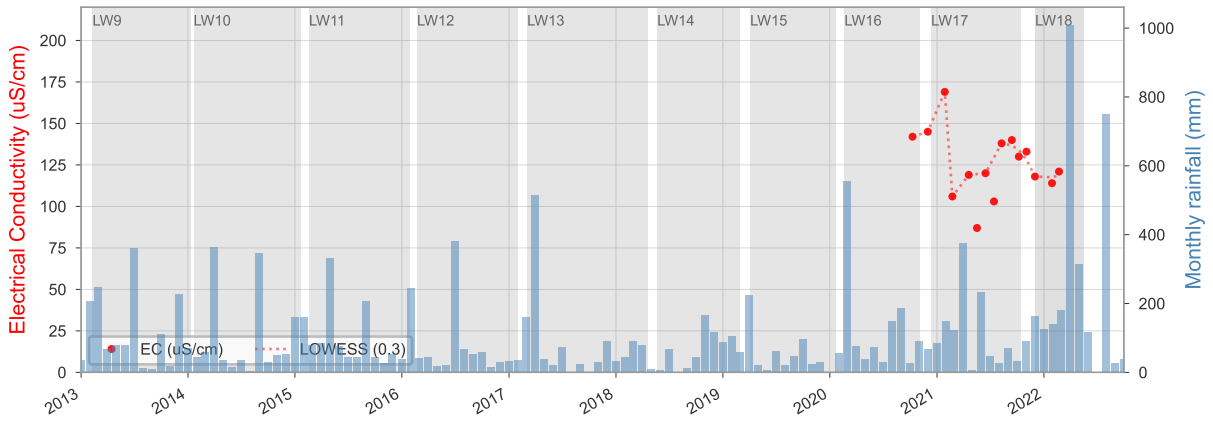
WC24A_POOL1



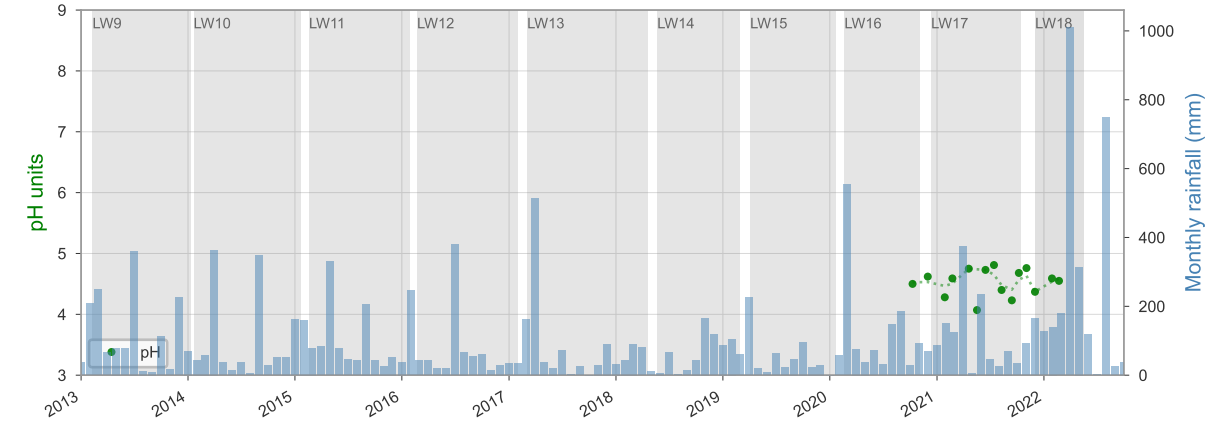
WC24_POOL10



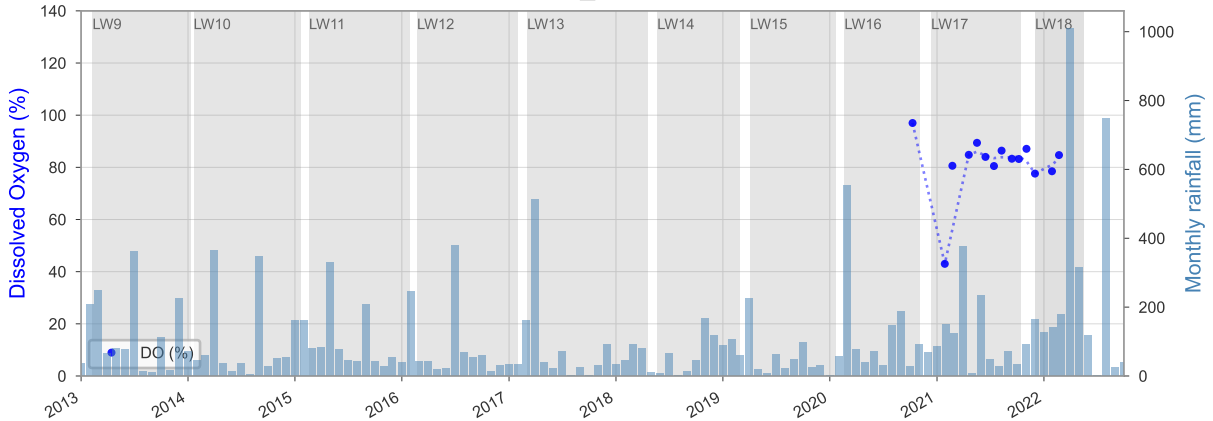
WC24_POOL10



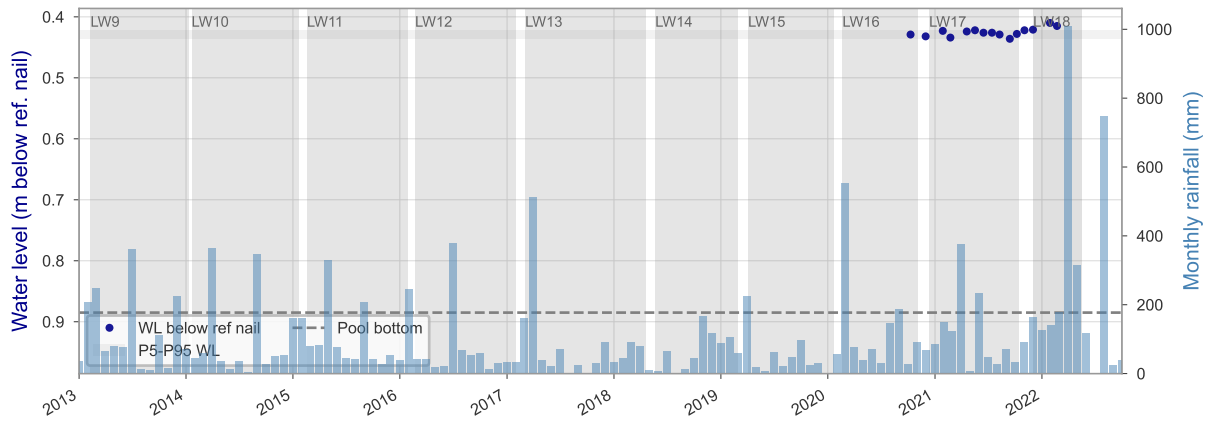
WC24_POOL10



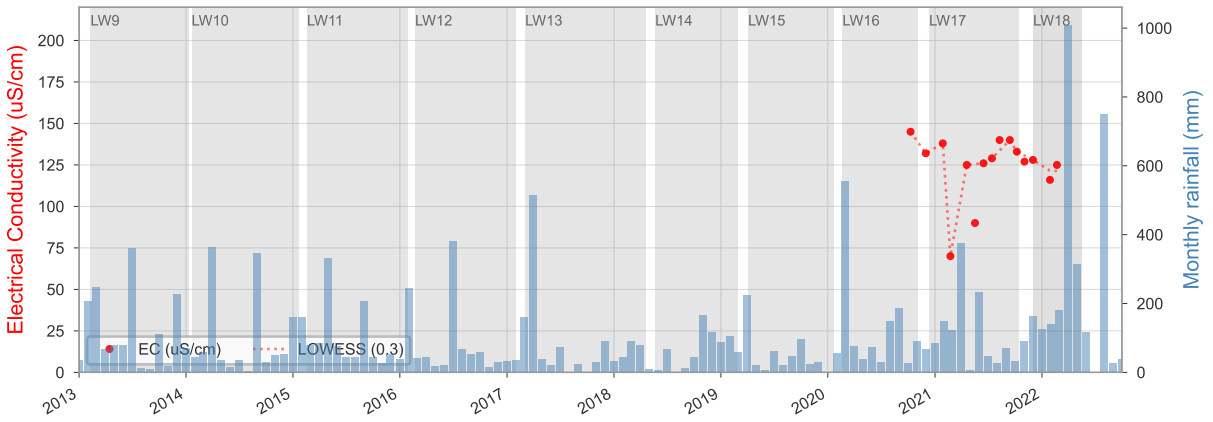
WC24_POOL10



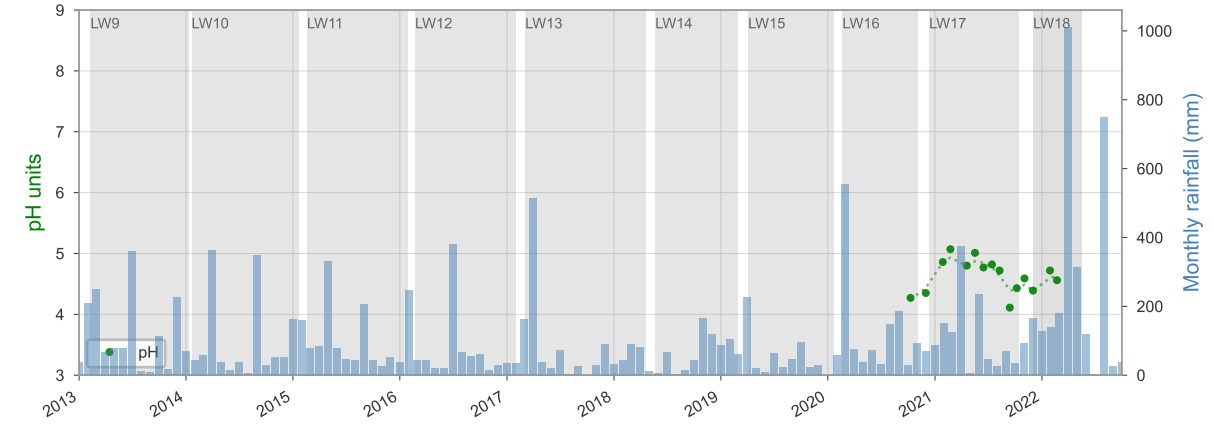
WC24_POOL22



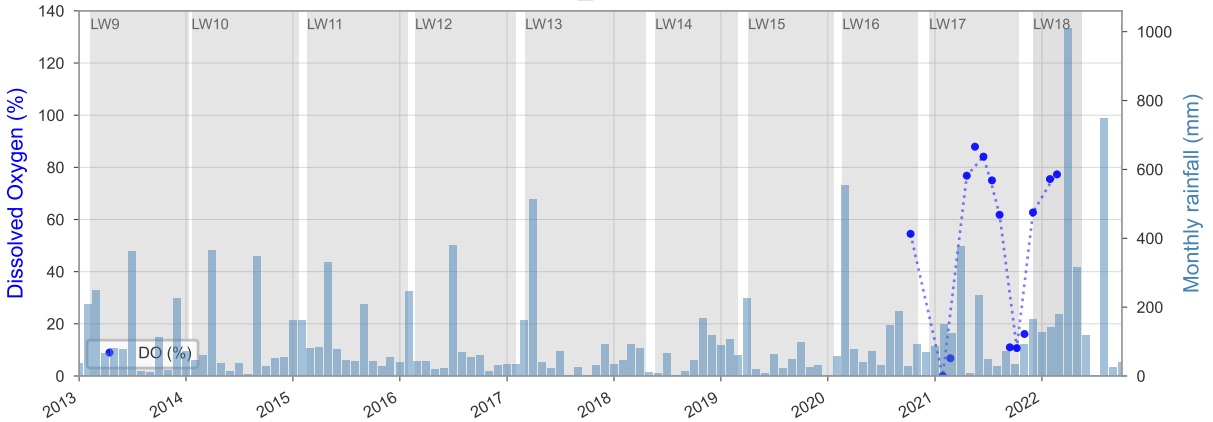
WC24_POOL22



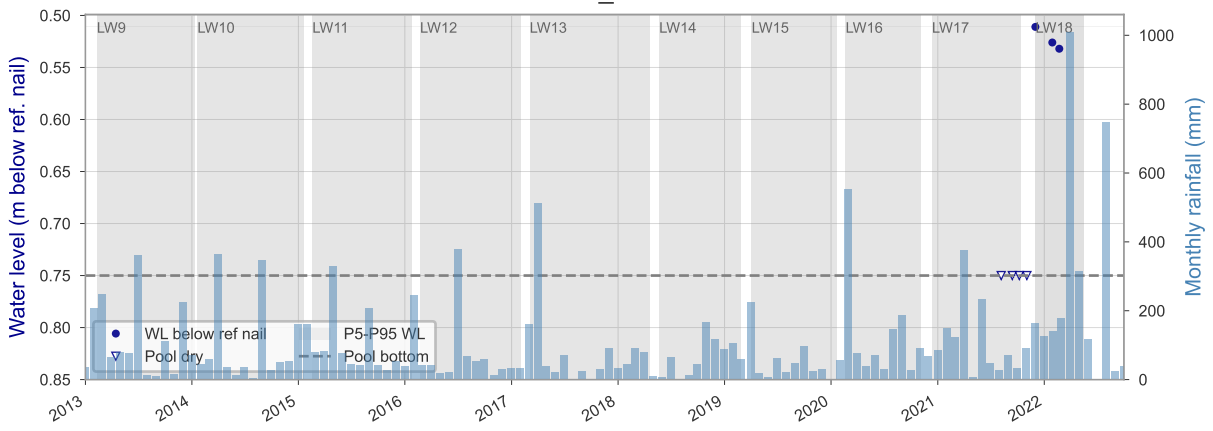
WC24_POOL22



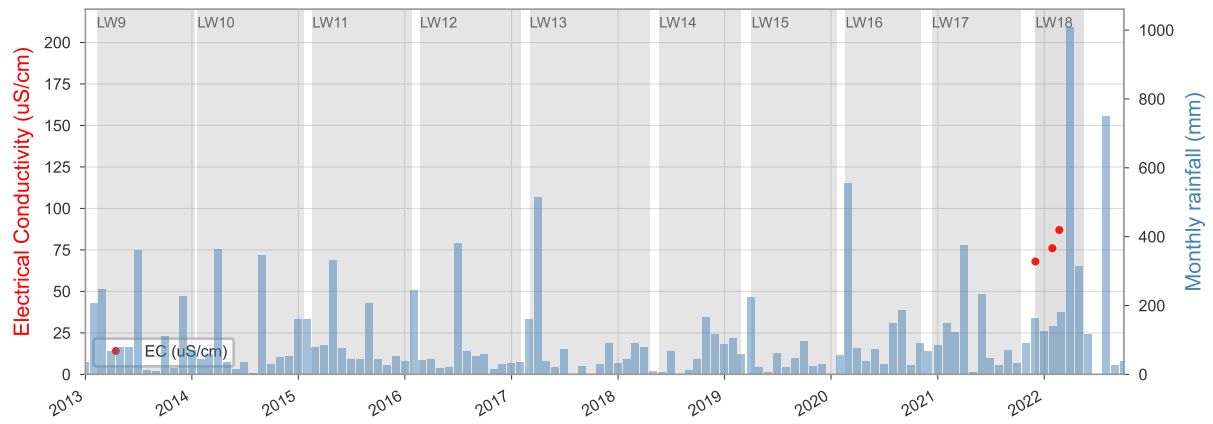
WC24_POOL22



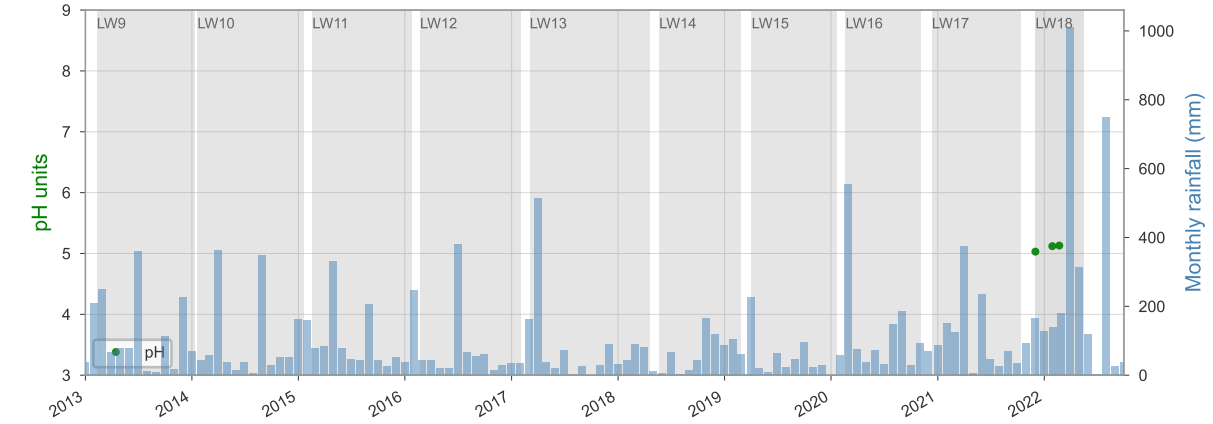
WC26A_POOL4



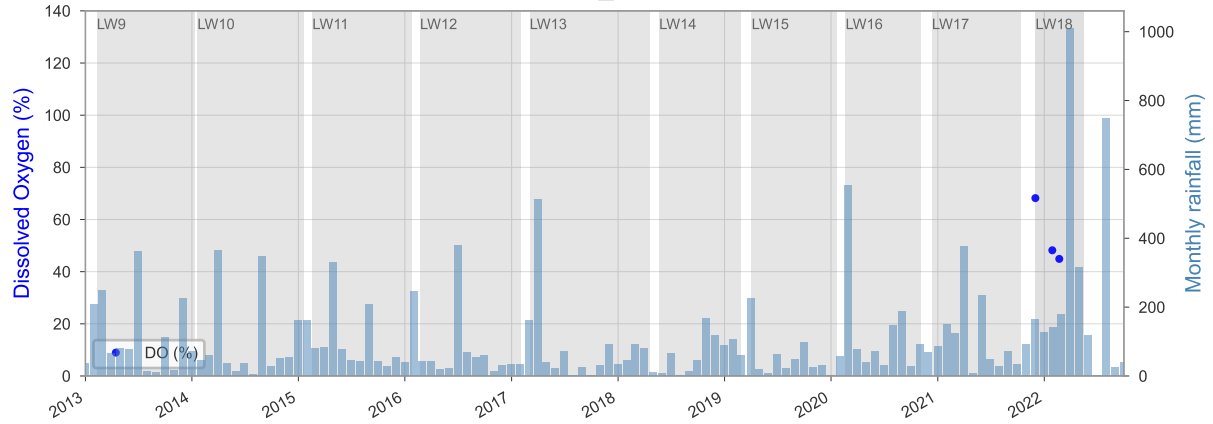
WC26A_POOL4



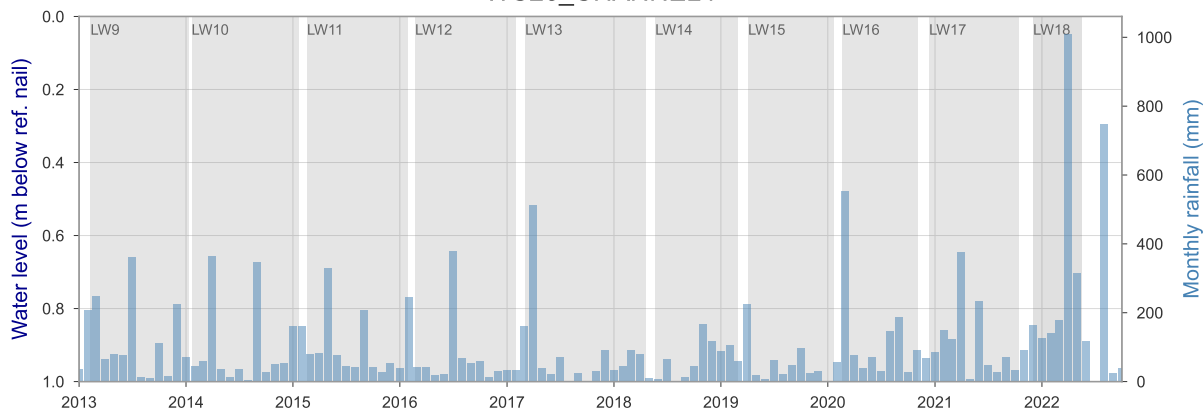
WC26A_POOL4



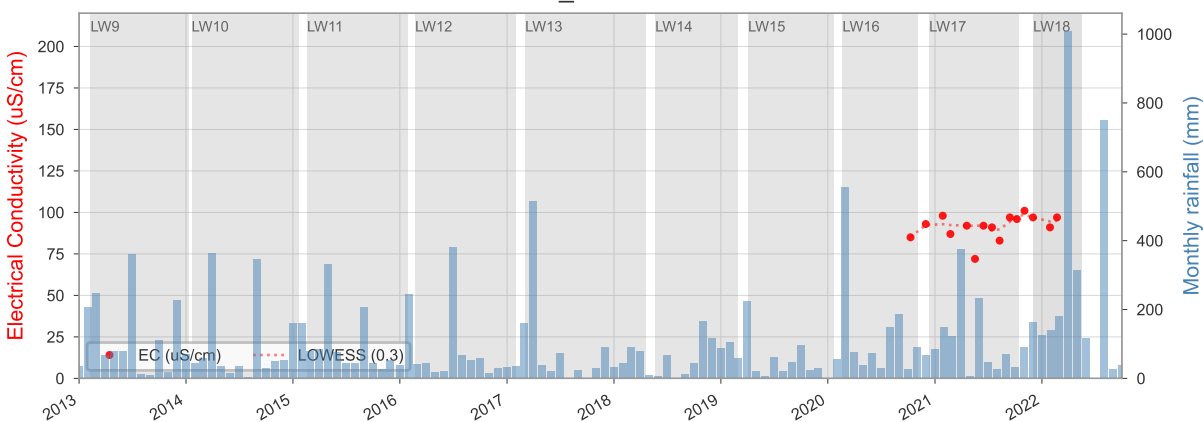
WC26A_POOL4



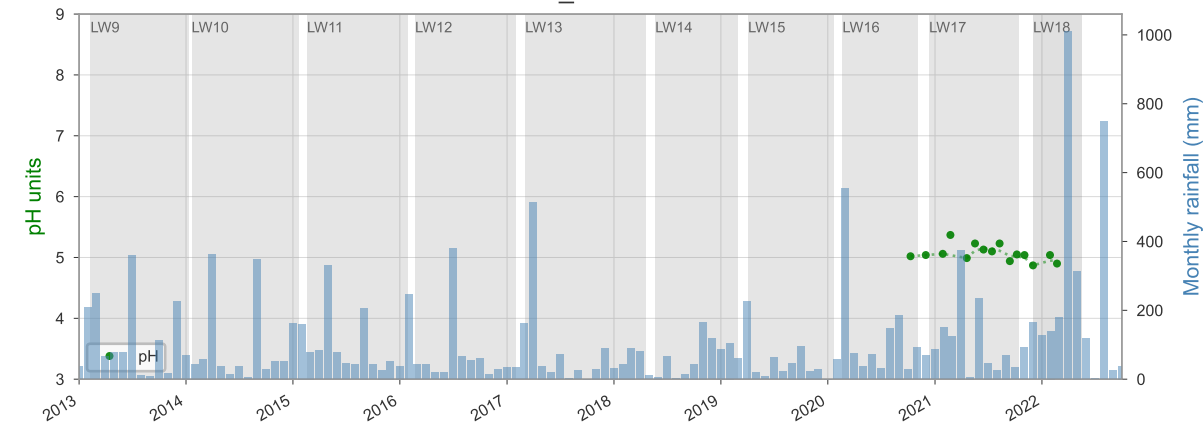
WC26_CHANNEL4



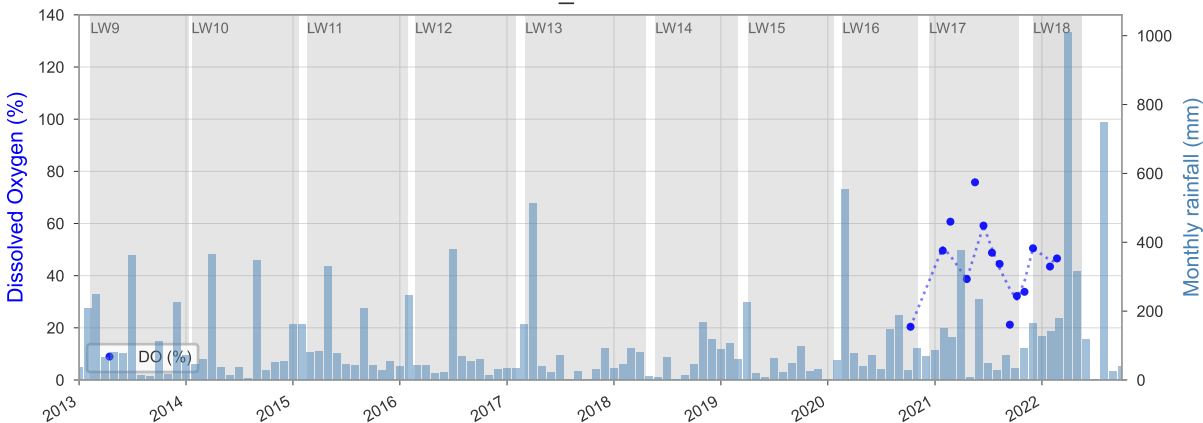
WC26_CHANNEL4



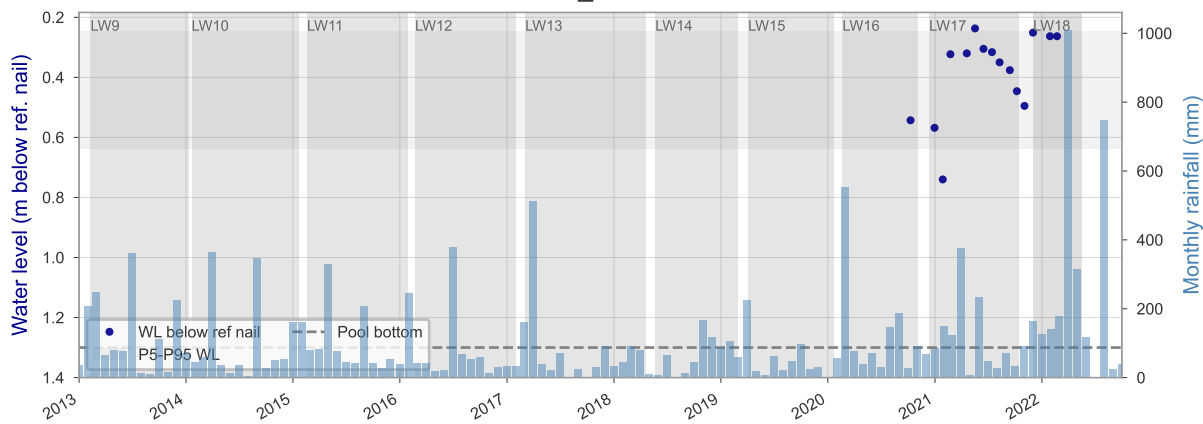
WC26_CHANNEL4



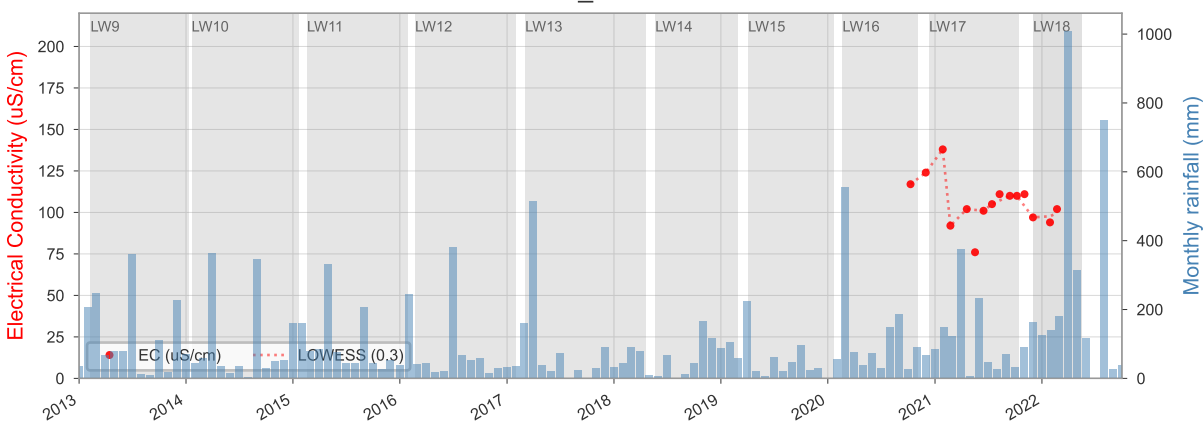
WC26_CHANNEL4



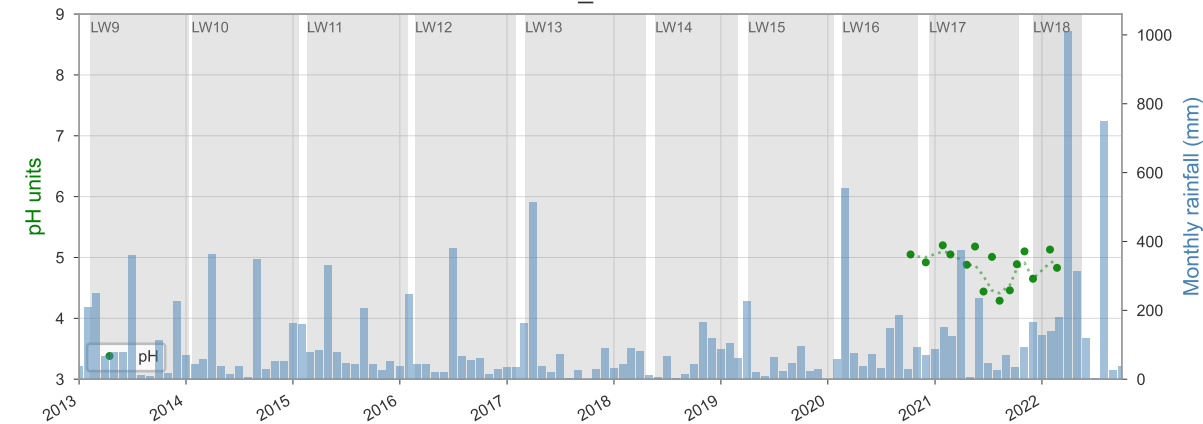
WC26_POOL14



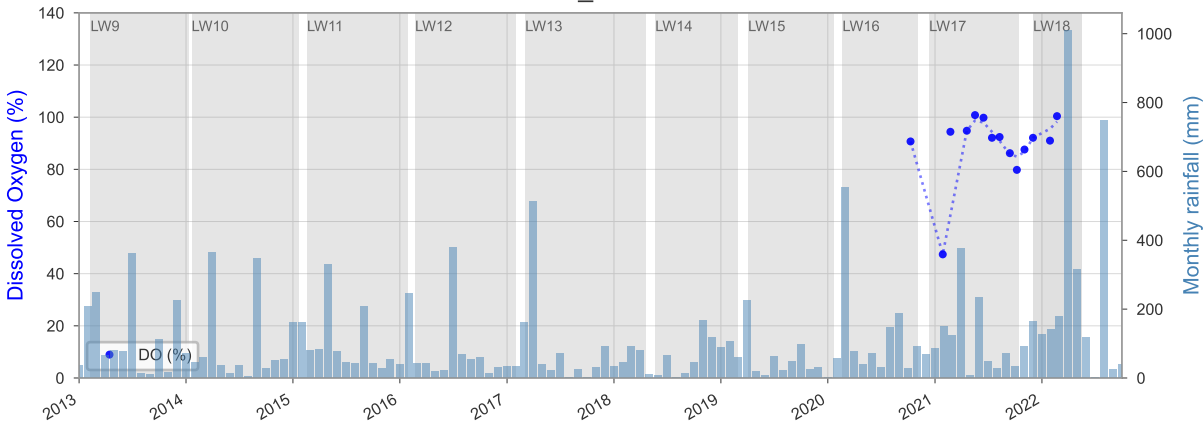
WC26_POOL14



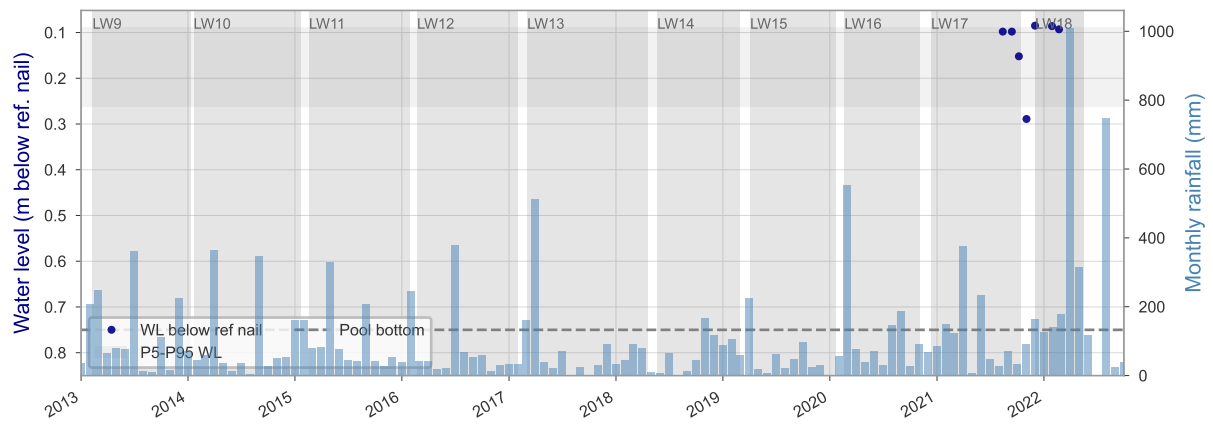
WC26_POOL14



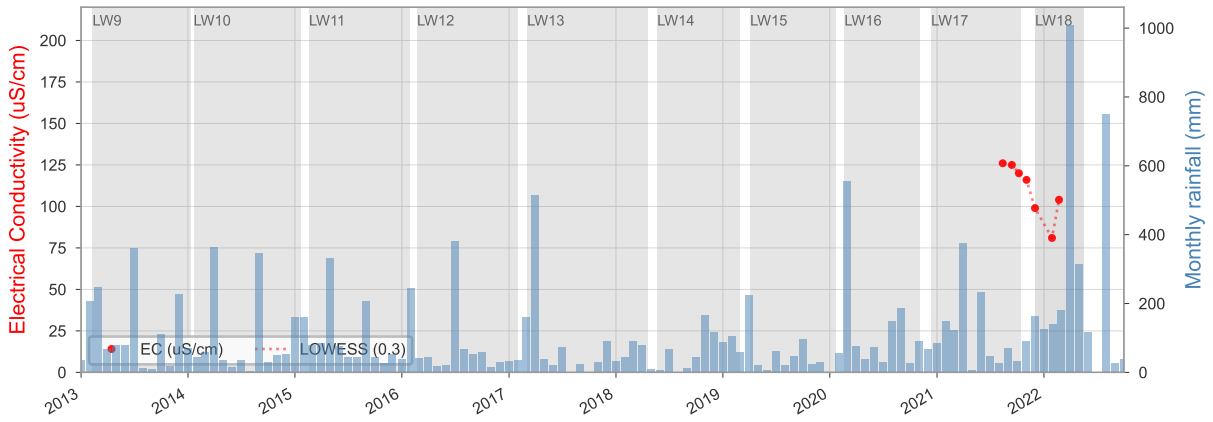
WC26_POOL14



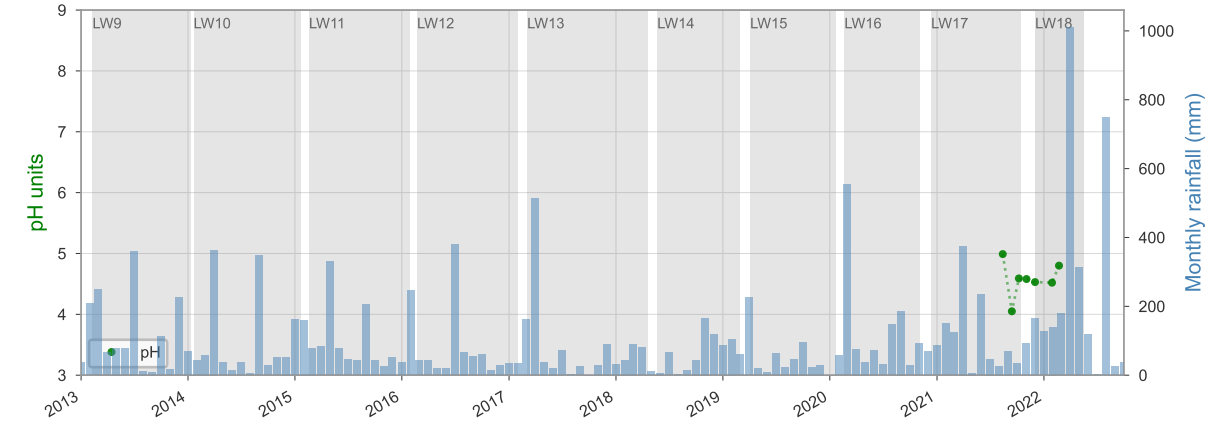
WC26_POOL19



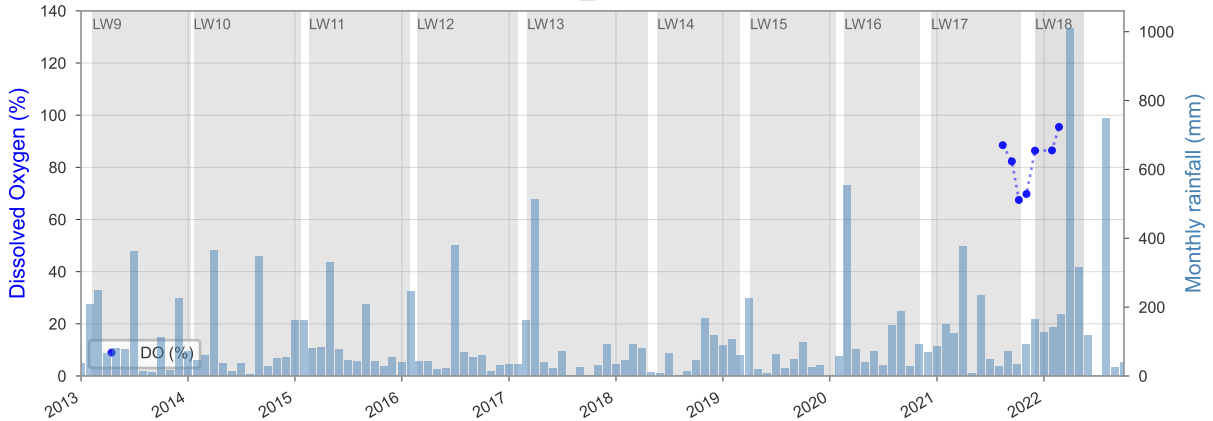
WC26_POOL19



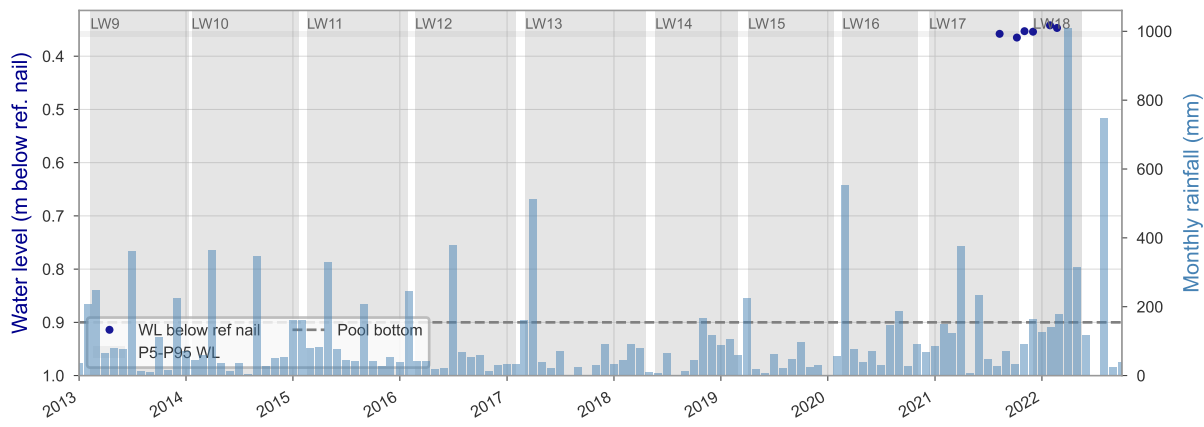
WC26_POOL19



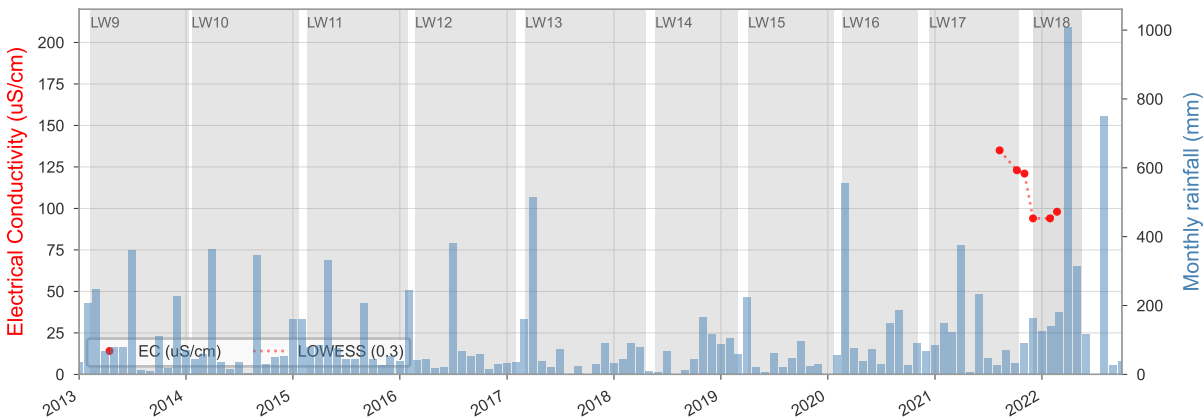
WC26_POOL19



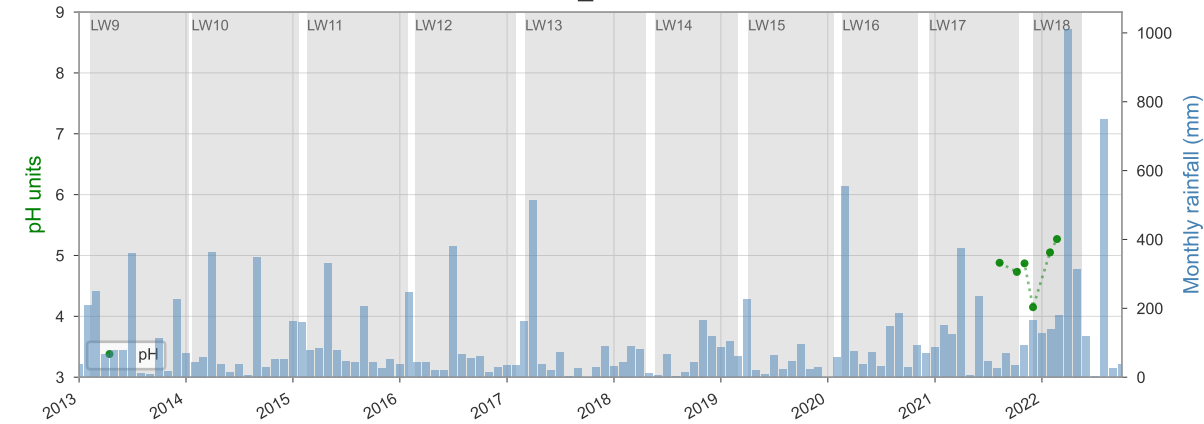
WC26_POOL32



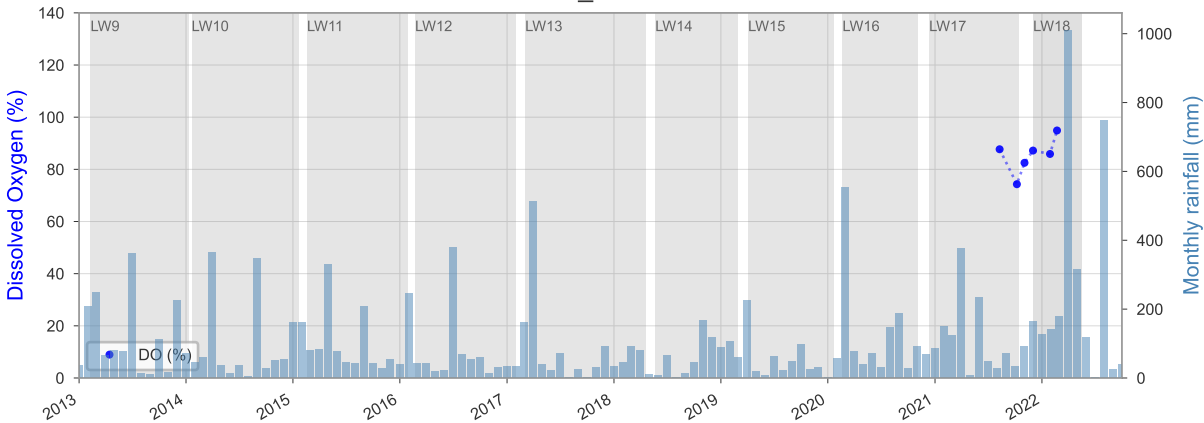
WC26_POOL32



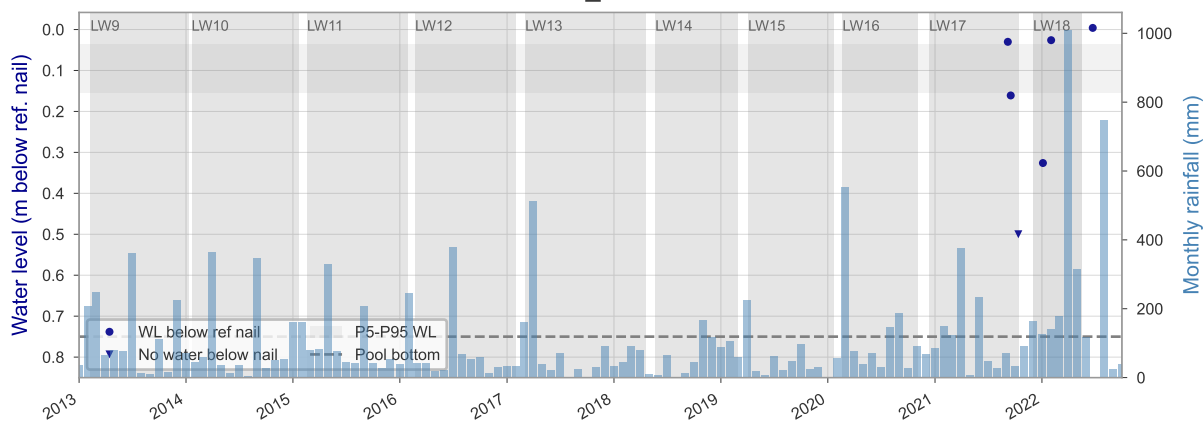
WC26_POOL32



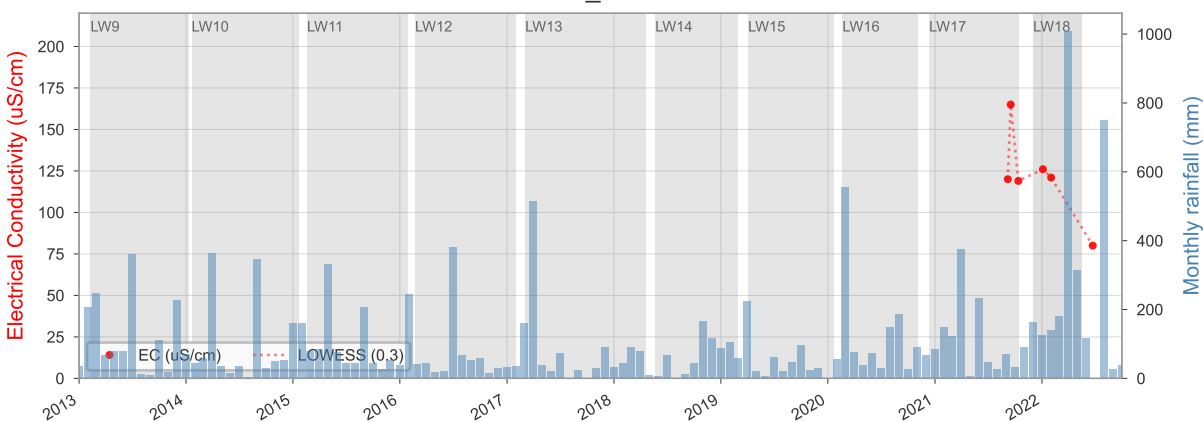
WC26_POOL32



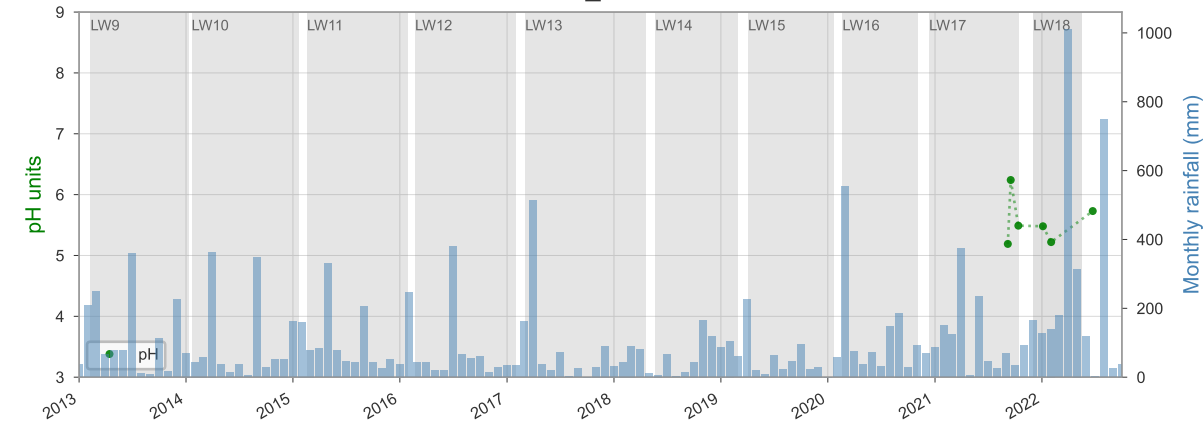
WC28_POOL5



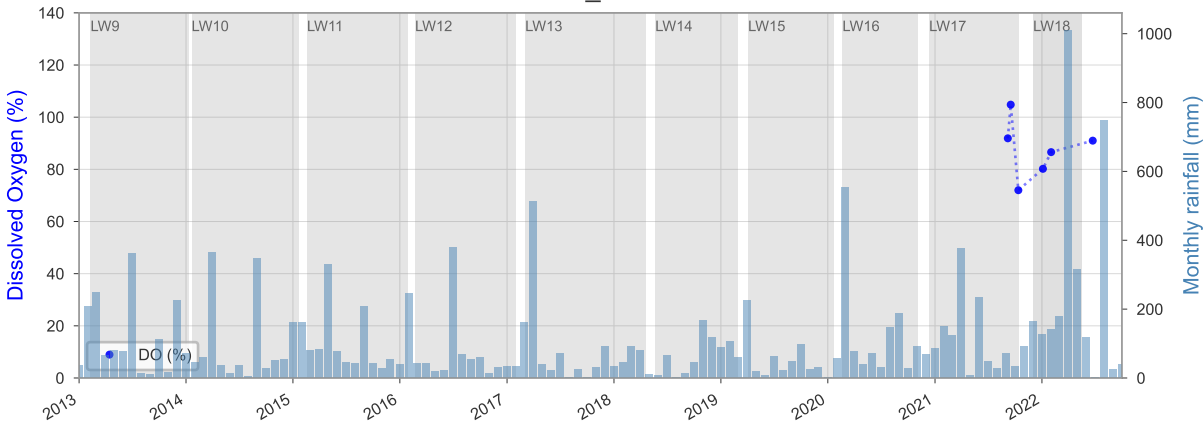
WC28_POOL5



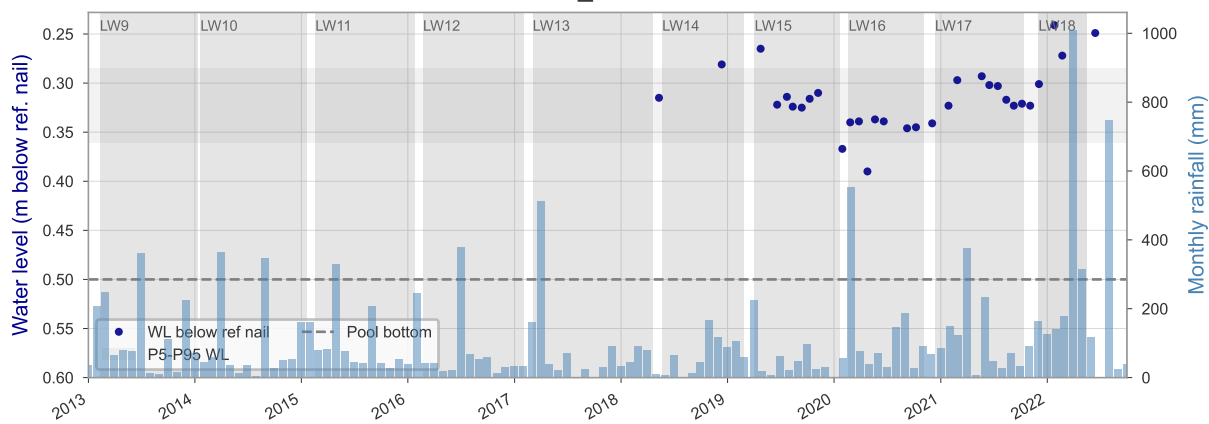
WC28_POOL5



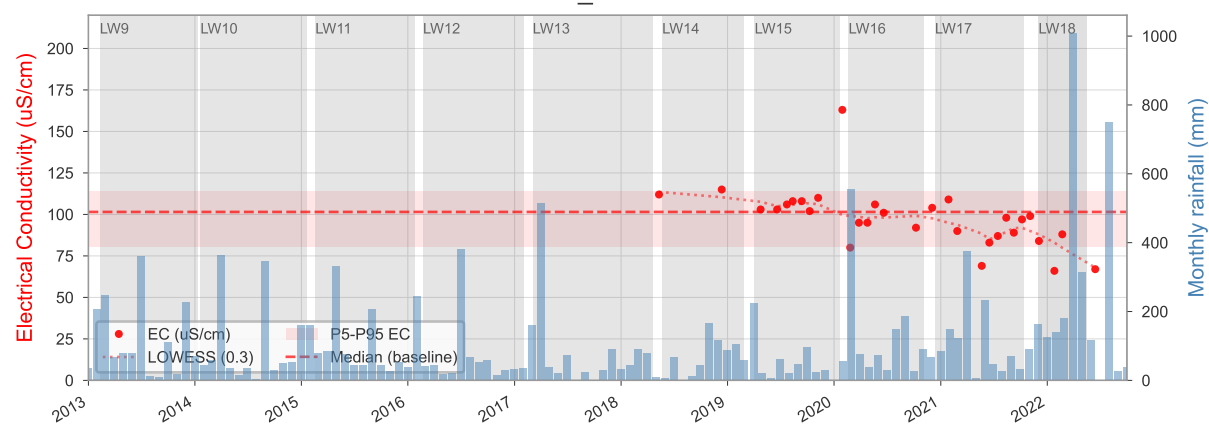
WC28_POOL5



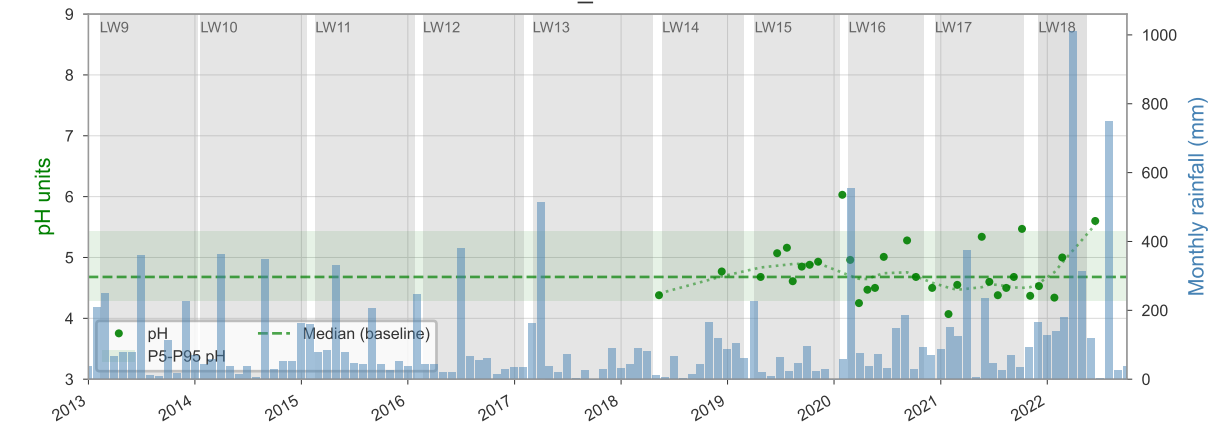
WC6_POOL10



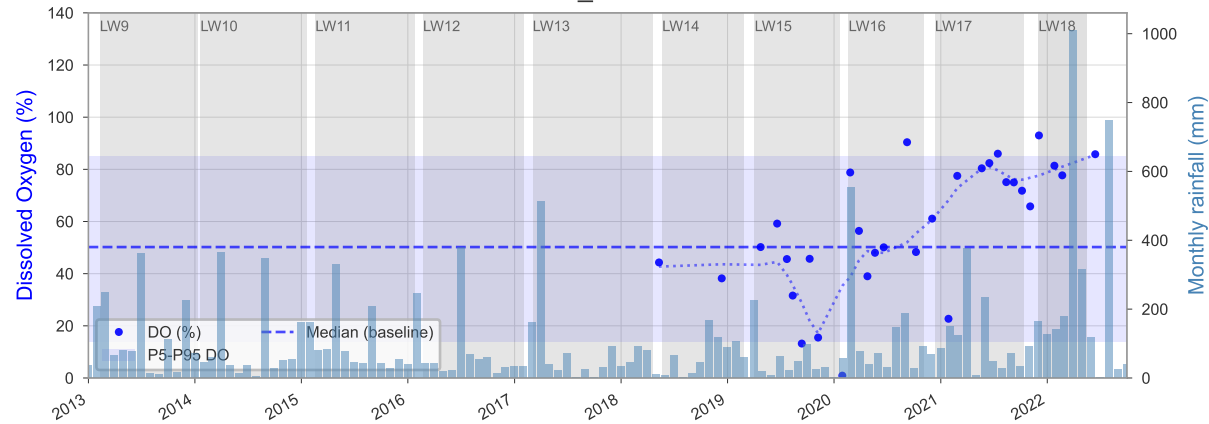
WC6_POOL10



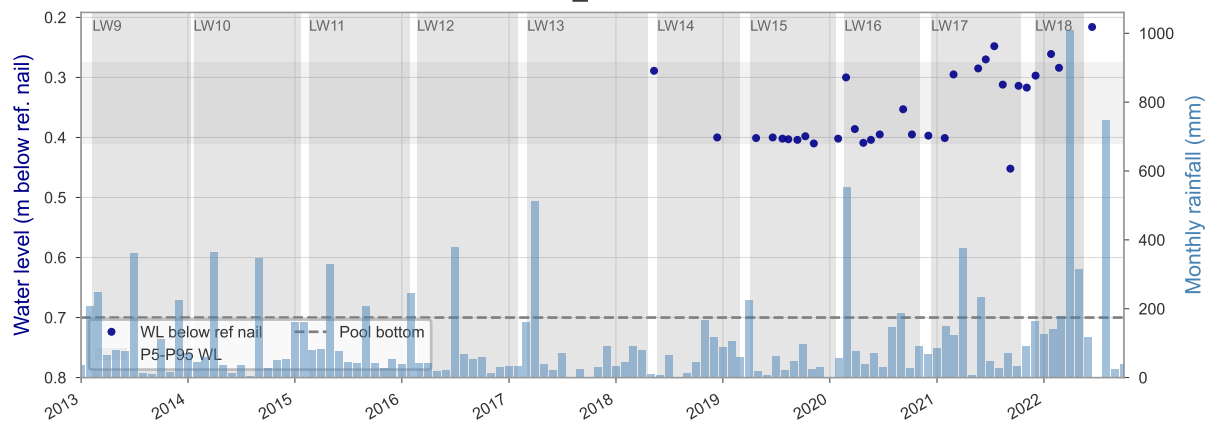
WC6_POOL10



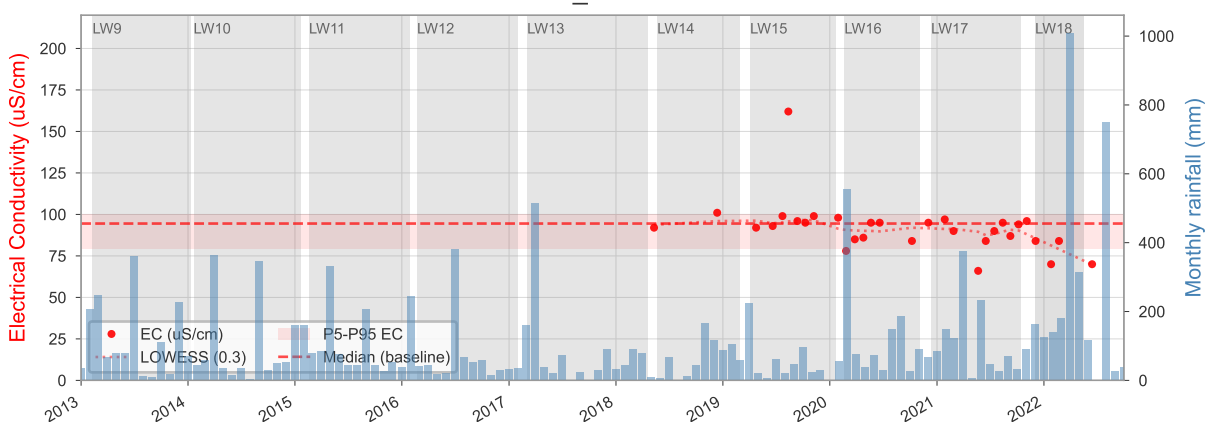
WC6_POOL10



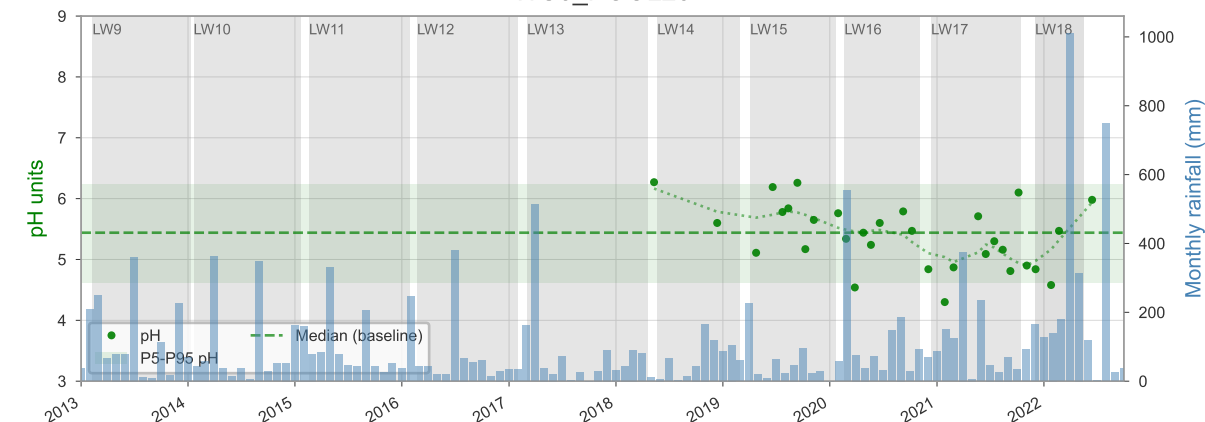
WC6_POOL20



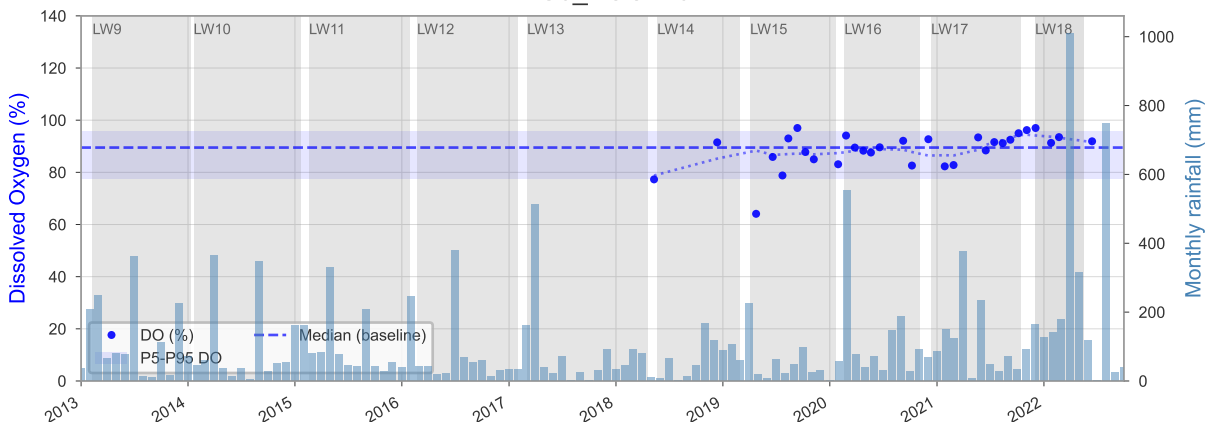
WC6_POOL20



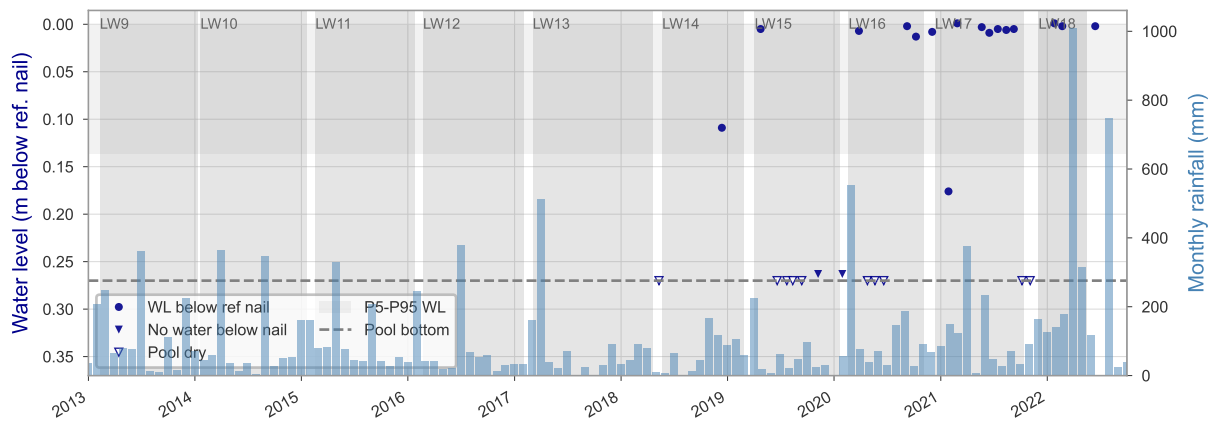
WC6_POOL20



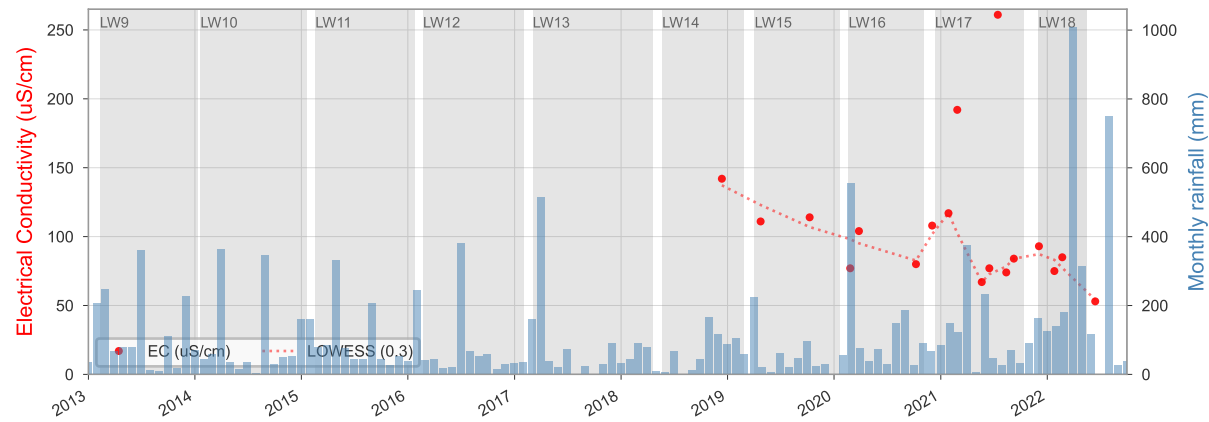
WC6_POOL20



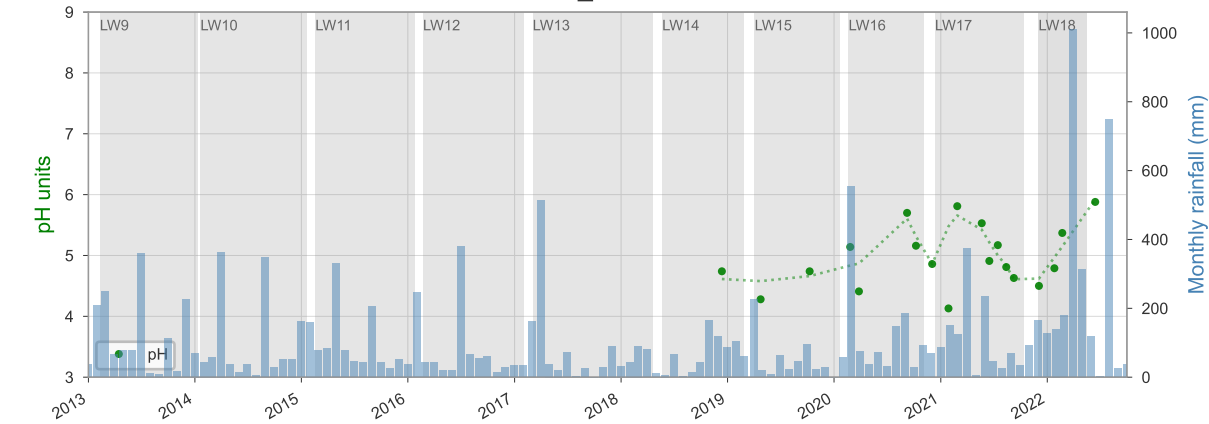
WC6_POOL30



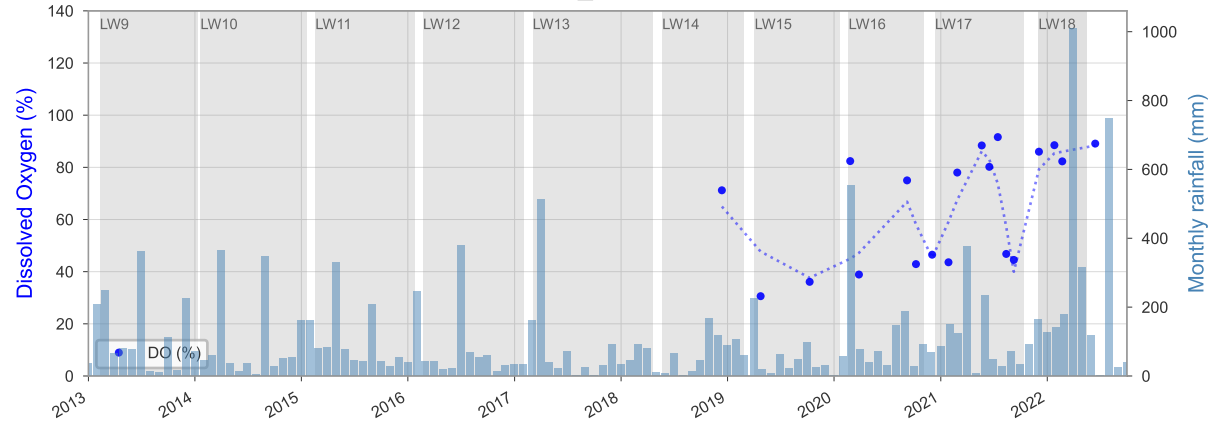
WC6_POOL30



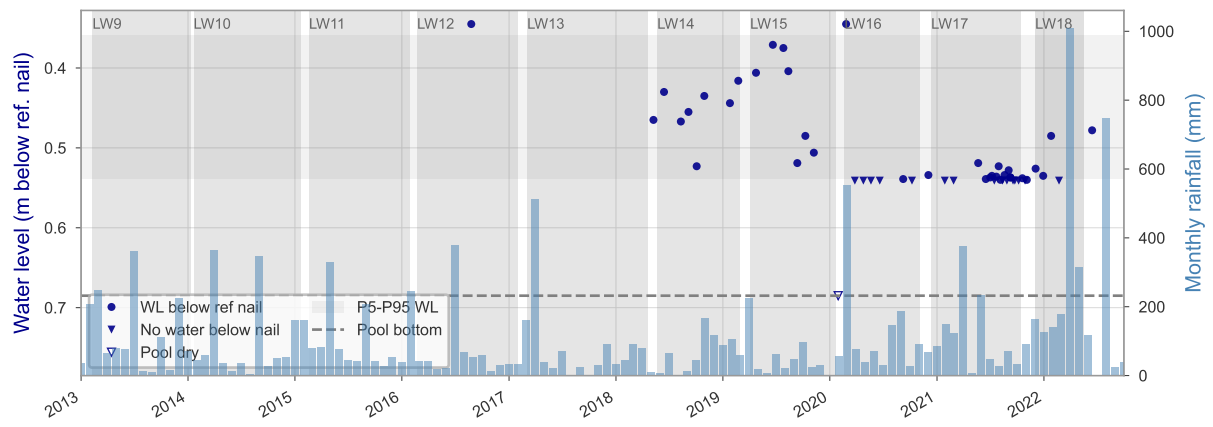
WC6_POOL30



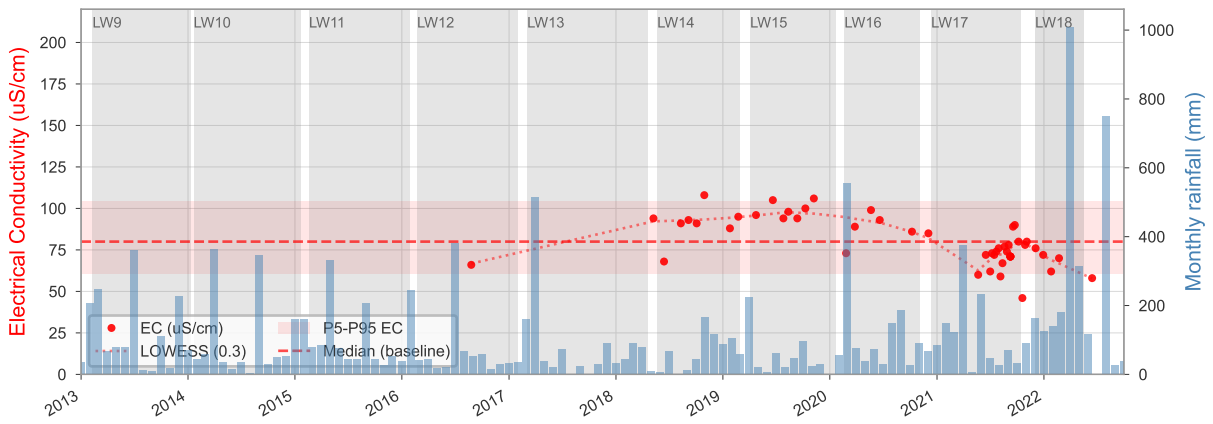
WC6_POOL30



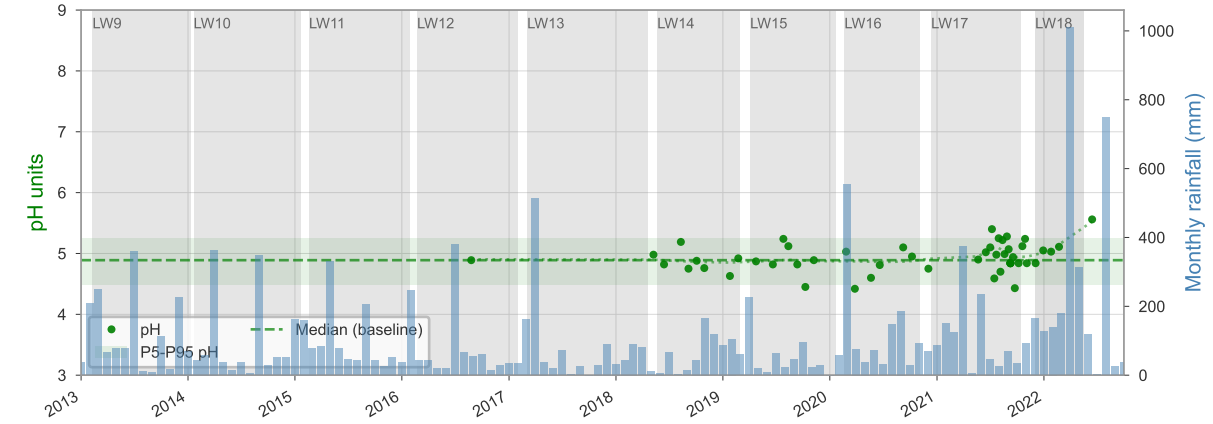
WC7_POOL1



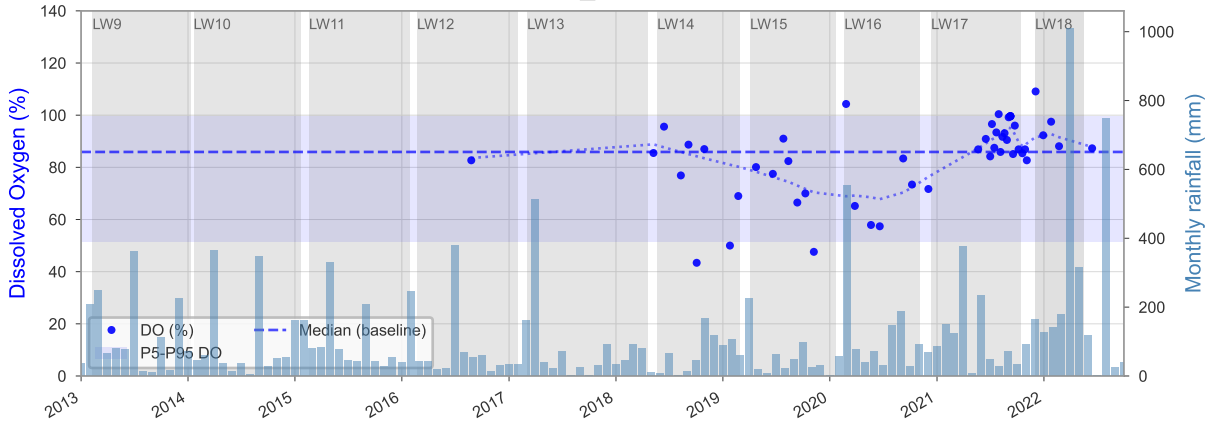
WC7_POOL1



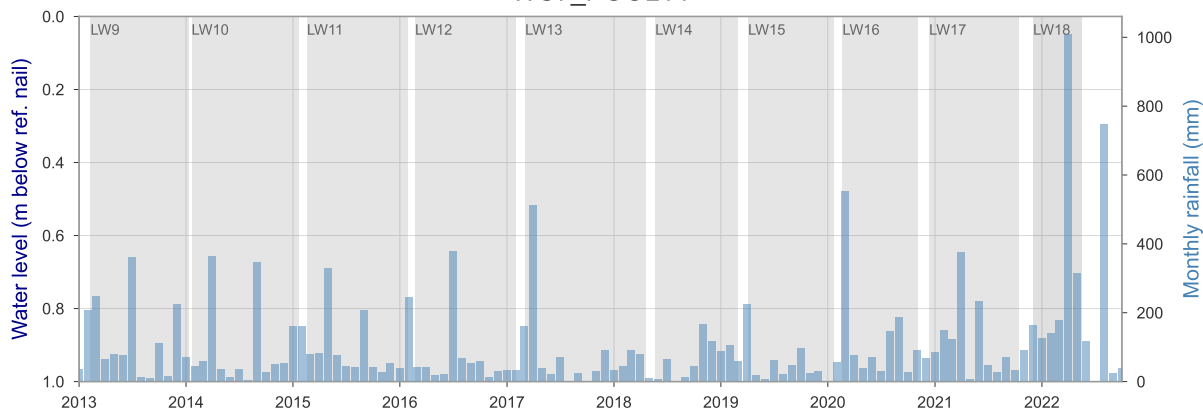
WC7_POOL1



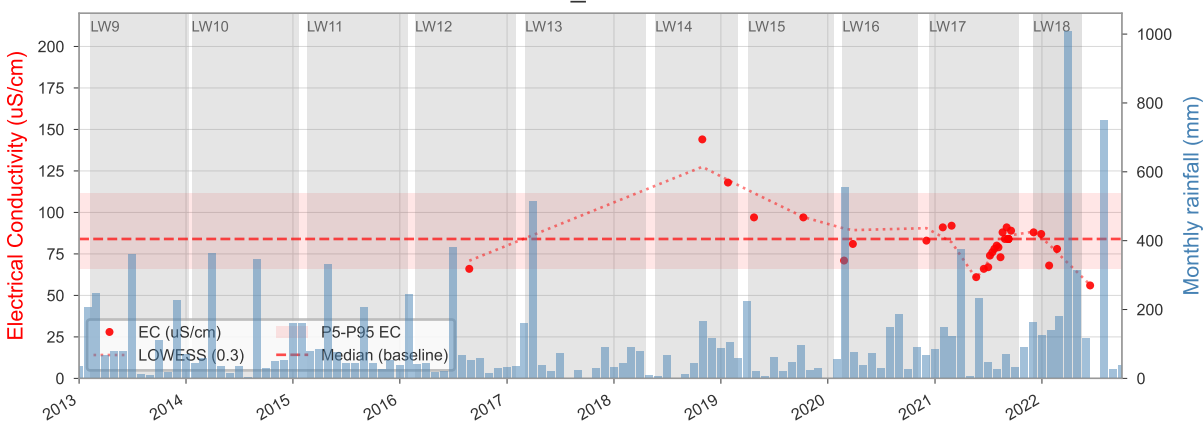
WC7_POOL1



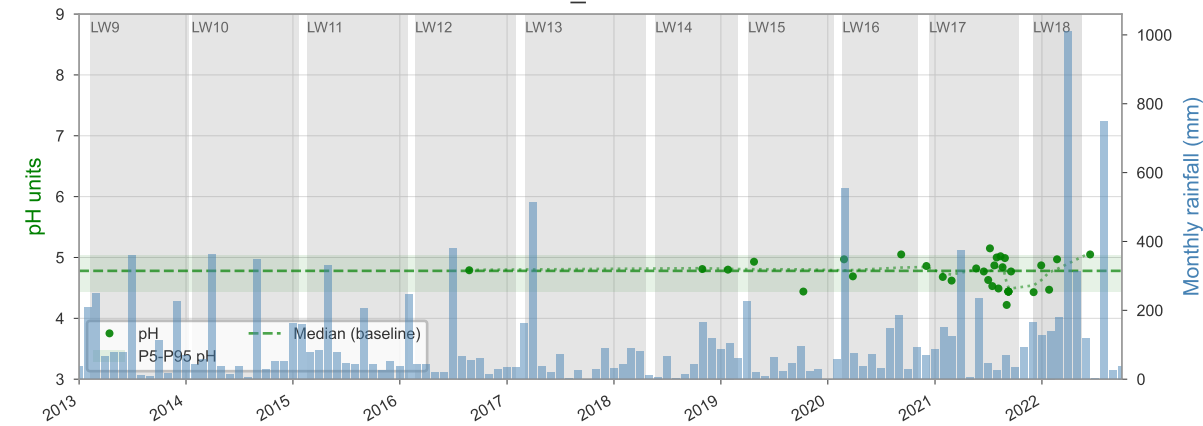
WC7_POOL14



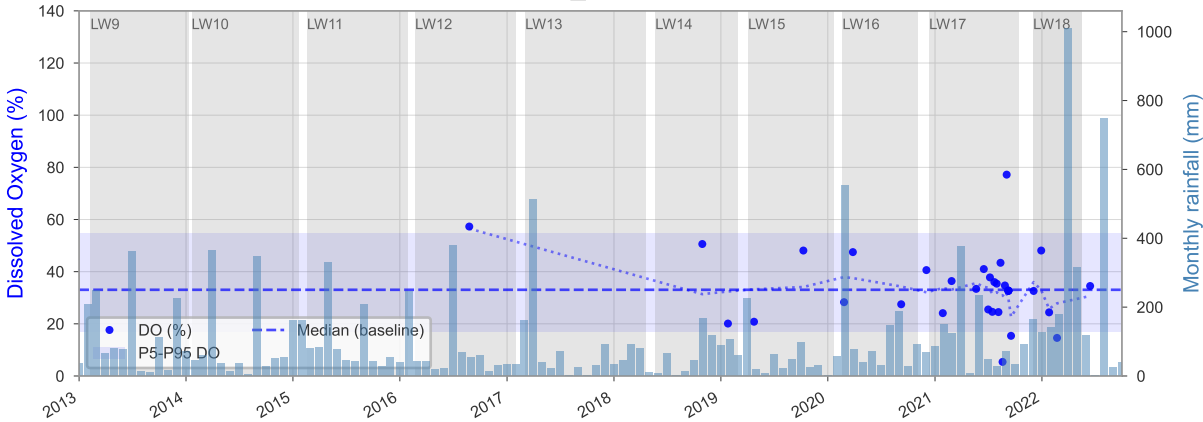
WC7_POOL14



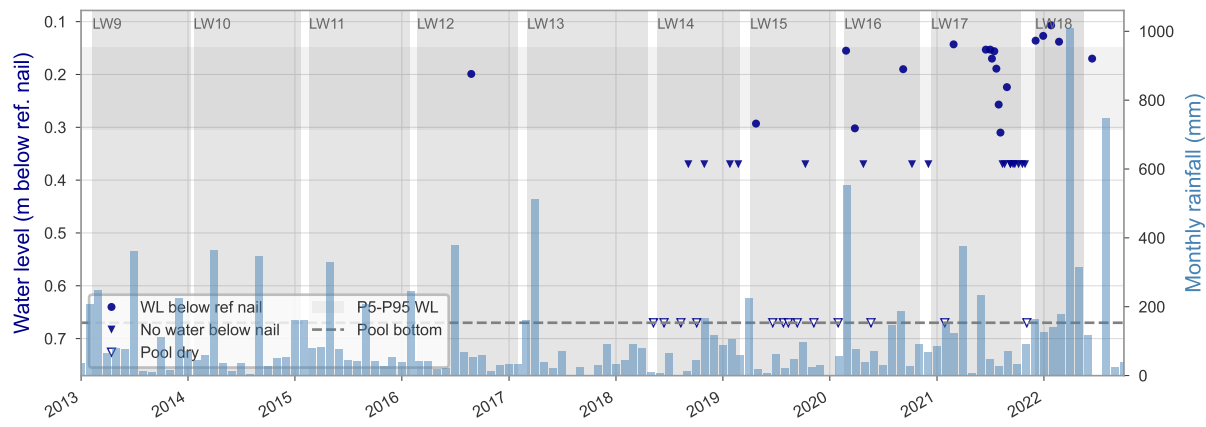
WC7_POOL14



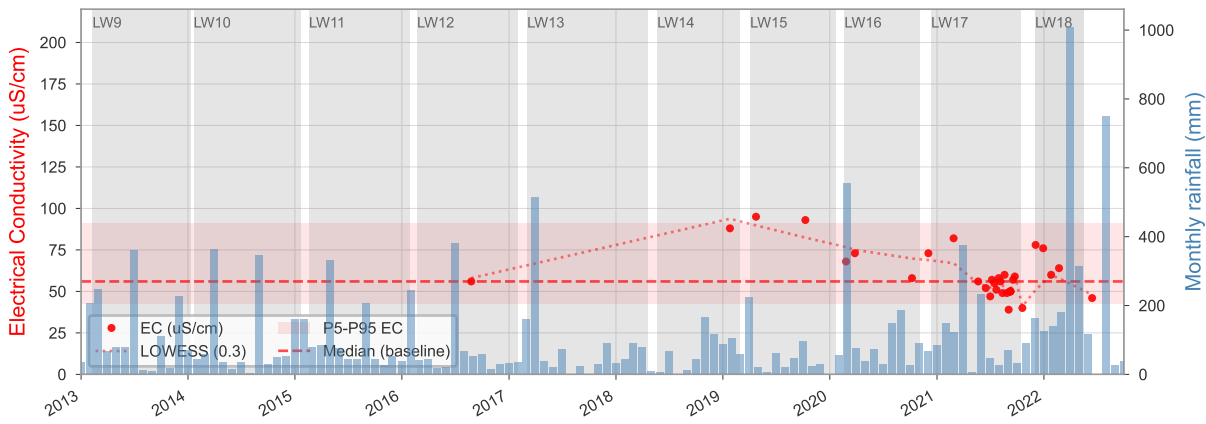
WC7_POOL14



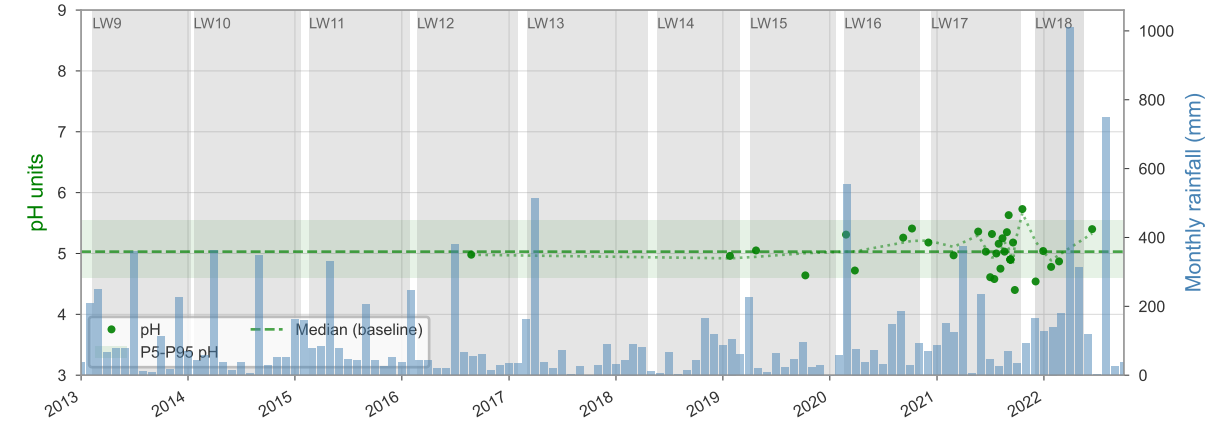
WC7_POOL9



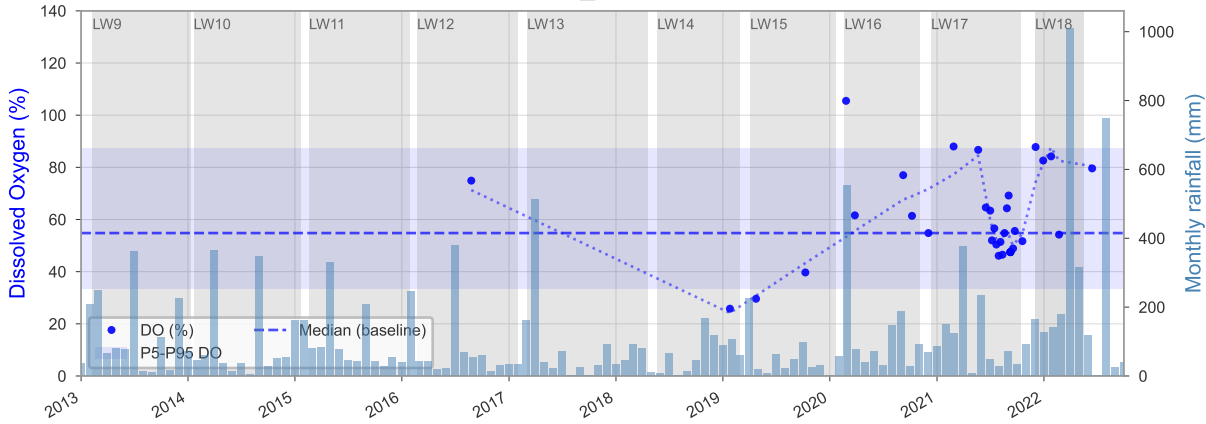
WC7_POOL9



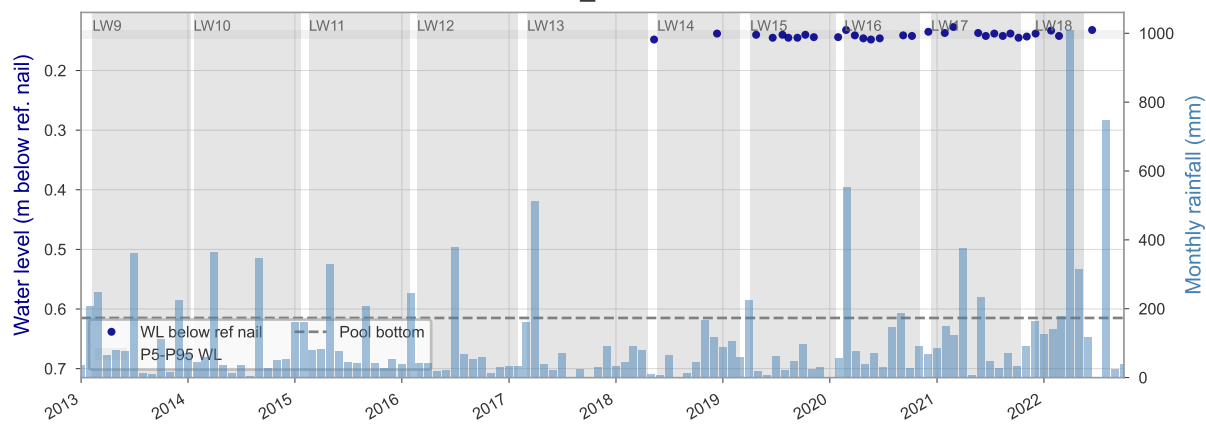
WC7_POOL9



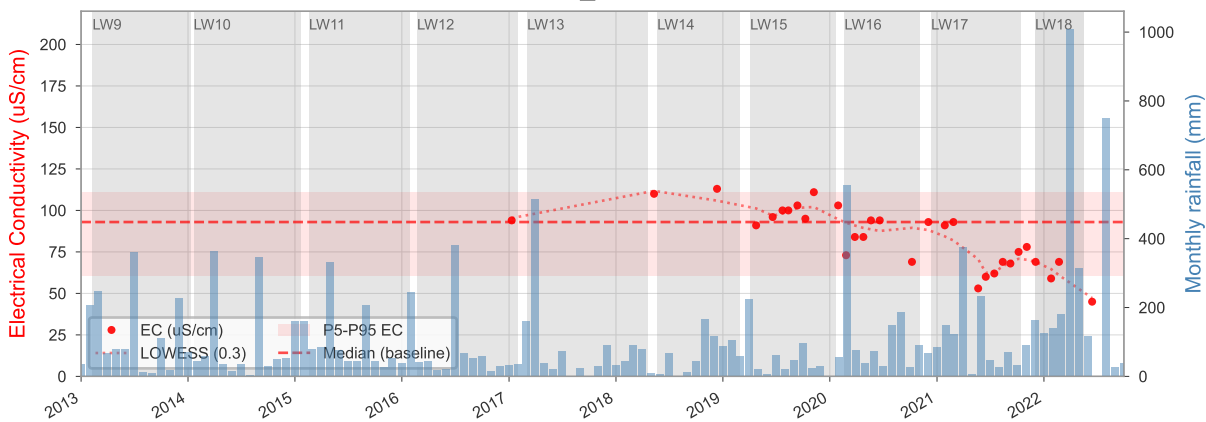
WC7_POOL9



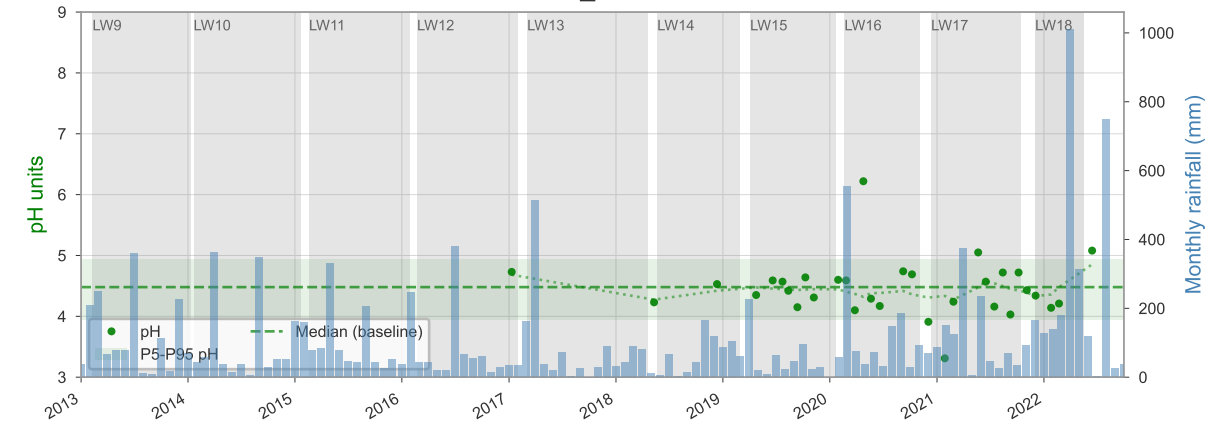
WC8_POOL1



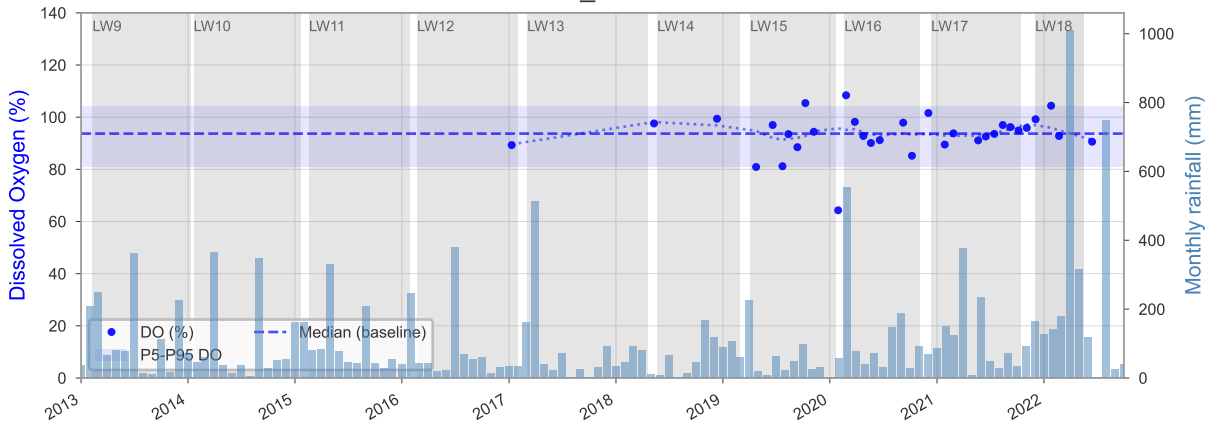
WC8_POOL1



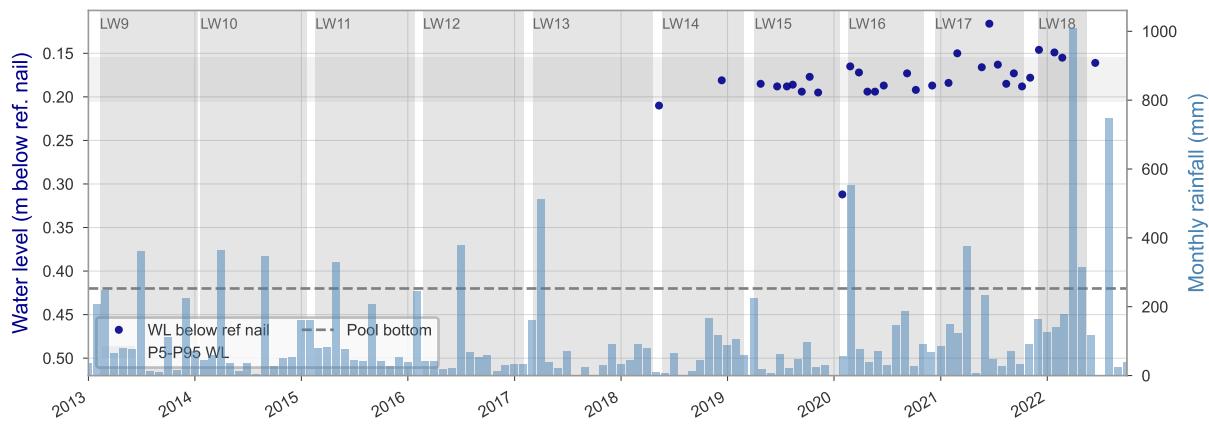
WC8_POOL1



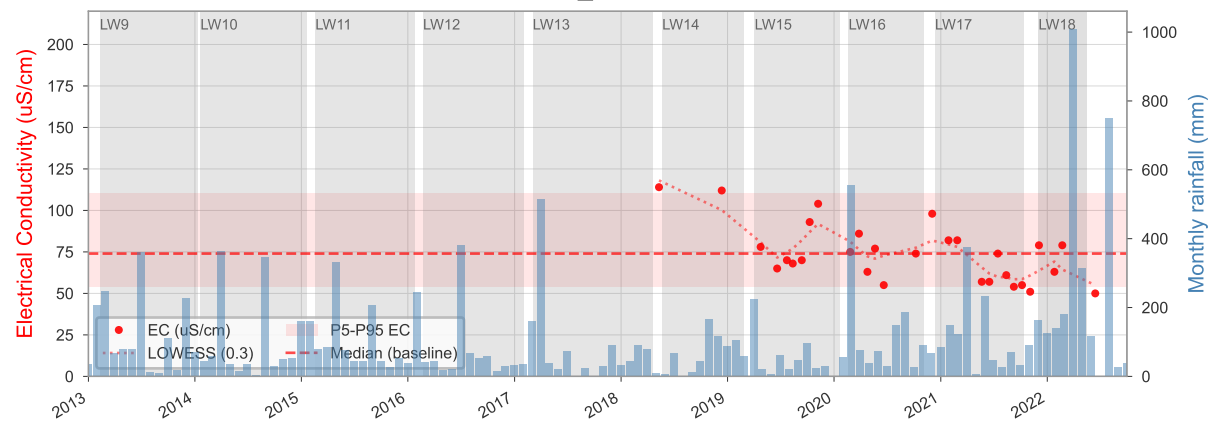
WC8_POOL1



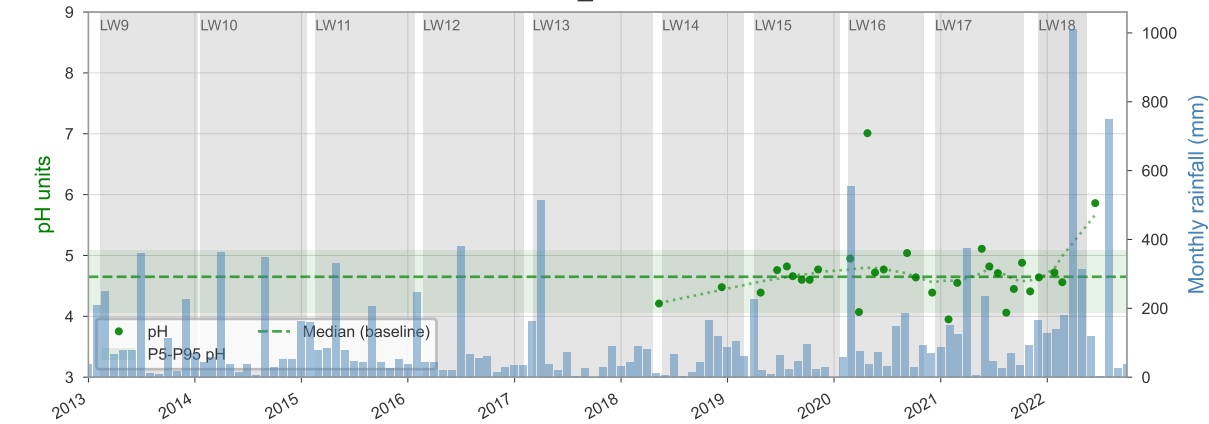
WC8_POOL10



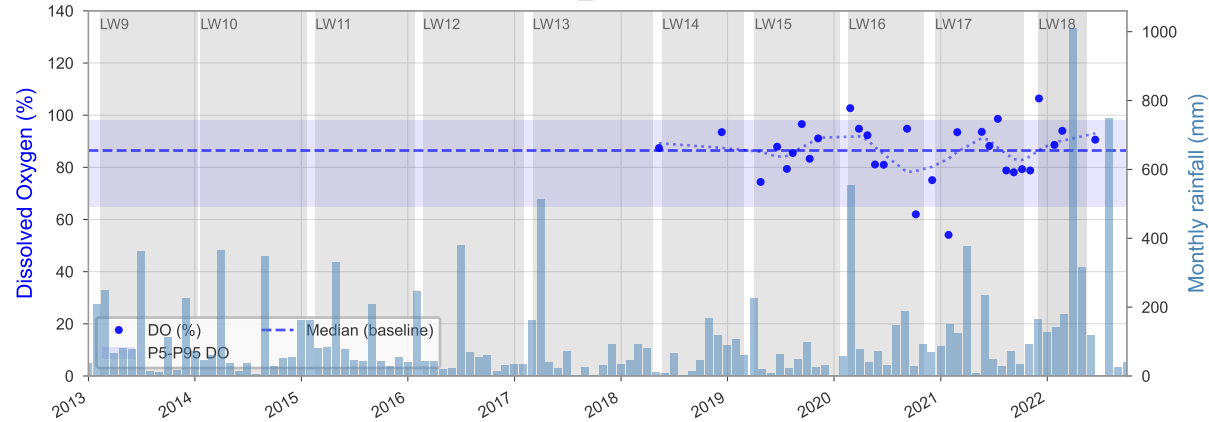
WC8_POOL10



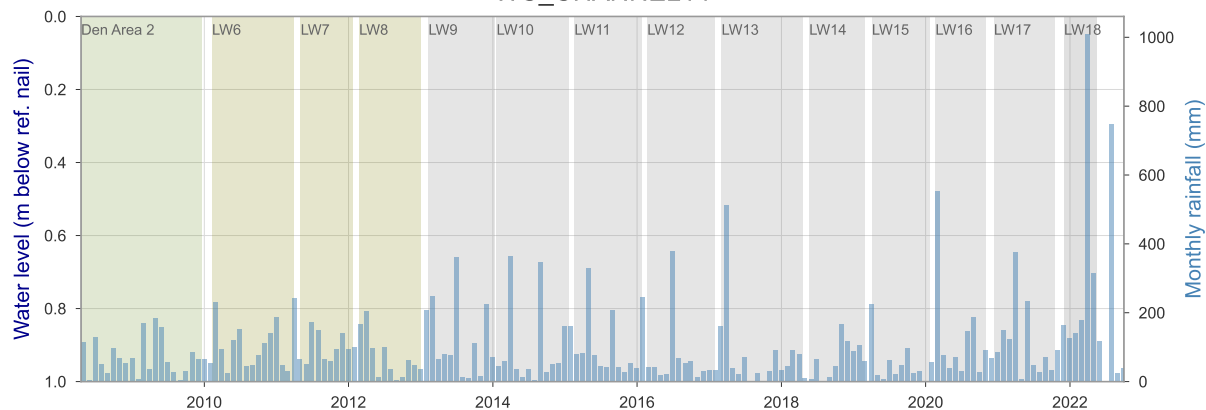
WC8_POOL10



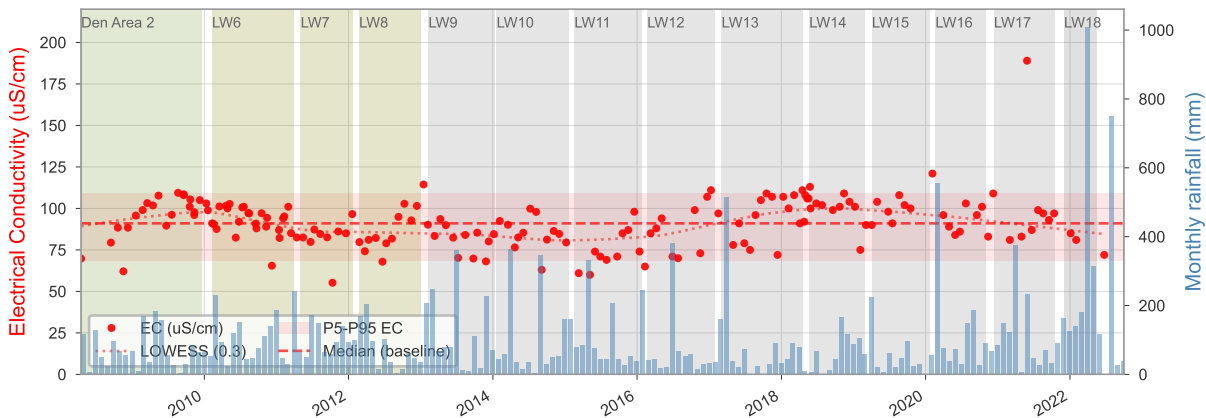
WC8_POOL10



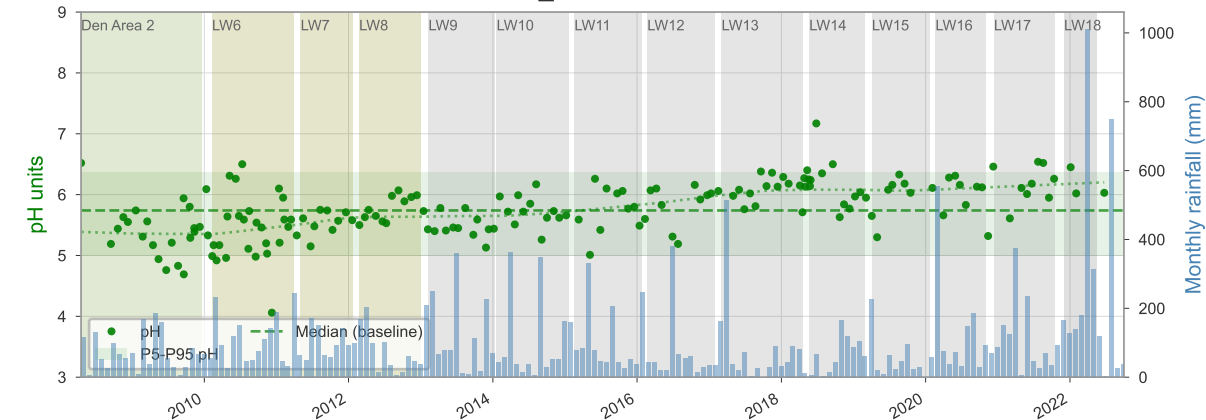
WC_CHANNEL14



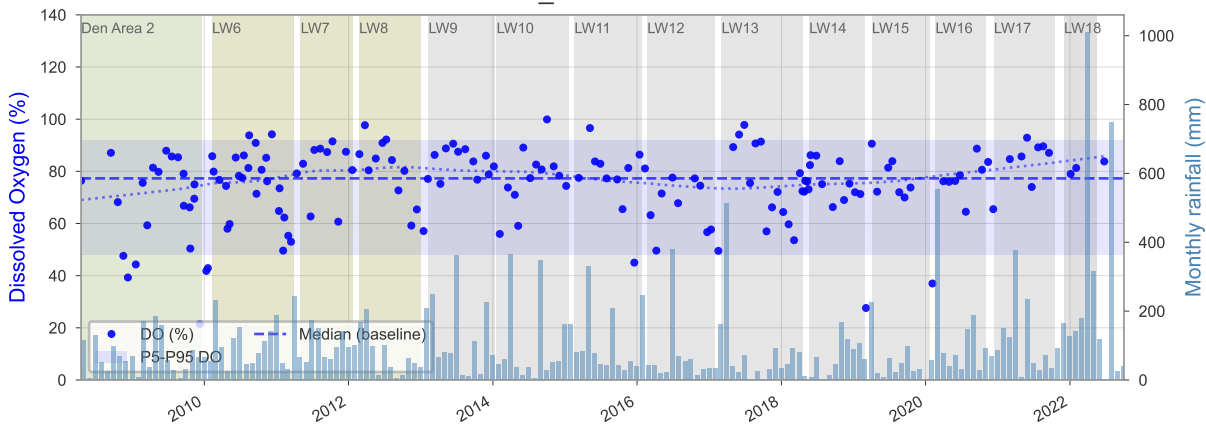
WC_CHANNEL14



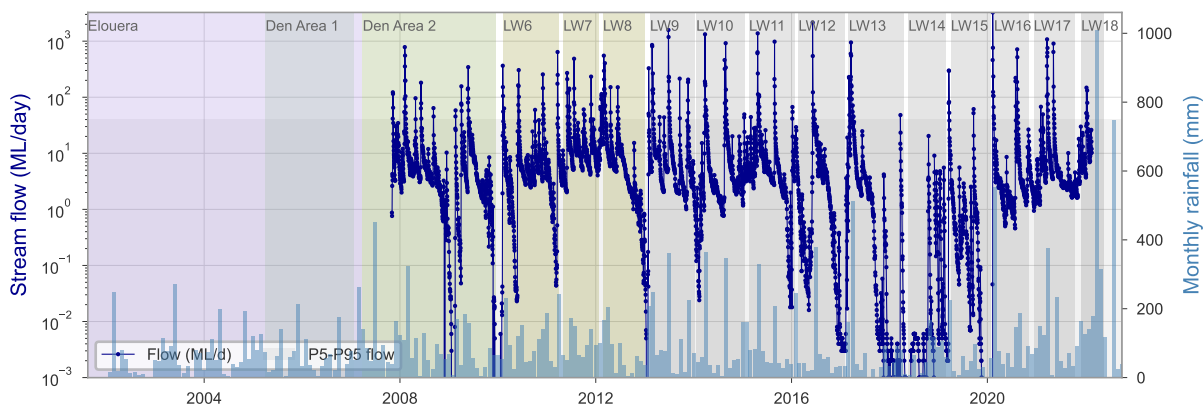
WC_CHANNEL14



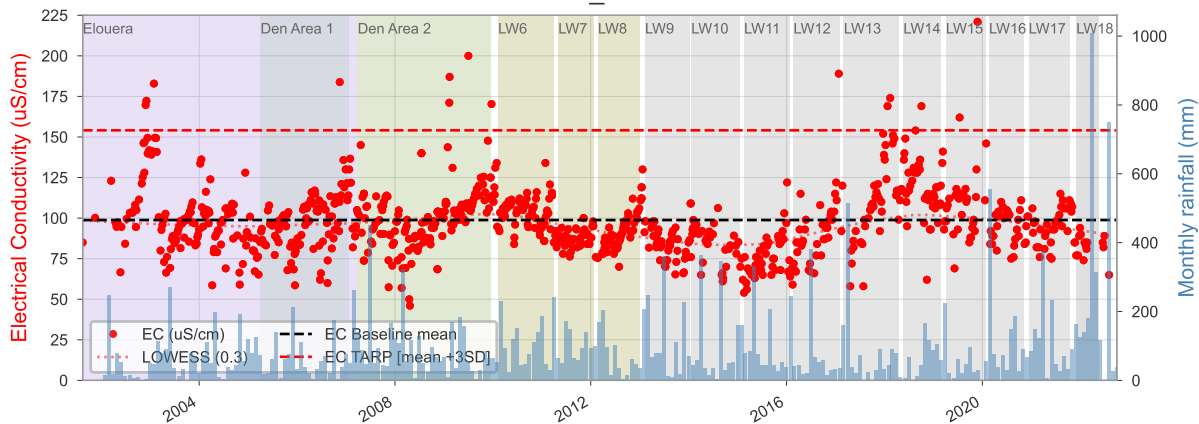
WC_CHANNEL14



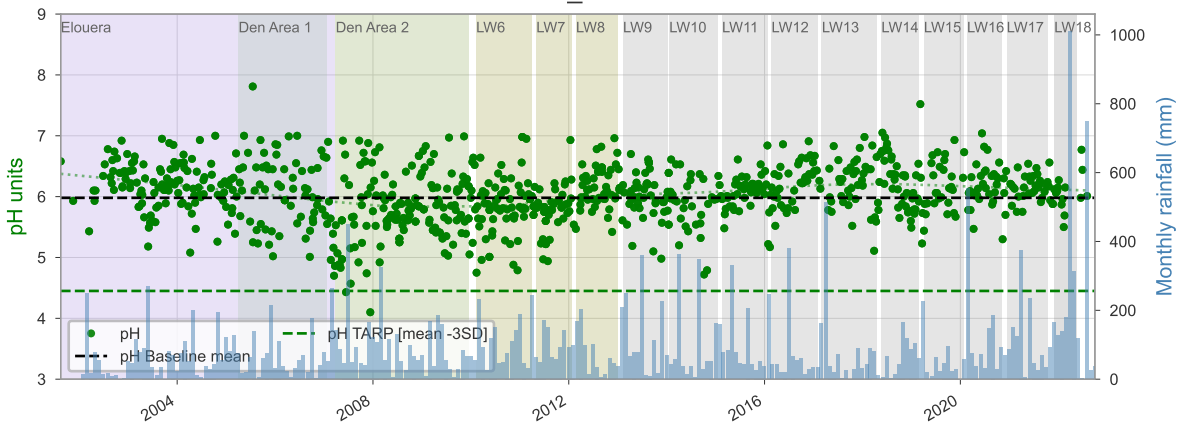
WWL



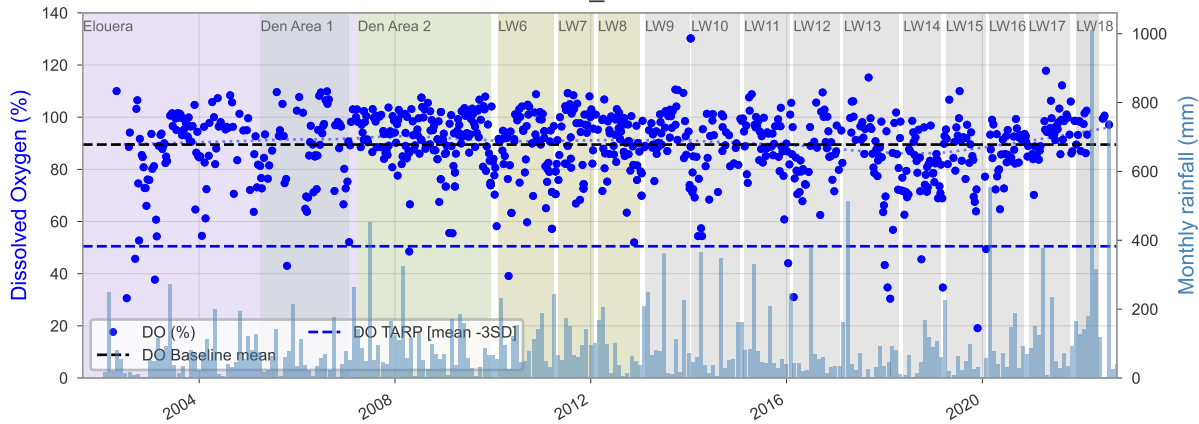
WC_FR6



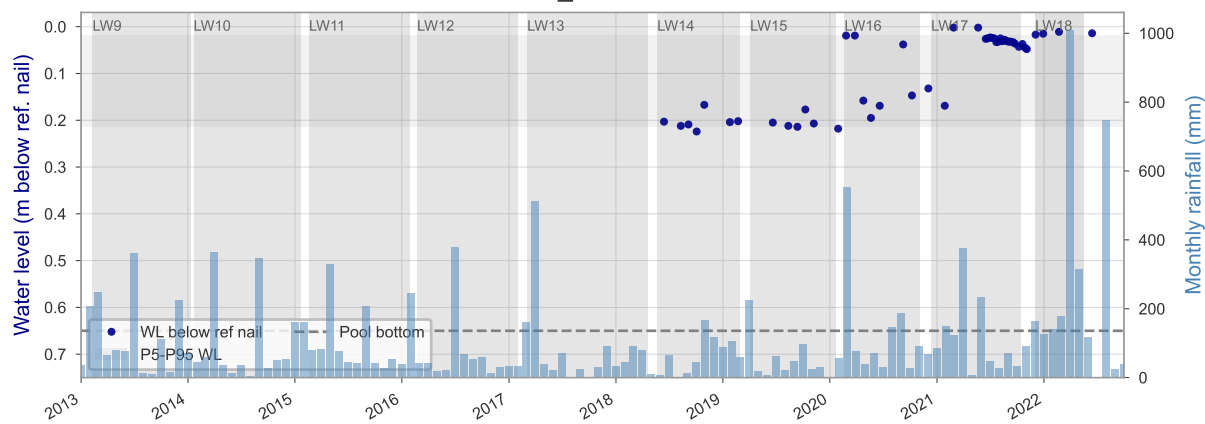
WC_FR6



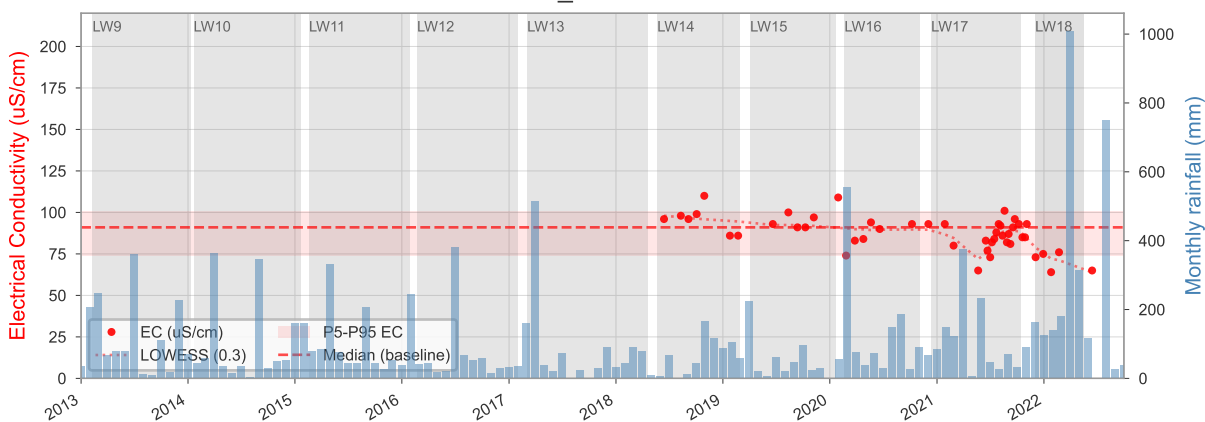
WC_FR6



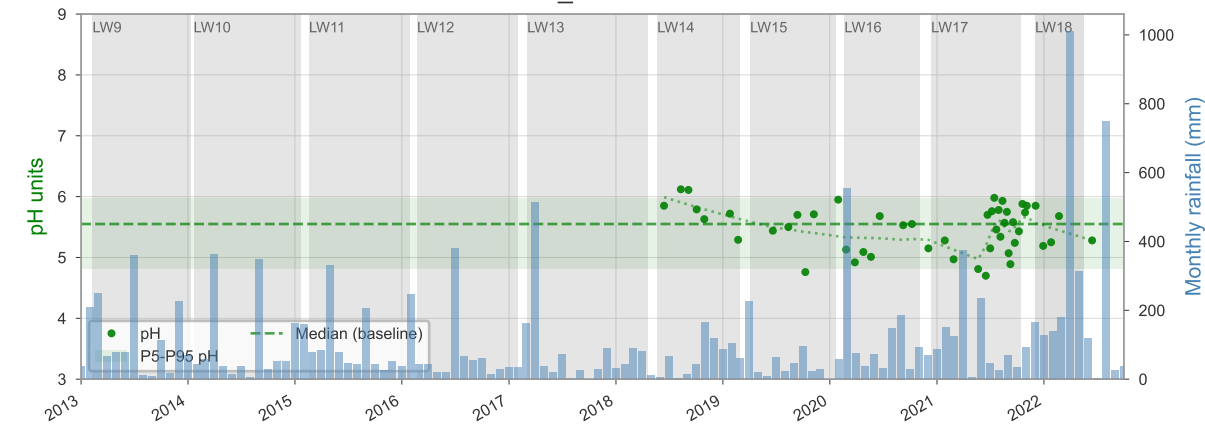
WC_POOL104



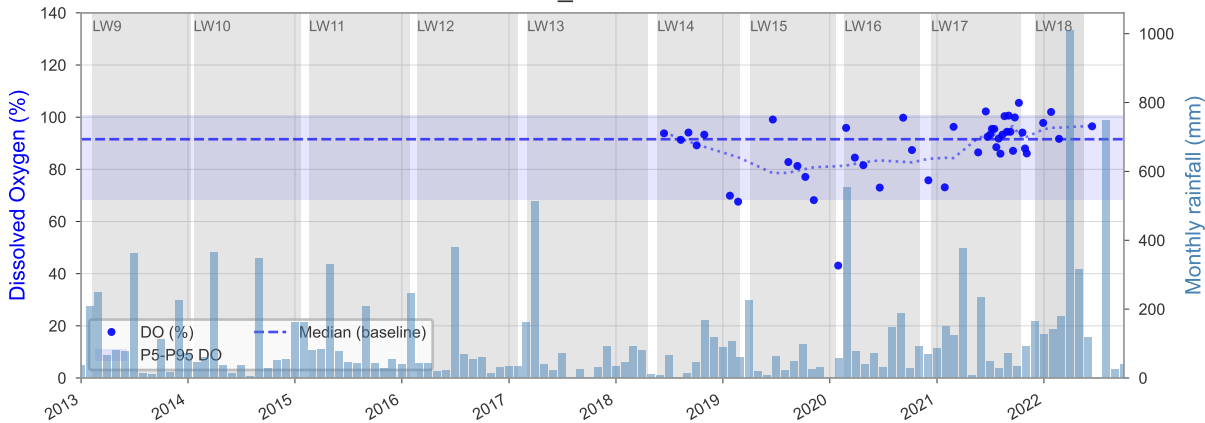
WC_POOL104



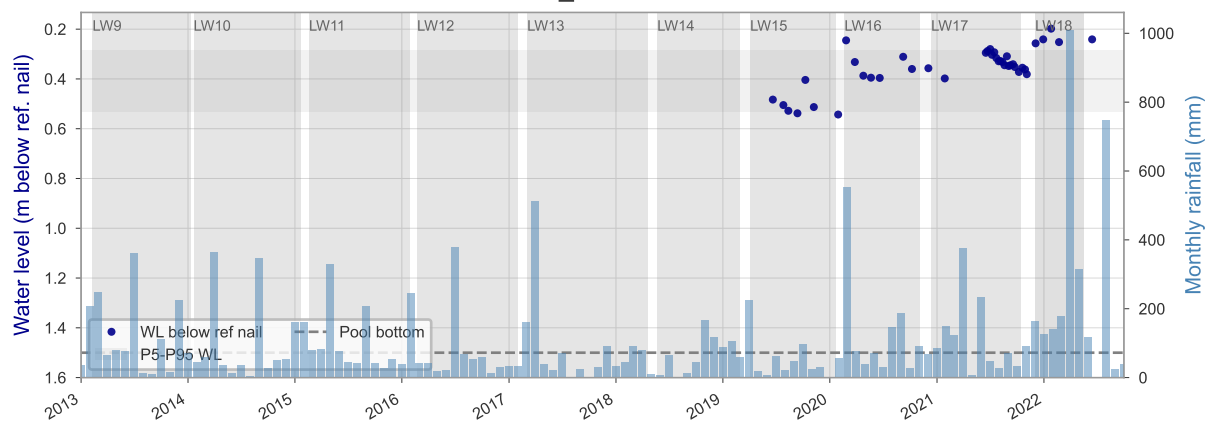
WC_POOL104



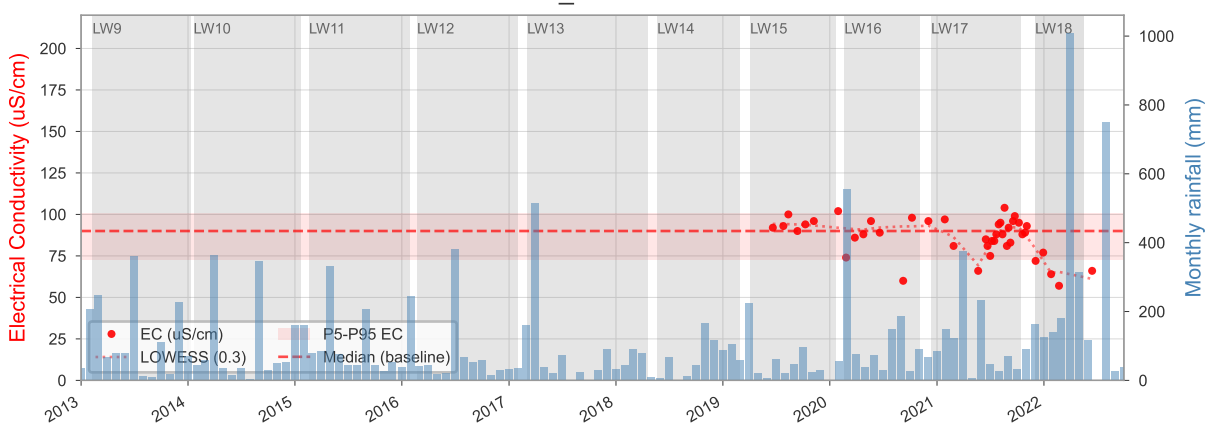
WC_POOL104



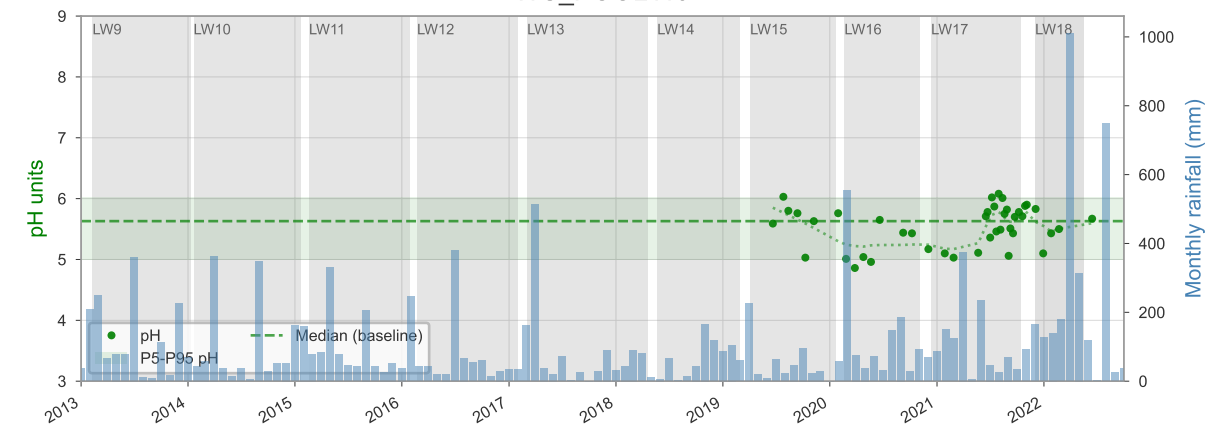
WC_POOL119



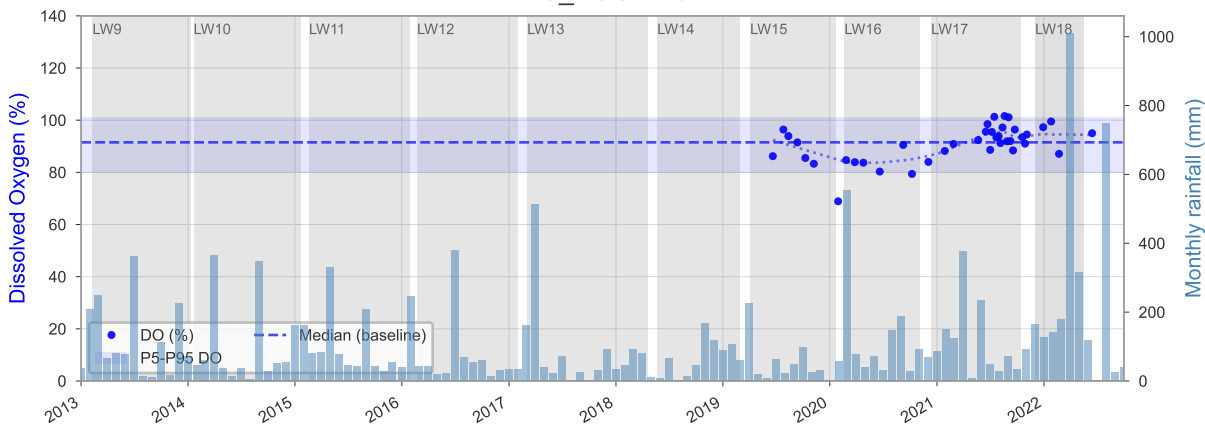
WC_POOL119



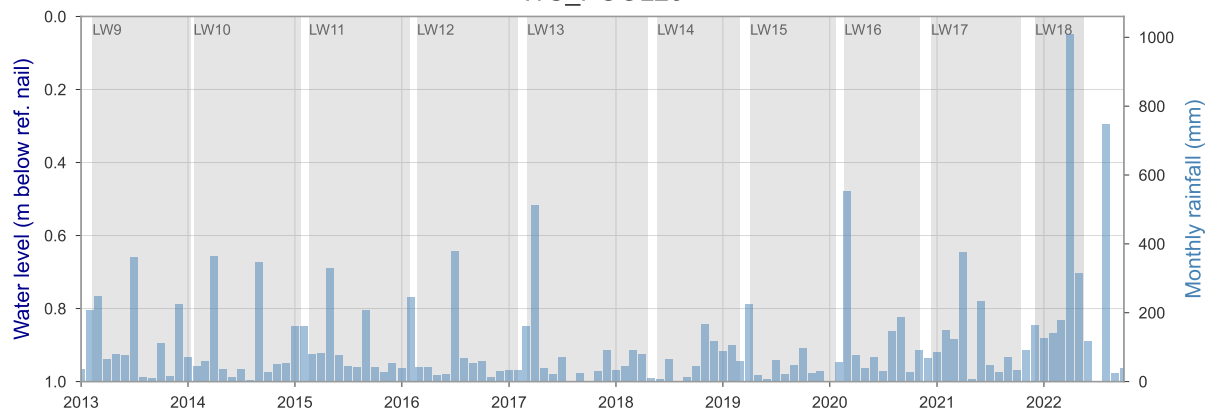
WC_POOL119



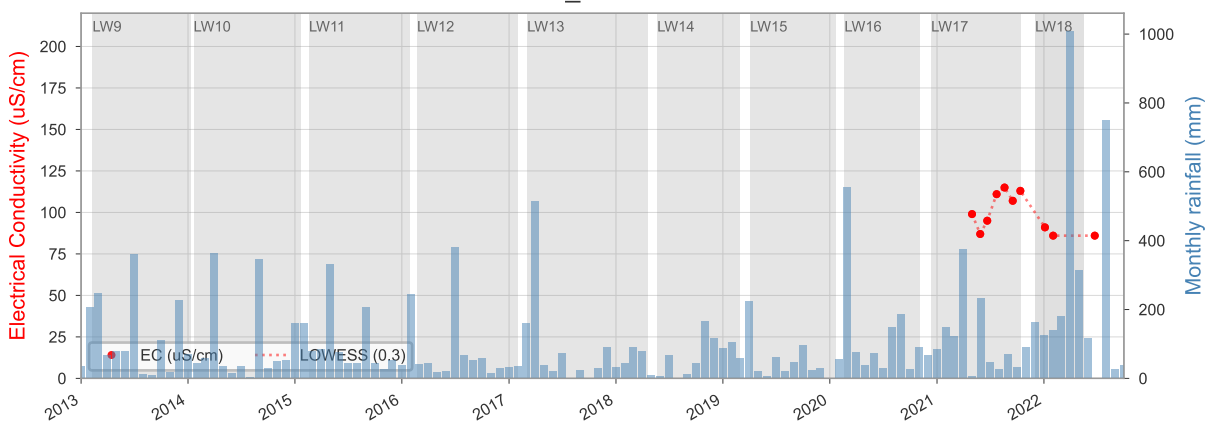
WC_POOL119



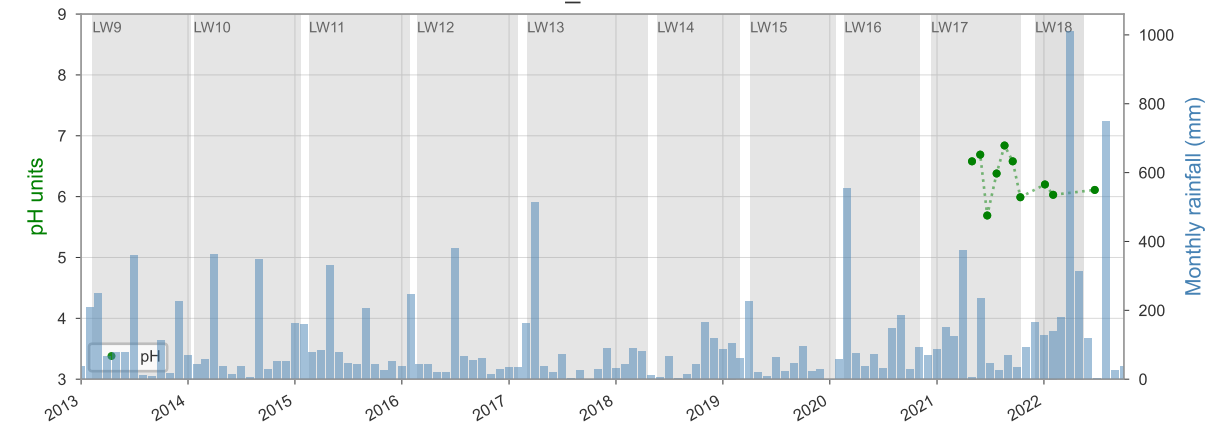
WC_POOL20



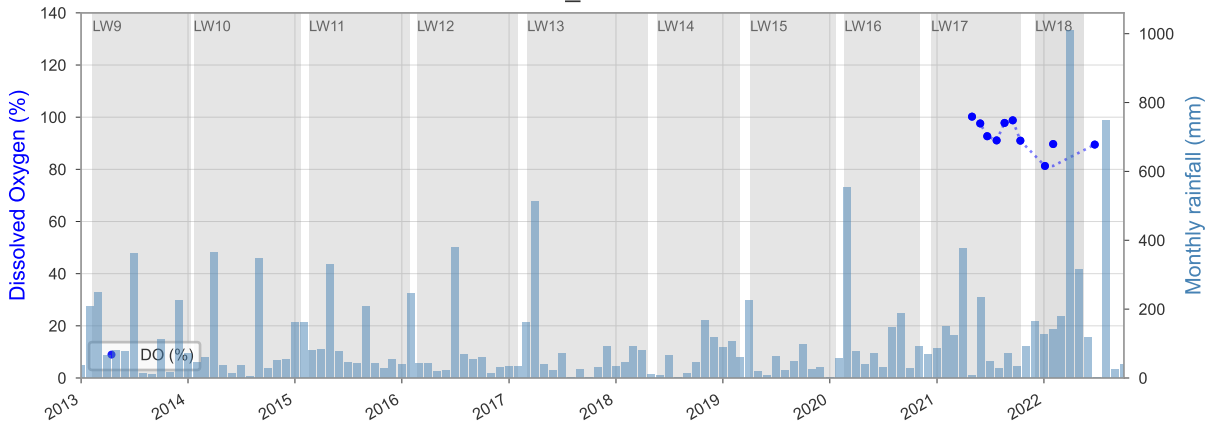
WC_POOL20



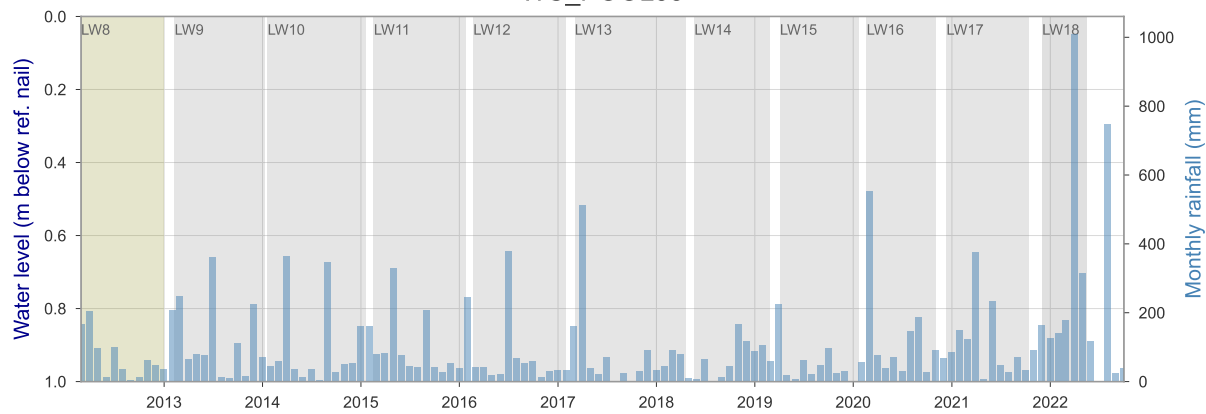
WC_POOL20



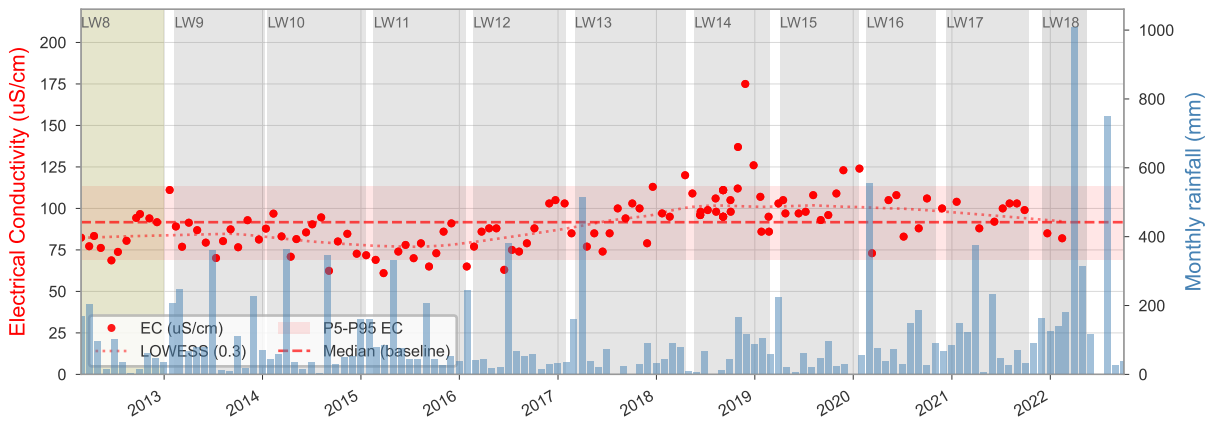
WC_POOL20



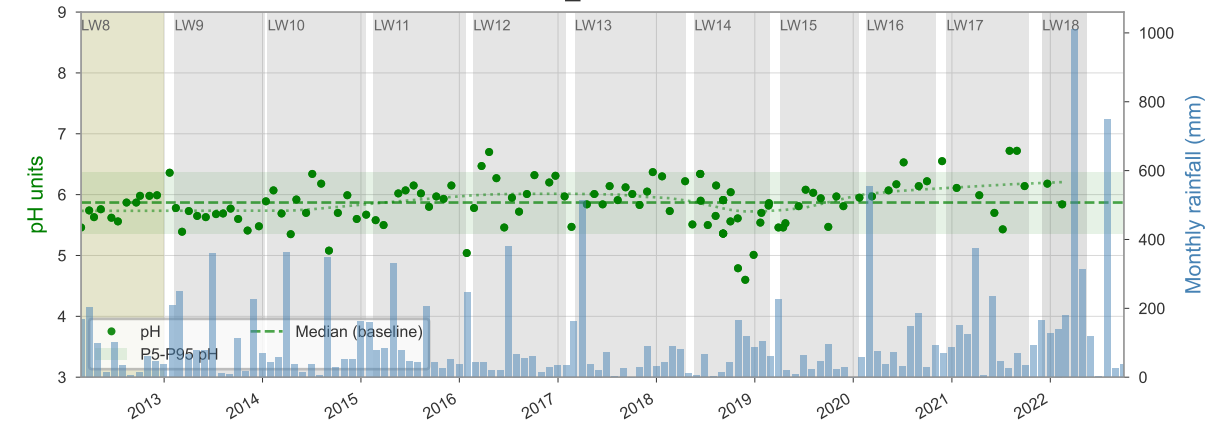
WC_POOL38



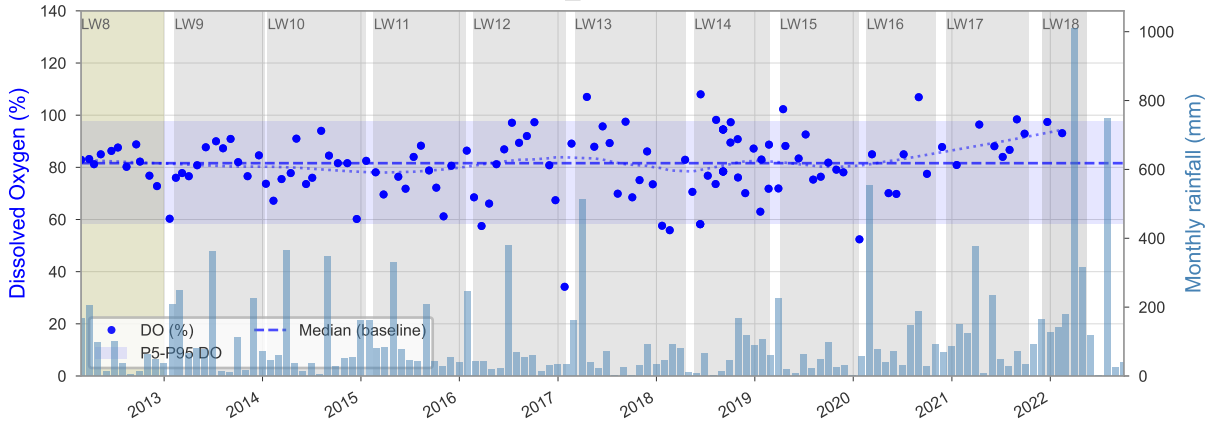
WC_POOL38



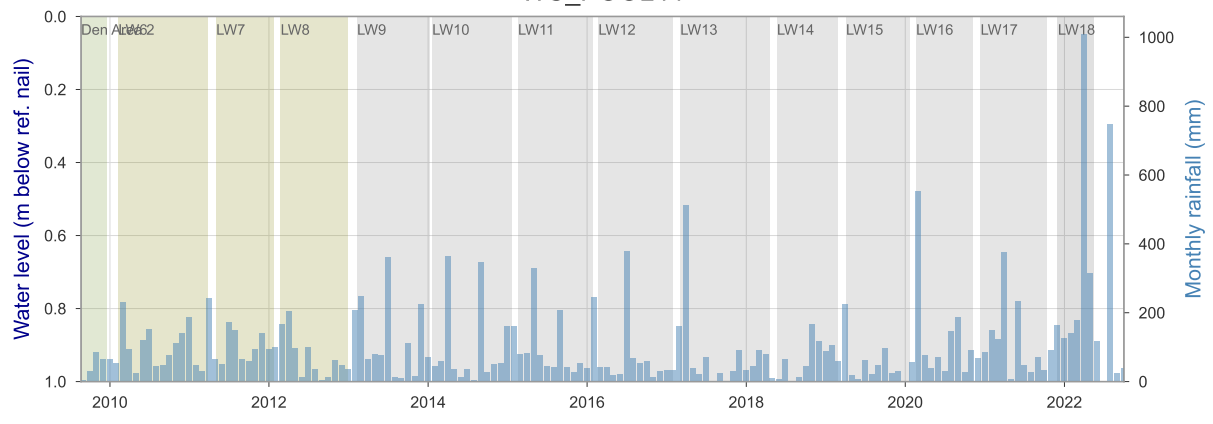
WC_POOL38



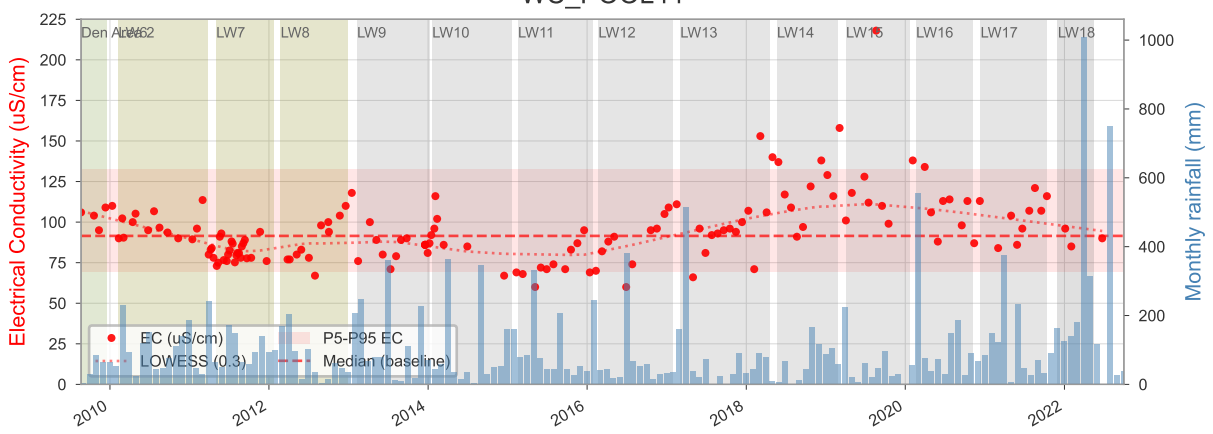
WC_POOL38



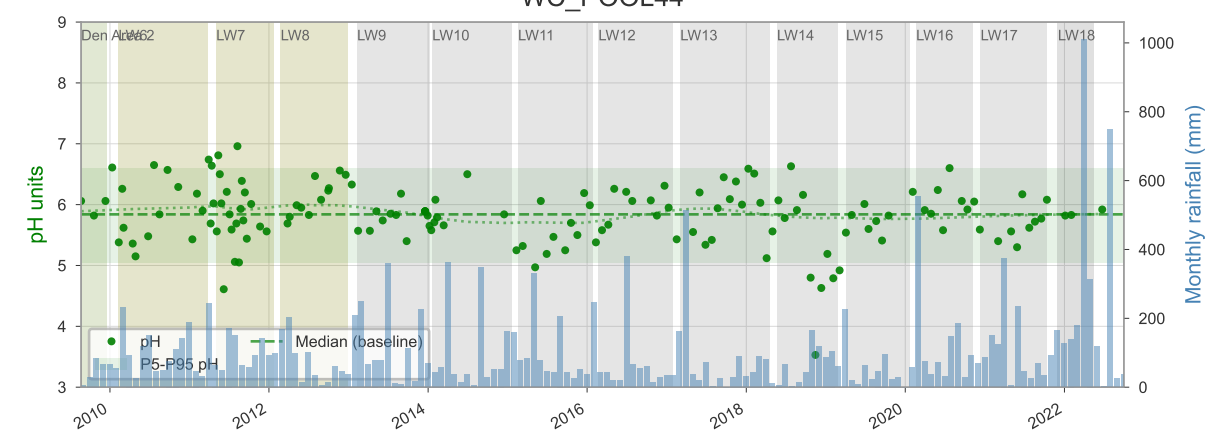
WC_POOL44



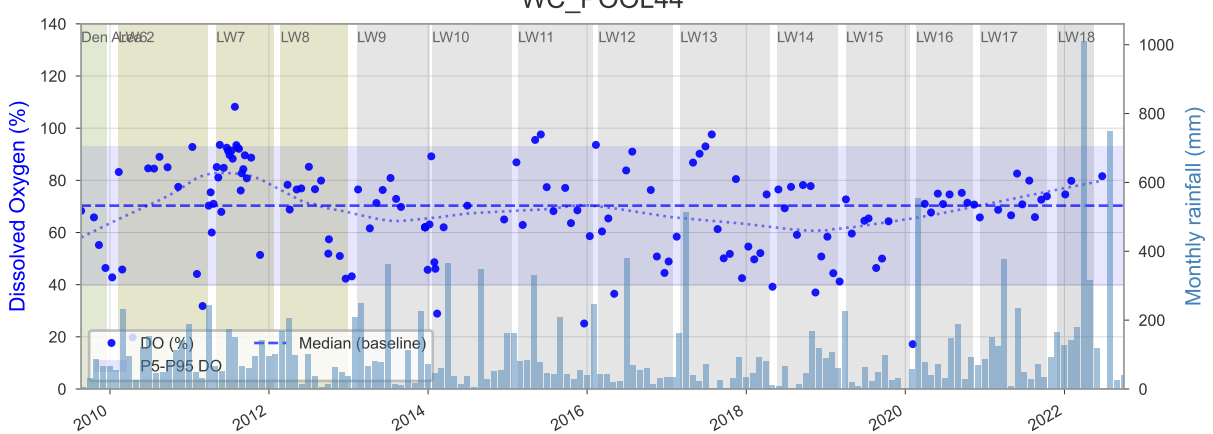
WC_POOL44



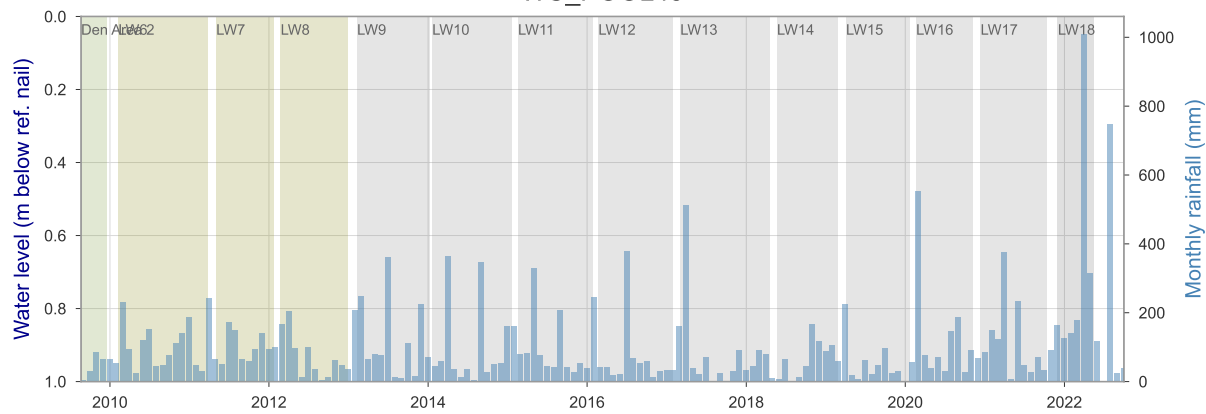
WC_POOL44



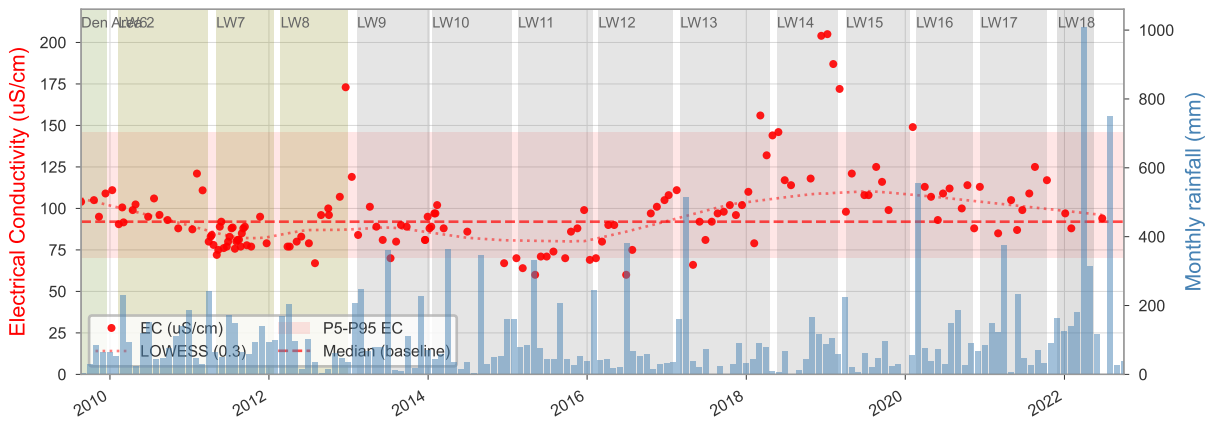
WC_POOL44



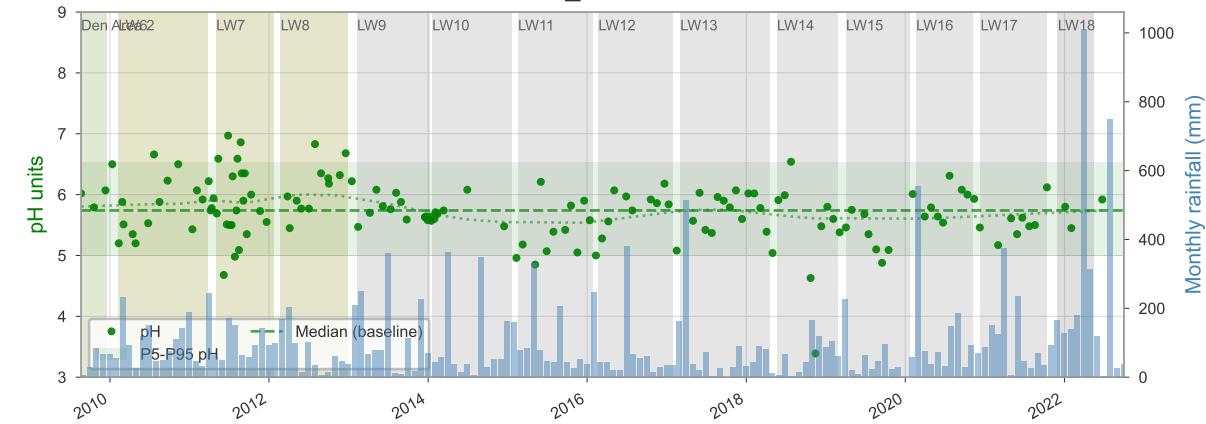
WC_POOL45



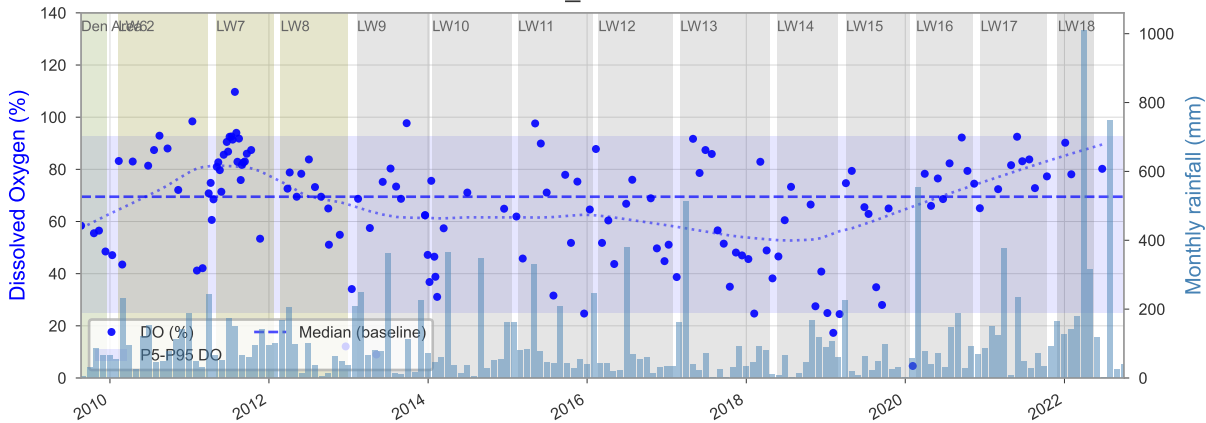
WC_POOL45



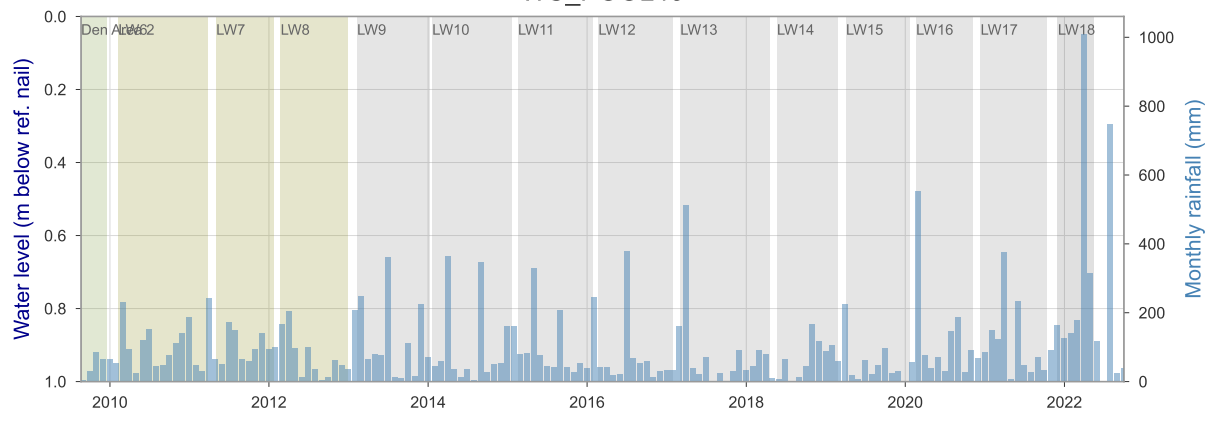
WC_POOL45



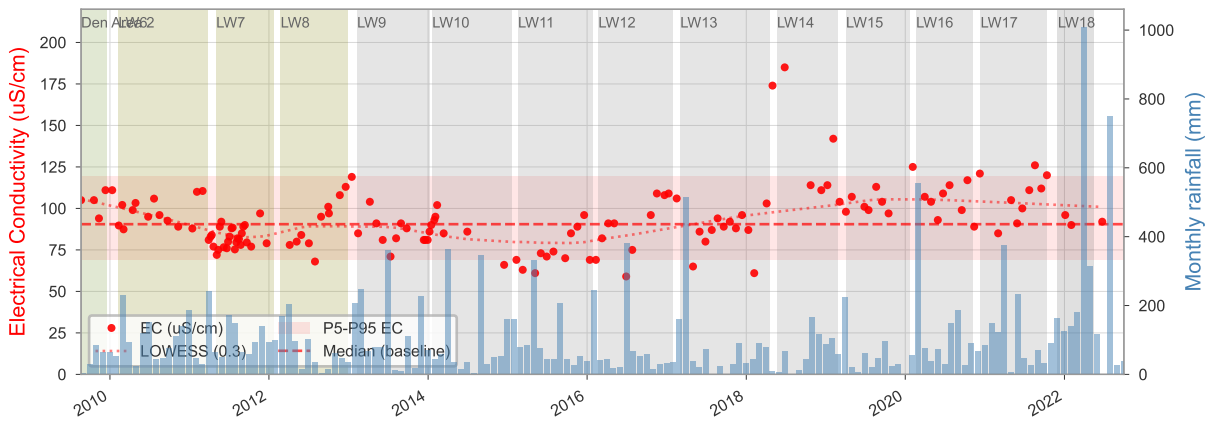
WC_POOL45



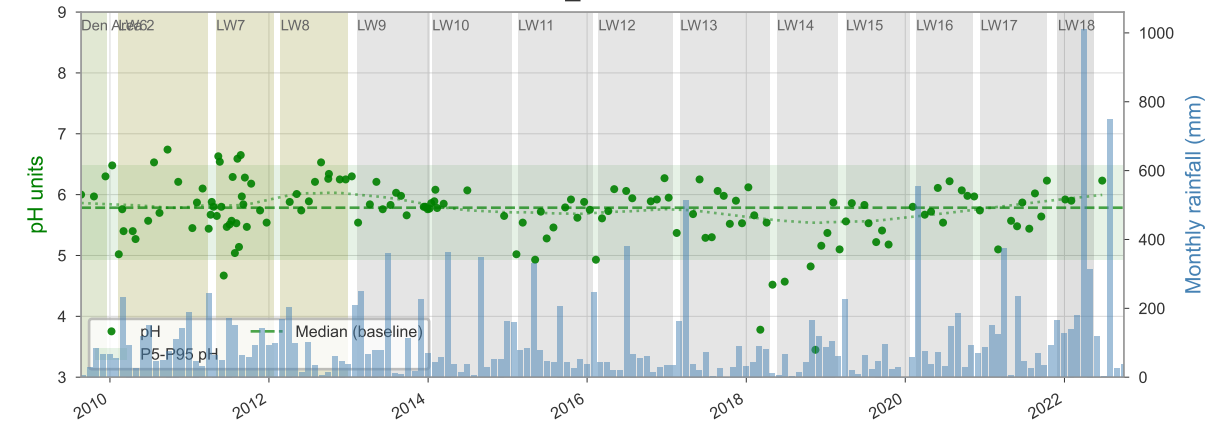
WC_POOL46



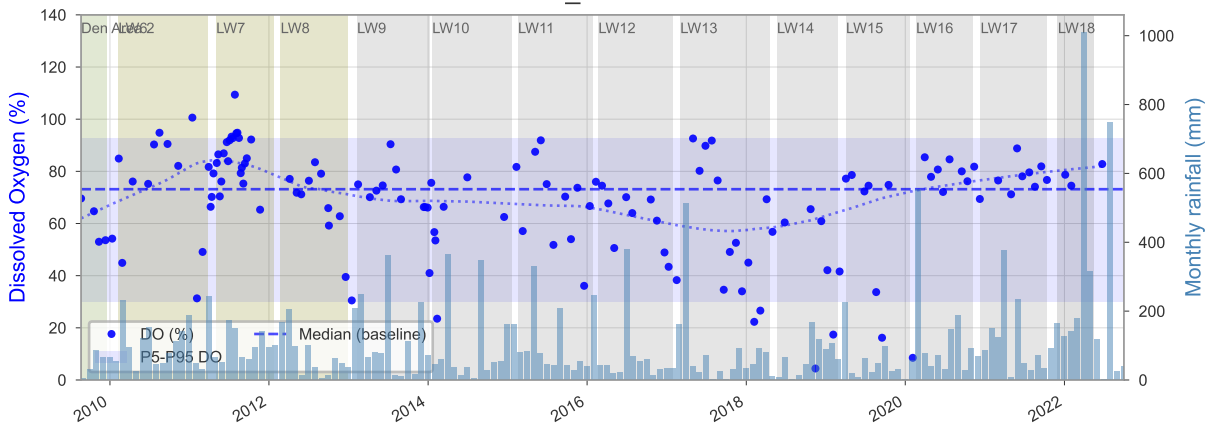
WC_POOL46



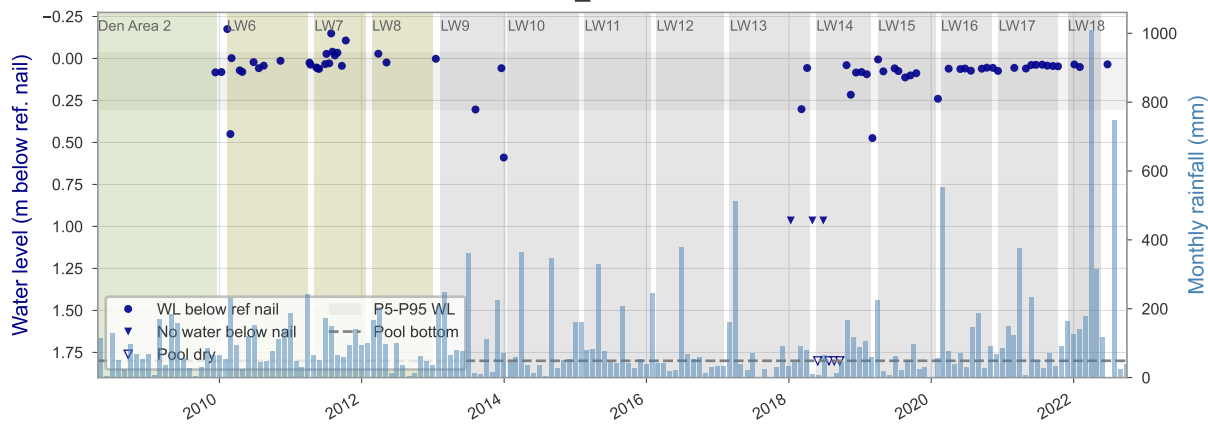
WC_POOL46



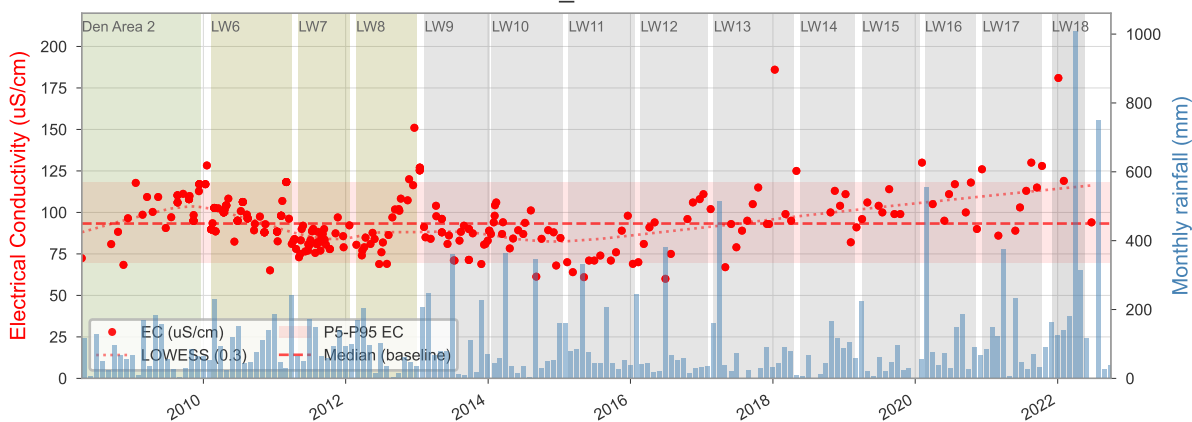
WC_POOL46



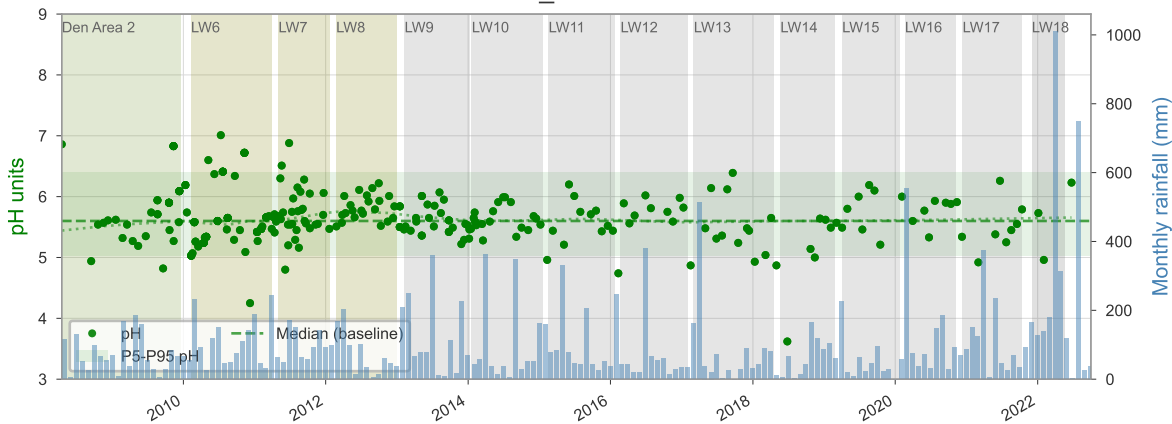
WC_POOL49



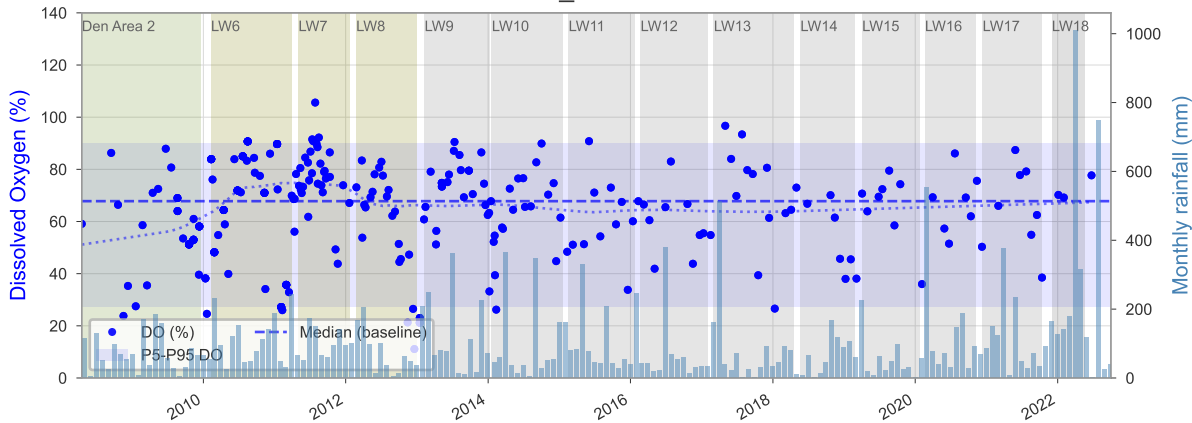
WC_POOL49



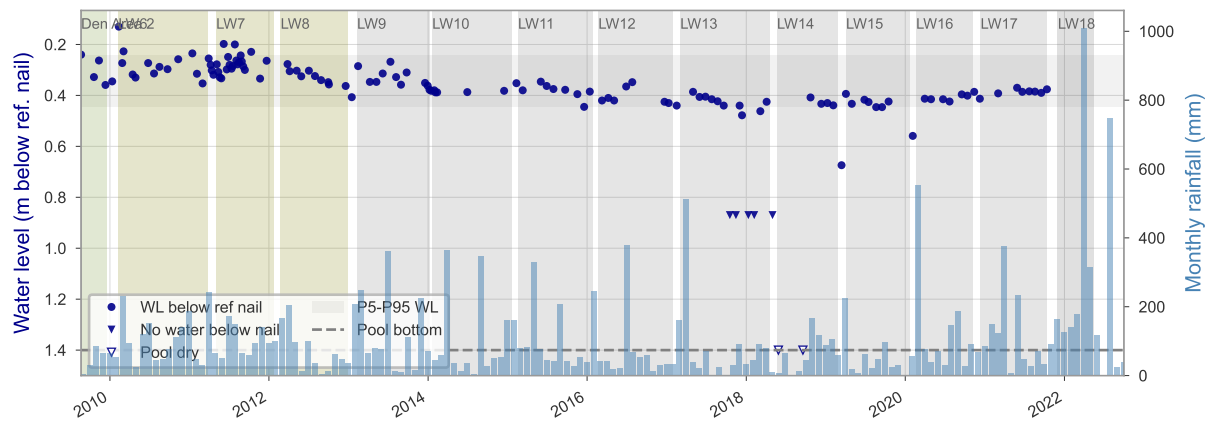
WC_POOL49



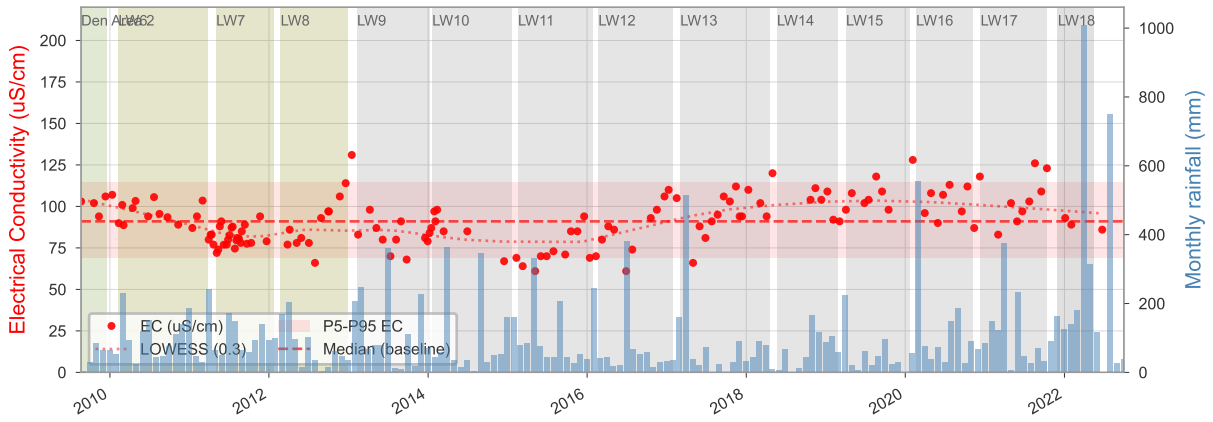
WC_POOL49



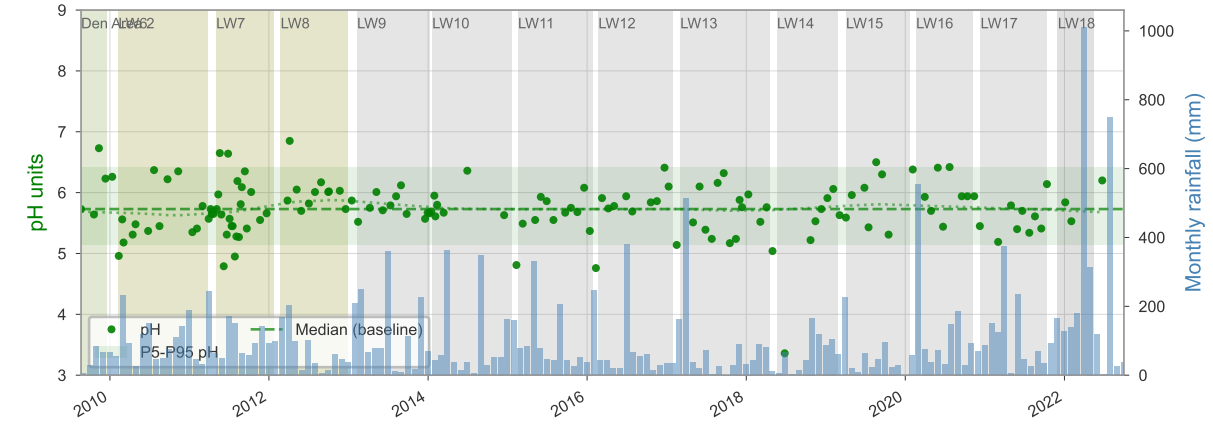
WC_POOL50



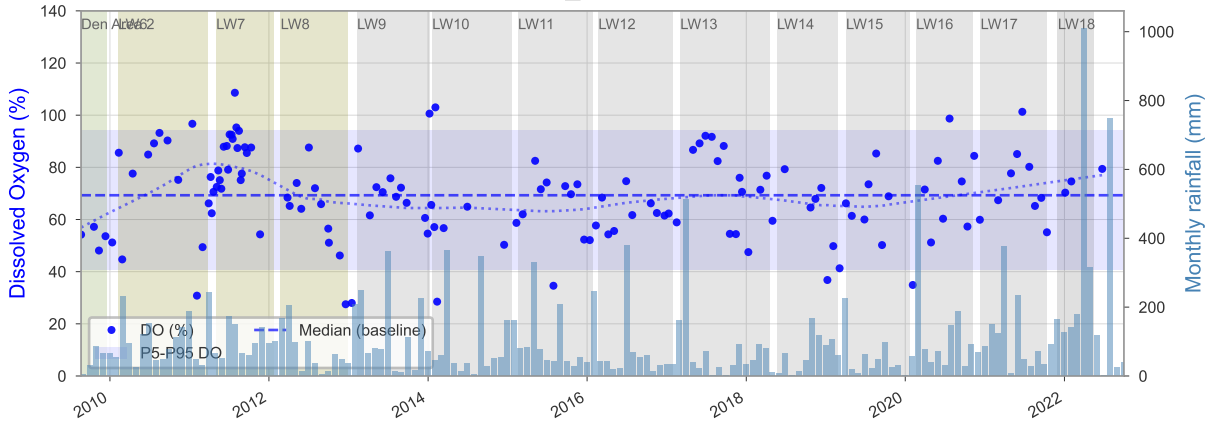
WC_POOL50



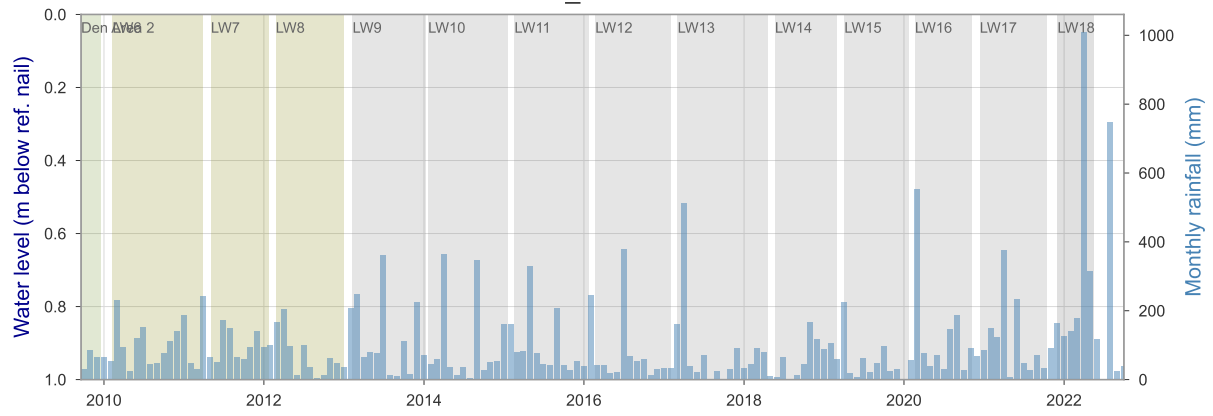
WC_POOL50



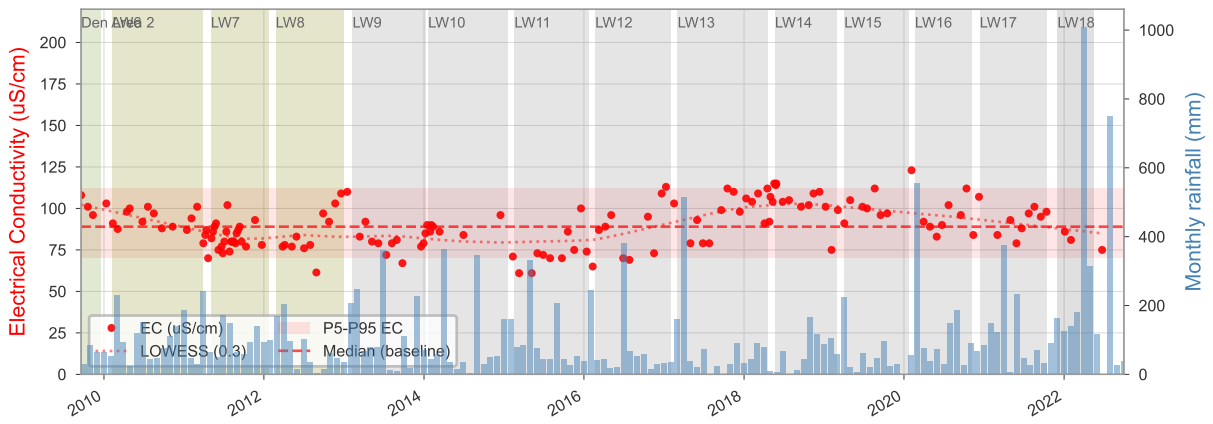
WC_POOL50



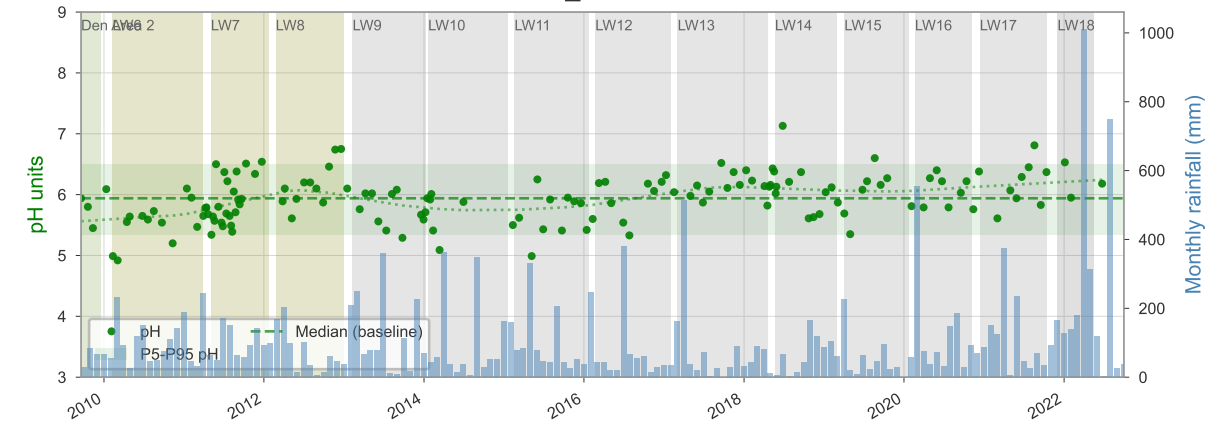
WC_POOL53



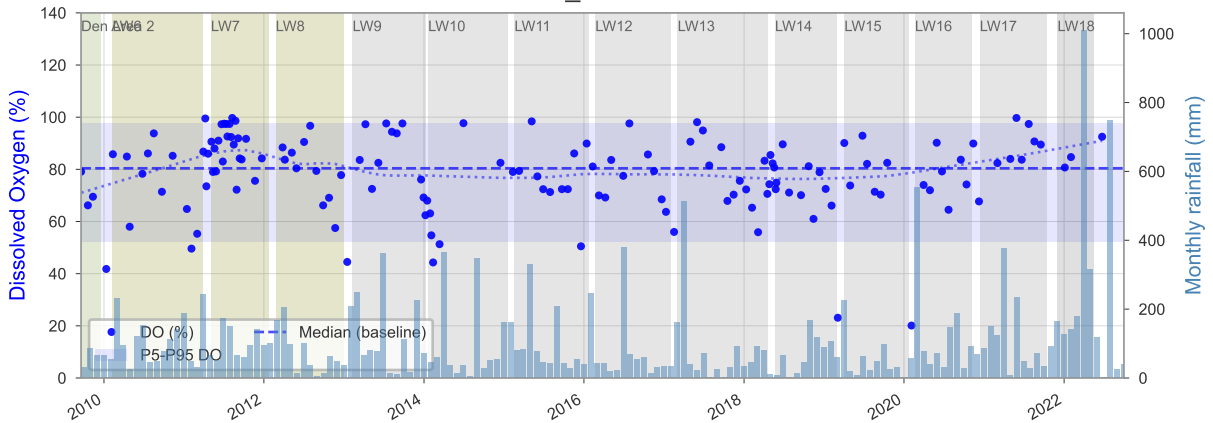
WC_POOL53



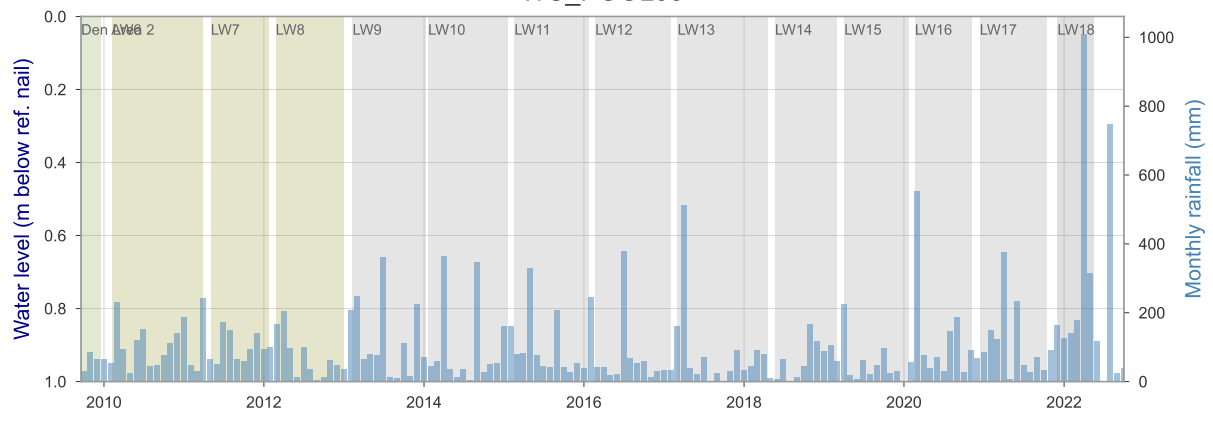
WC_POOL53



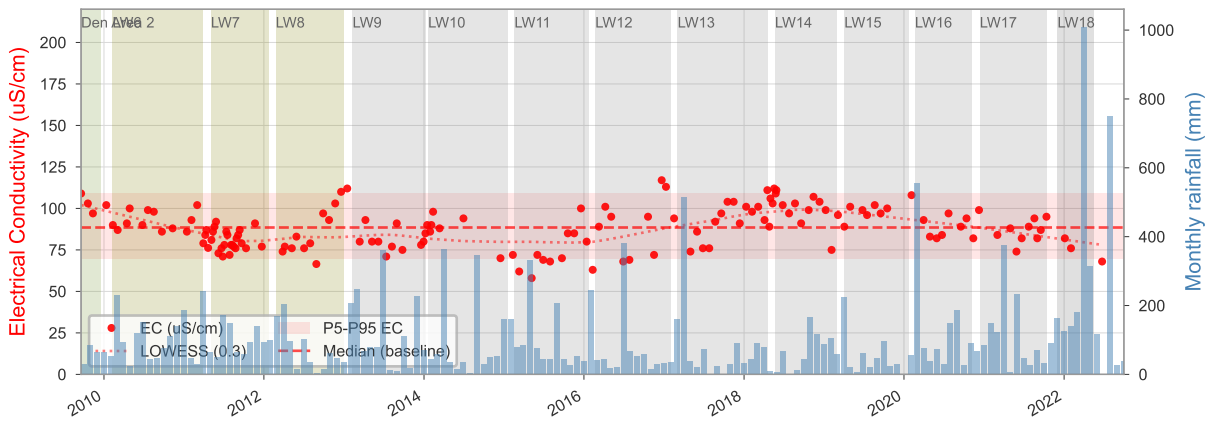
WC_POOL53



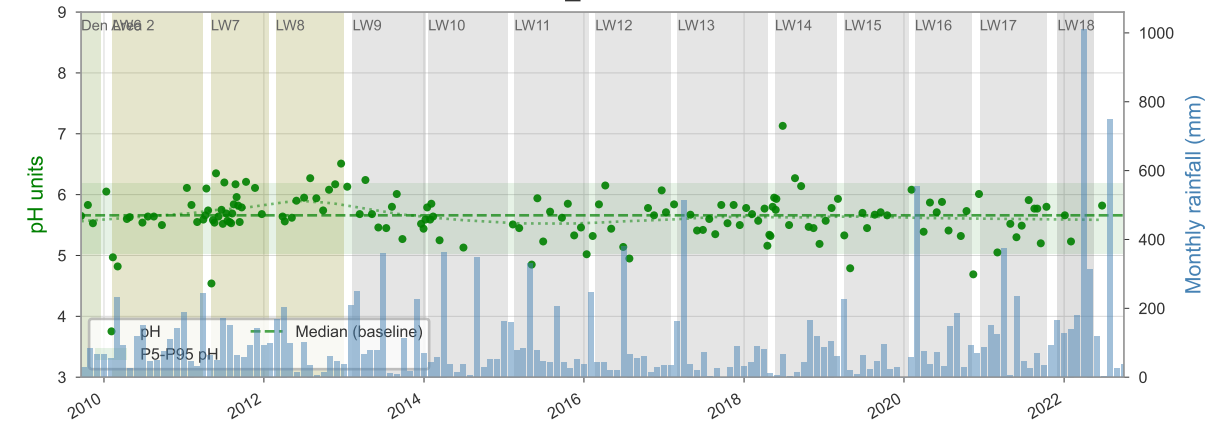
WC_POOL55



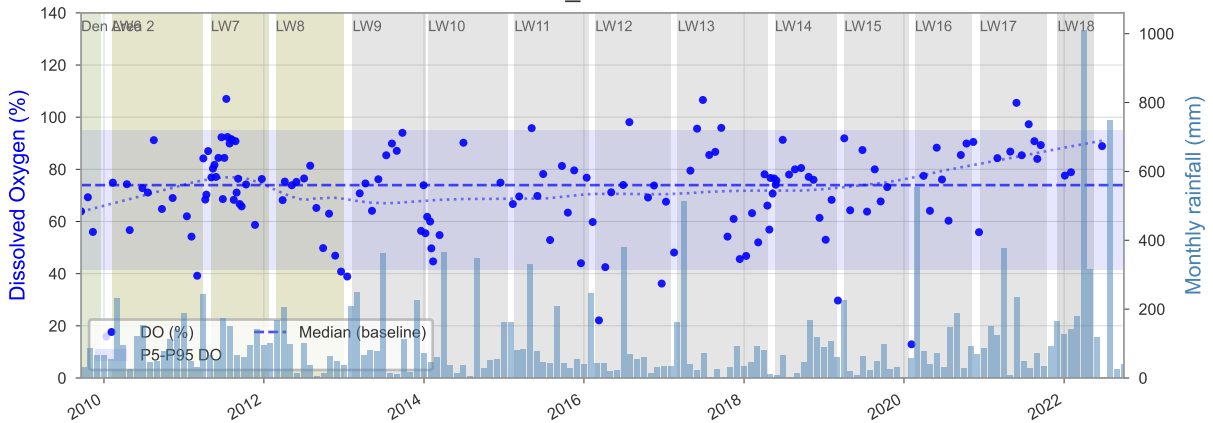
WC_POOL55



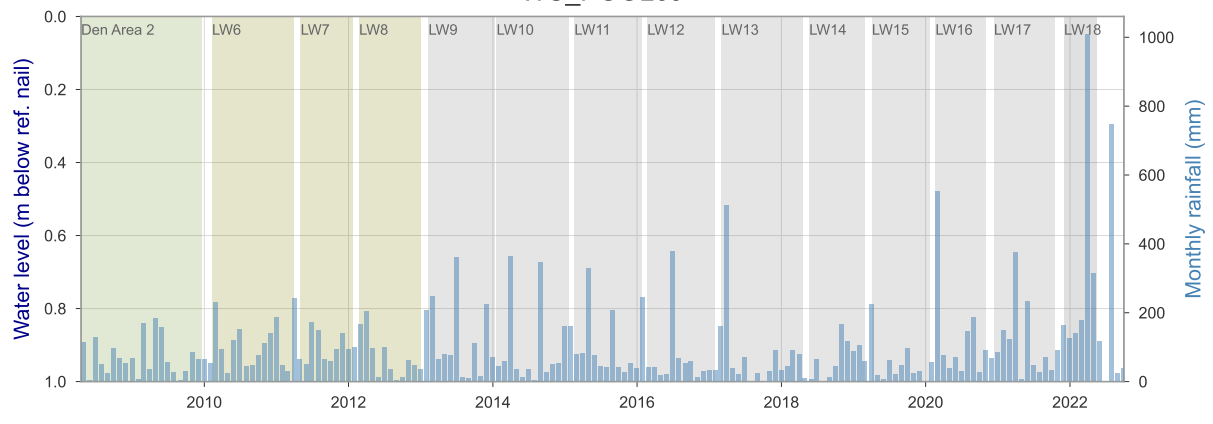
WC_POOL55



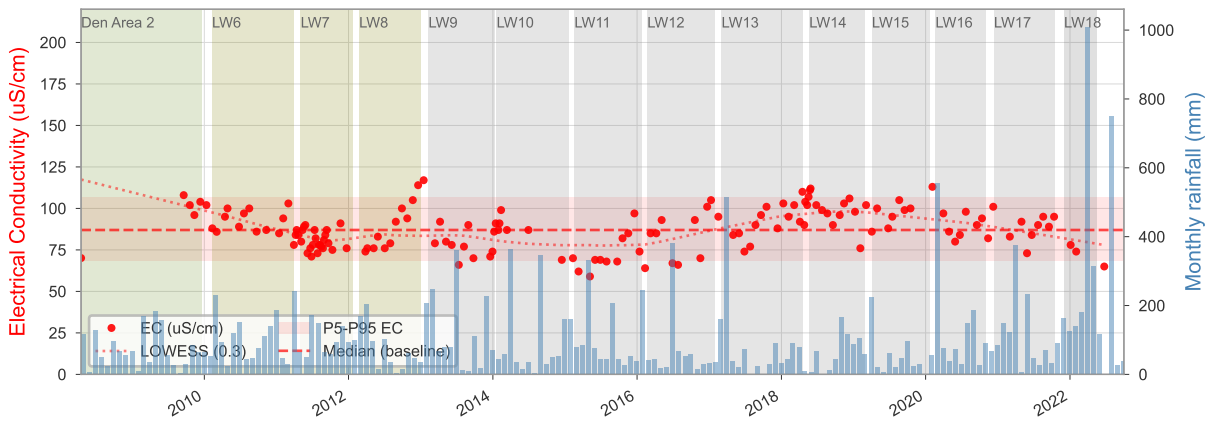
WC_POOL55



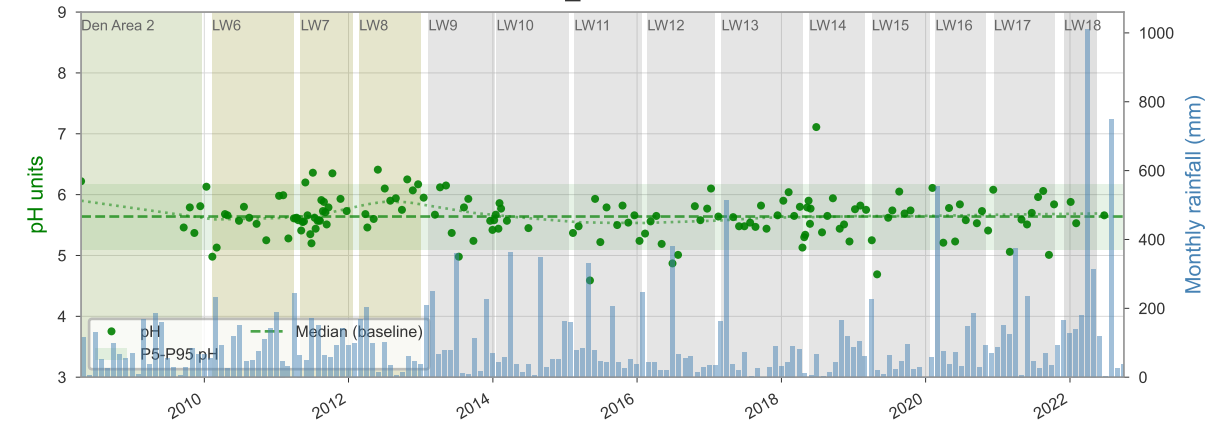
WC_POOL69



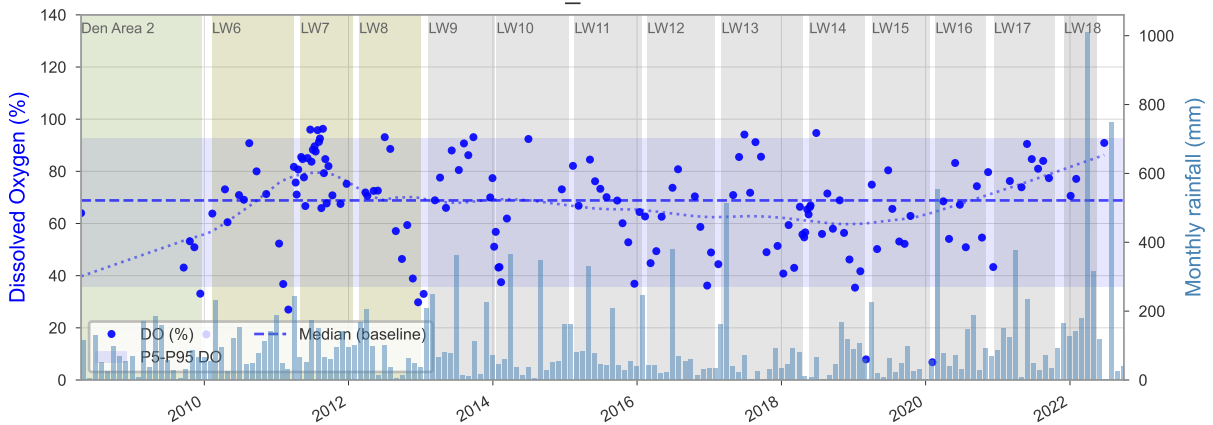
WC_POOL69



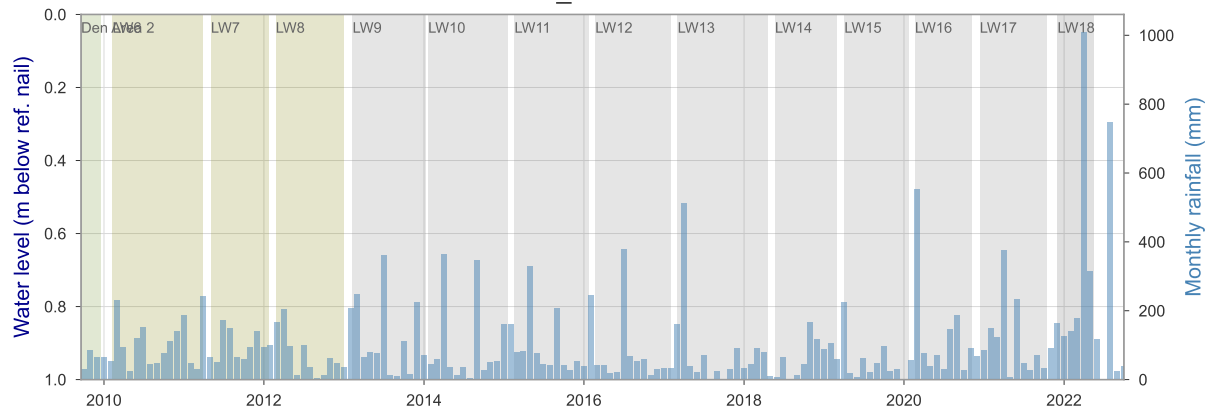
WC_POOL69



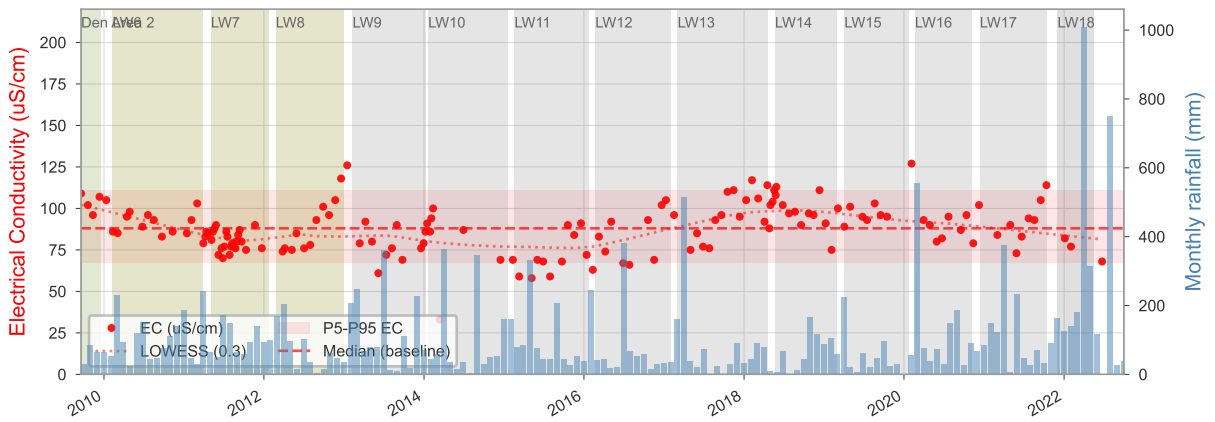
WC_POOL69



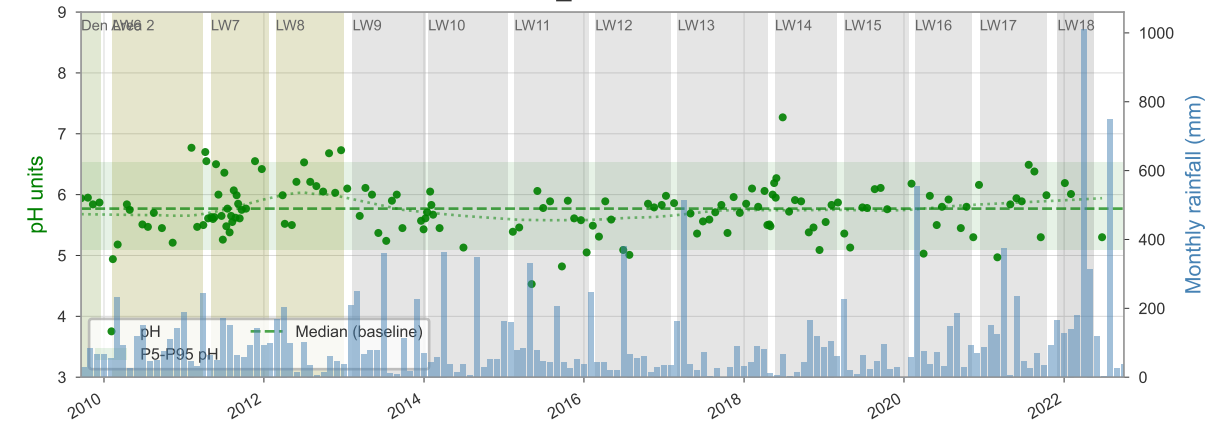
WC_POOL72A



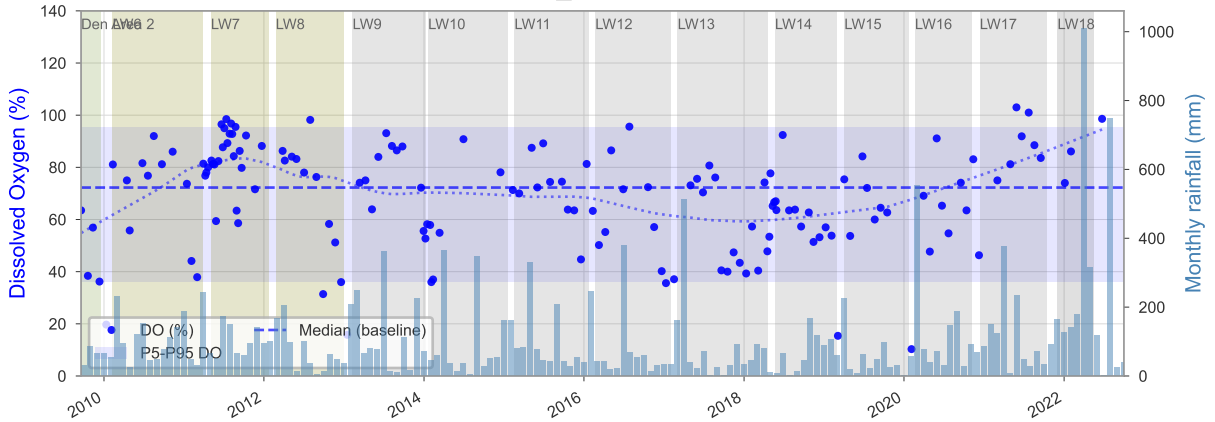
WC_POOL72A



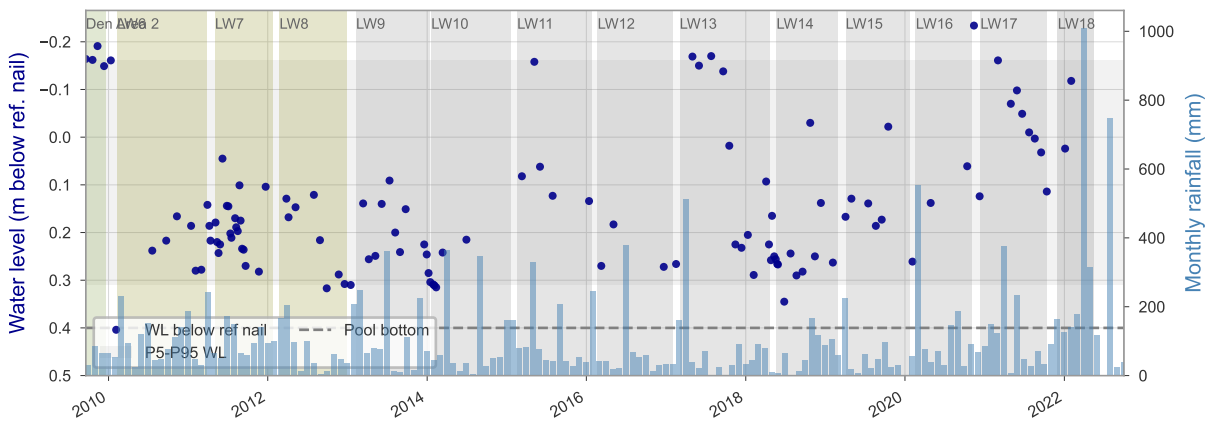
WC_POOL72A



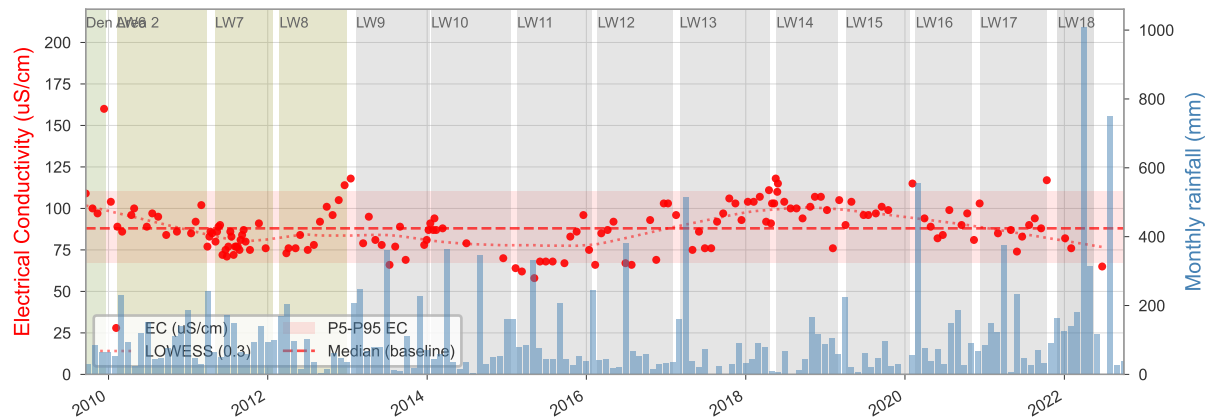
WC_POOL72A



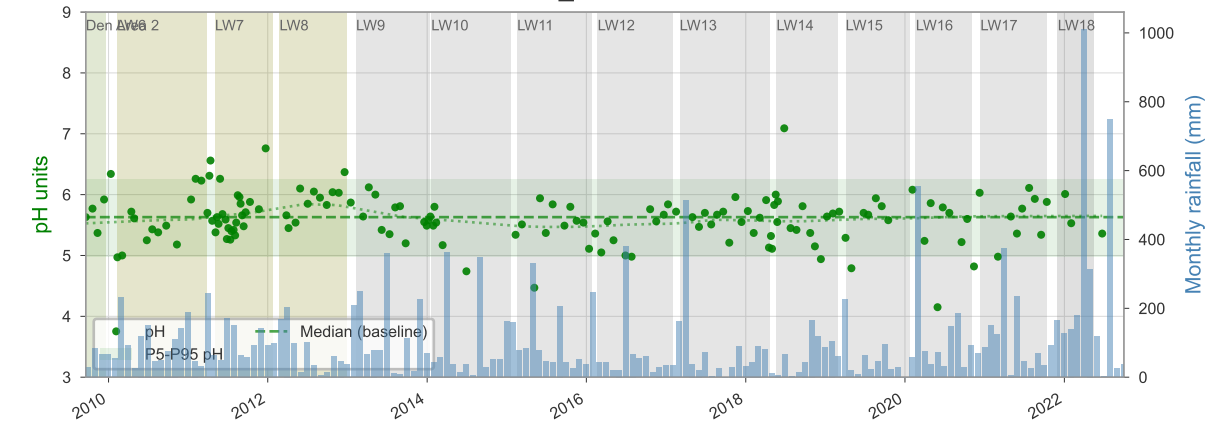
WC_POOL72B



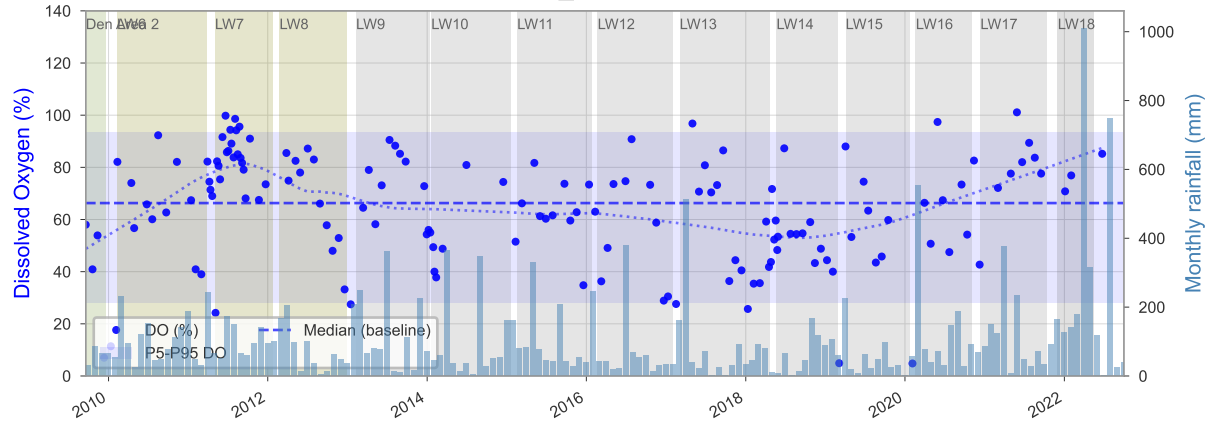
WC_POOL72B



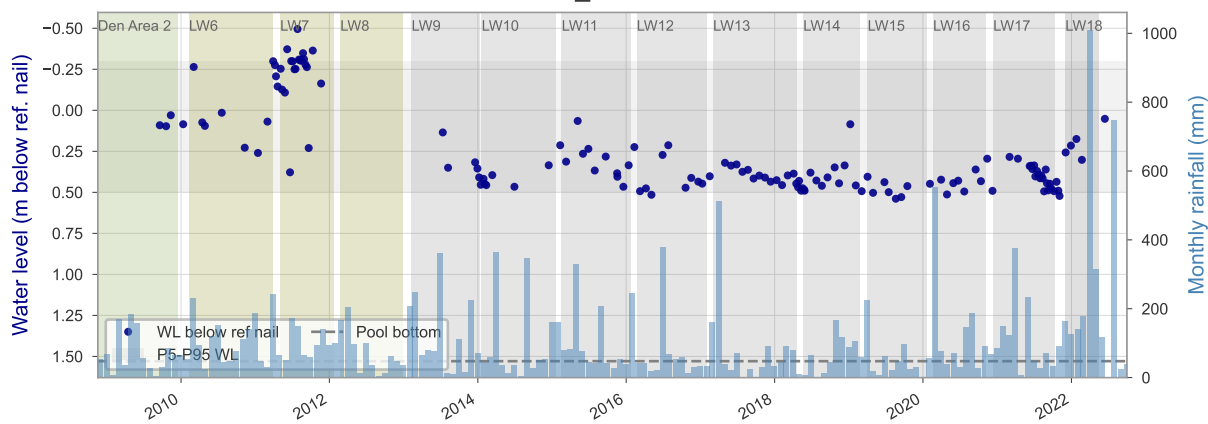
WC_POOL72B



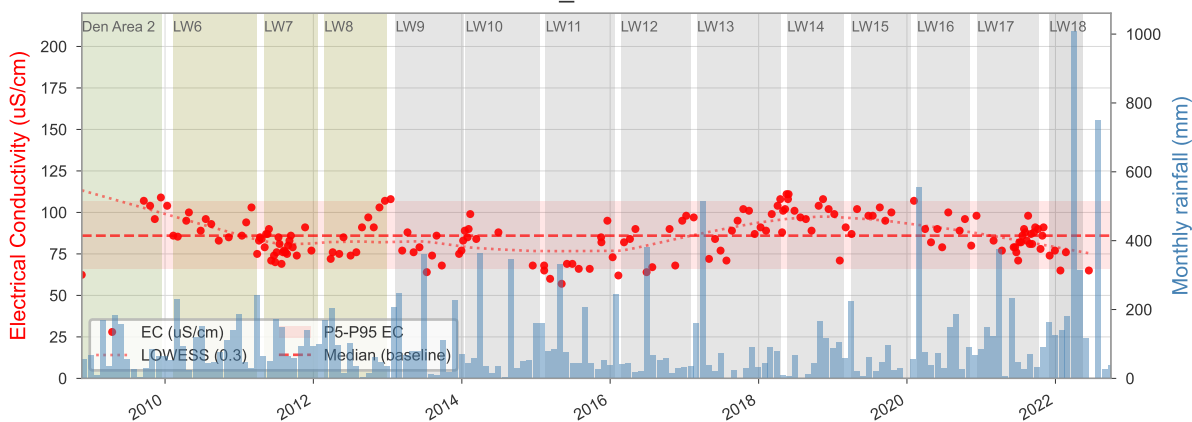
WC_POOL72B



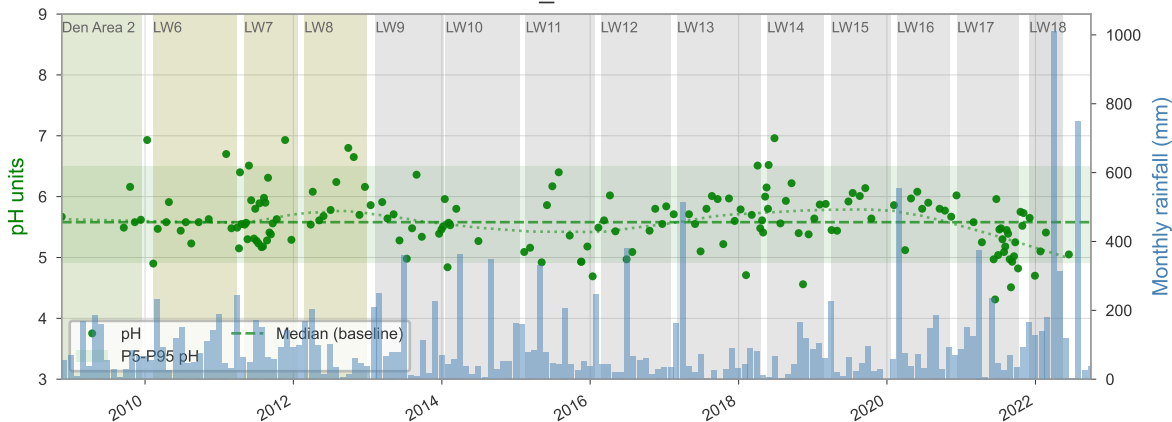
WC_POOL87



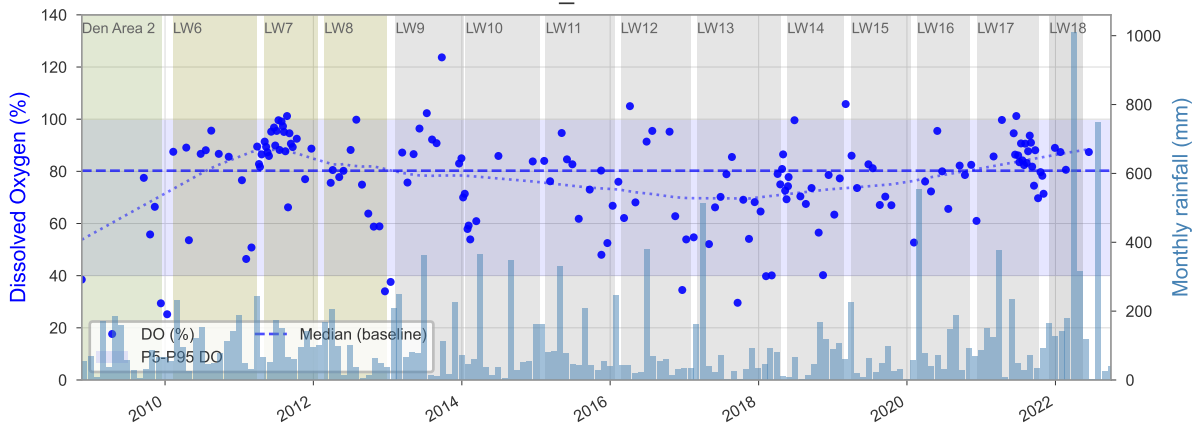
WC_POOL87



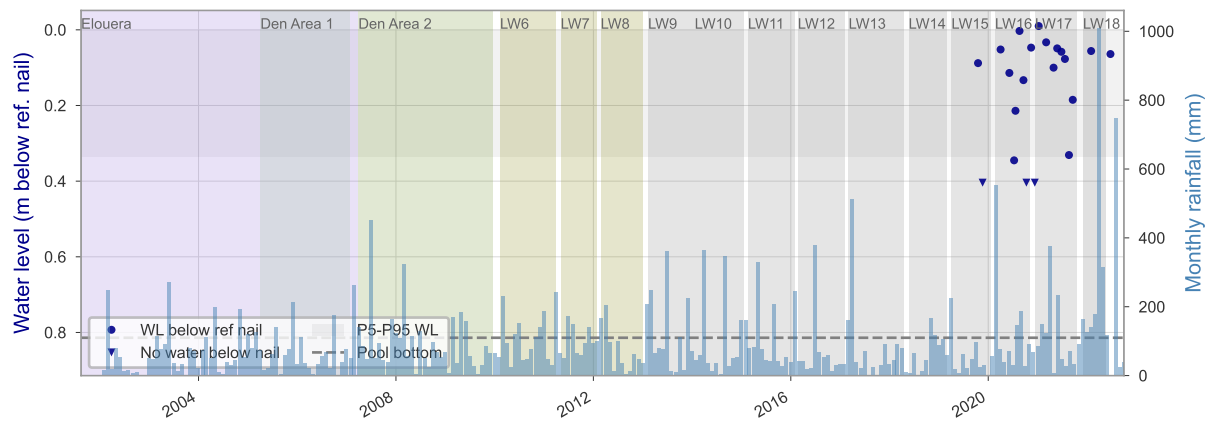
WC_POOL87



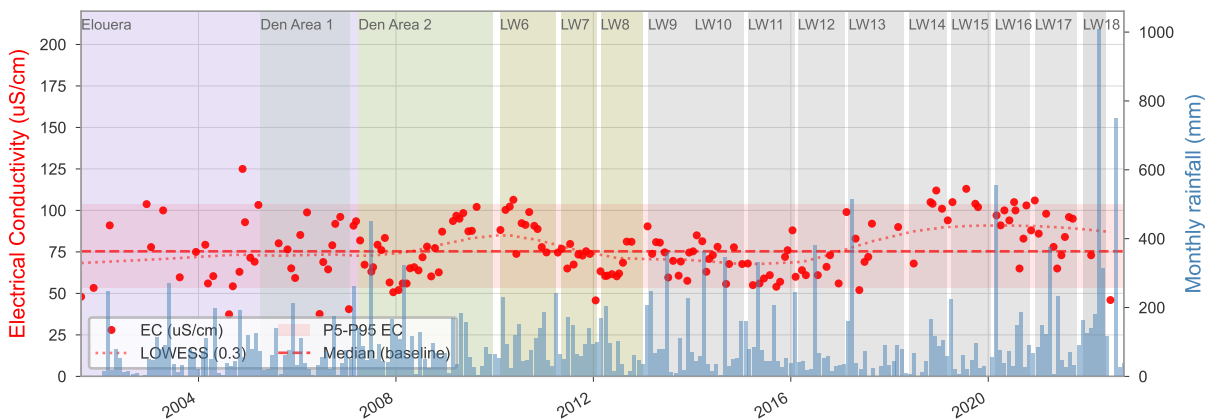
WC_POOL87



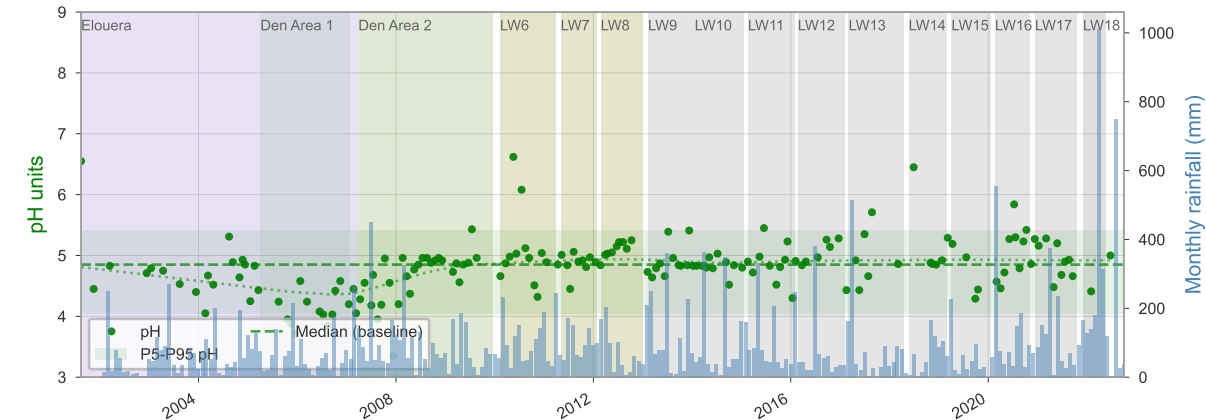
WWU1



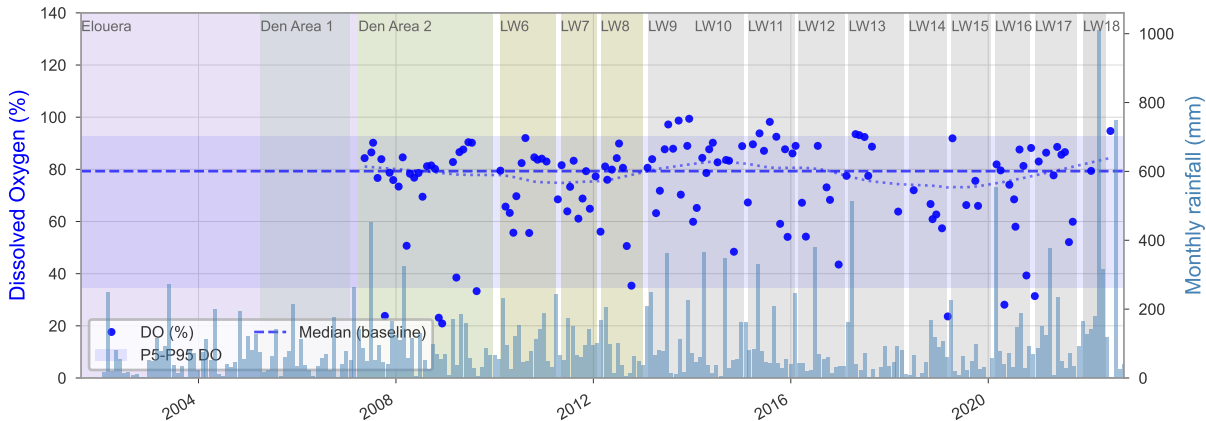
WWU1



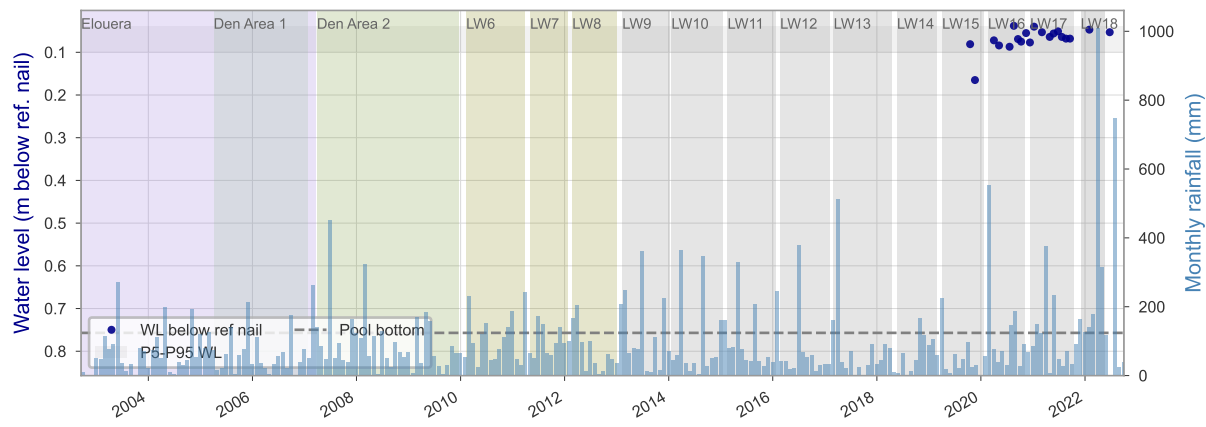
WWU1



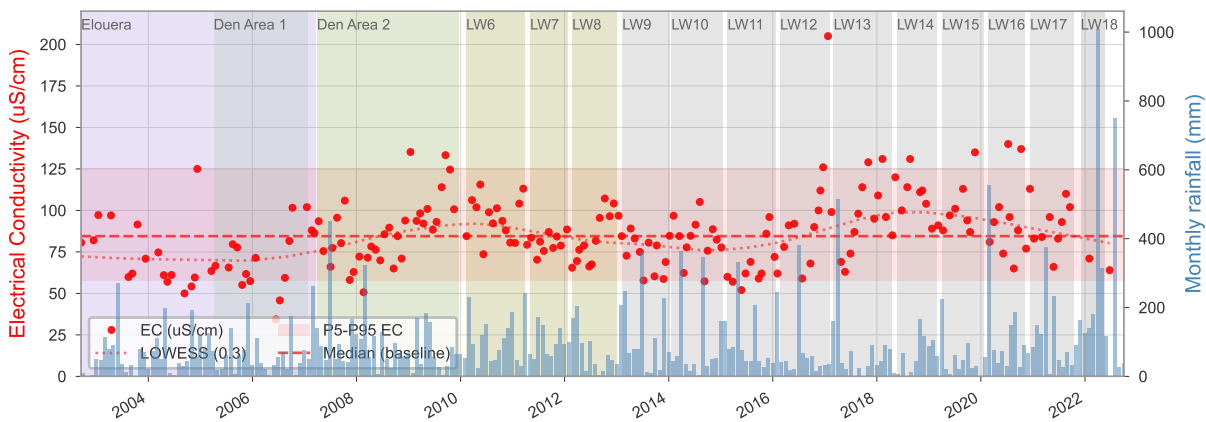
WWU1



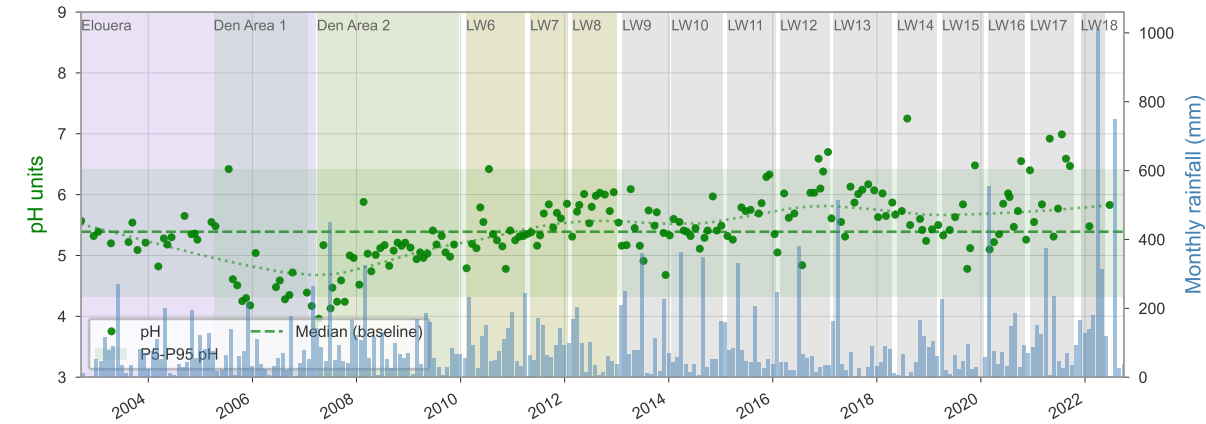
WWU4



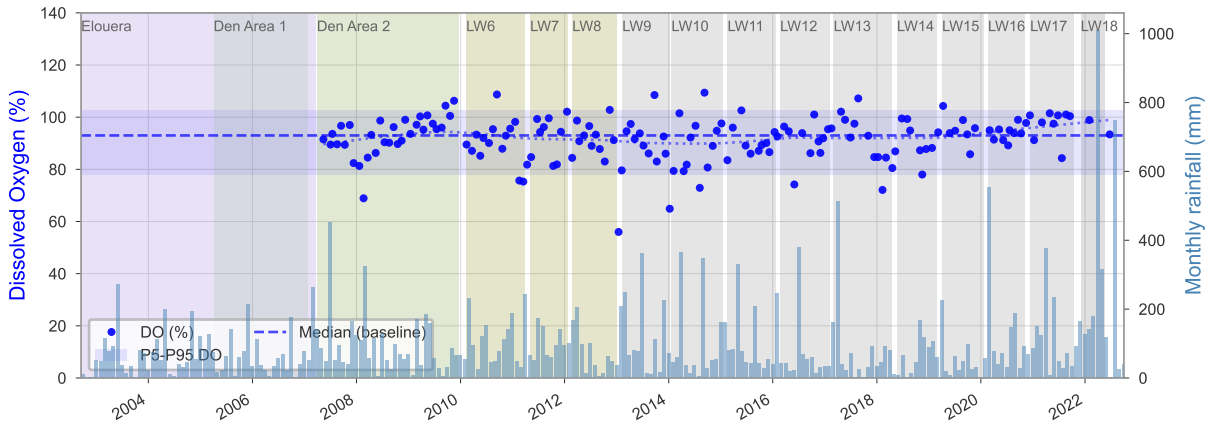
WWU4



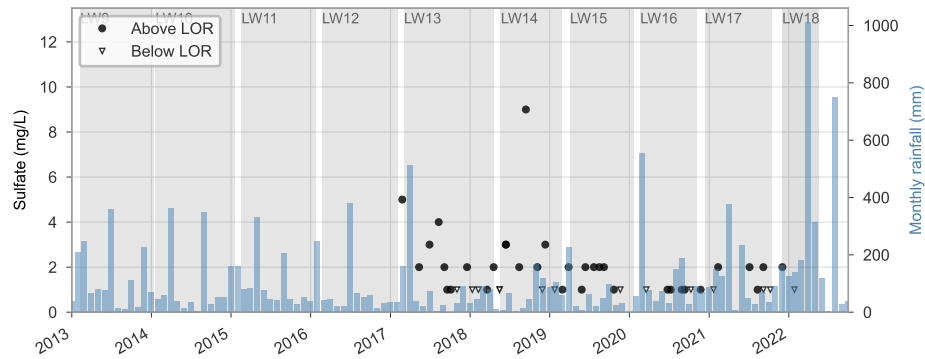
WWU4



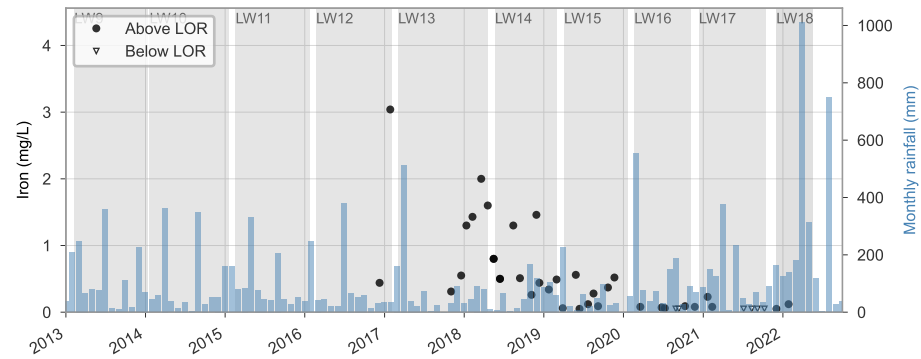
WWU4



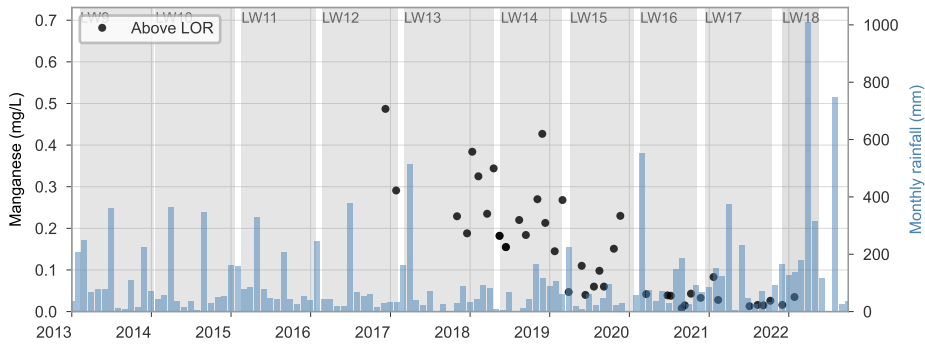
AR19_S1



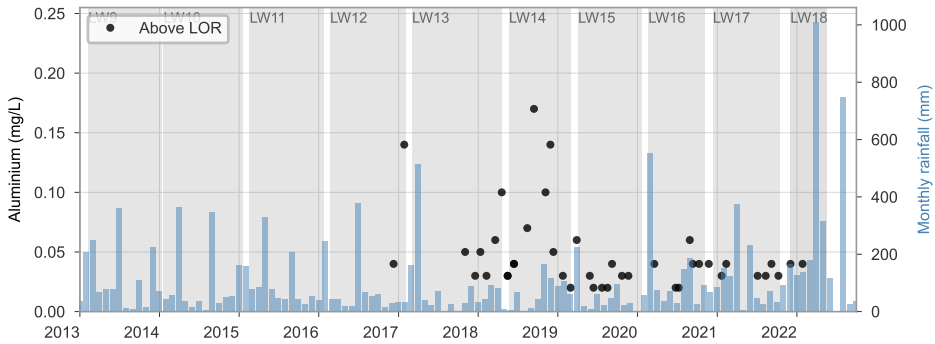
AR19_S1



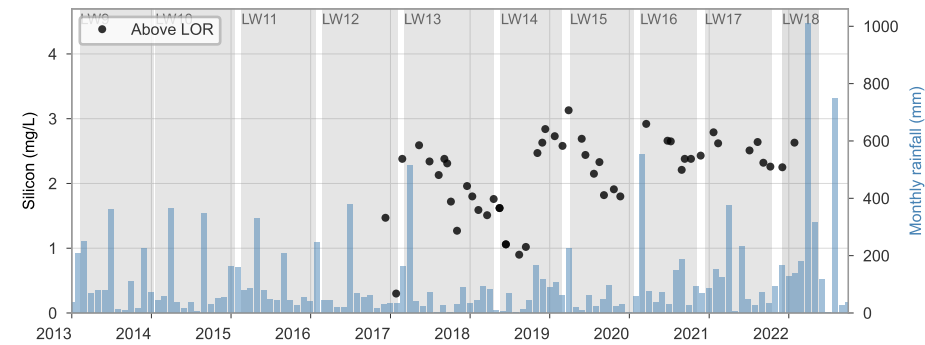
AR19_S1



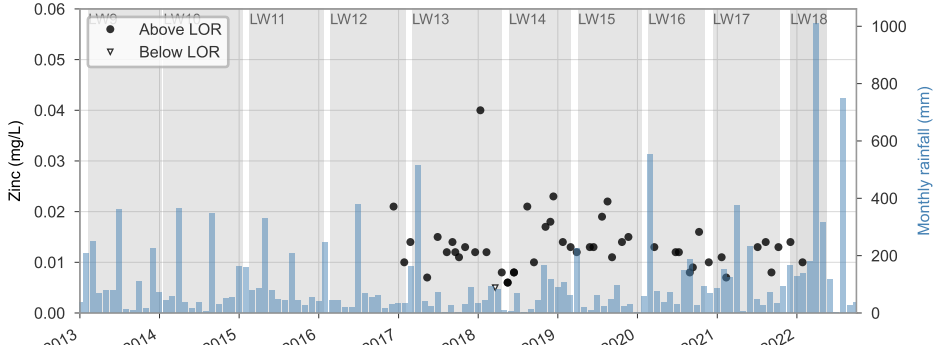
AR19_S1



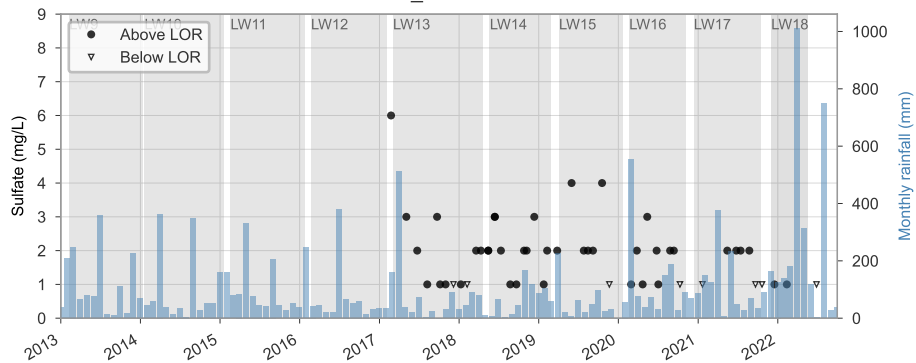
AR19_S1



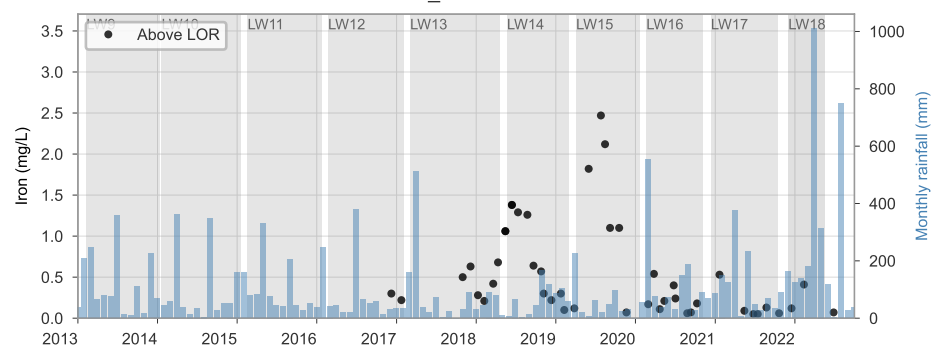
AR19_S1



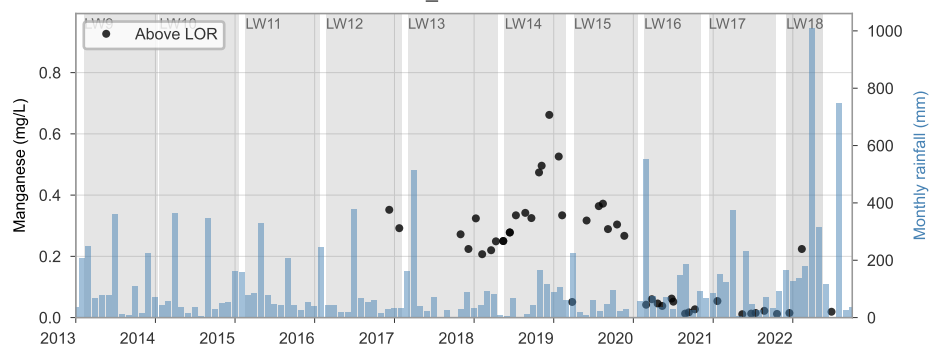
AR31_ROCKBAR1



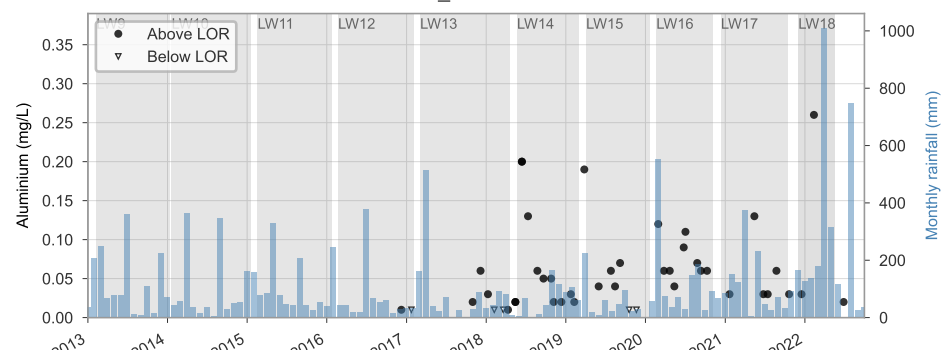
AR31_ROCKBAR1



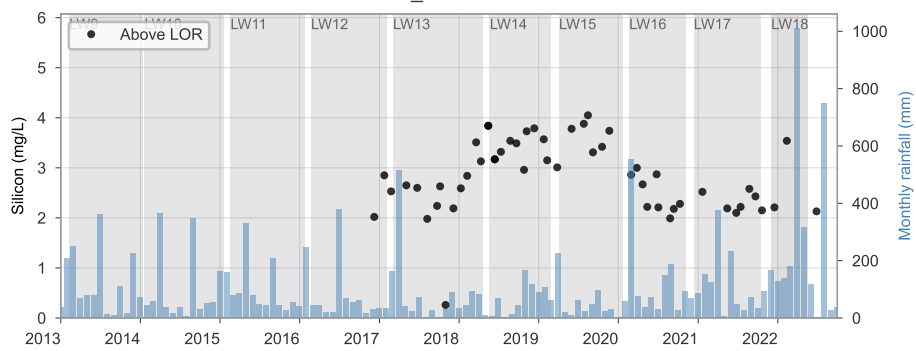
AR31_ROCKBAR1



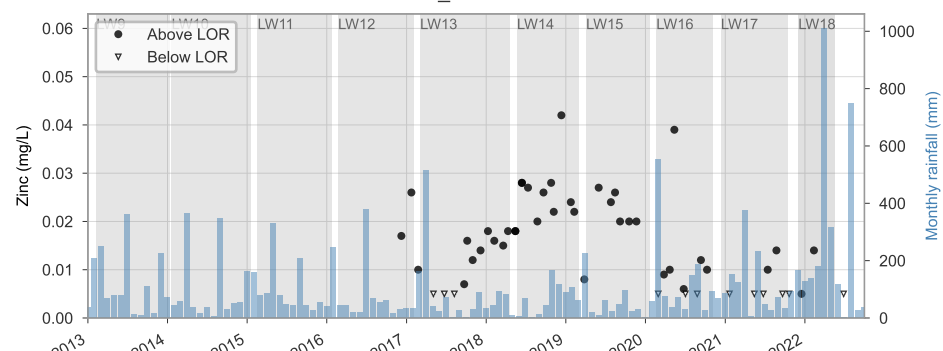
AR31_ROCKBAR1

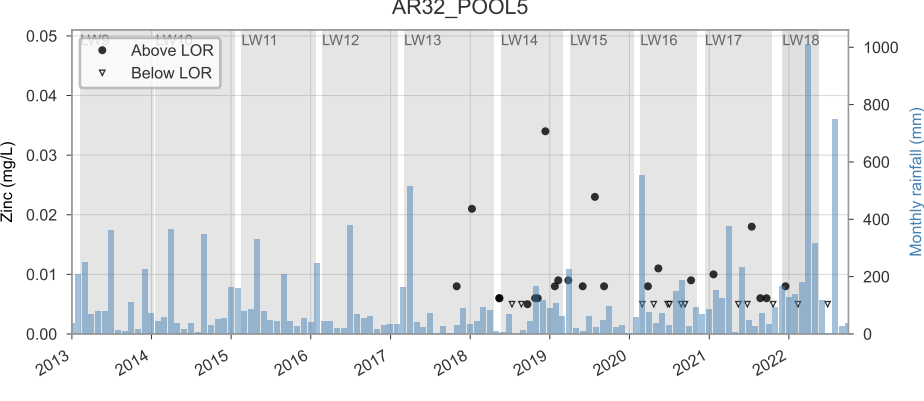
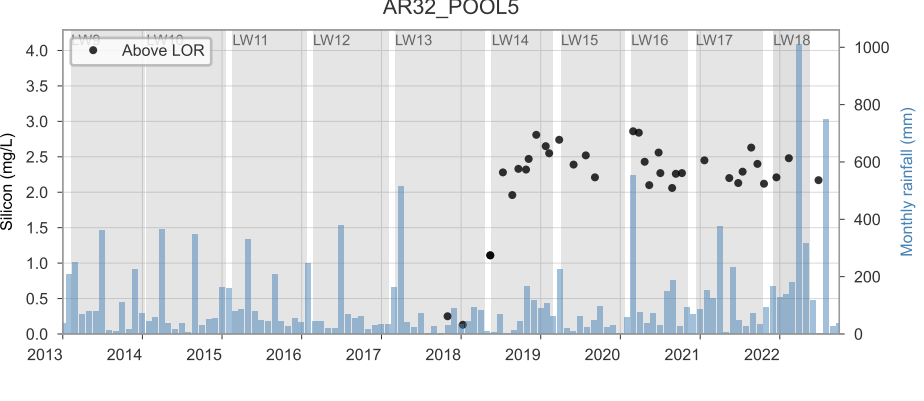
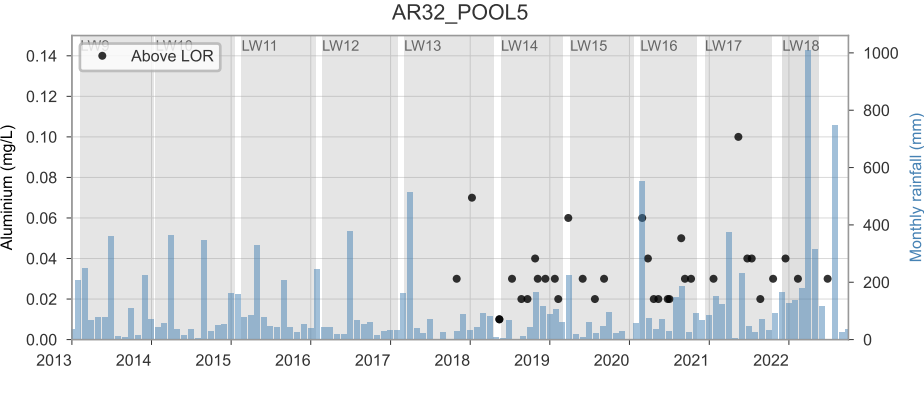
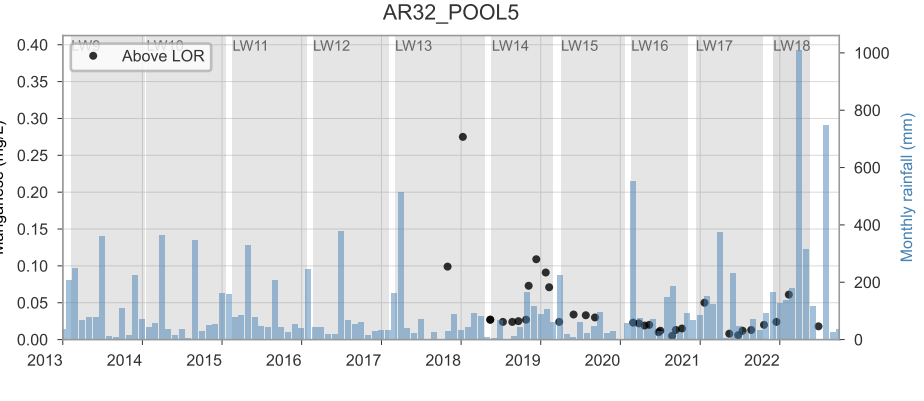
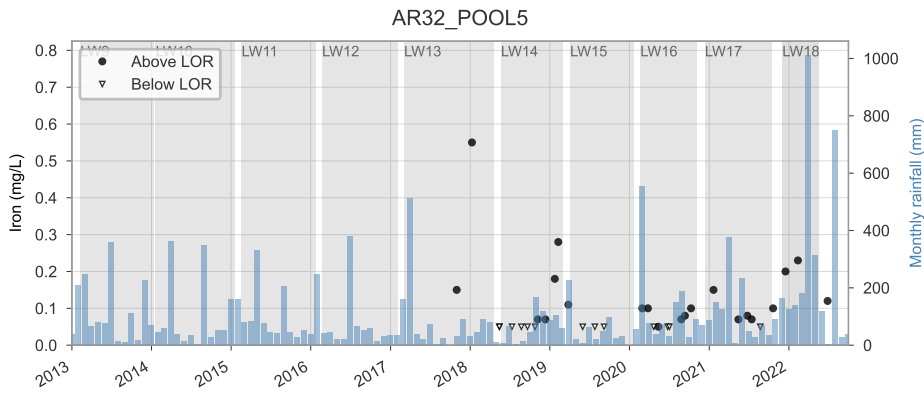
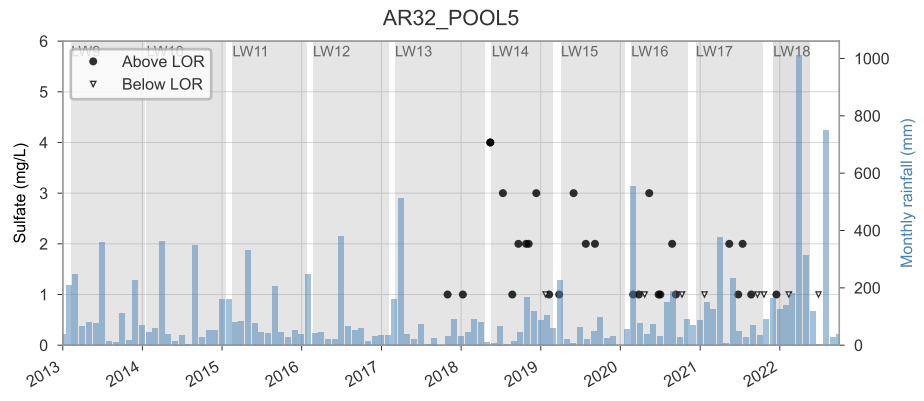


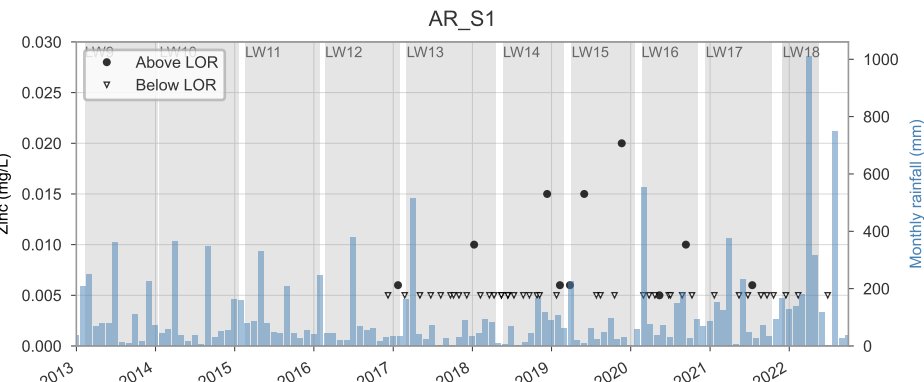
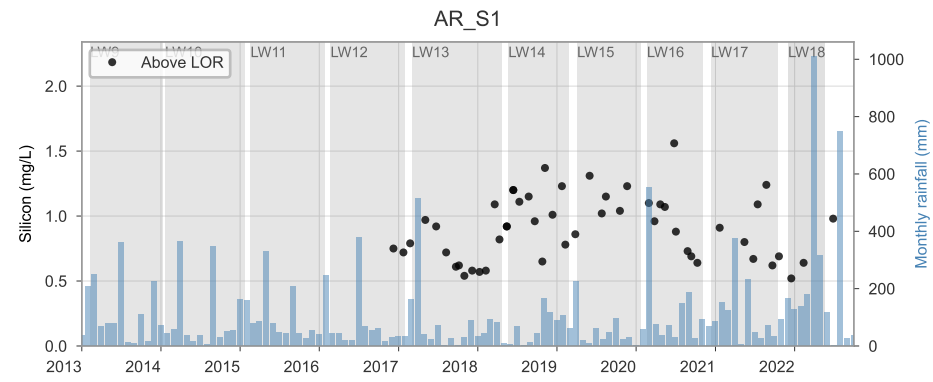
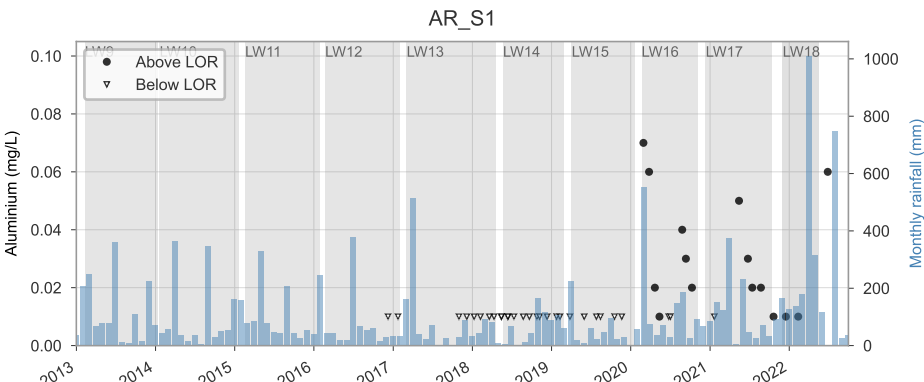
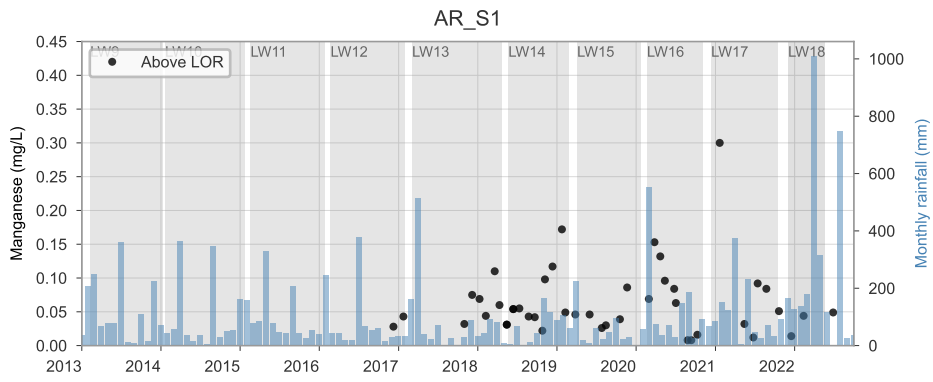
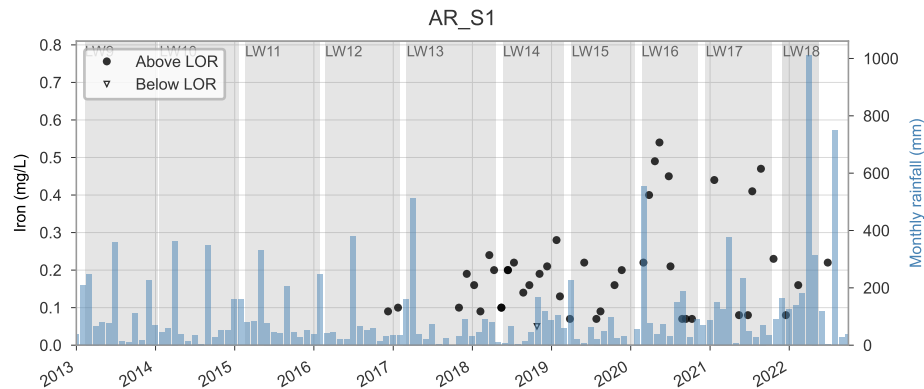
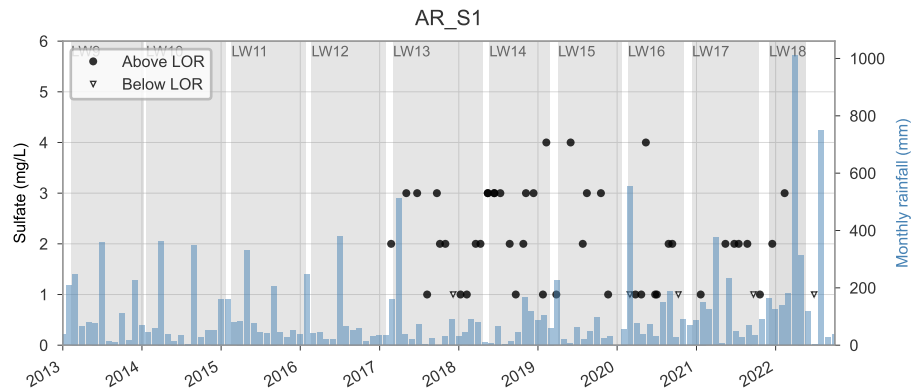
AR31_ROCKBAR1



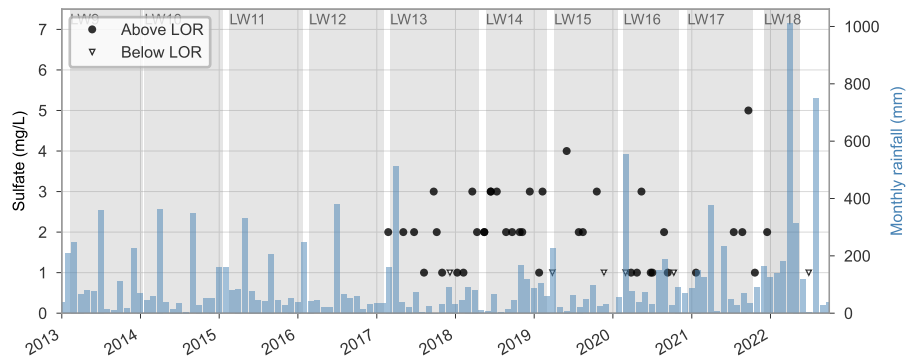
AR31_ROCKBAR1



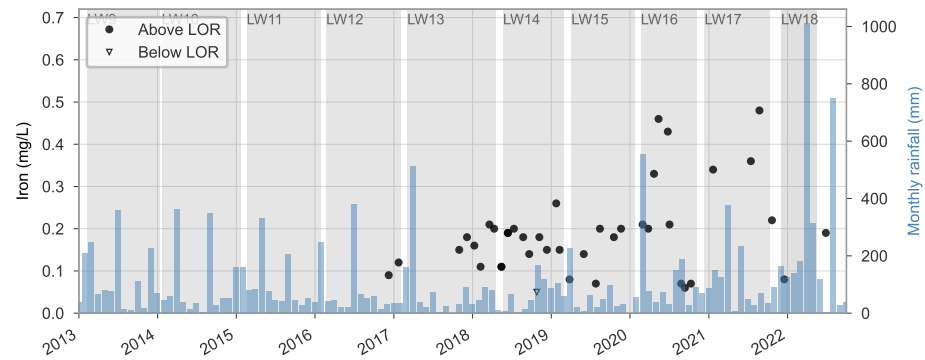




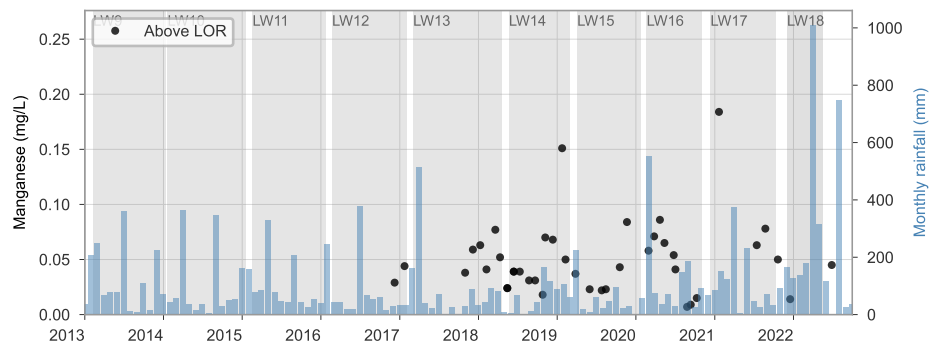
AR_S2



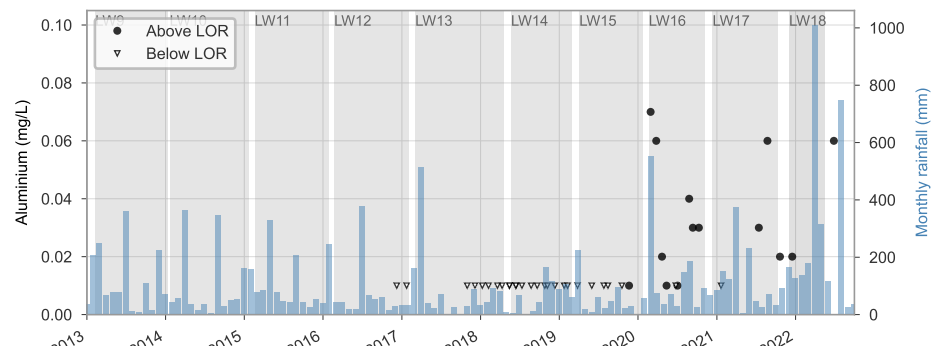
AR_S2



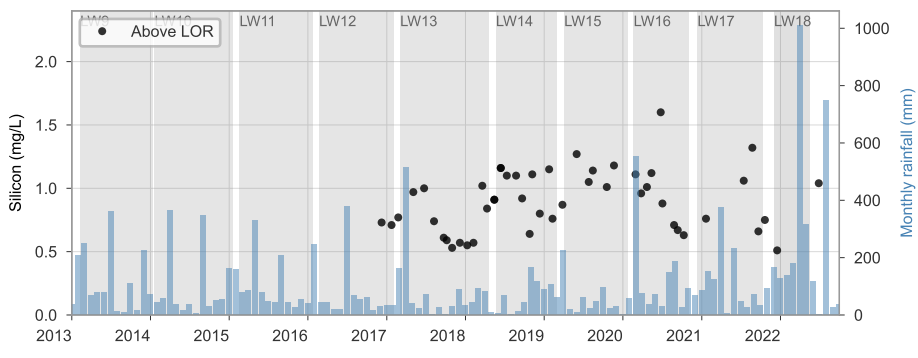
AR_S2



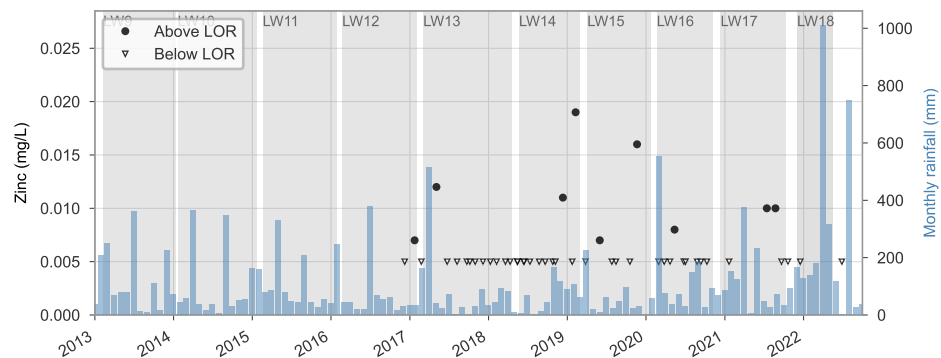
AR_S2

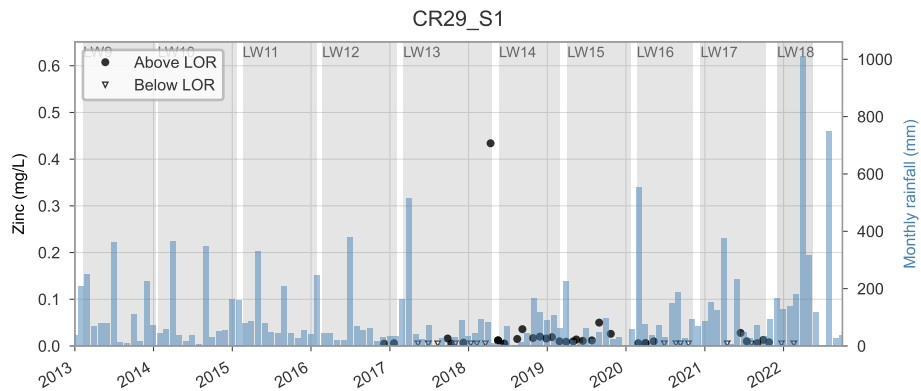
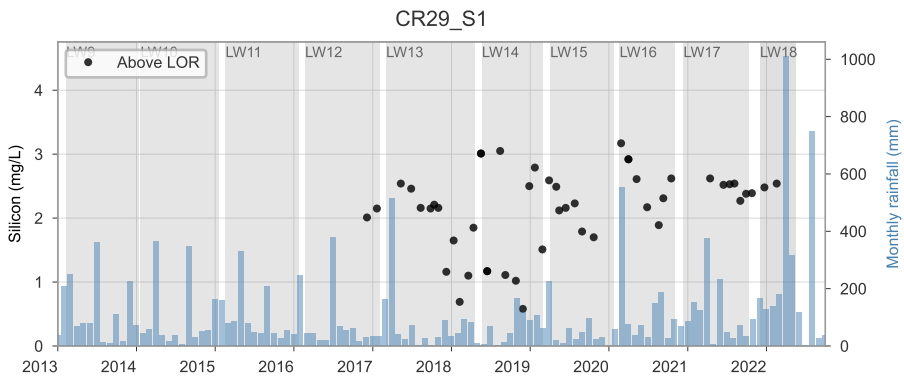
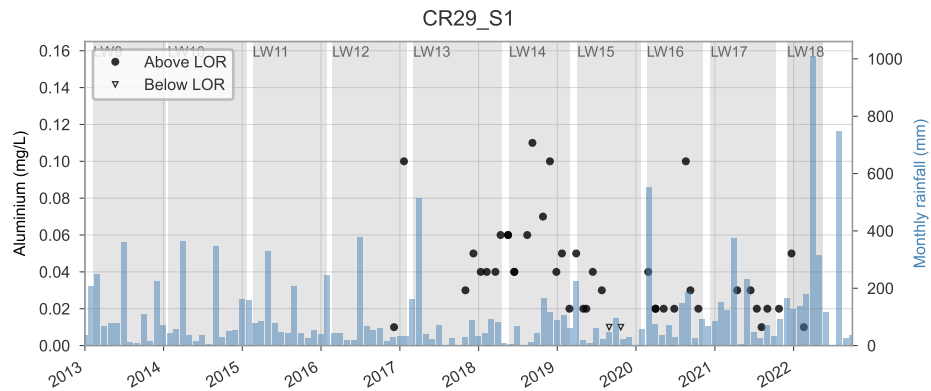
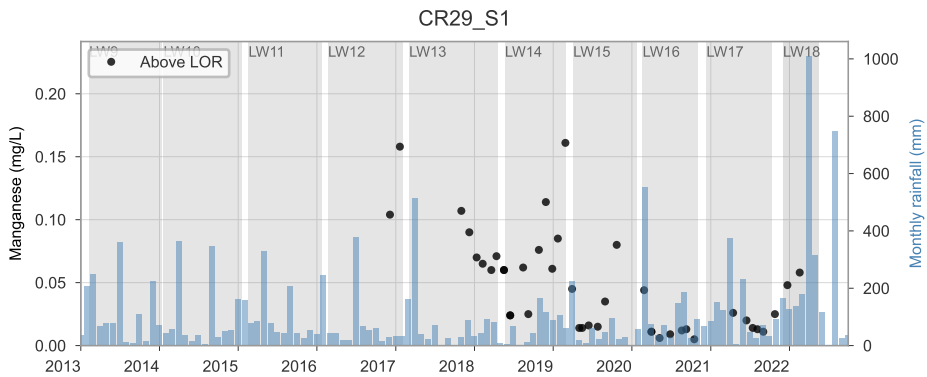
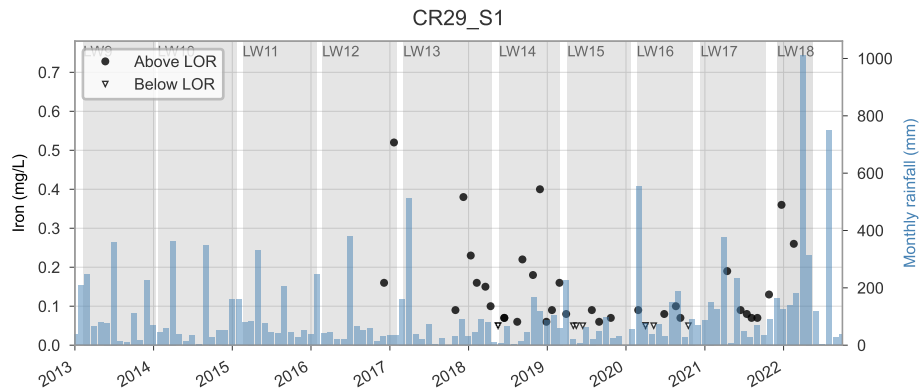
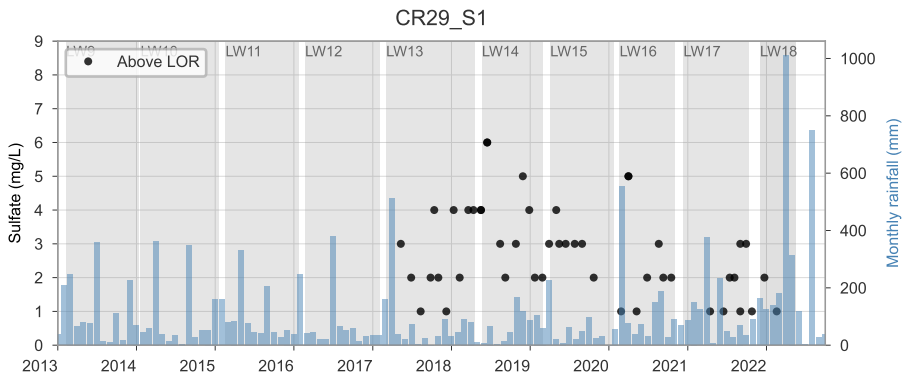


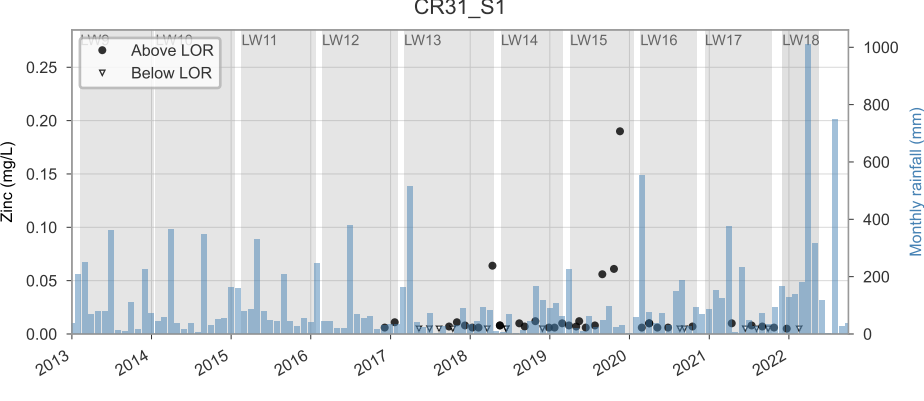
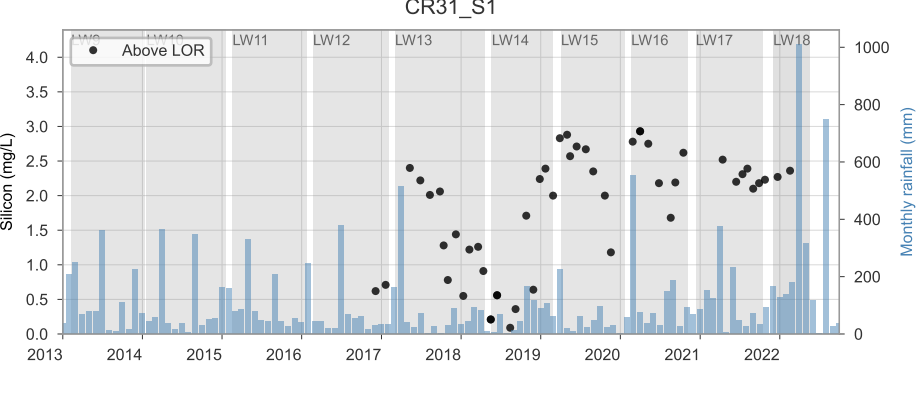
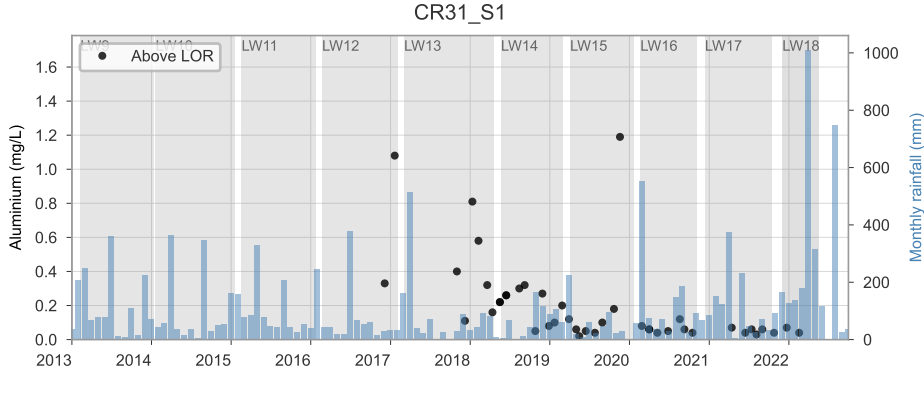
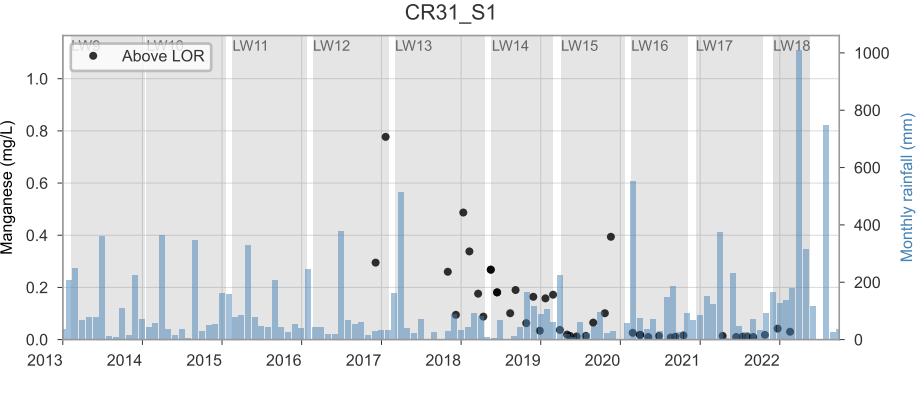
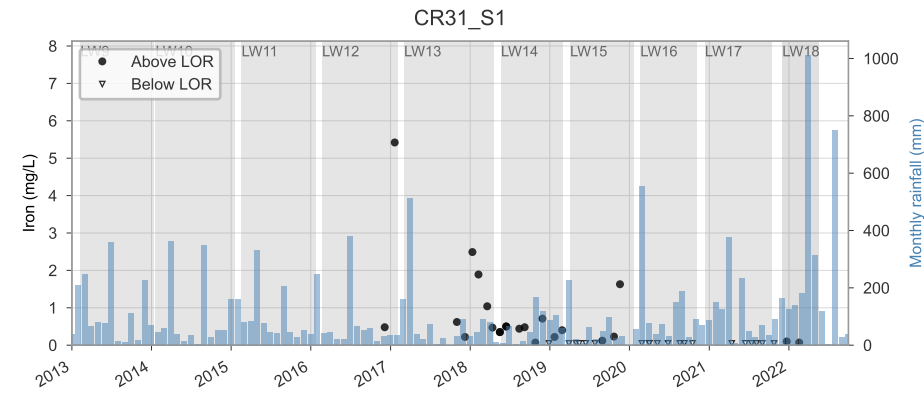
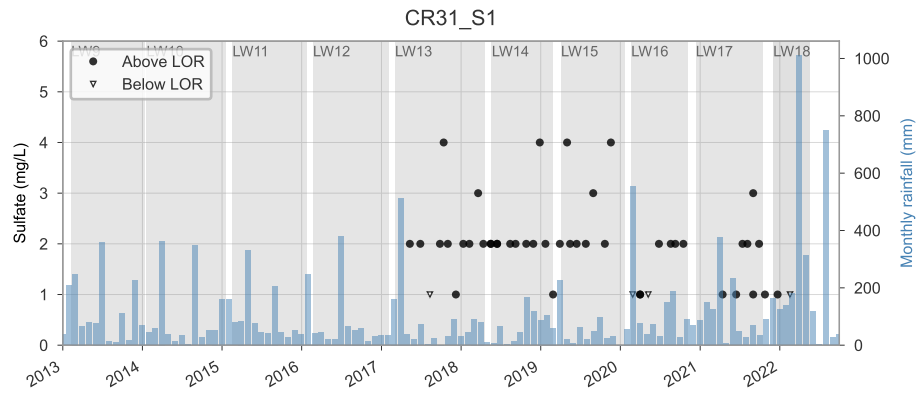
AR_S2



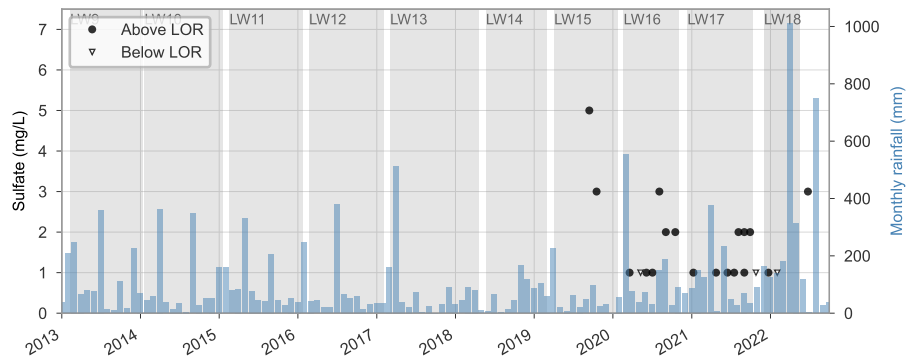
AR_S2



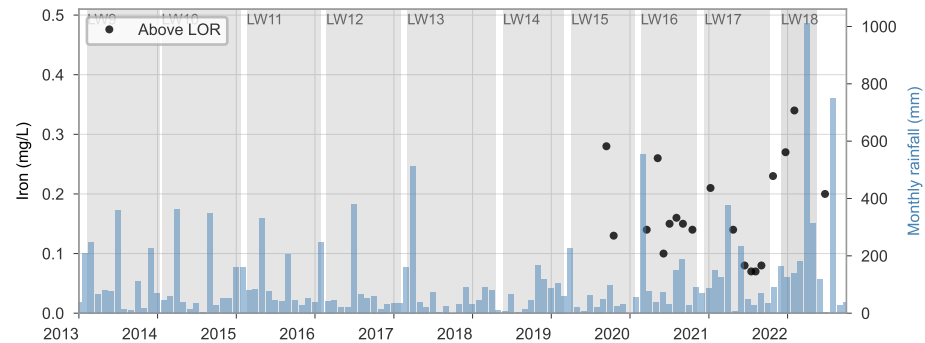




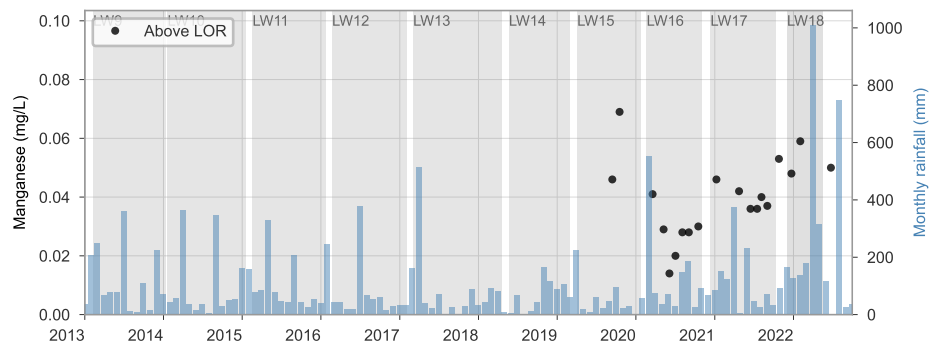
CR36_S1



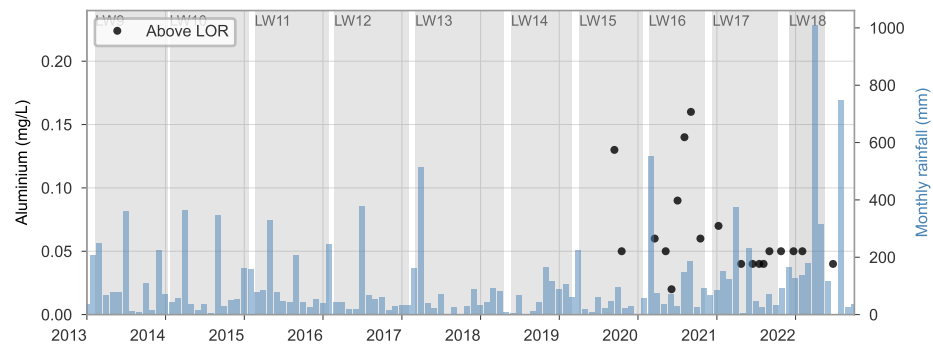
CR36_S1



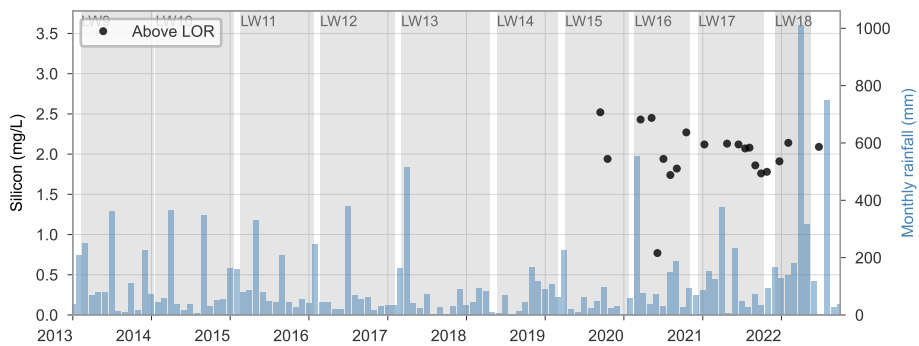
CR36_S1



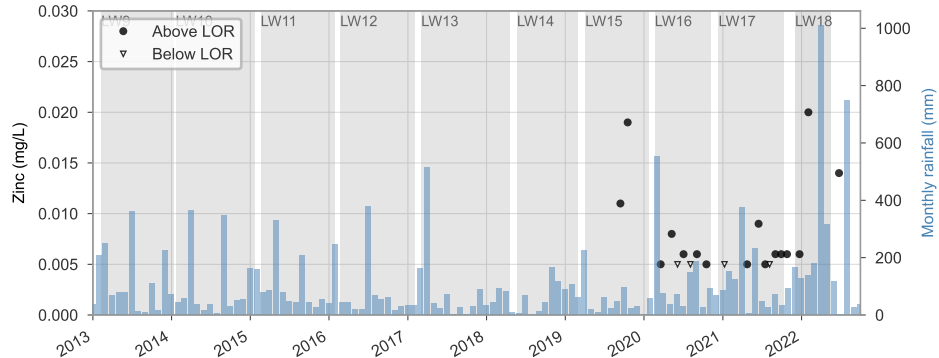
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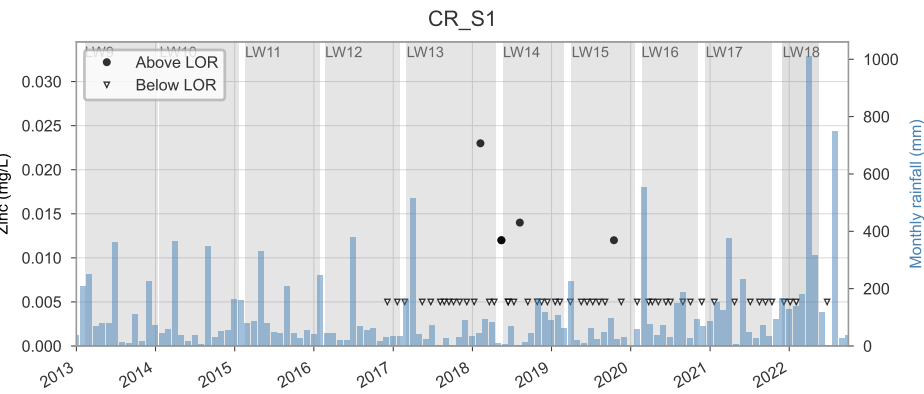
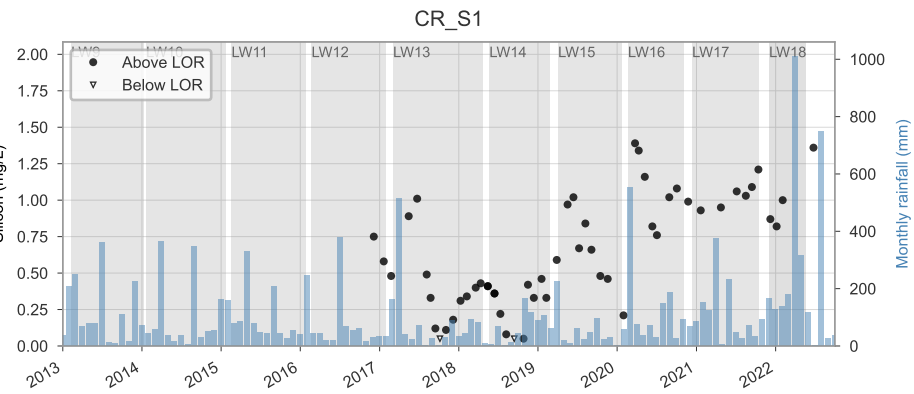
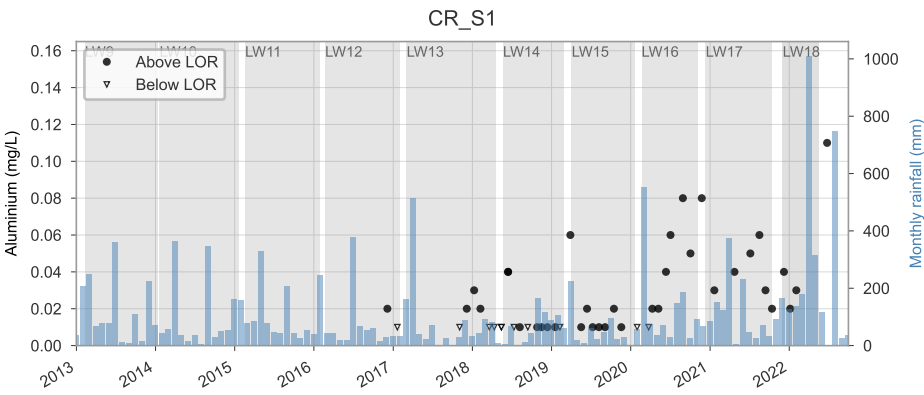
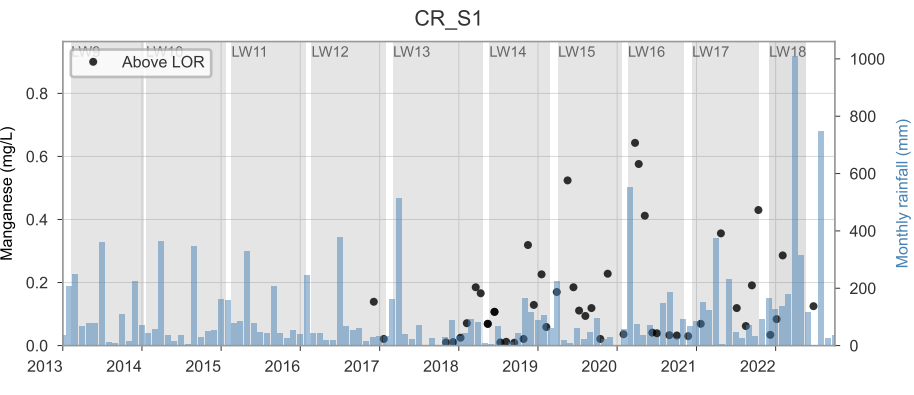
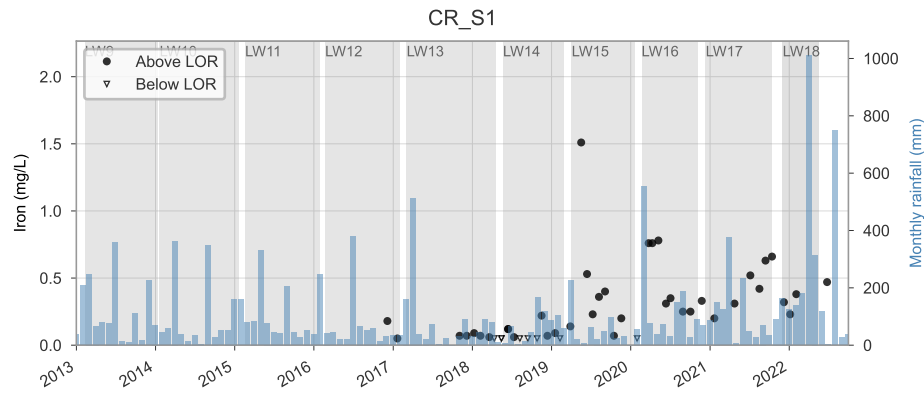
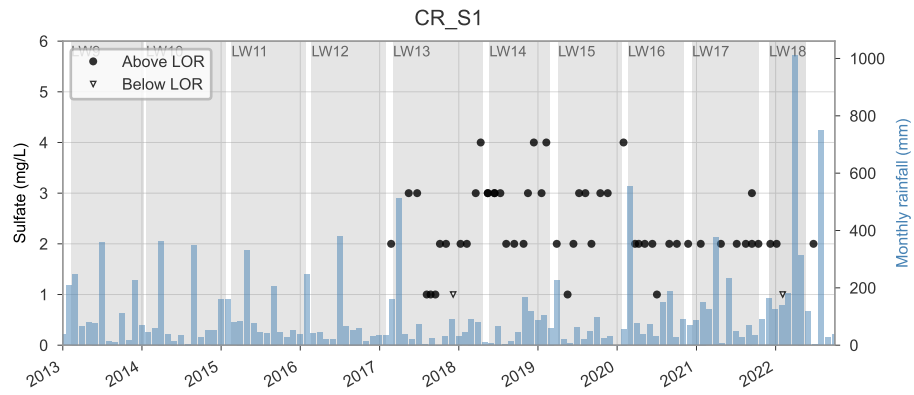


CR36_S1

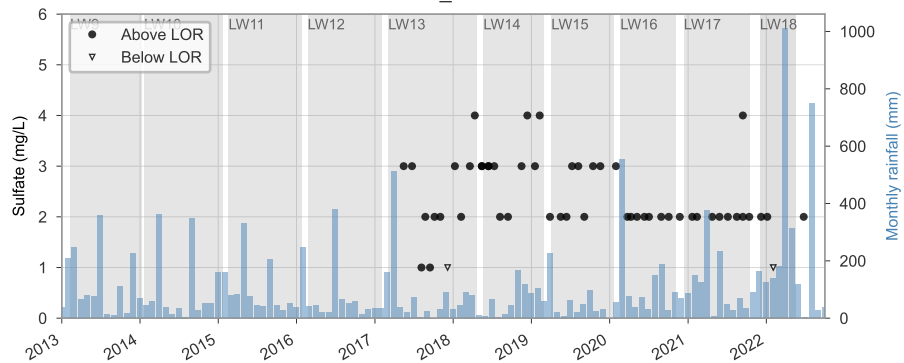


CR36_S1

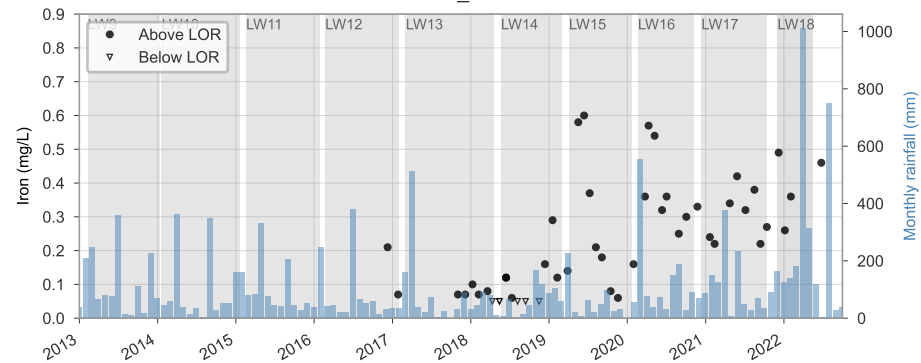




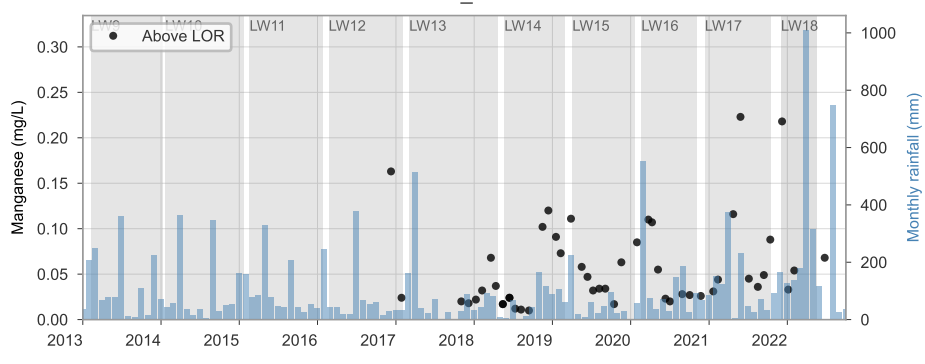
CR_S2



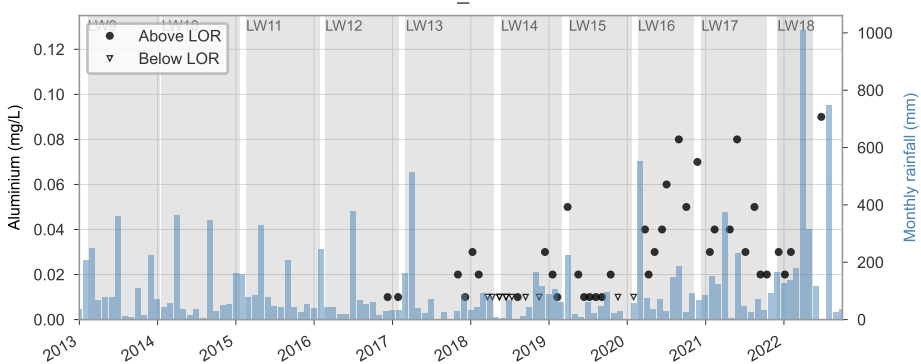
CR_S2



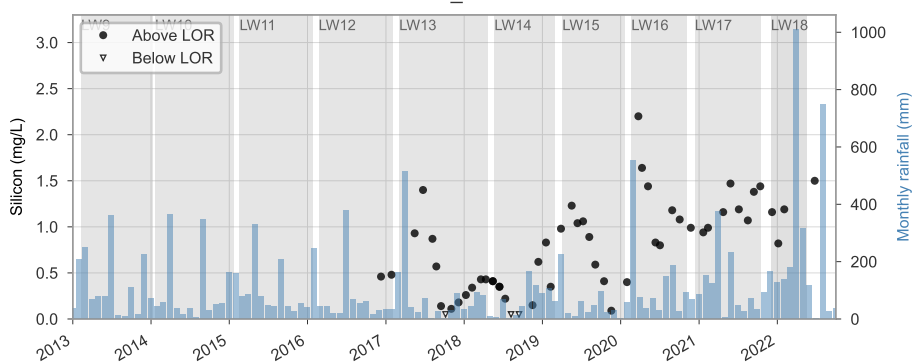
CR_S2



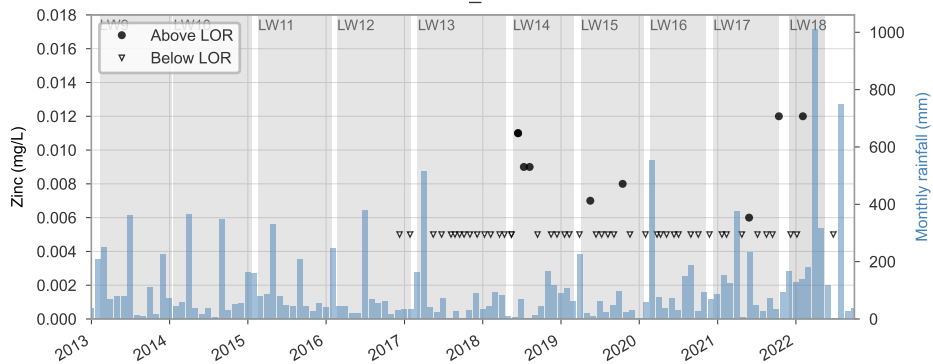
CR_S2



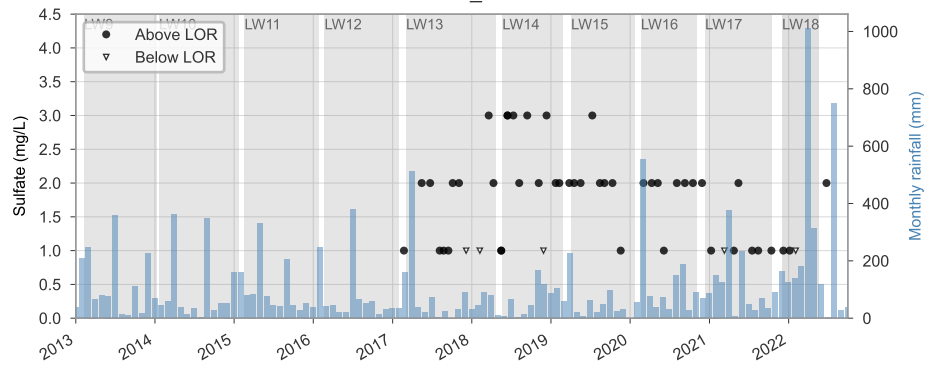
CR_S2



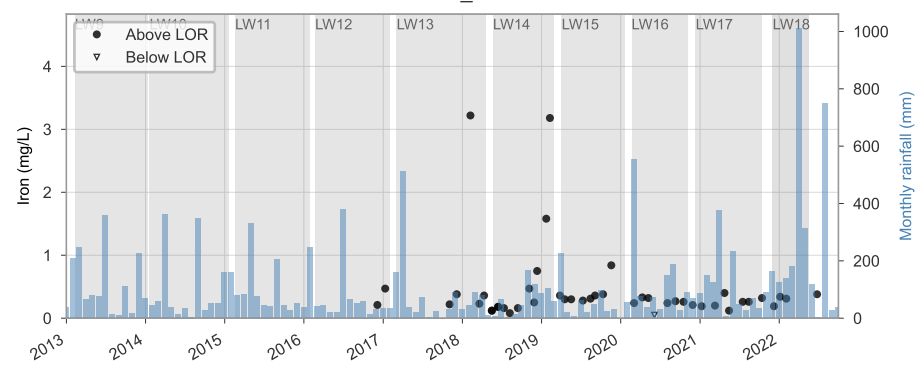
CR_S2



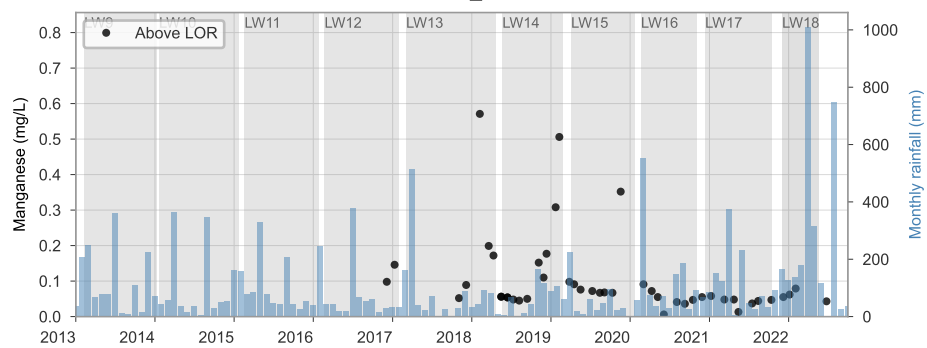
DC10_S1



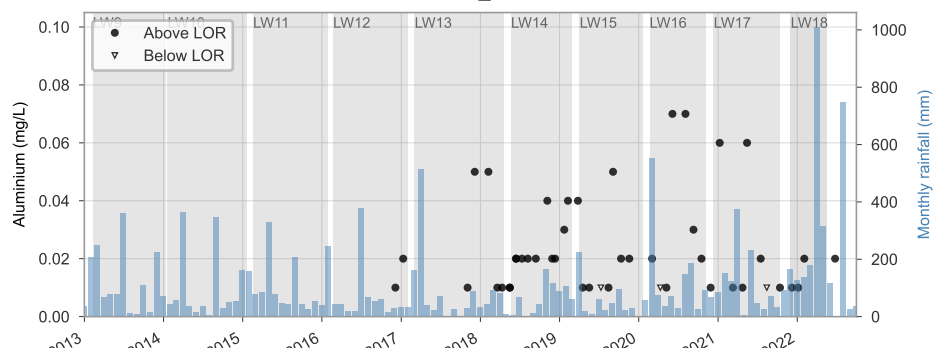
DC10_S1



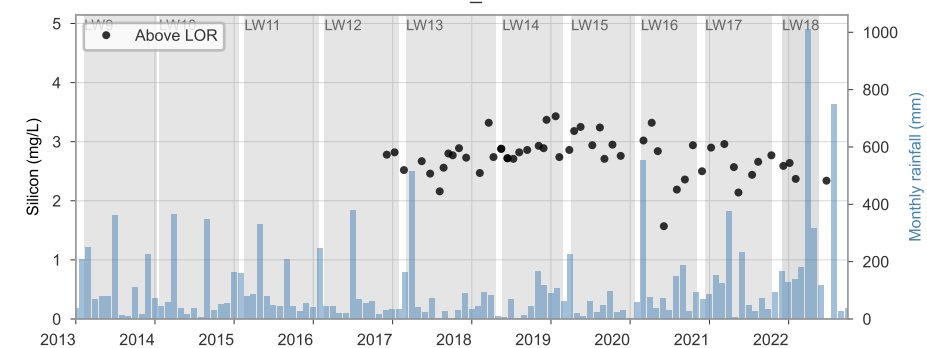
DC10_S1



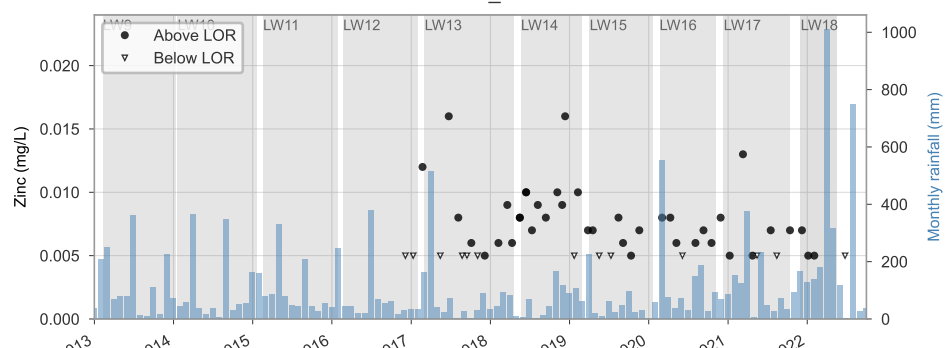
DC10_S1



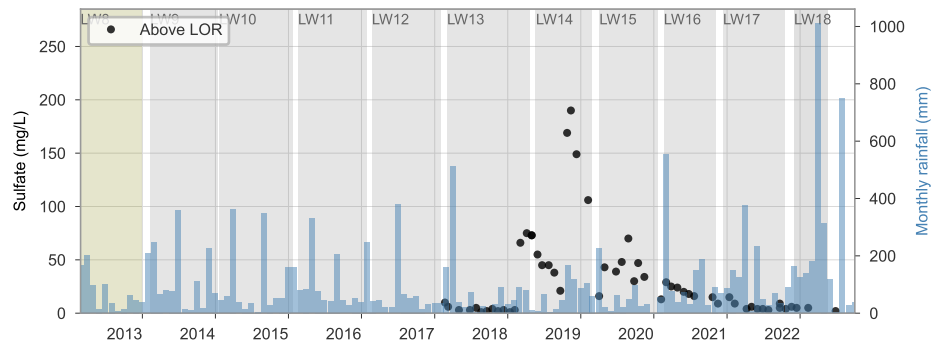
DC10_S1



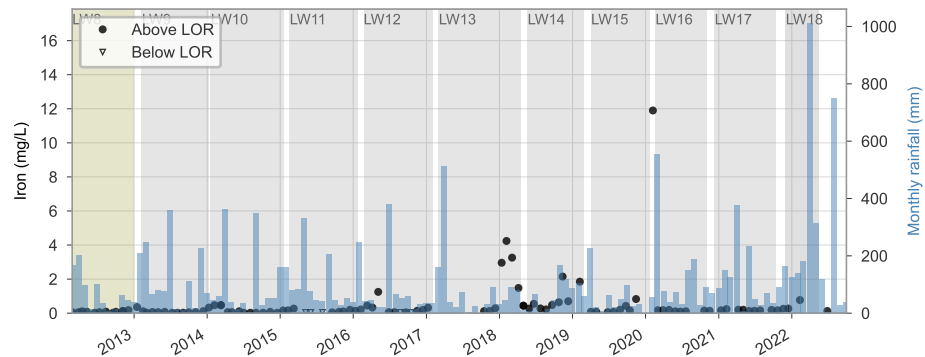
DC10_S1



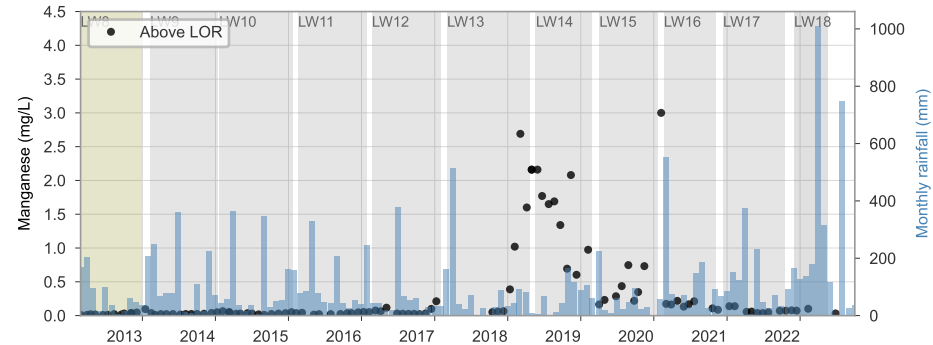
DC13_POOL2B



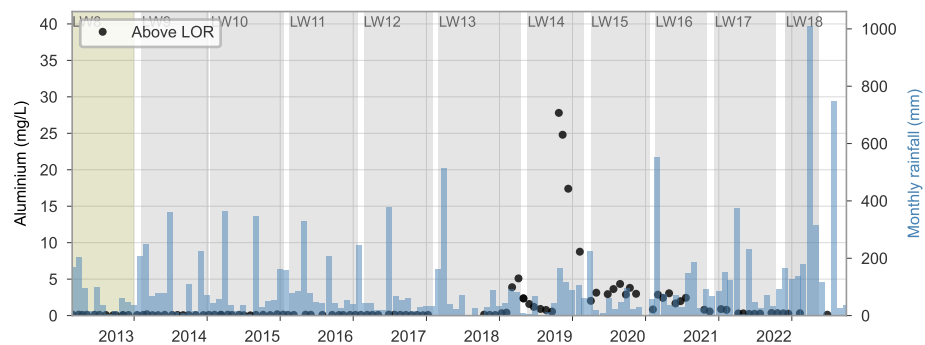
DC13_POOL2B



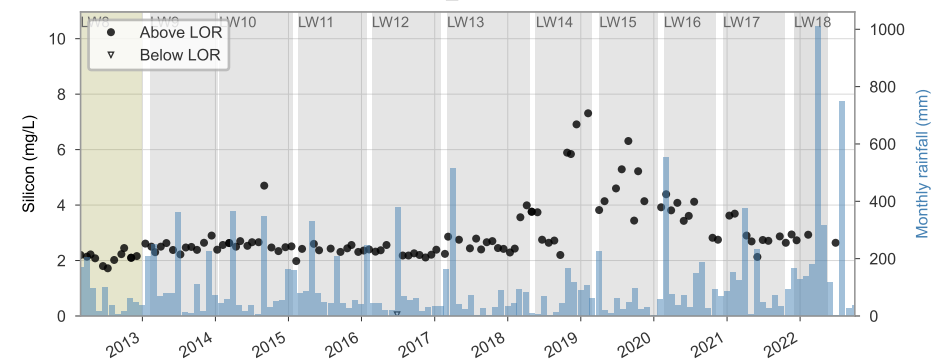
DC13_POOL2B



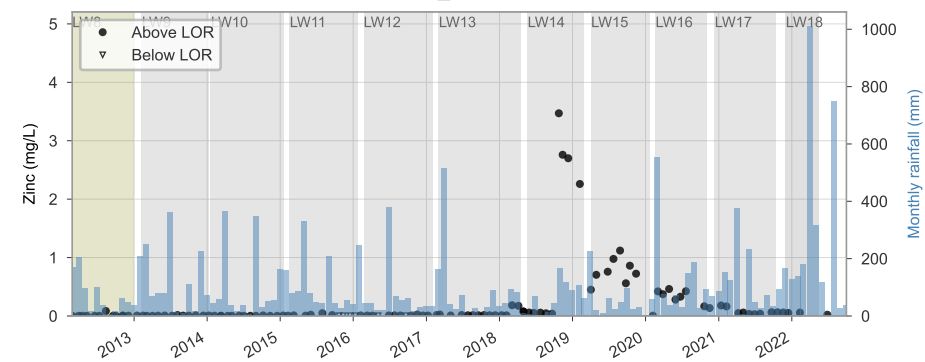
DC13_POOL2B



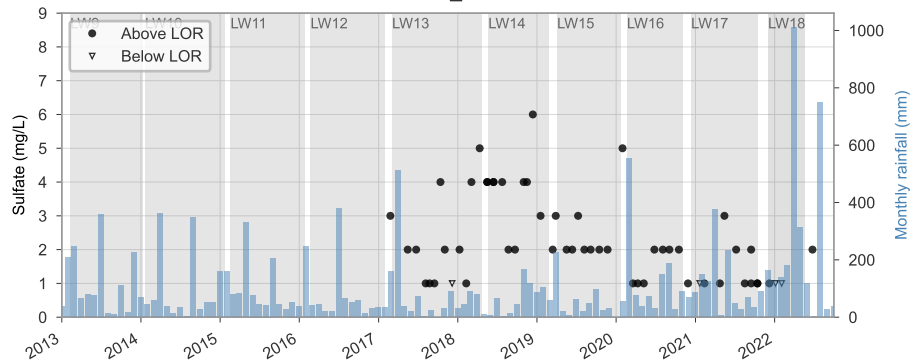
DC13_POOL2B



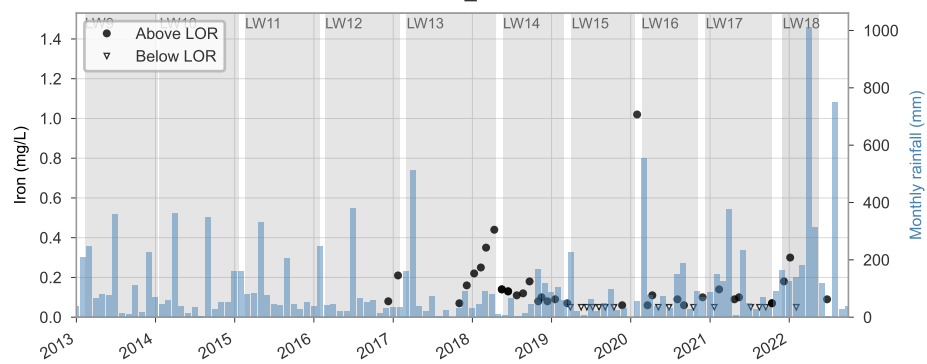
DC13_POOL2B



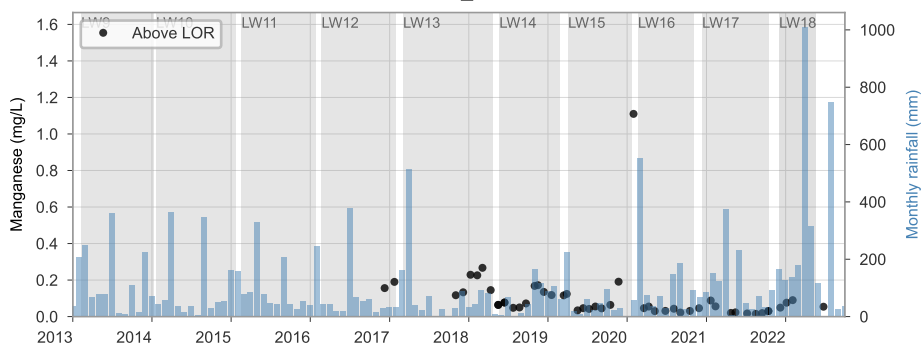
DC8_S1



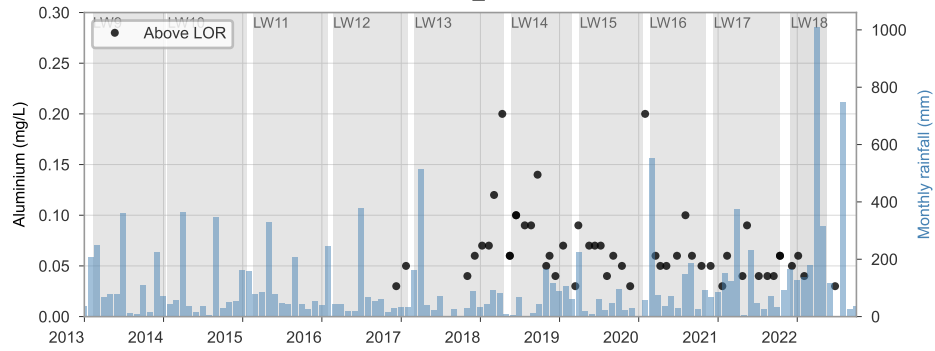
DC8_S1



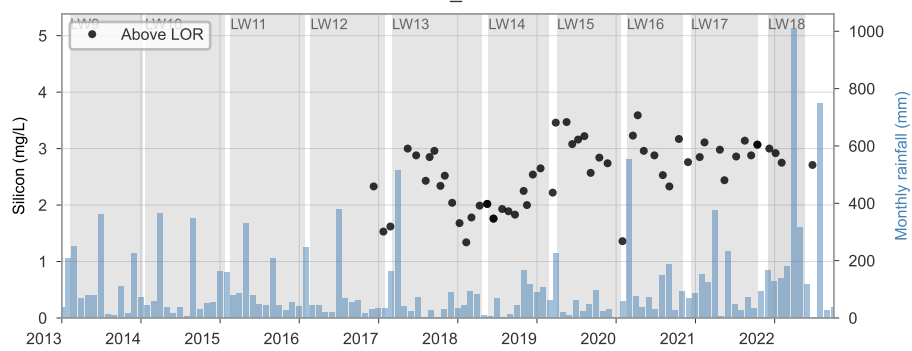
DC8_S1



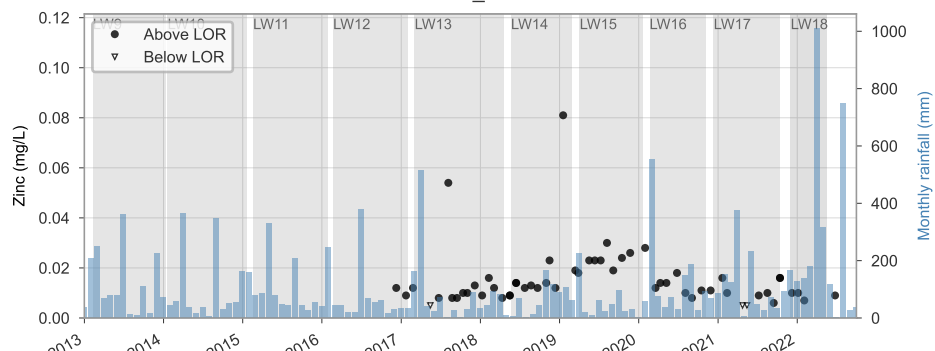
DC8_S1



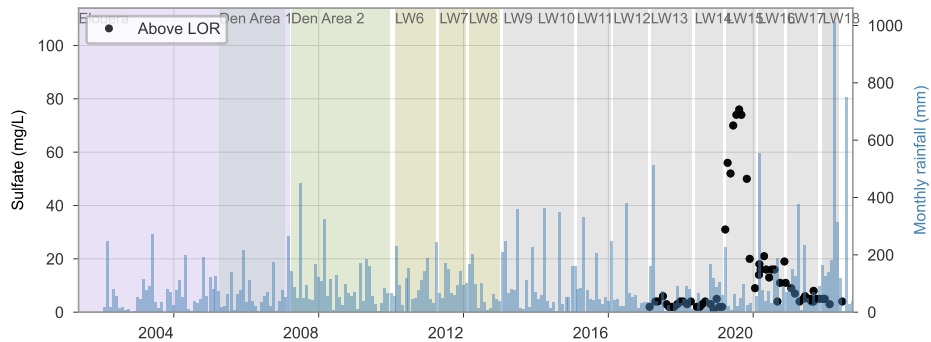
DC8_S1



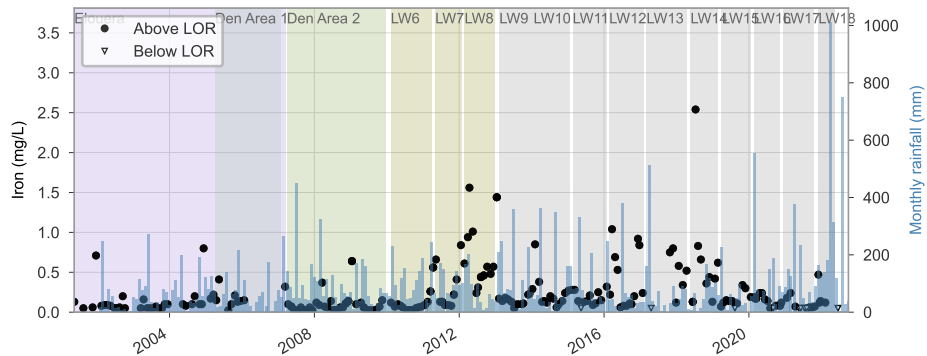
DC8_S1



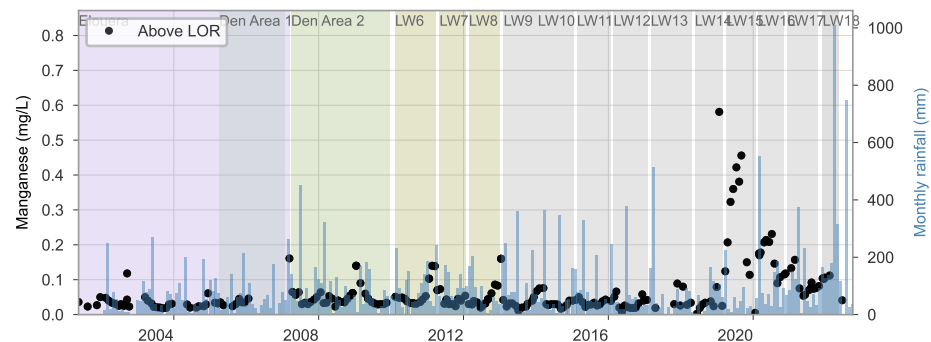
DCC_FR6



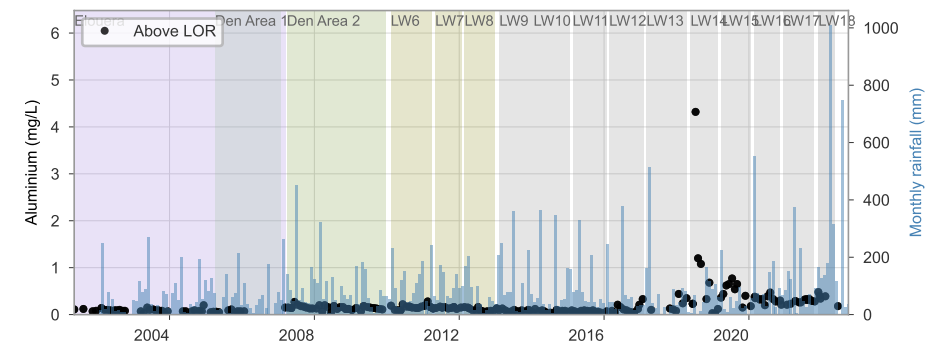
DCC_FR6



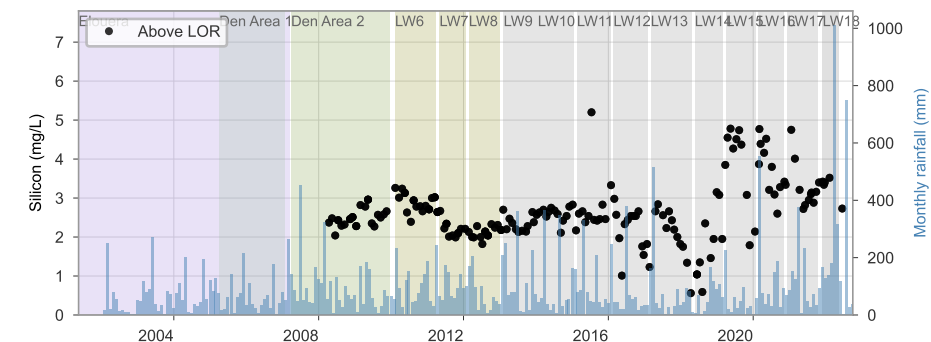
DCC_FR6



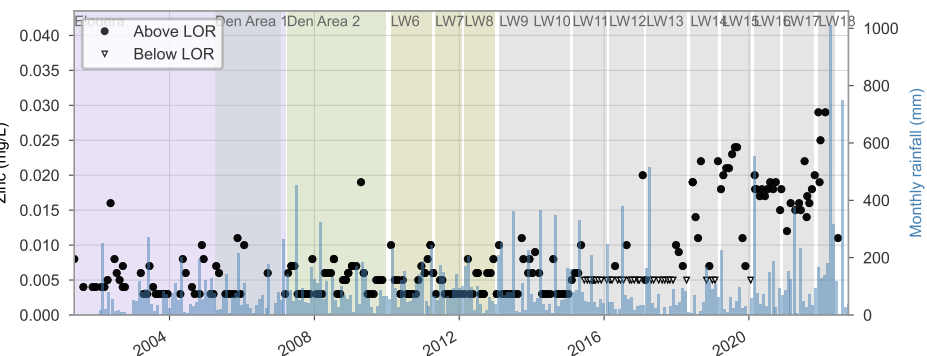
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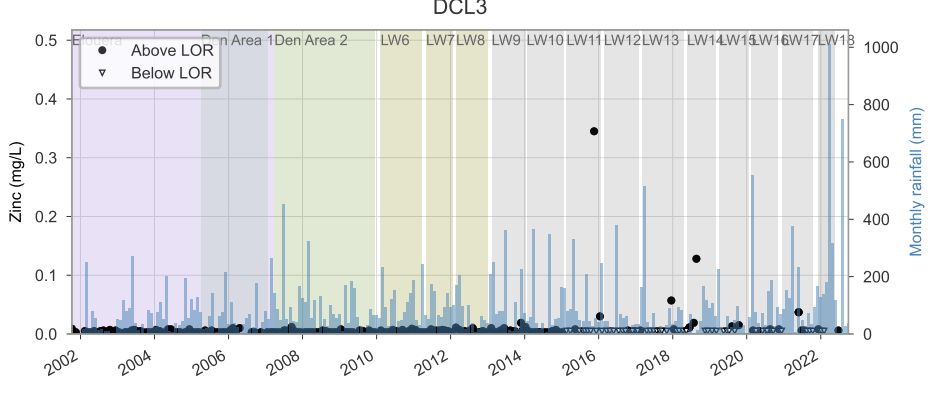
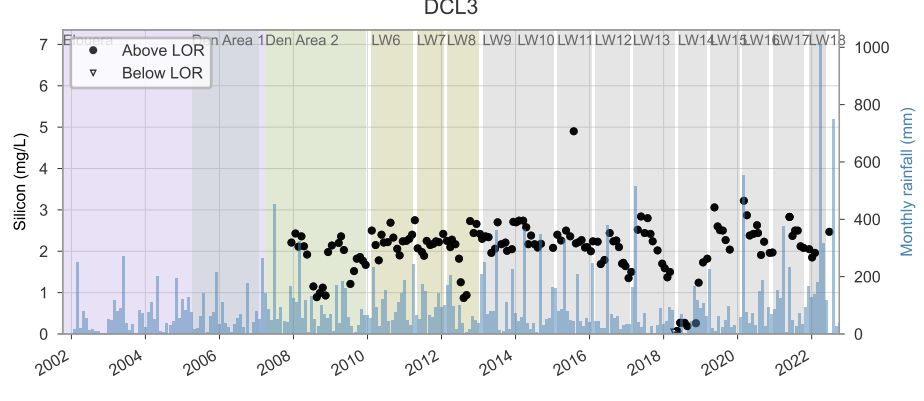
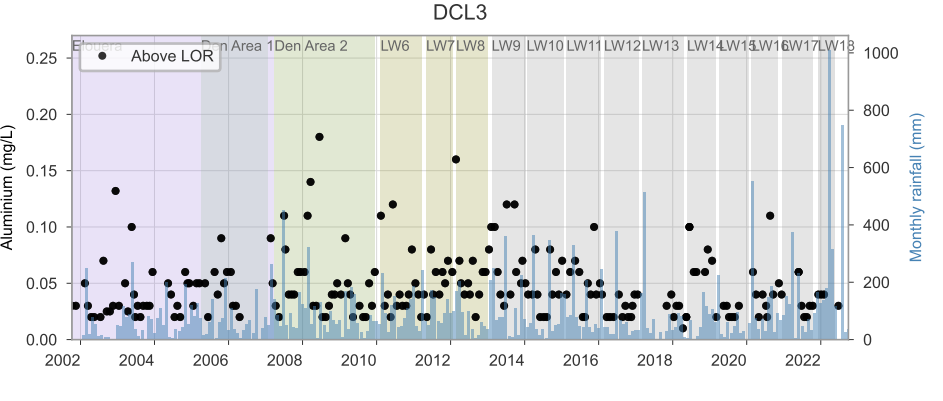
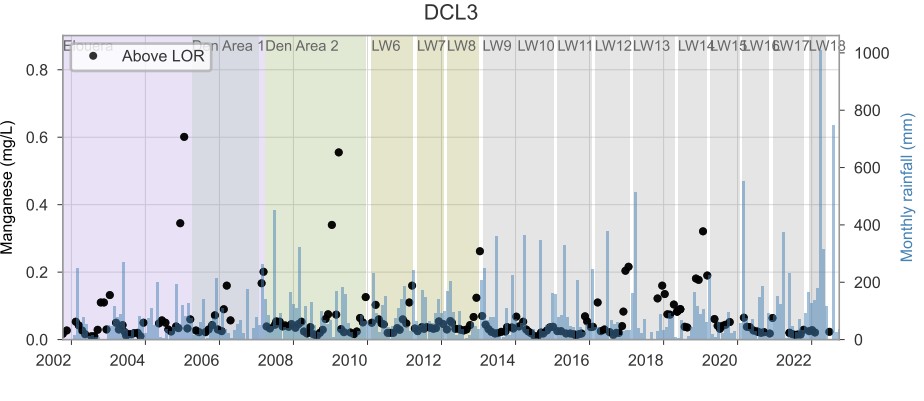
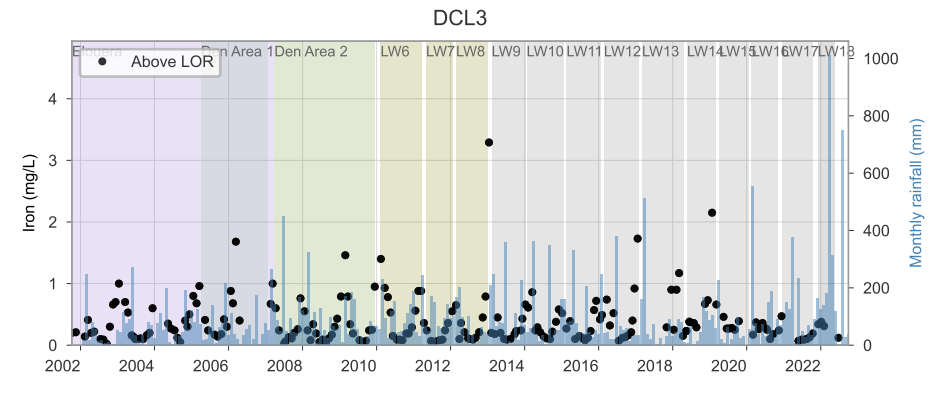
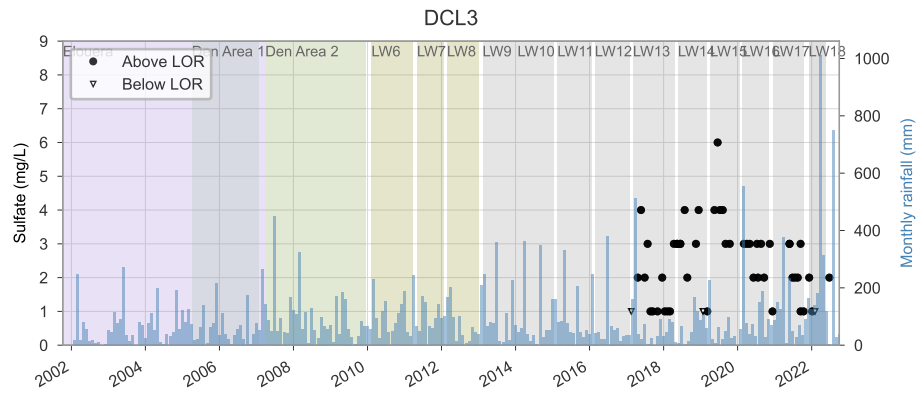


DCC_FR6

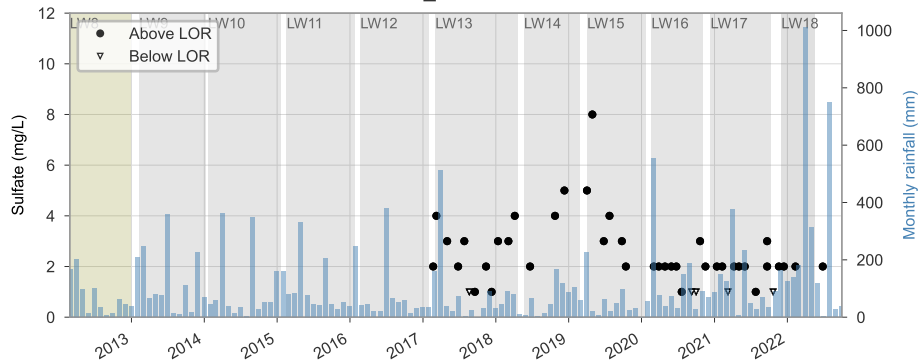


DCC_FR6

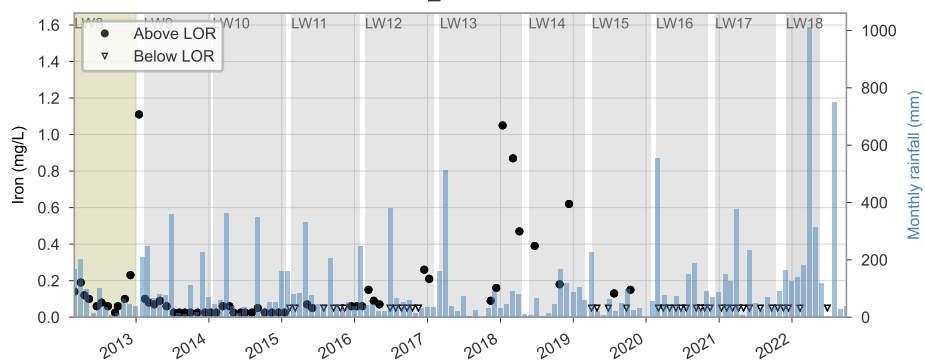




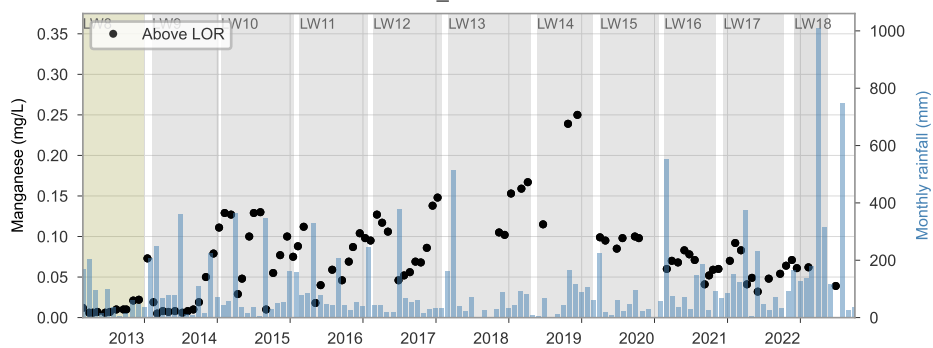
DC_POOL22



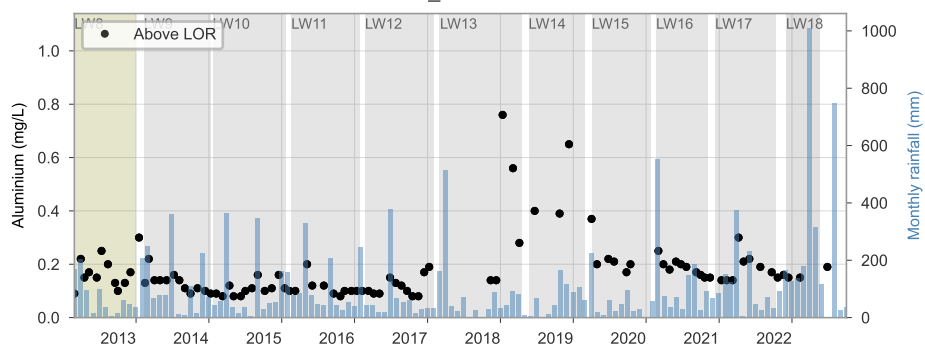
DC_POOL22



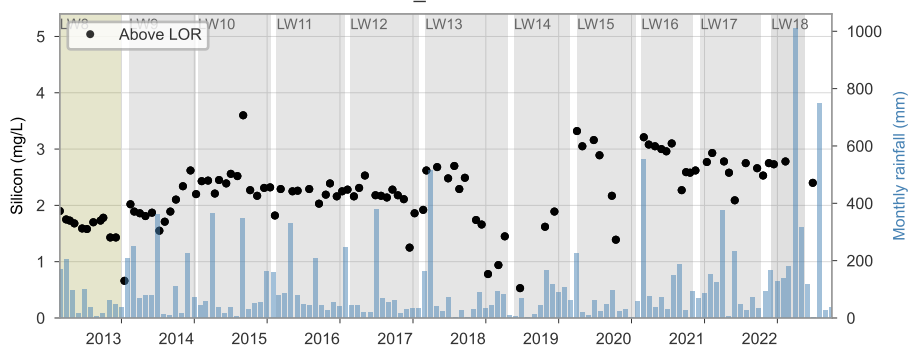
DC_POOL22



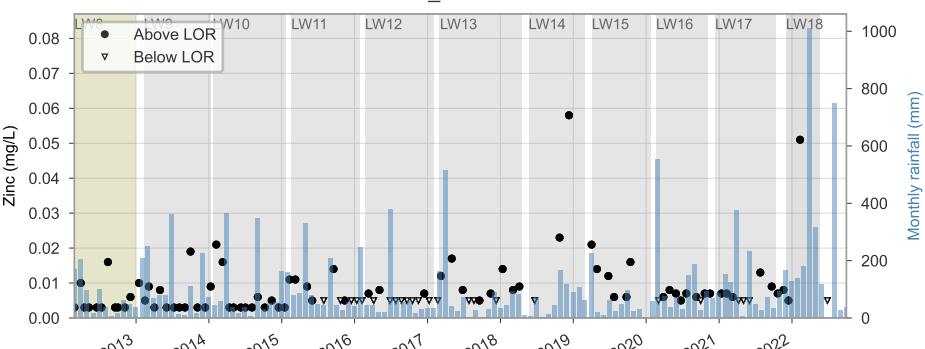
DC_POOL22

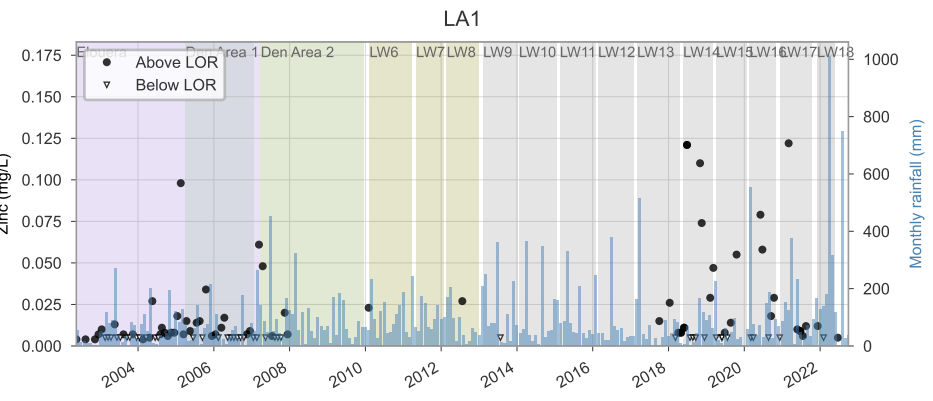
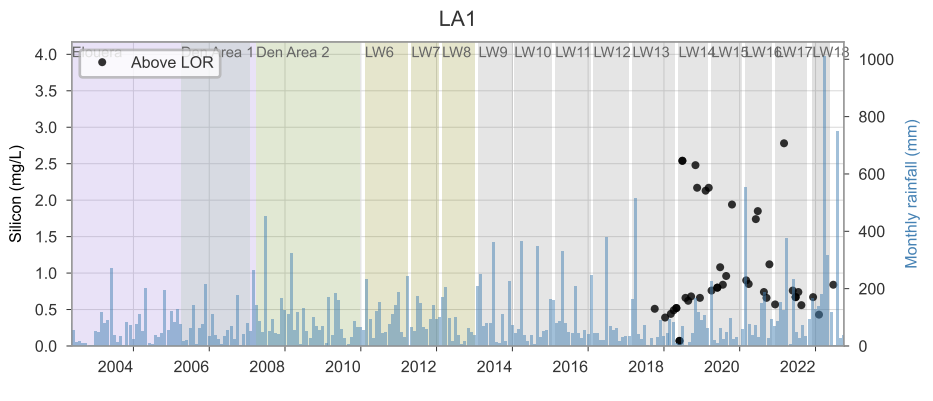
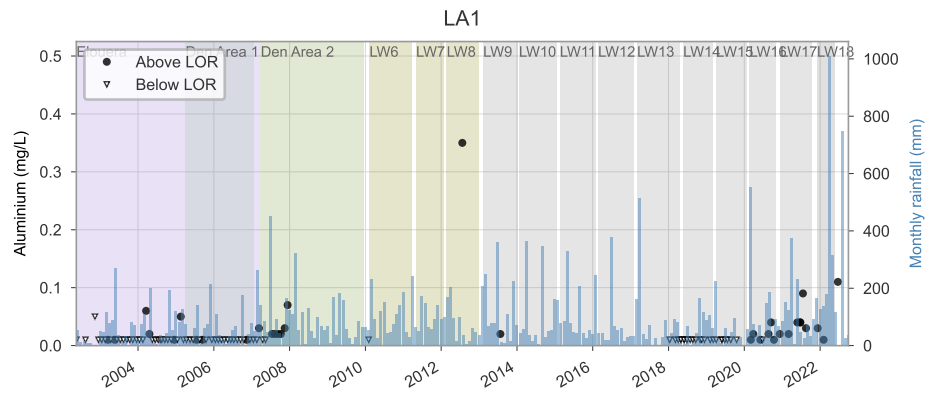
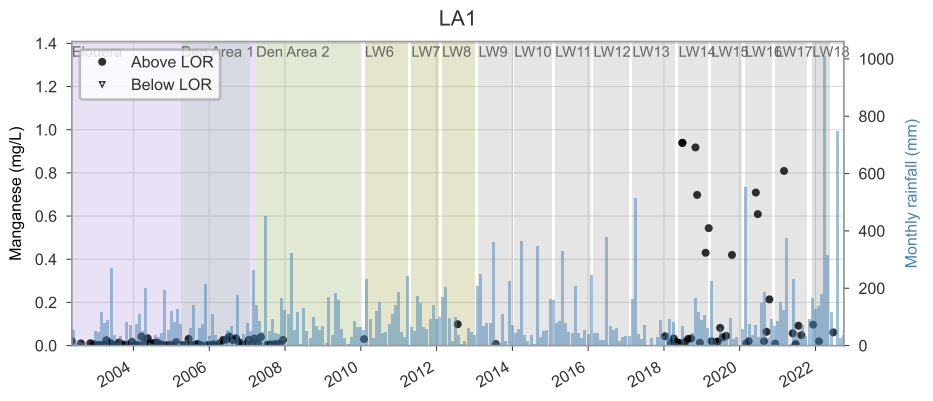
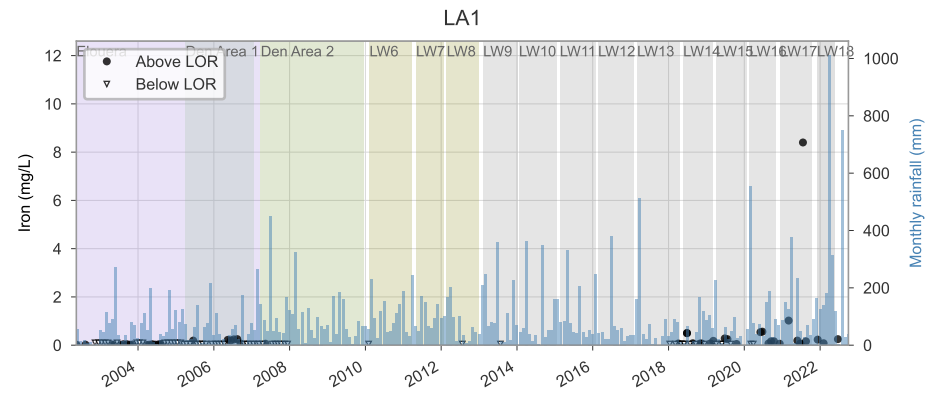
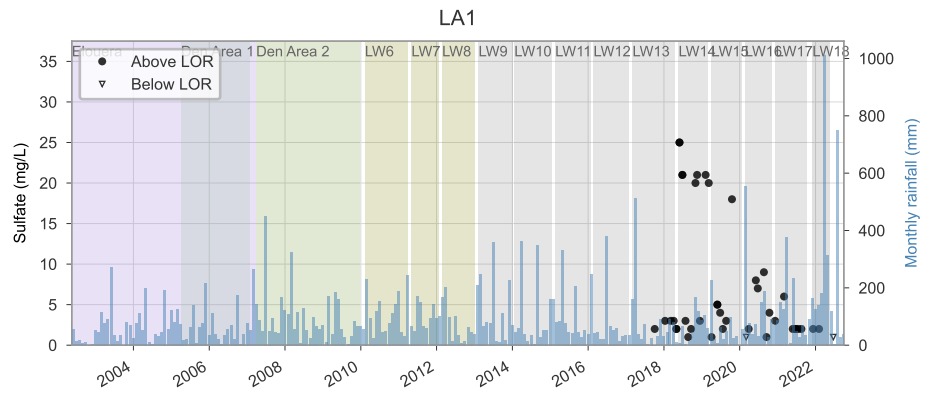


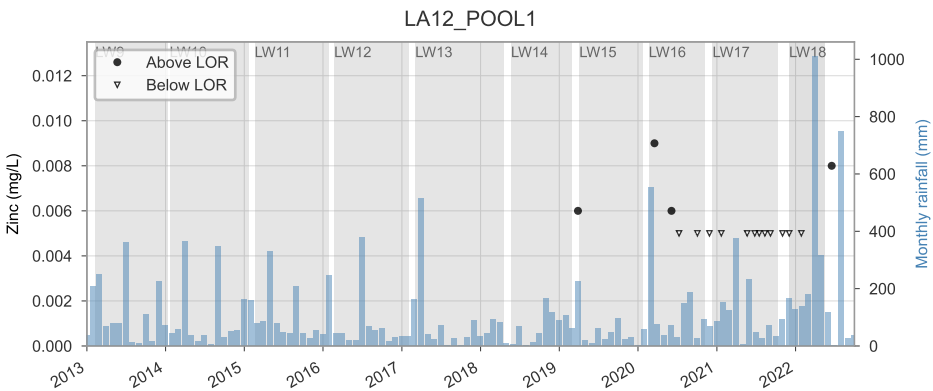
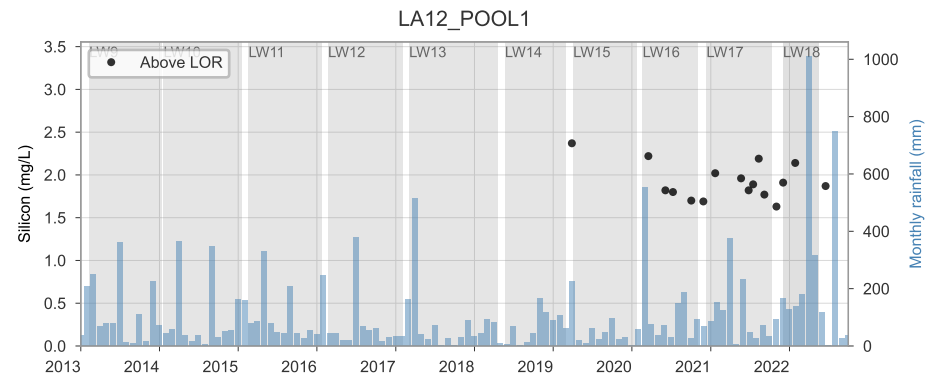
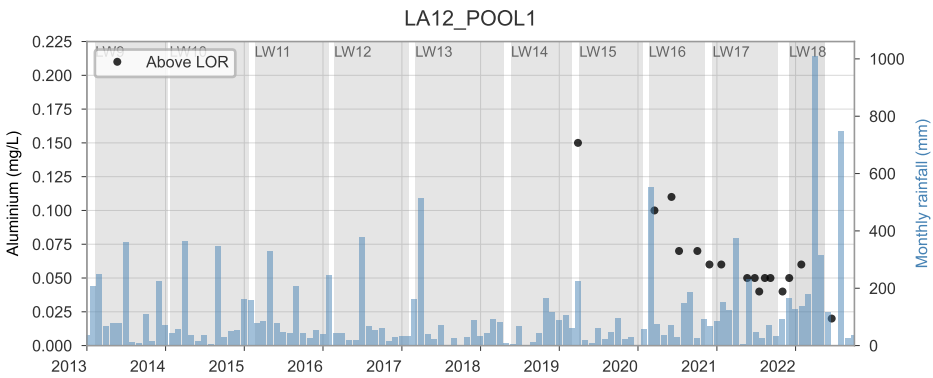
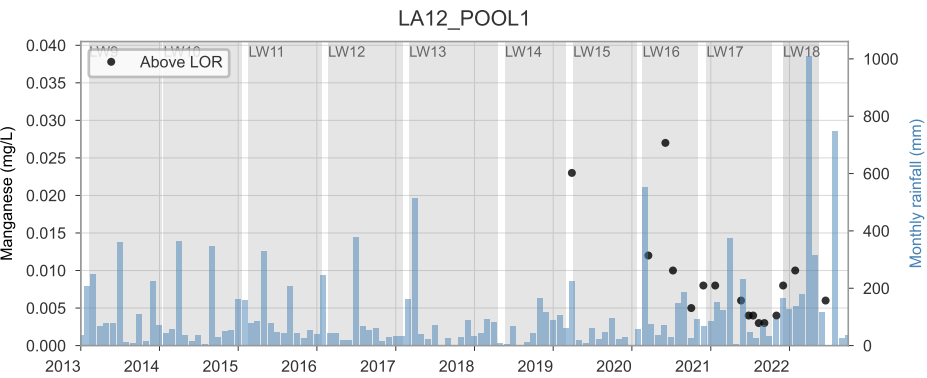
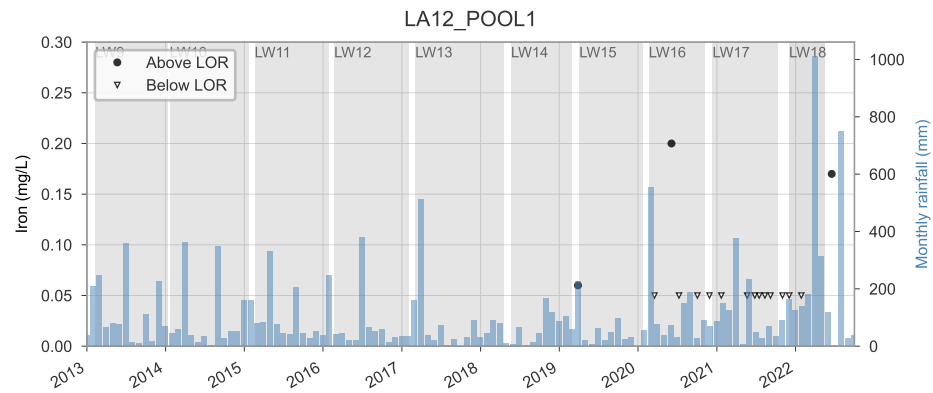
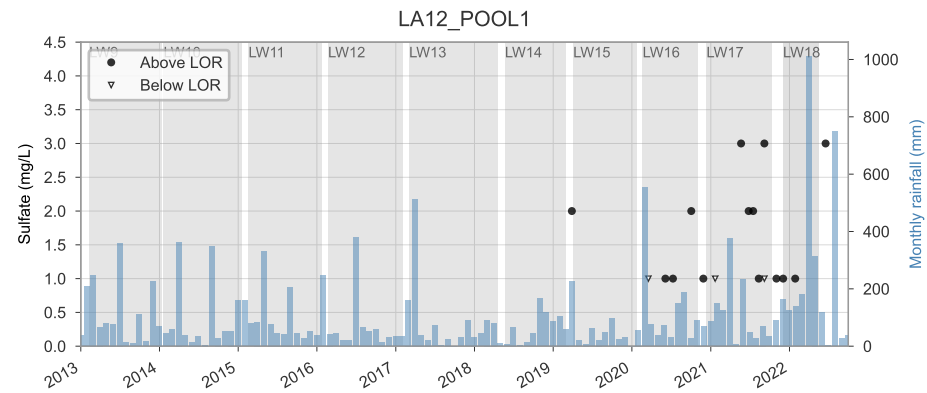
DC_POOL22



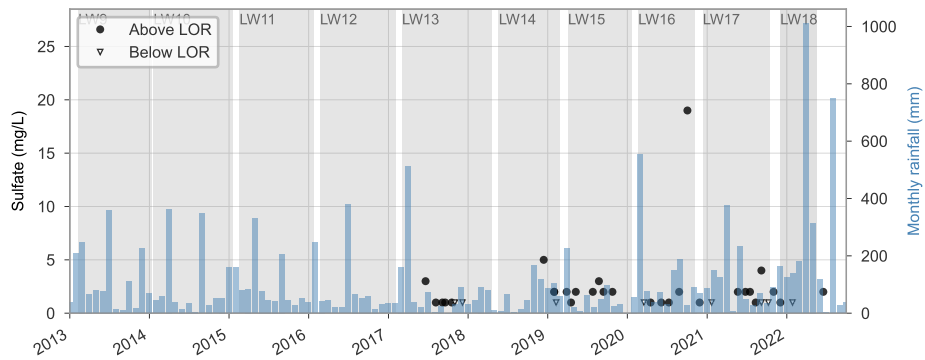
DC_POOL22



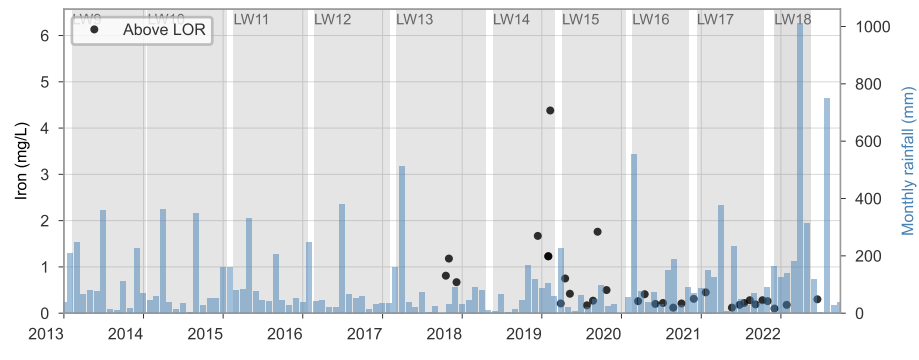




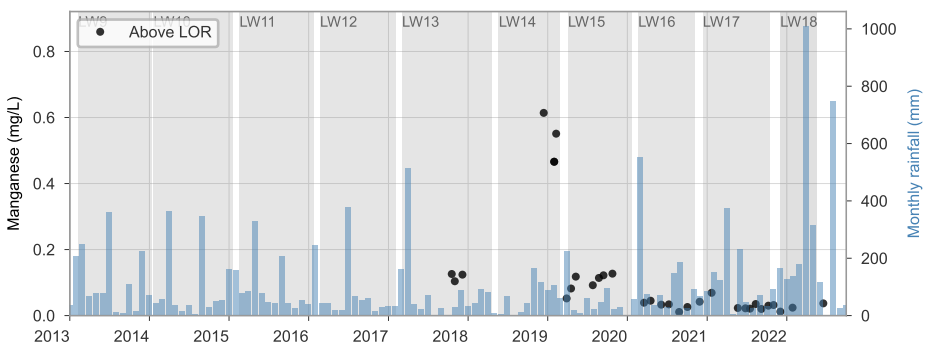
LA13A_S1



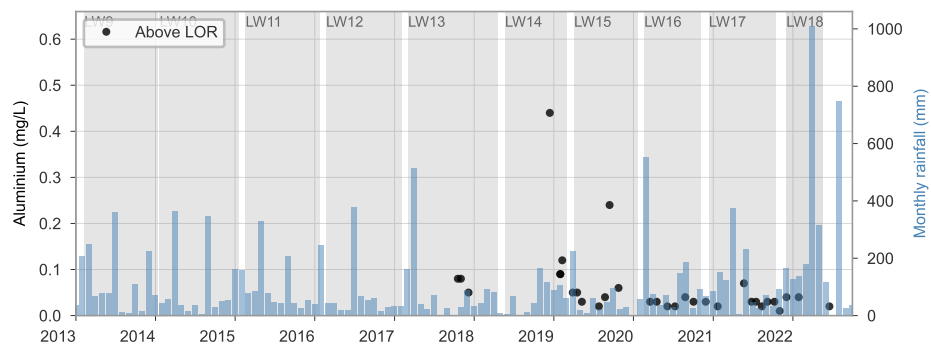
LA13A_S1



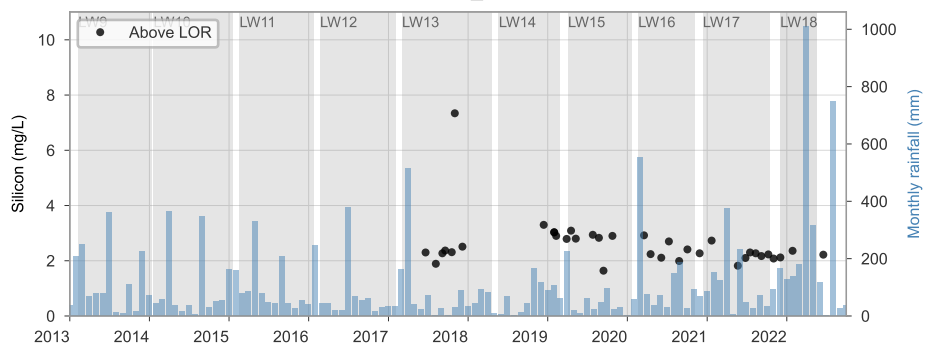
LA13A_S1



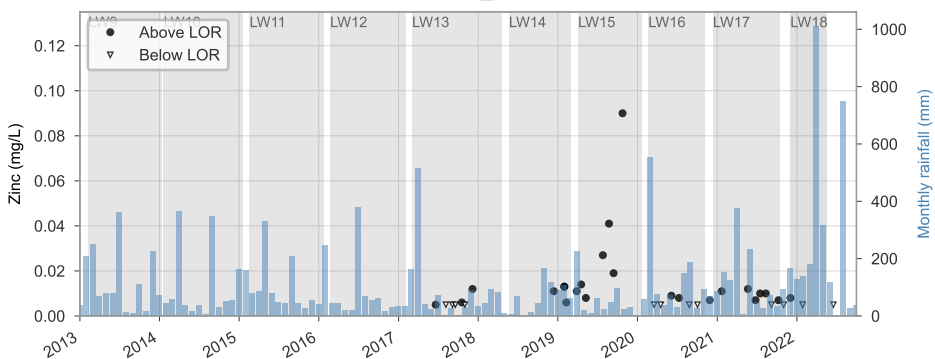
LA13A_S1



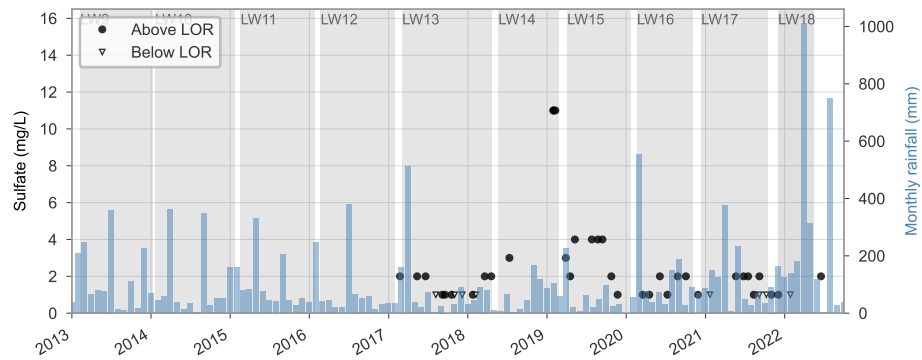
LA13A_S1



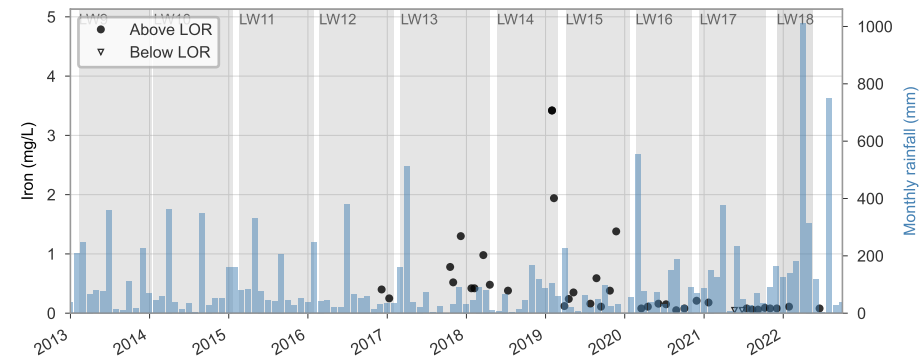
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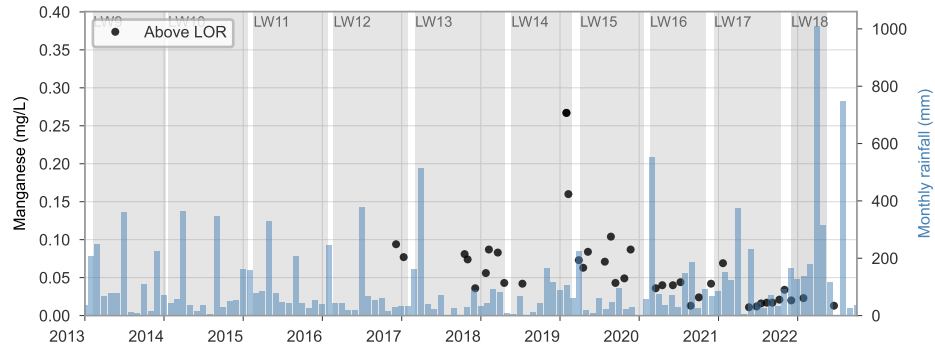
LA13_S1



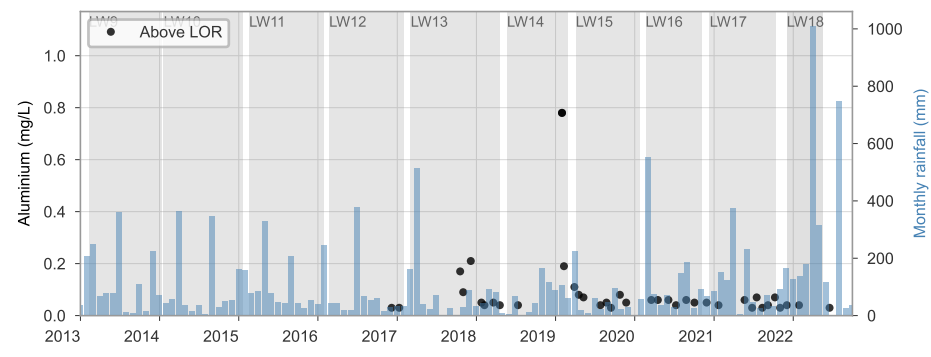
LA13_S1



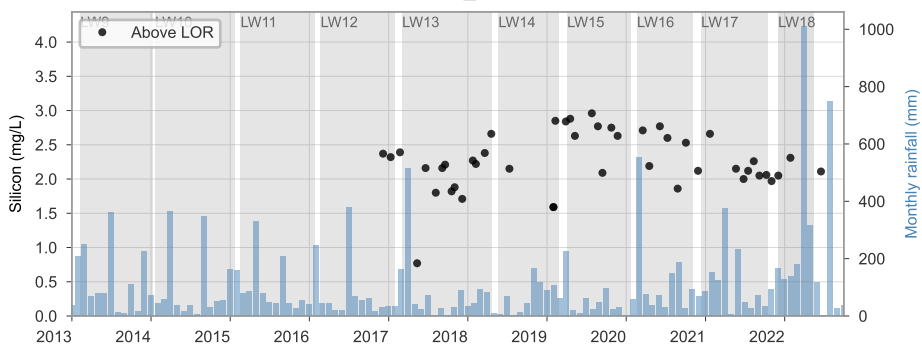
LA13_S1



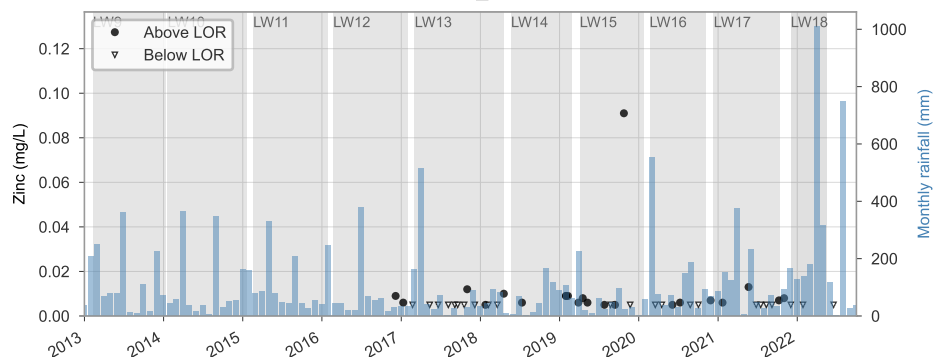
LA13_S1

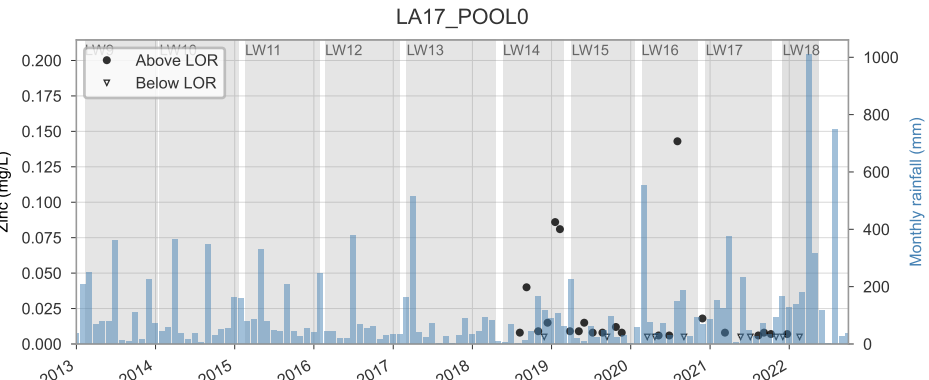
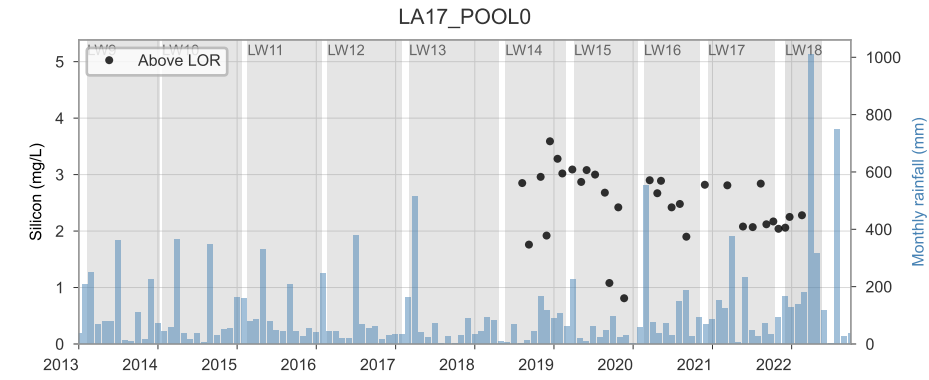
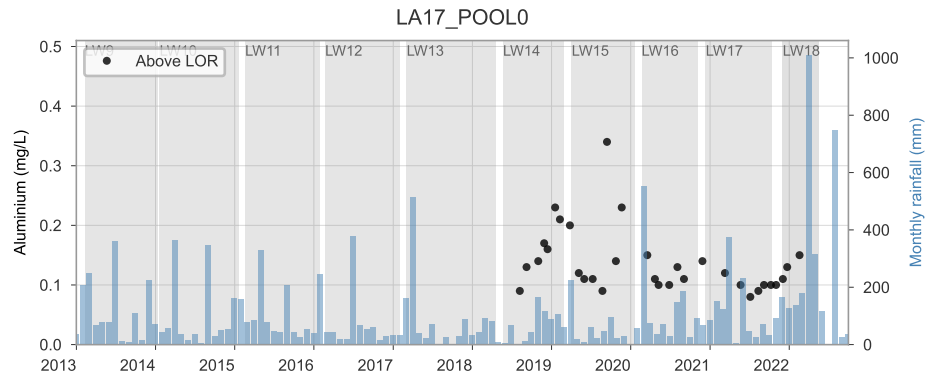
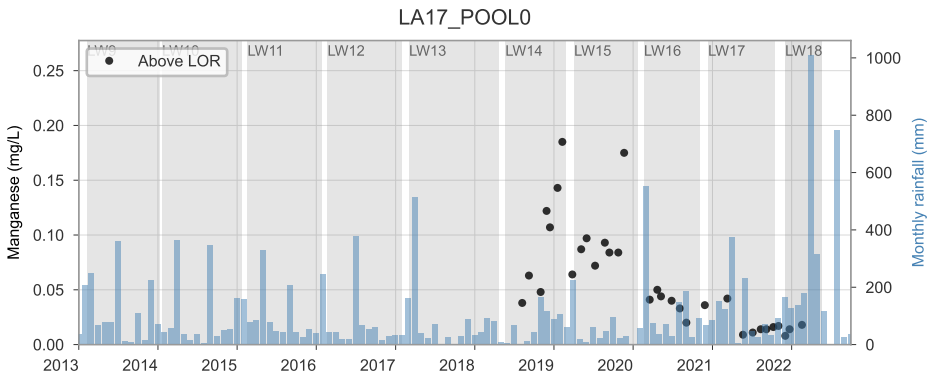
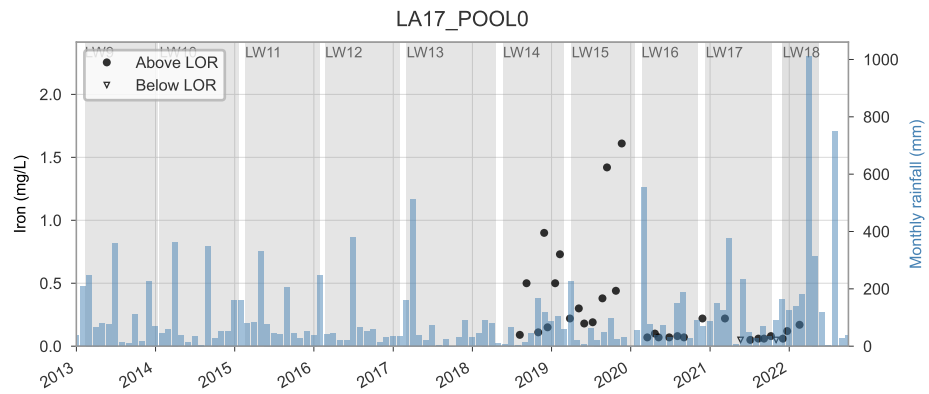
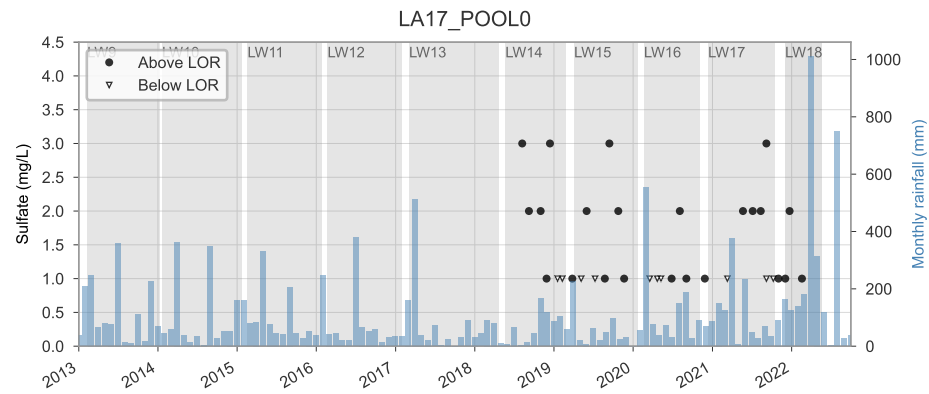


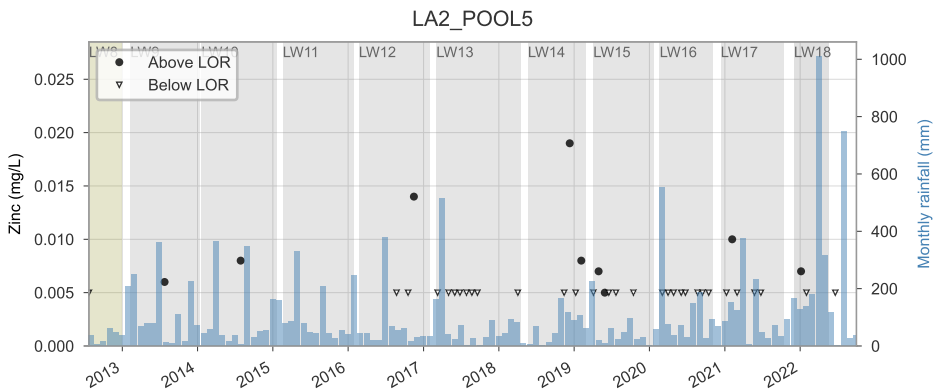
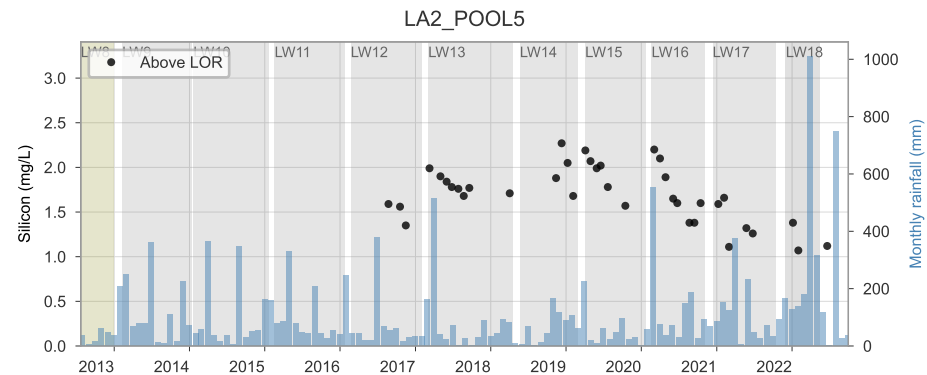
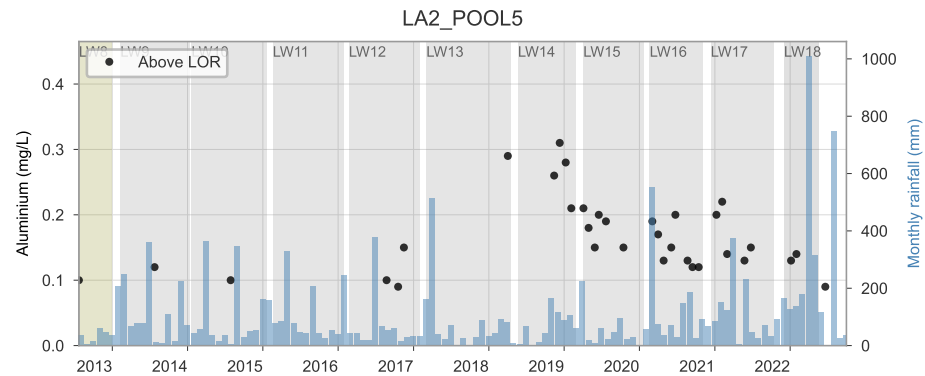
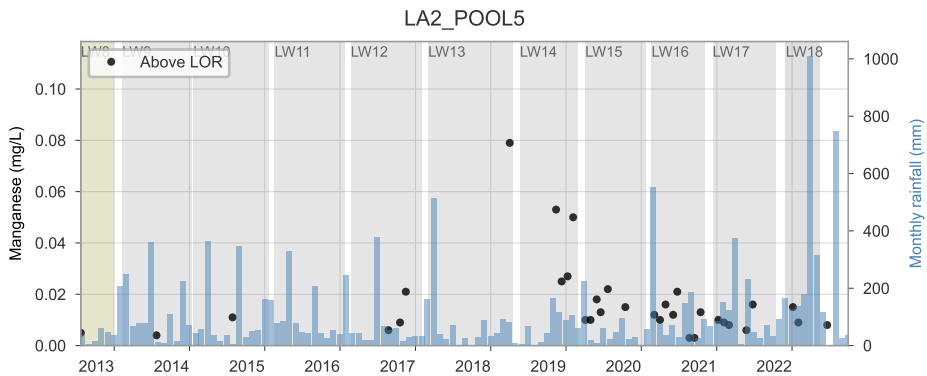
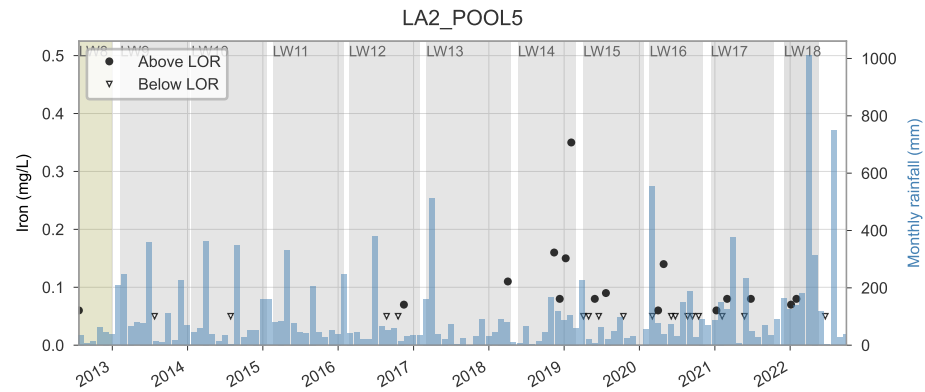
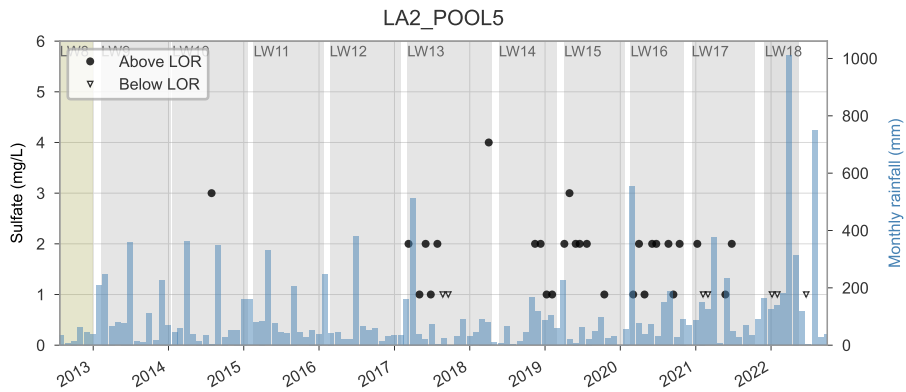
LA13_S1

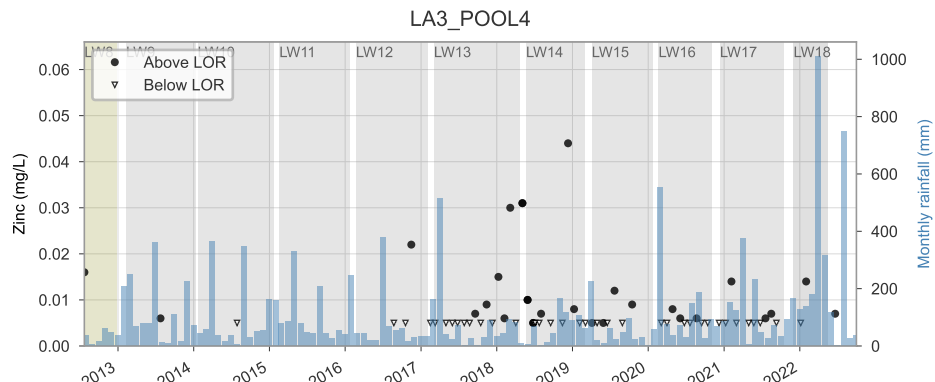
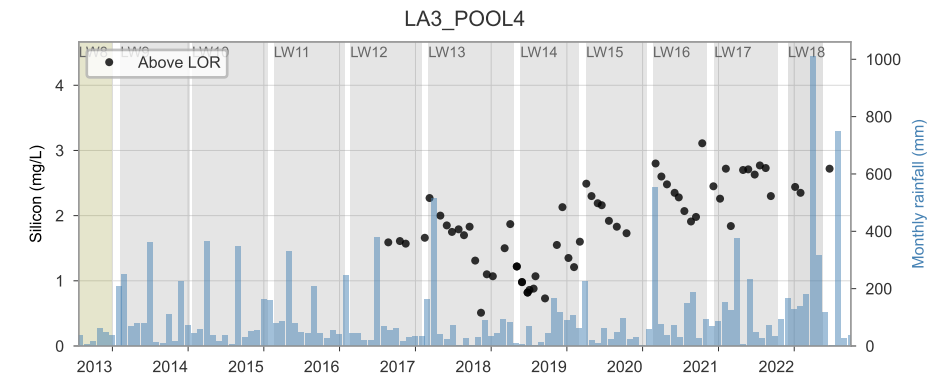
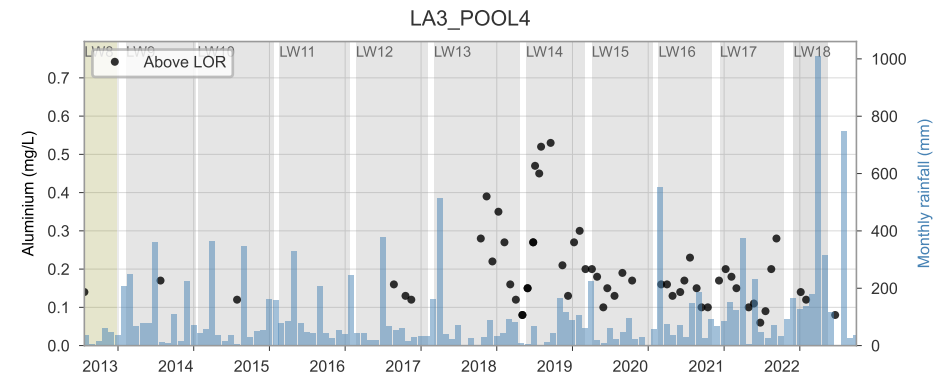
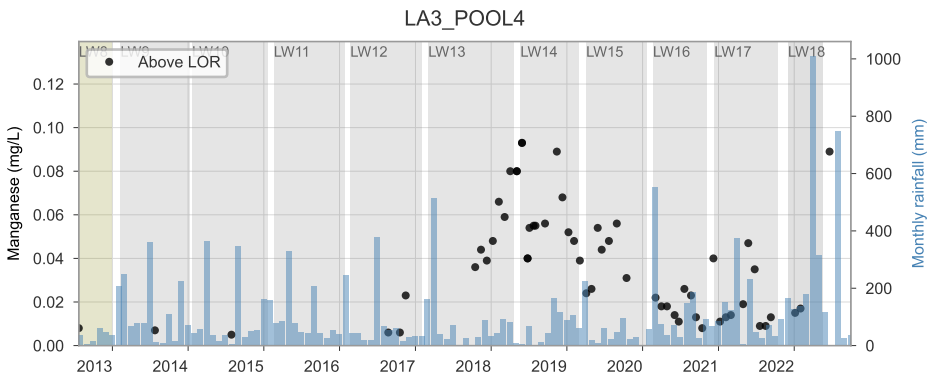
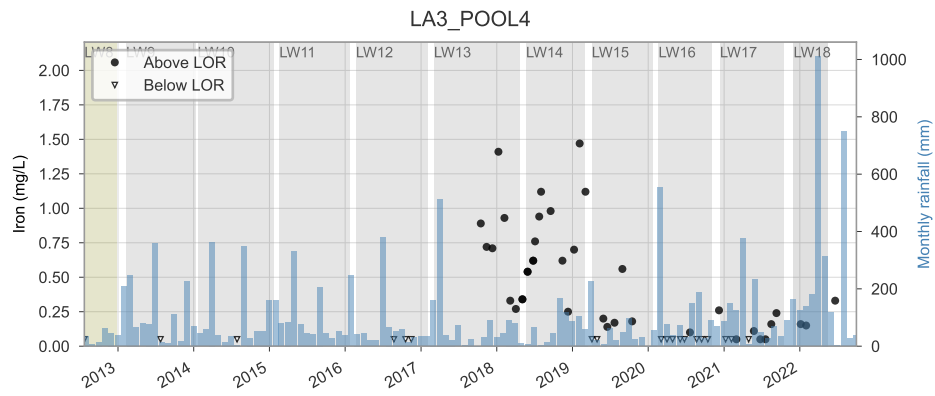
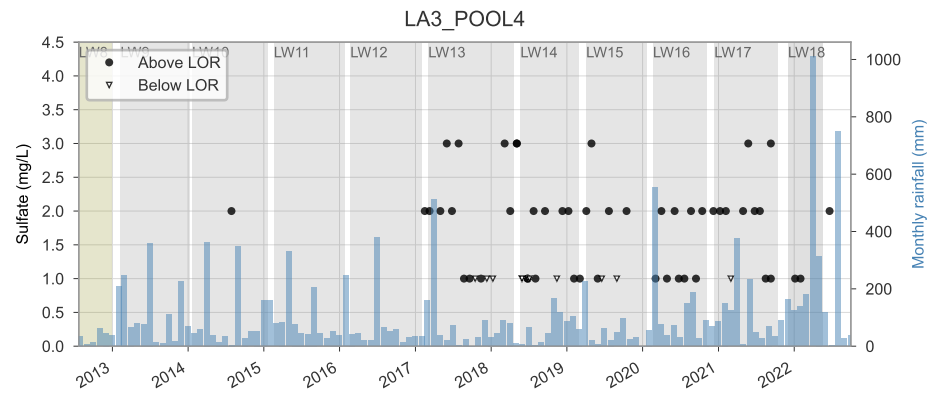


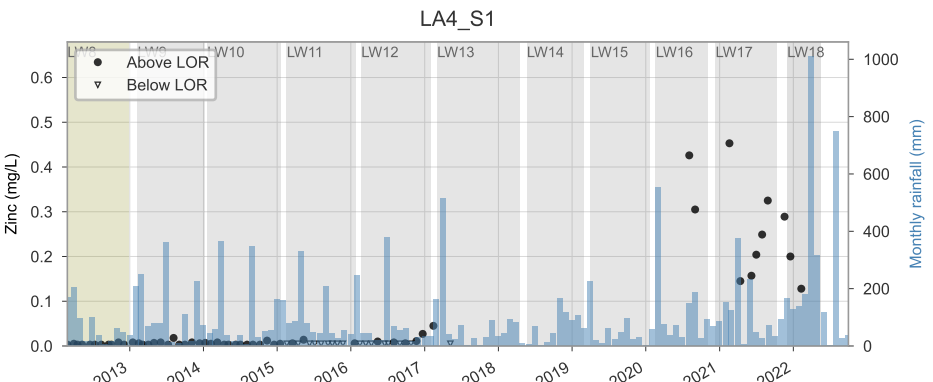
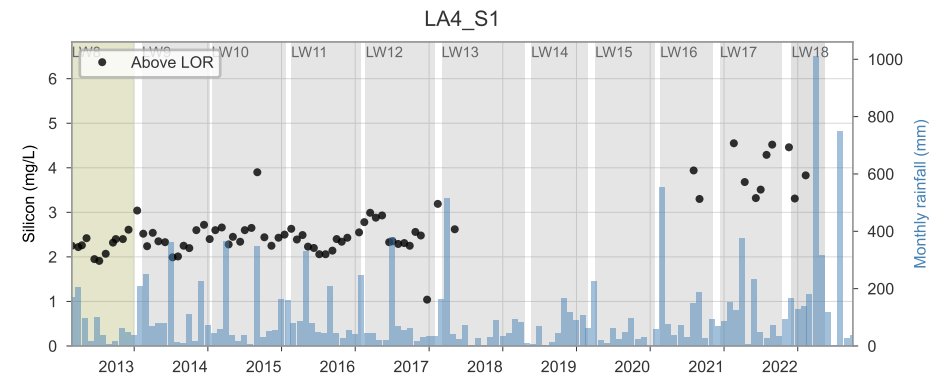
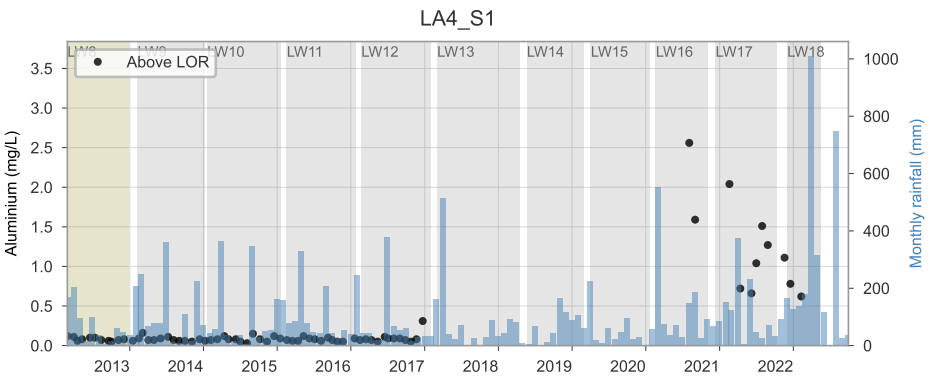
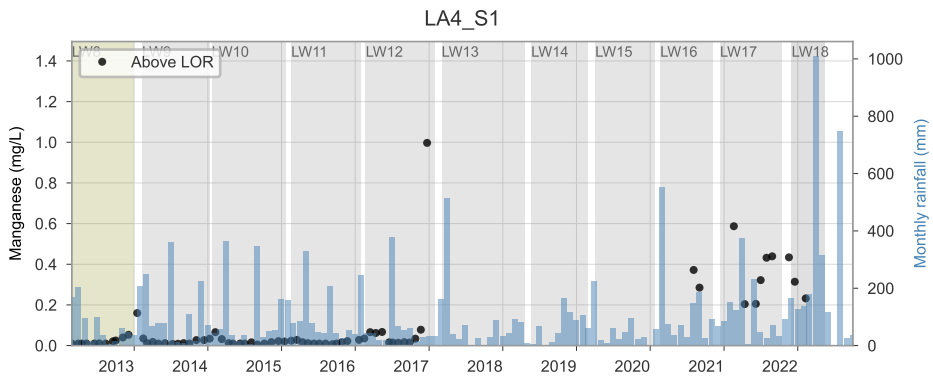
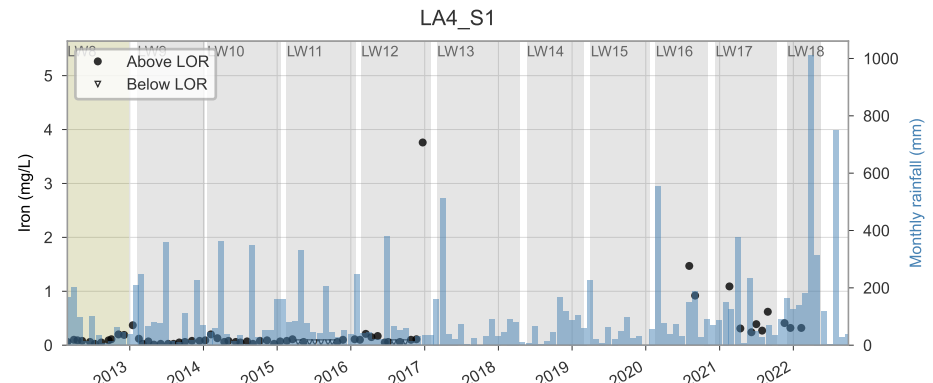
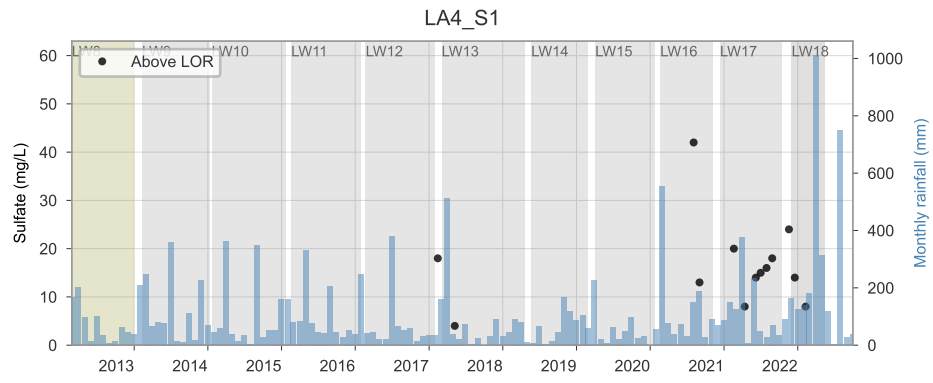
LA13_S1



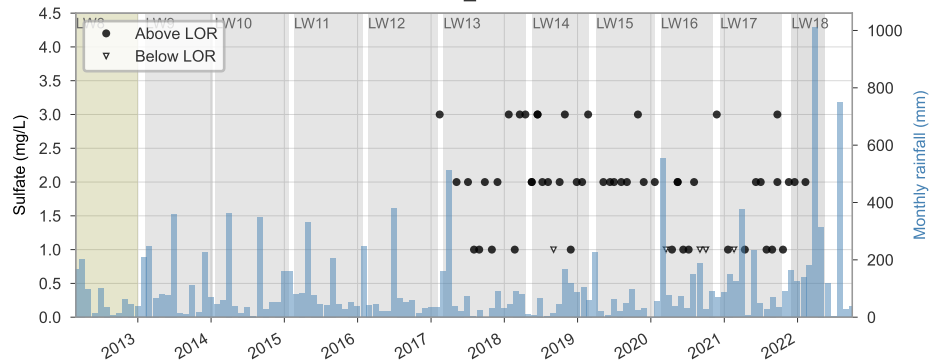




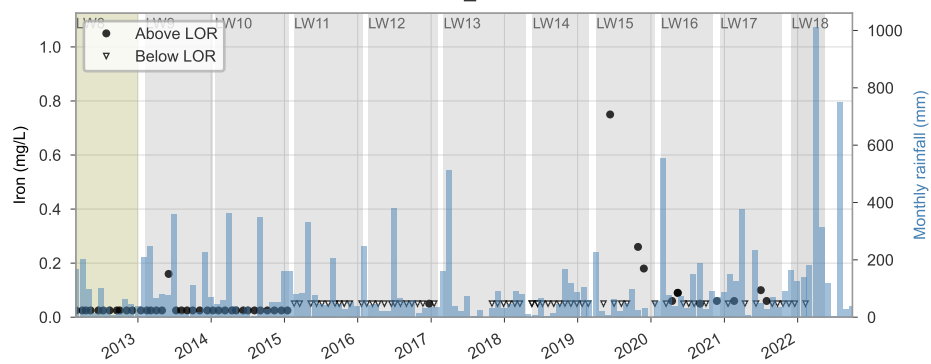




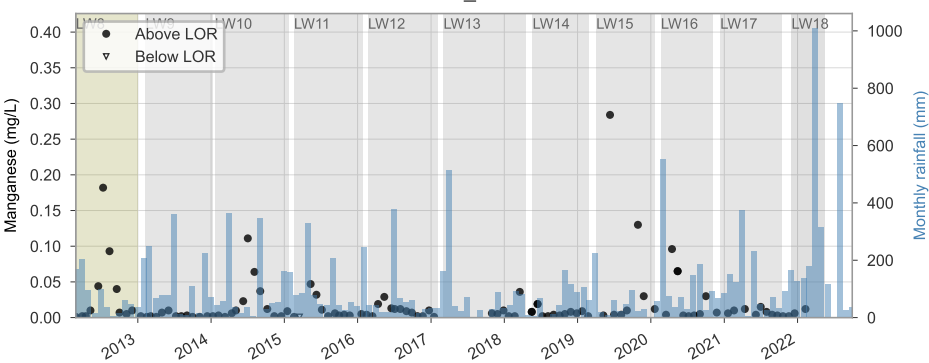
LA4_S2



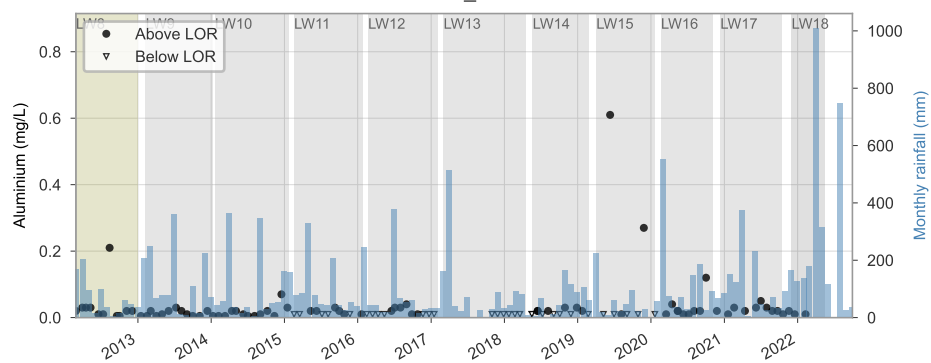
LA4_S2



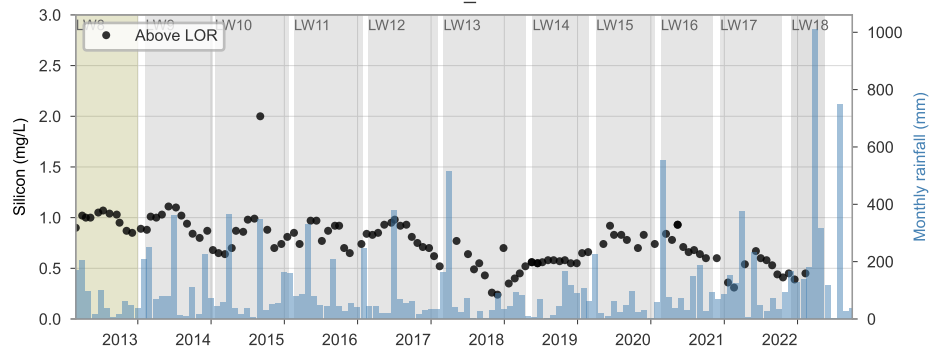
LA4_S2



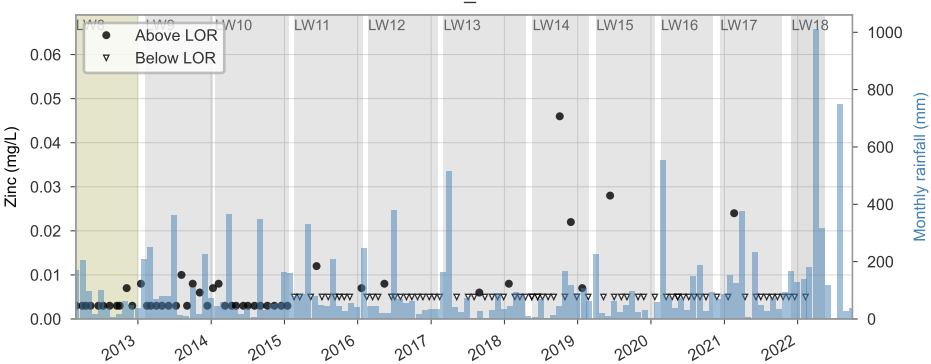
LA4_S2



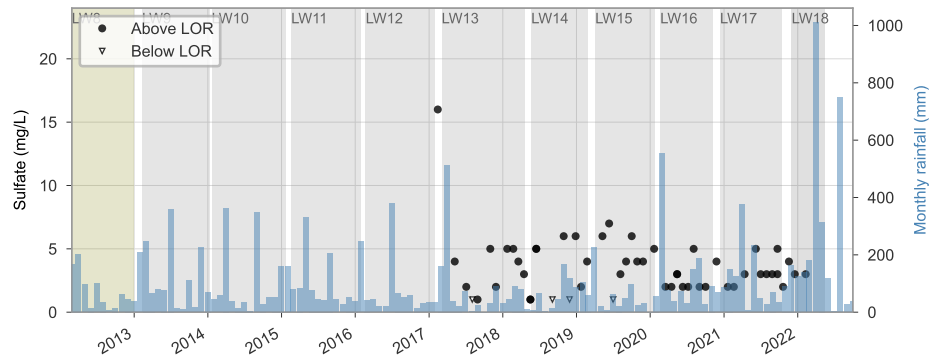
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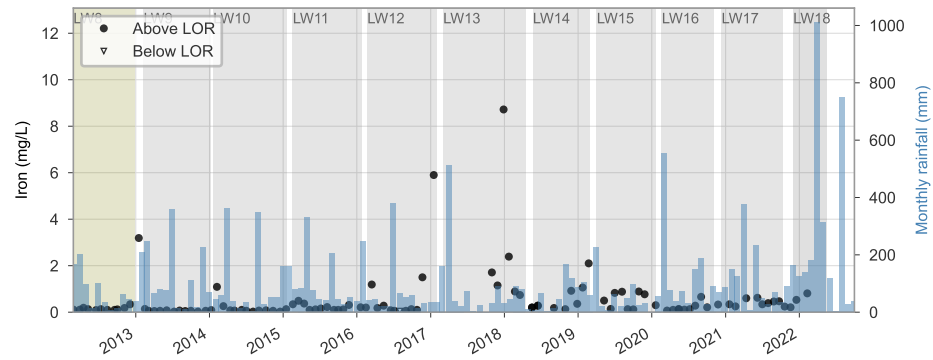
LA4_S2



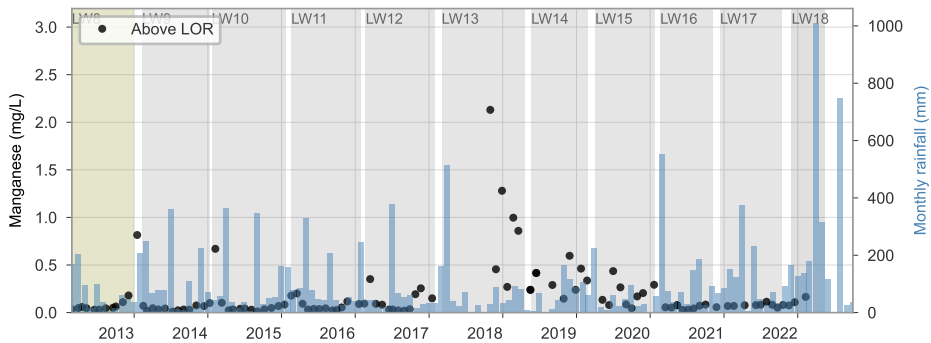
LA5_S1



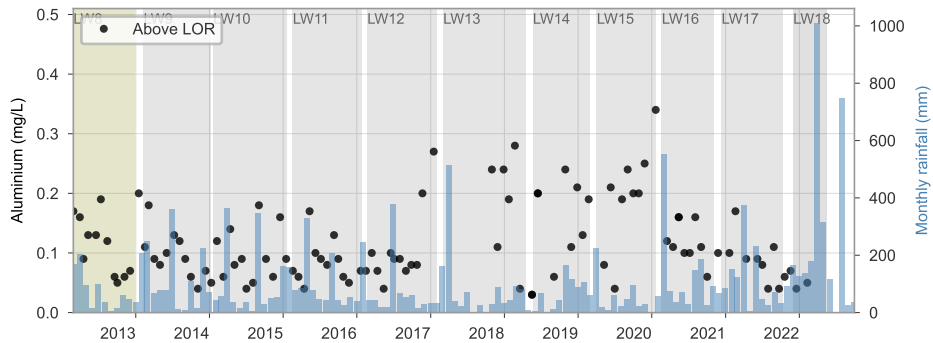
LA5_S1



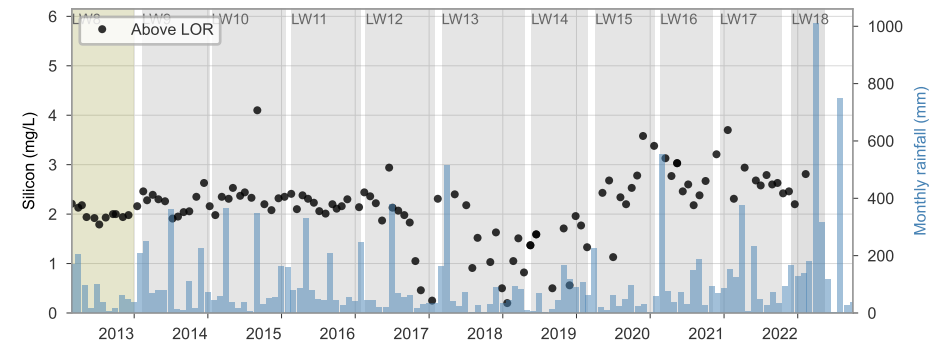
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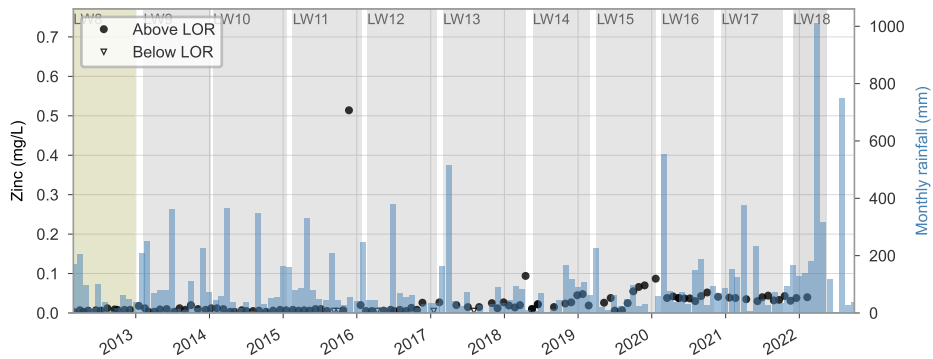
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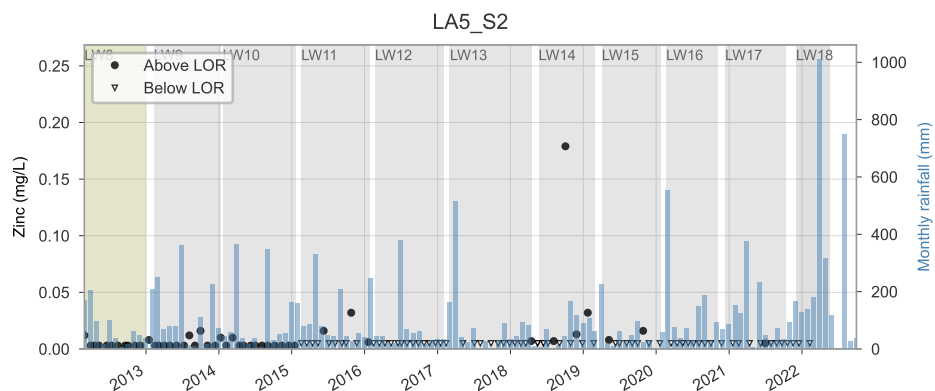
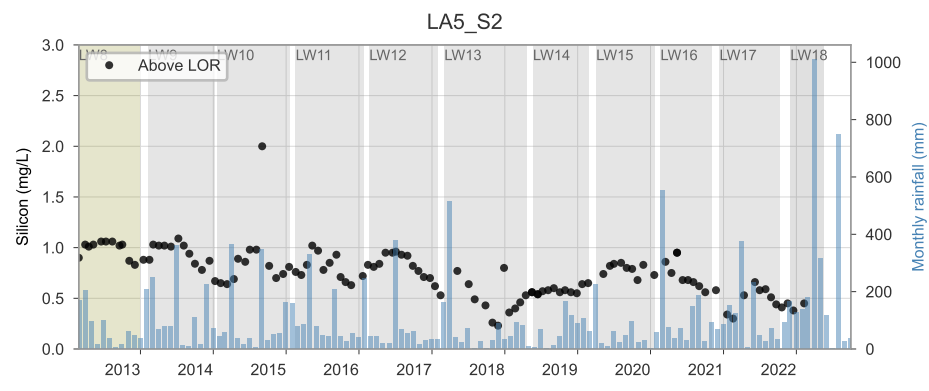
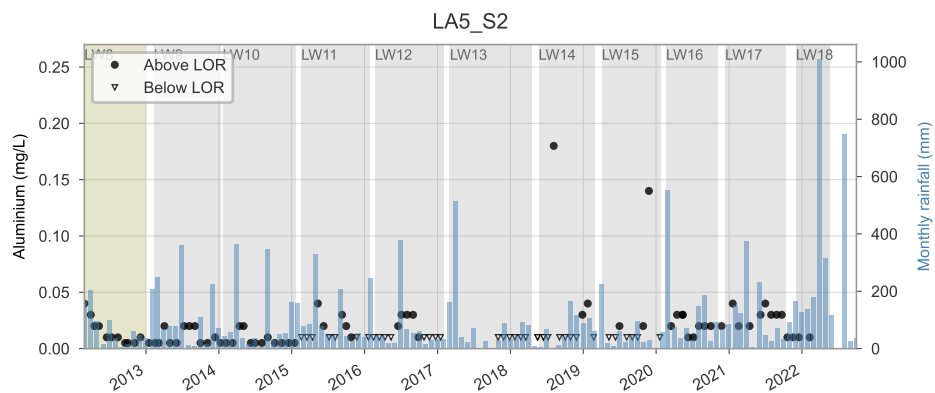
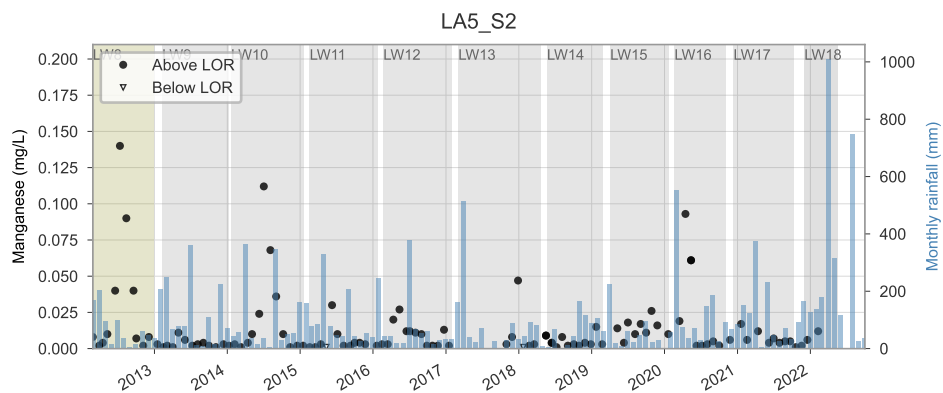
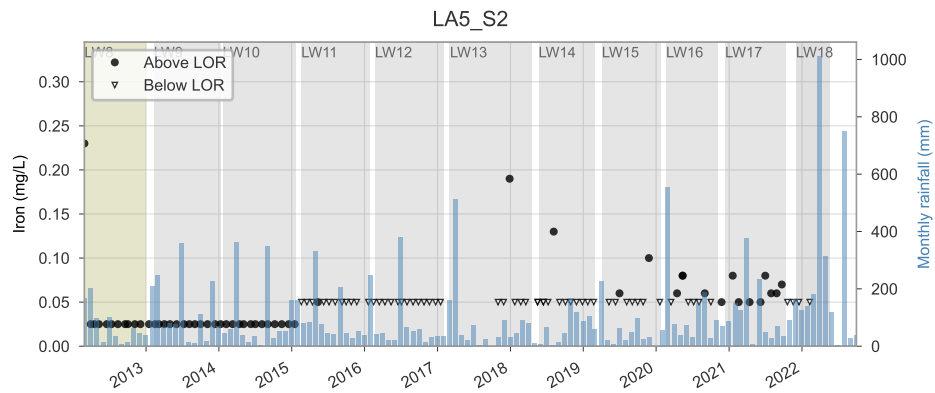
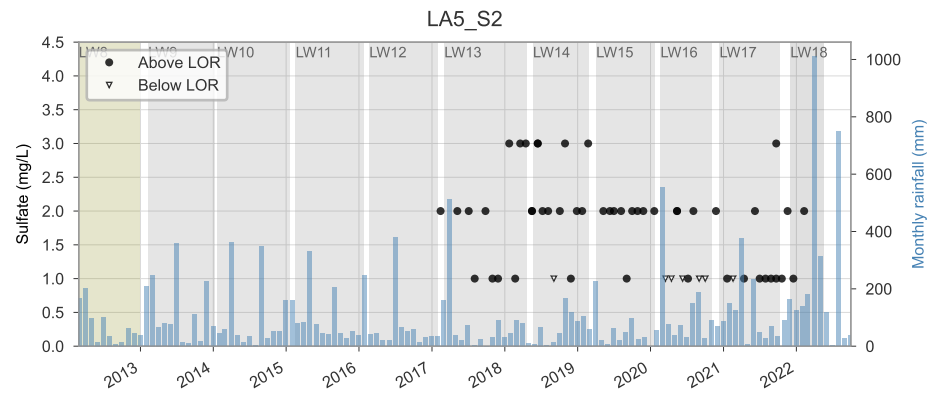


LA5_S1

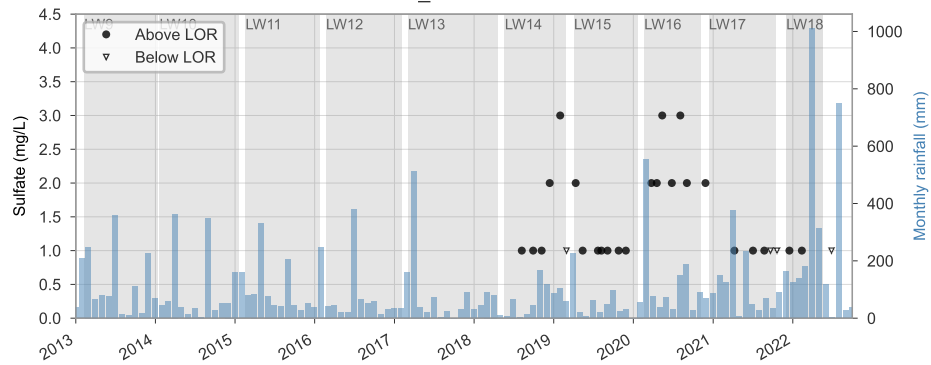


LA5_S1

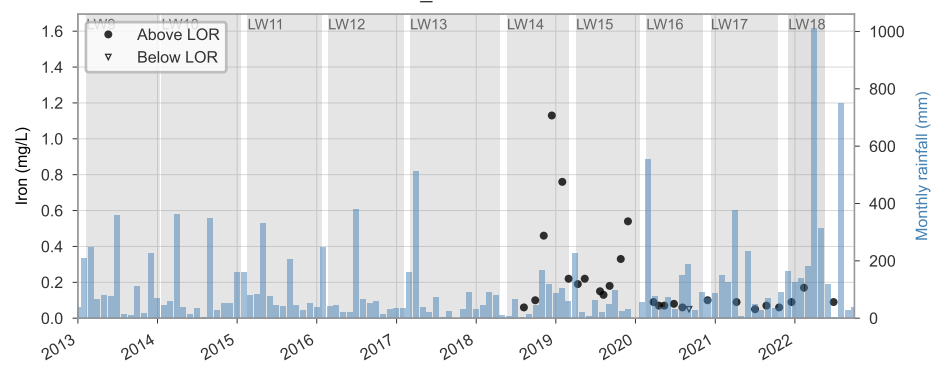




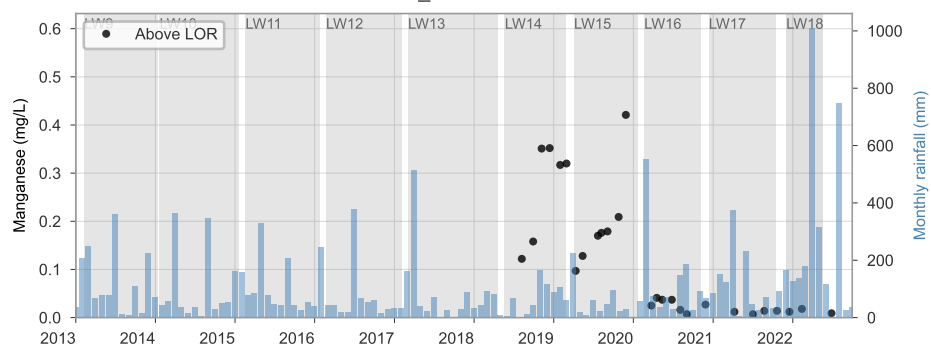
LA8_ROCKBAR1



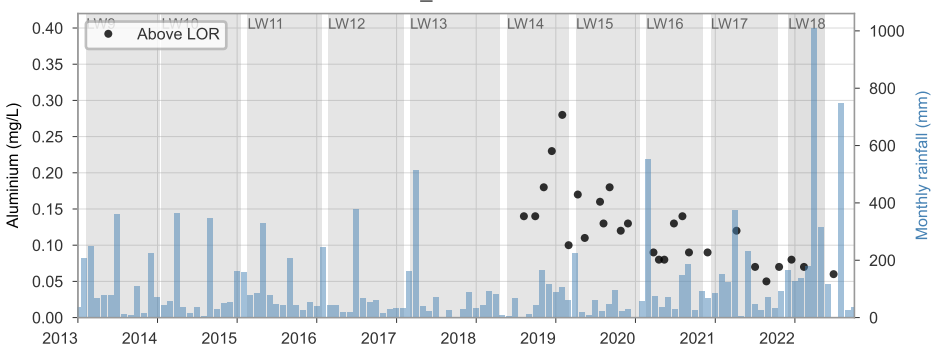
LA8_ROCKBAR1



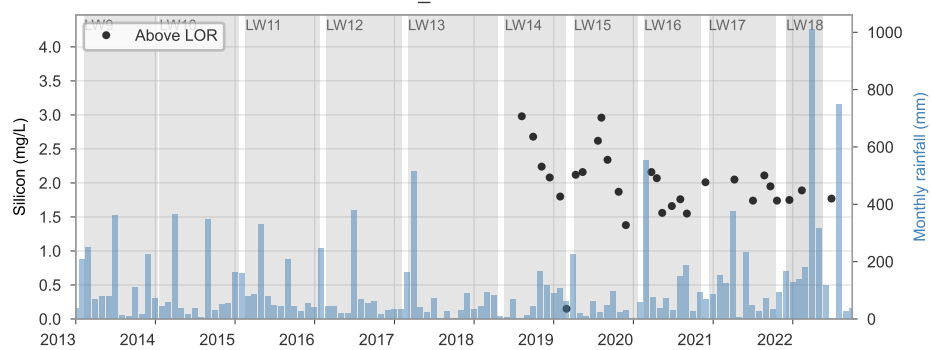
LA8_ROCKBAR1



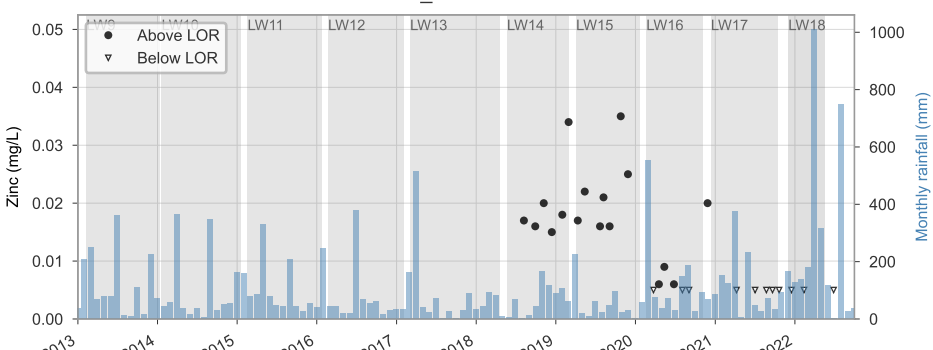
LA8_ROCKBAR1

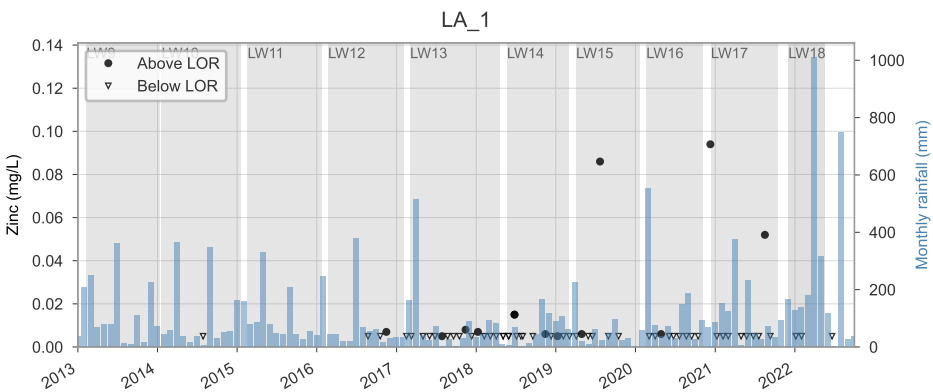
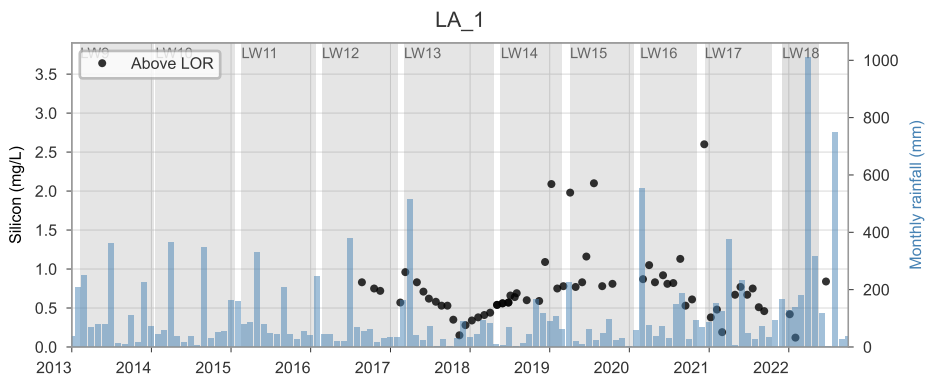
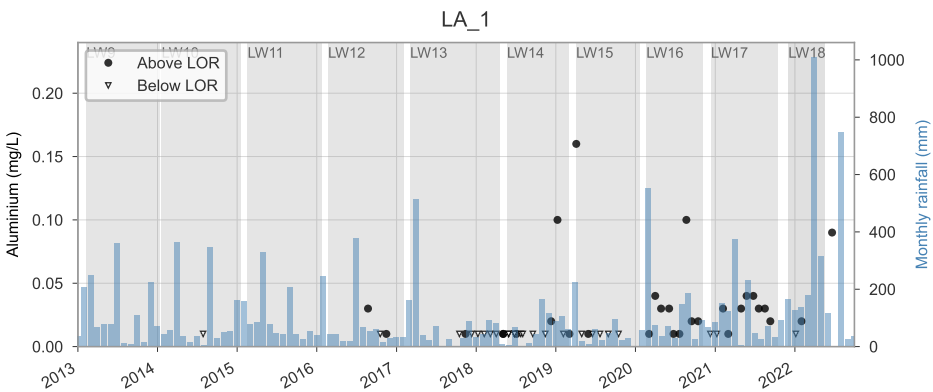
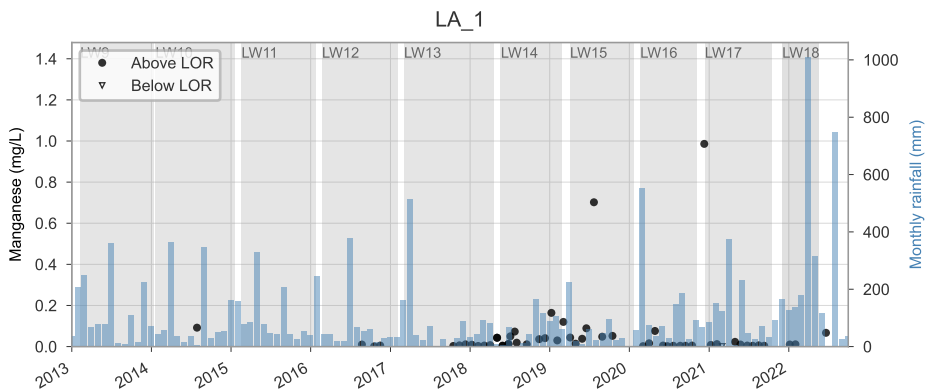
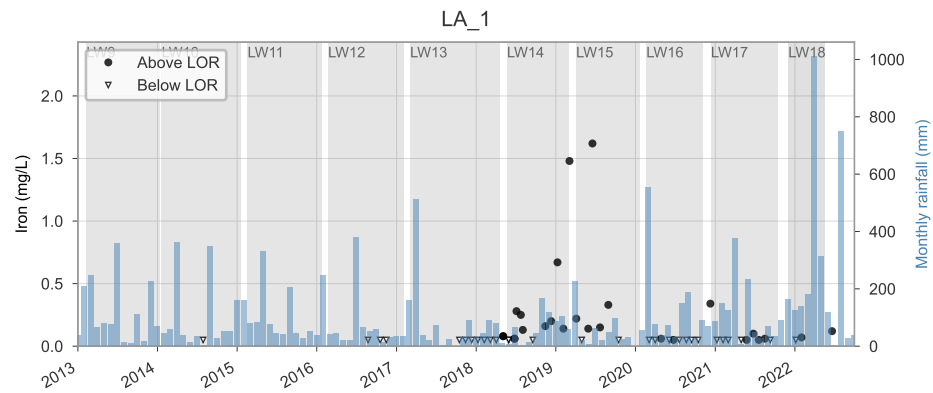
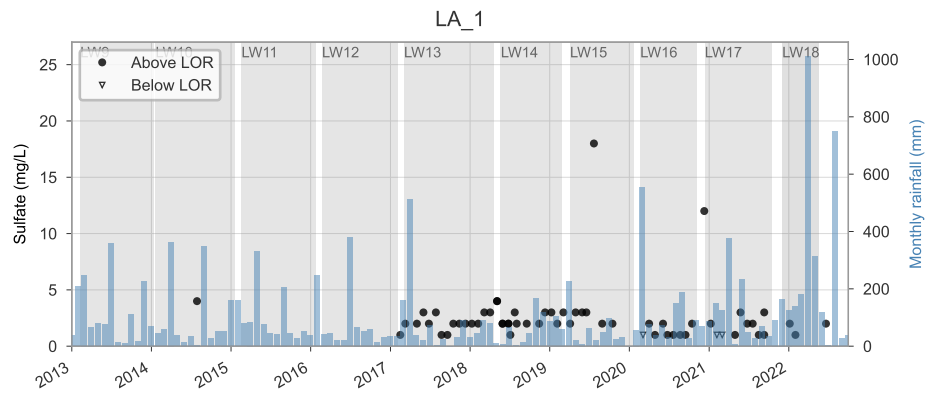


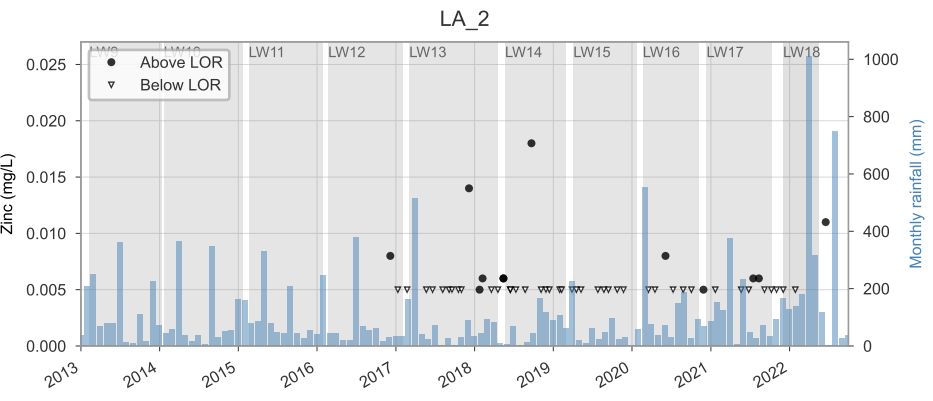
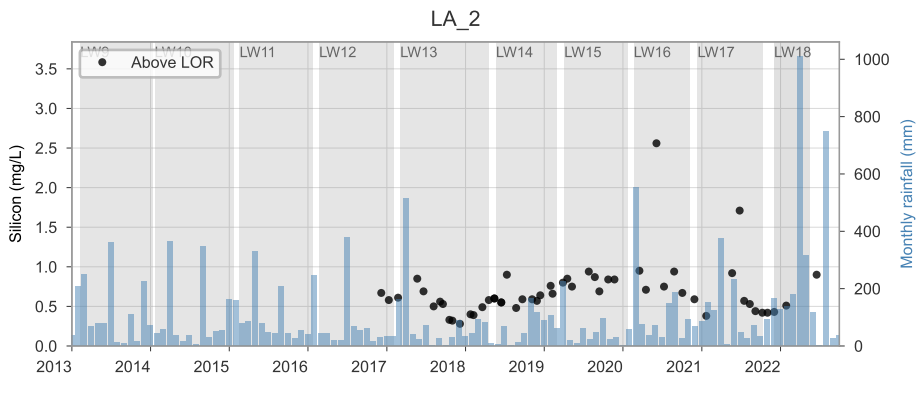
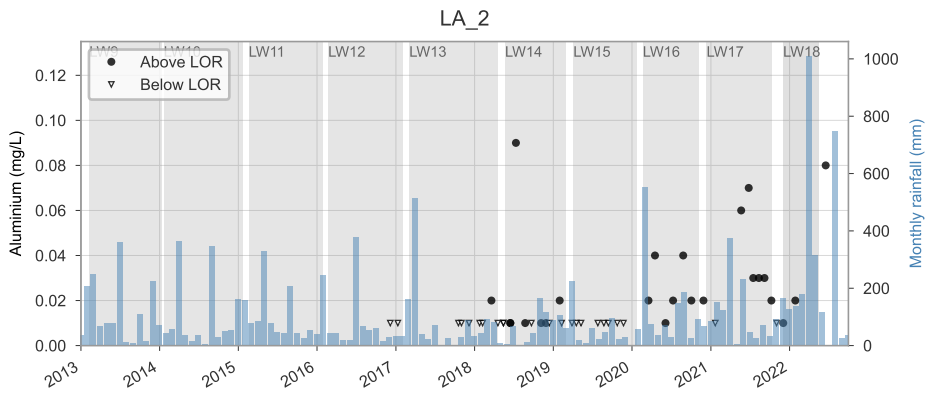
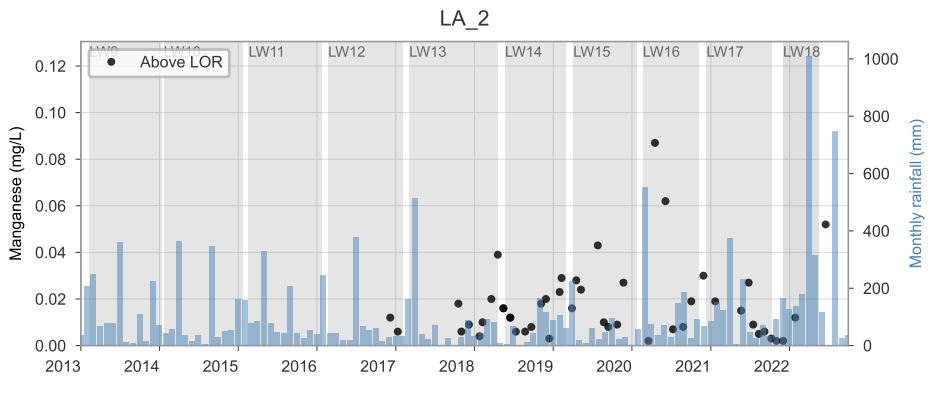
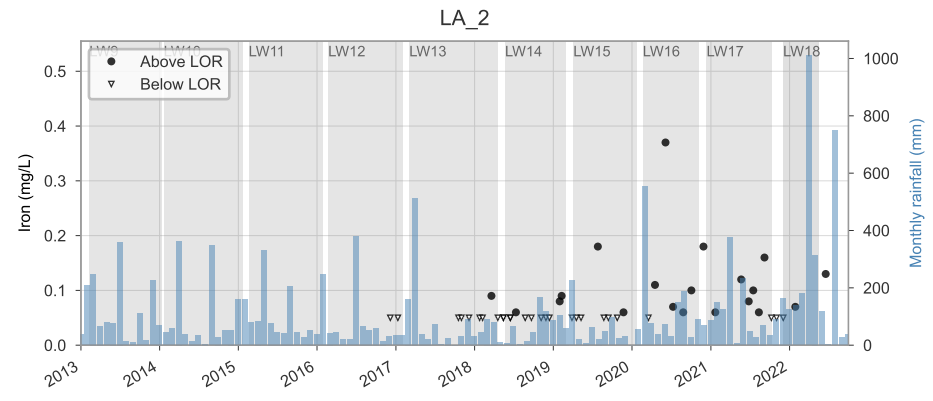
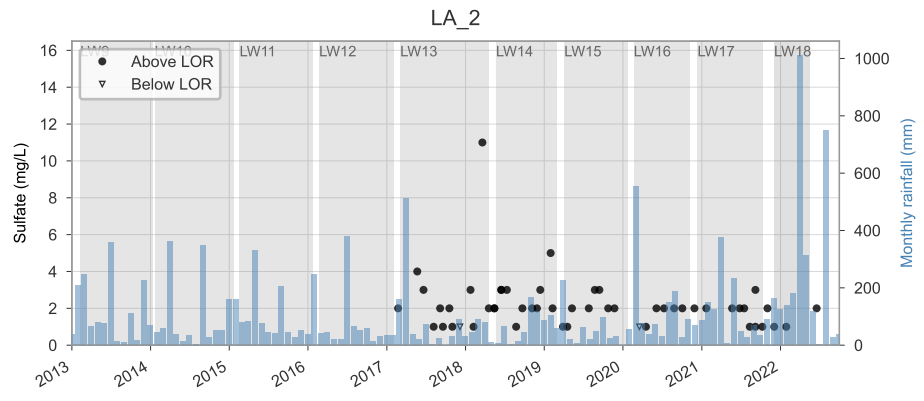
LA8_ROCKBAR1

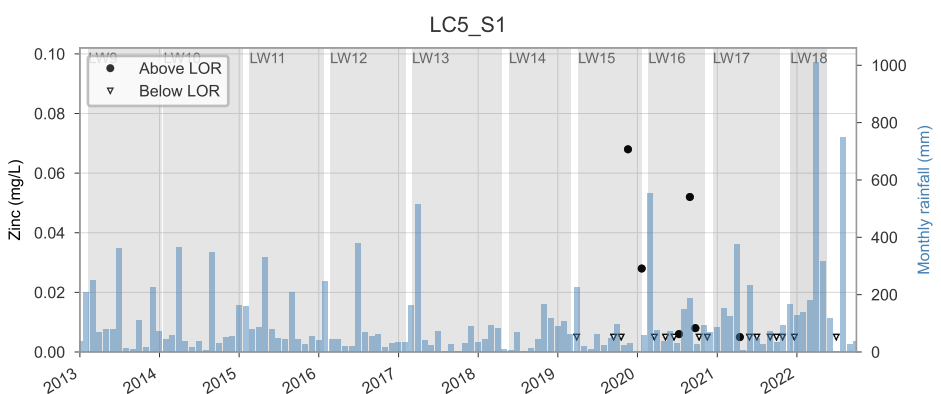
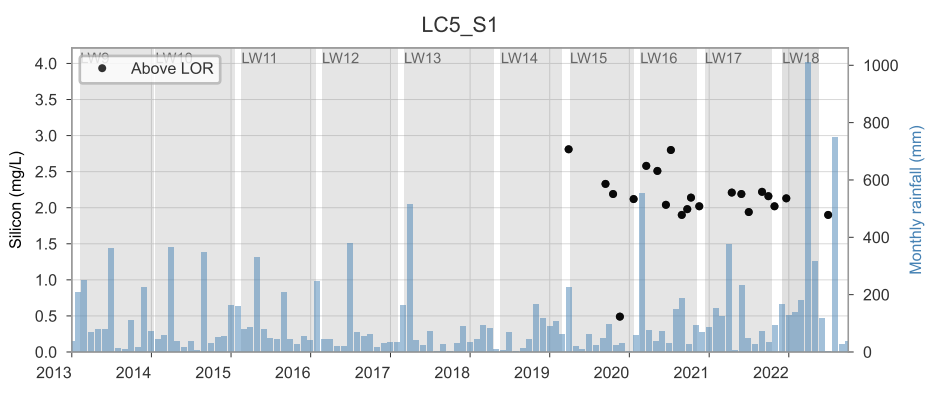
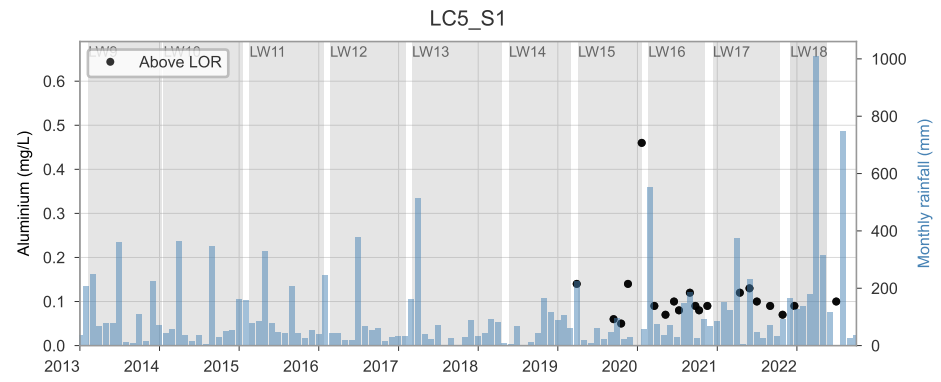
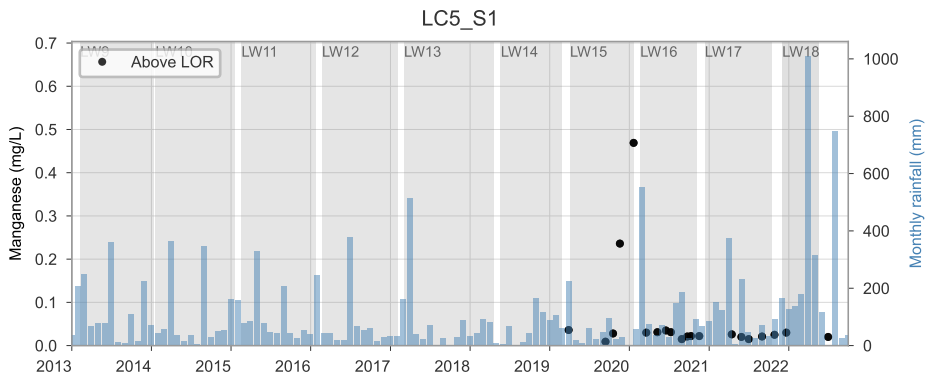
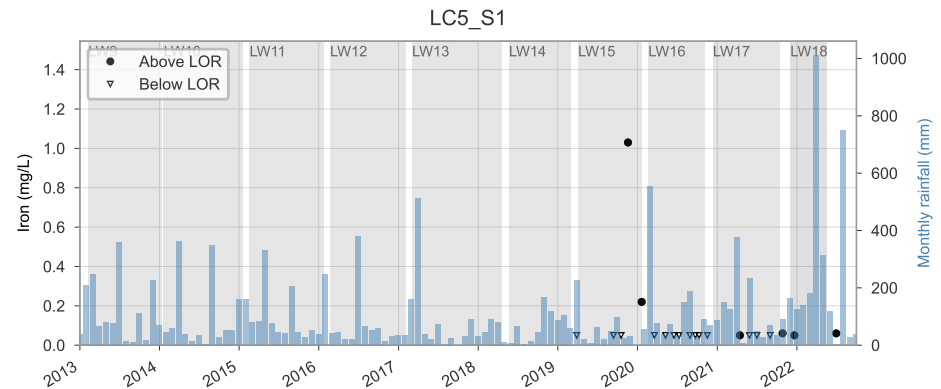
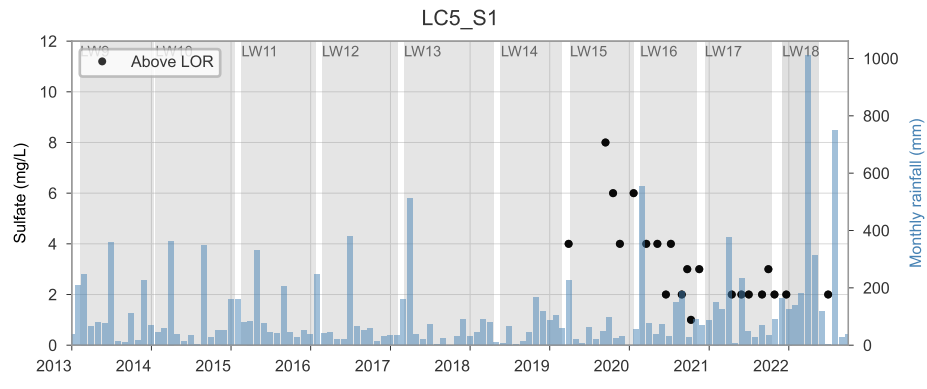


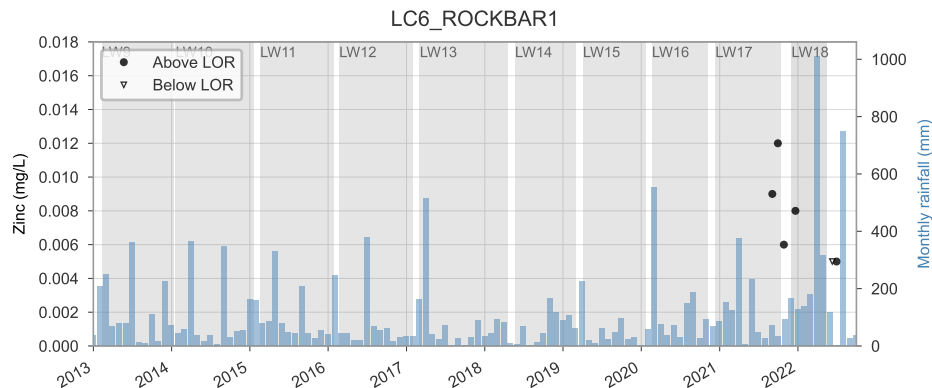
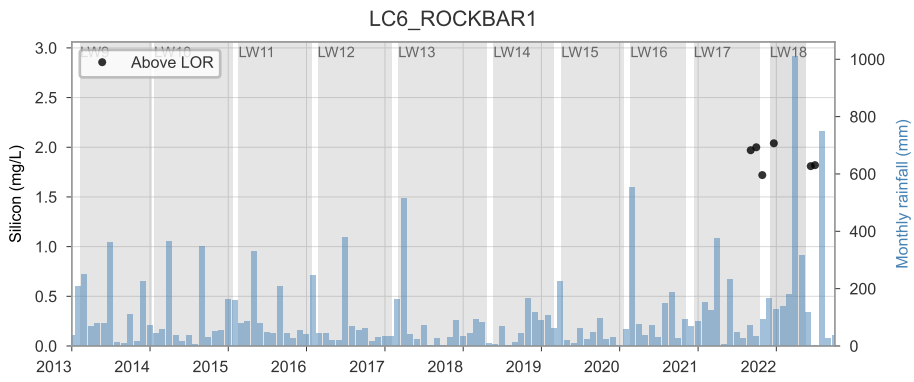
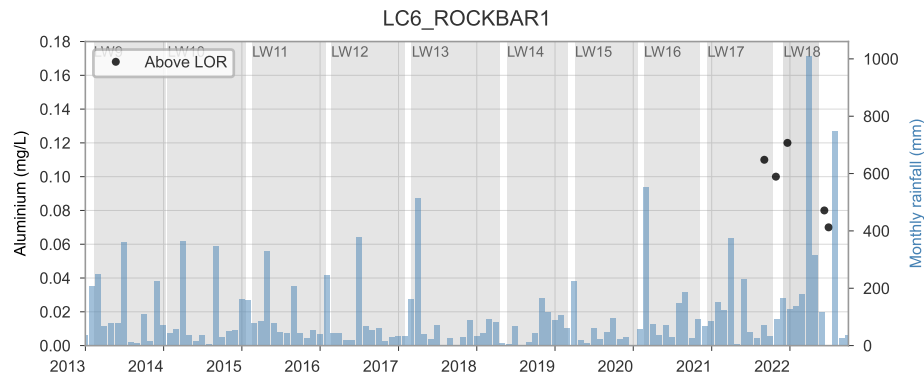
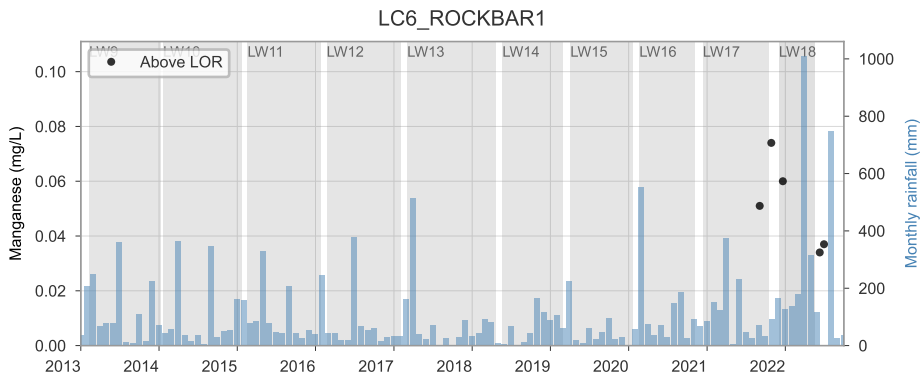
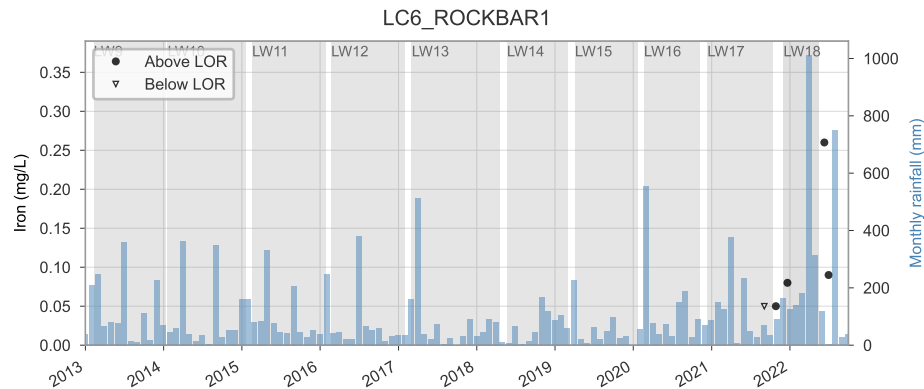
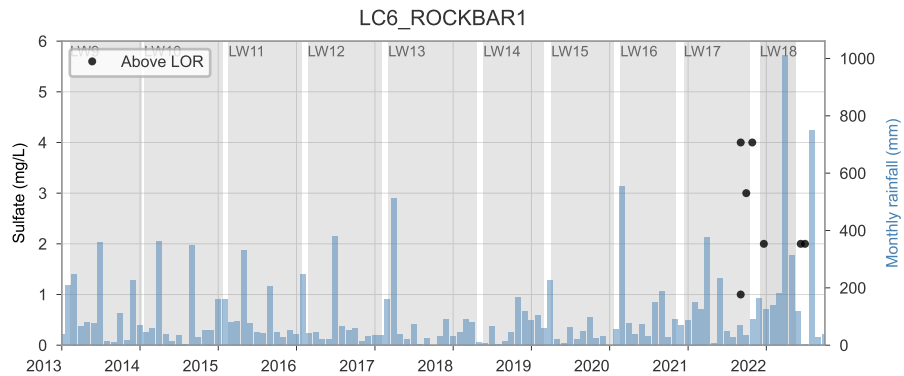
LA8_ROCKBAR1



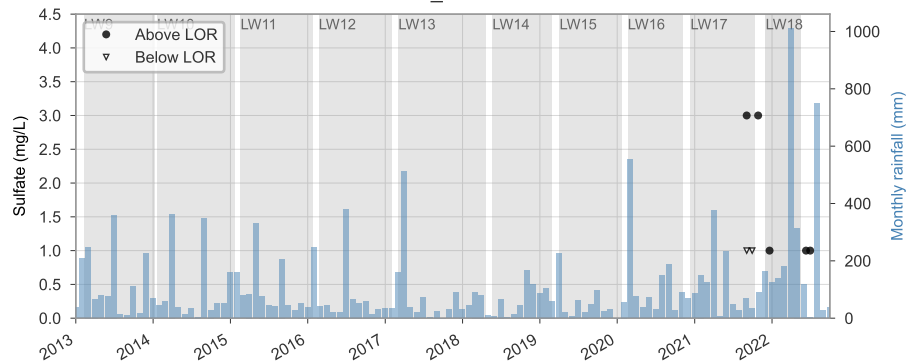




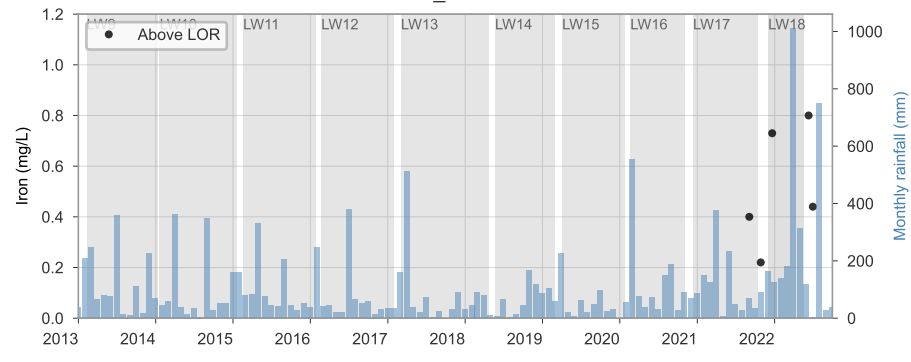




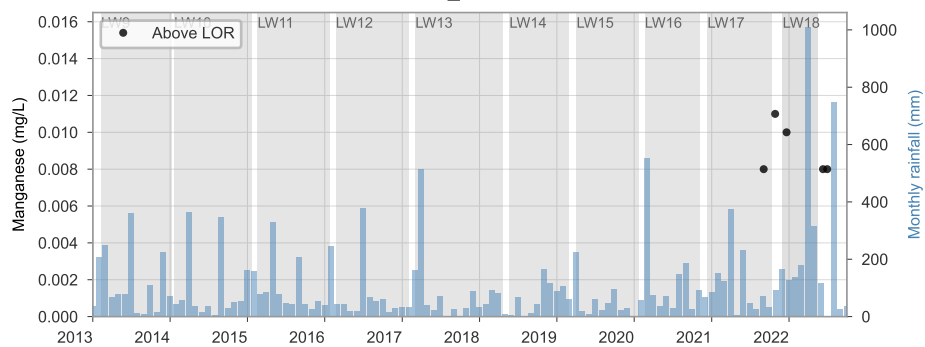
LC7_POOL2



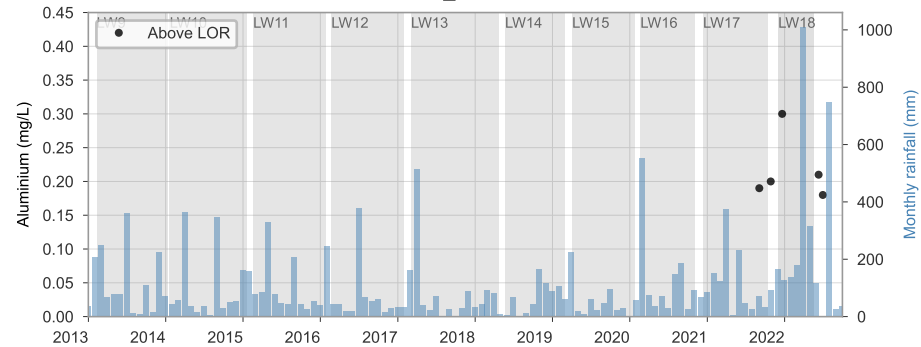
LC7_POOL2



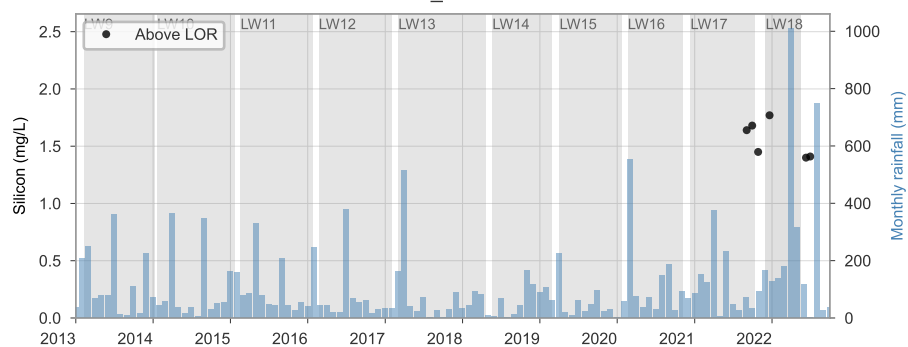
LC7_POOL2



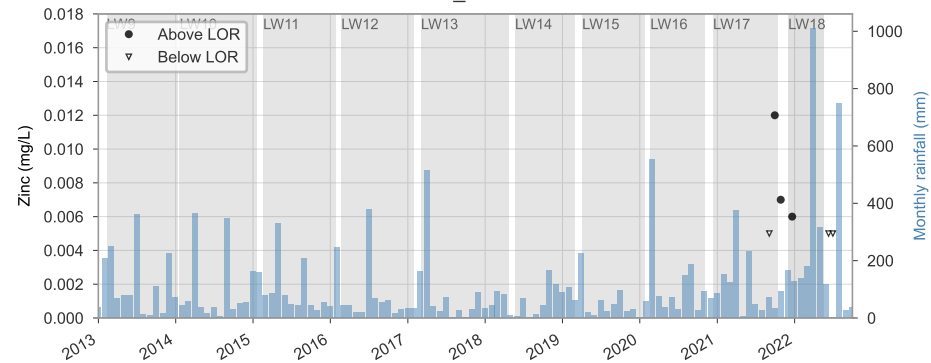
LC7_POOL2

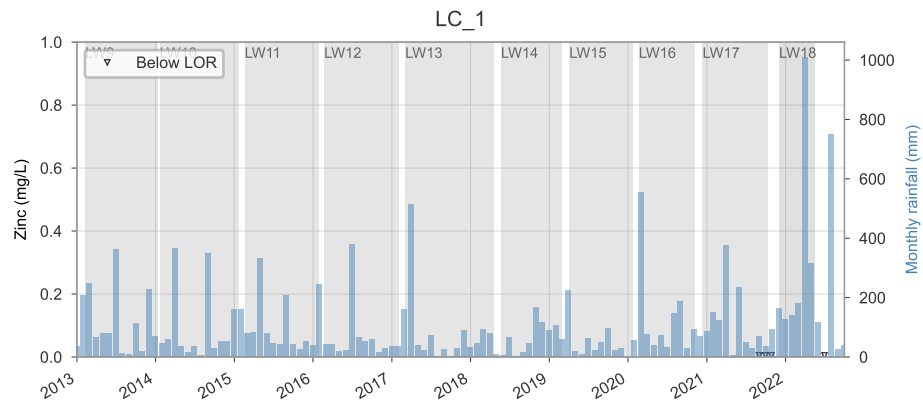
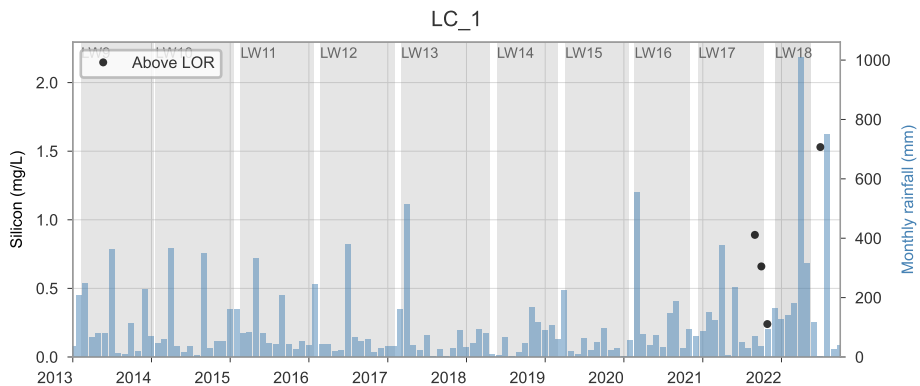
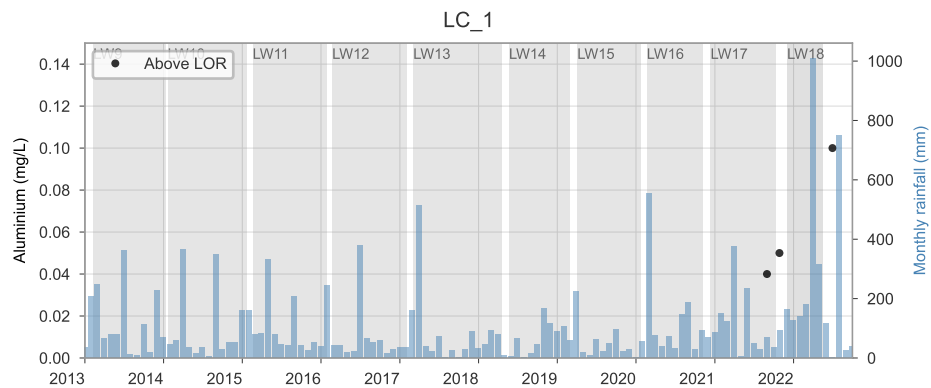
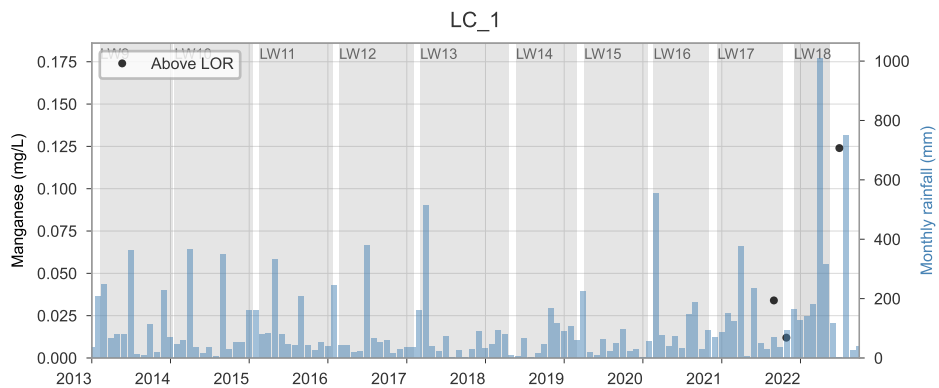
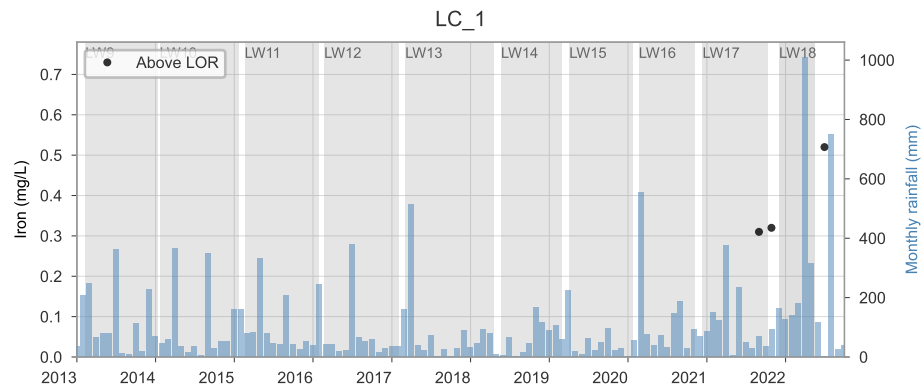
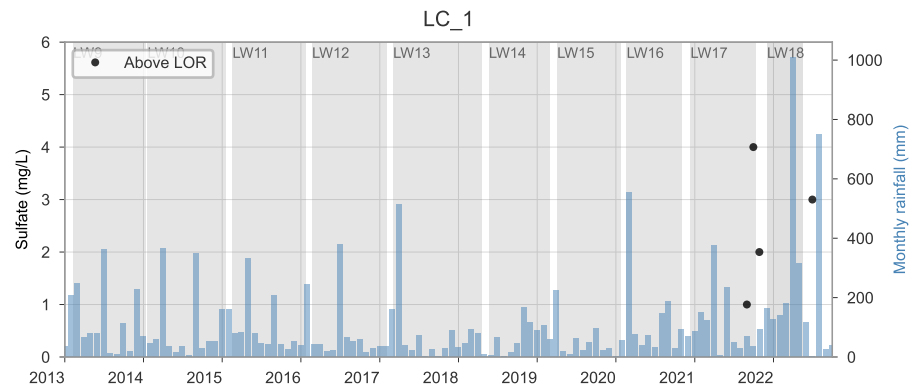


LC7_POOL2

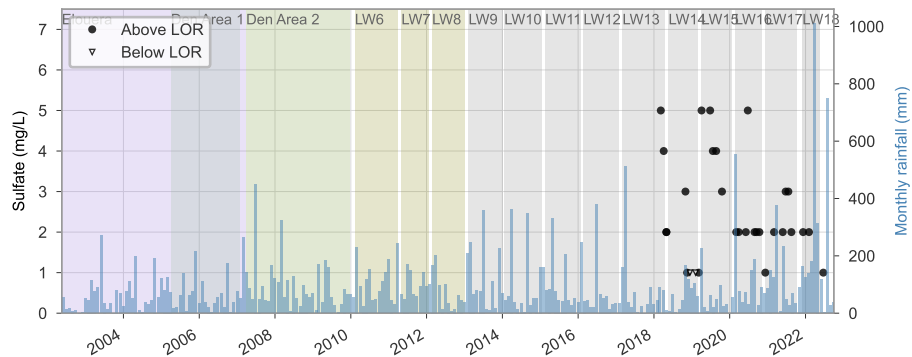


LC7_POOL2

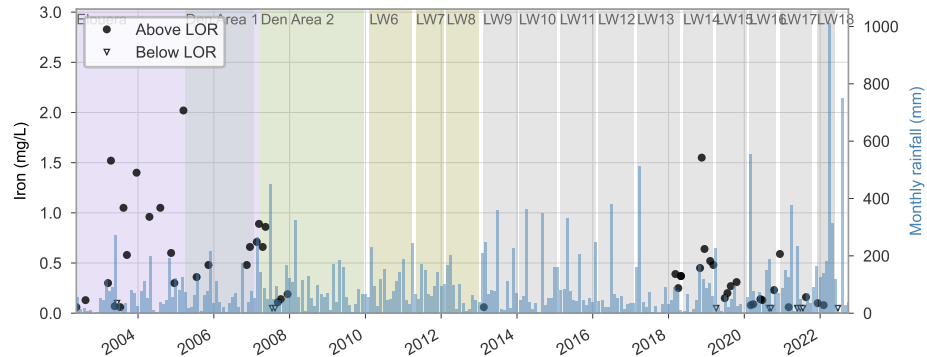




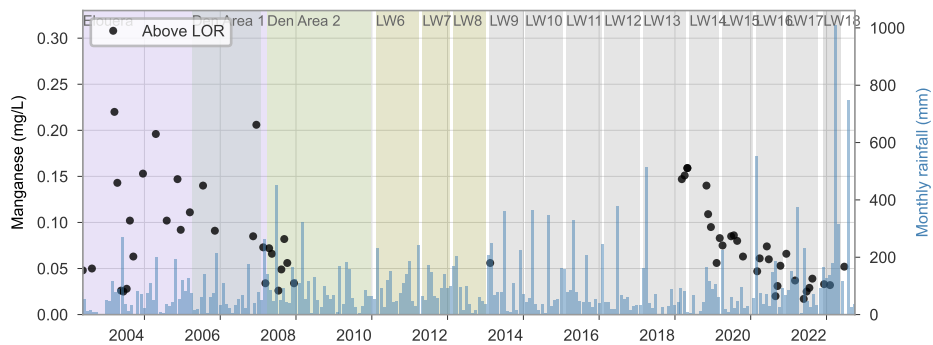
ND1_POOL2



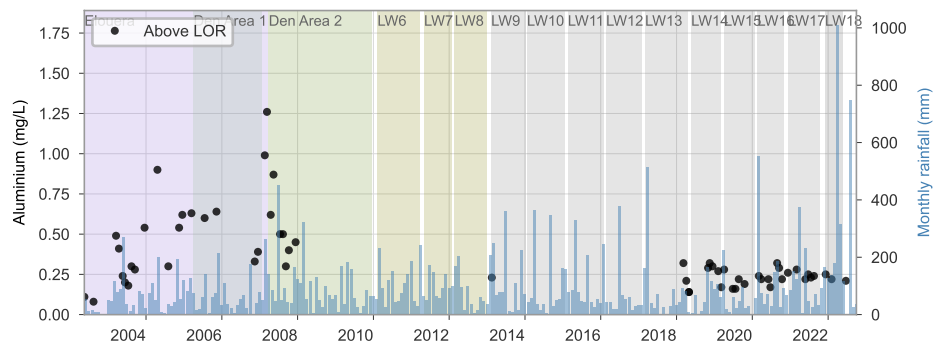
ND1_POOL2



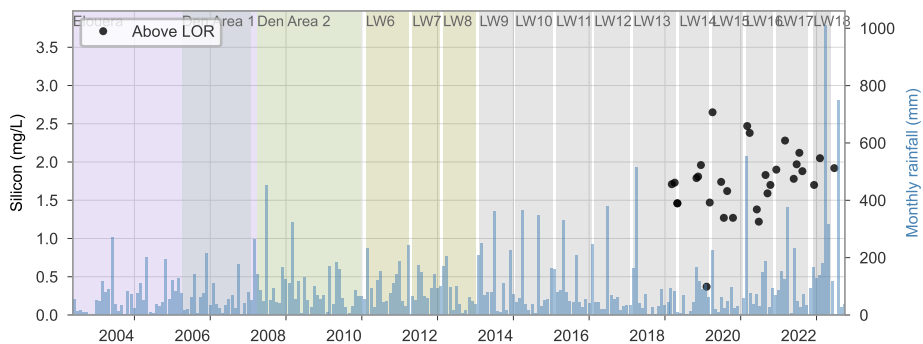
ND1_POOL2



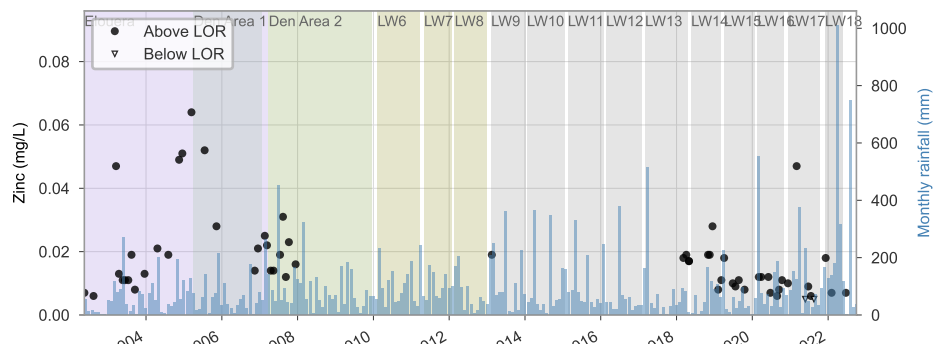
ND1_POOL2

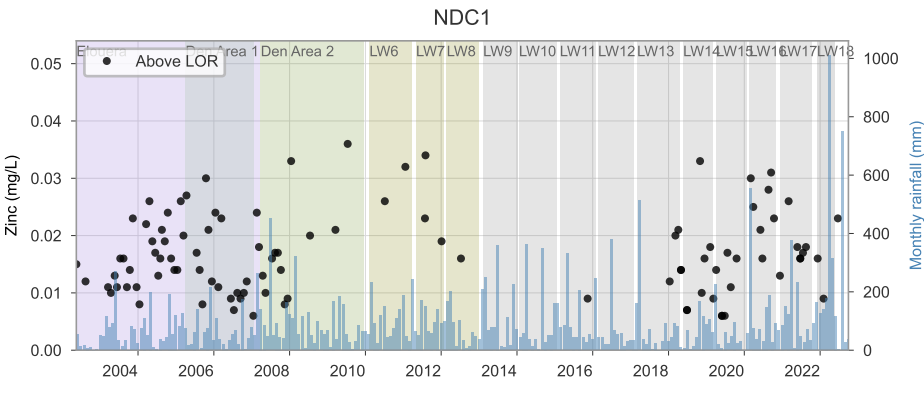
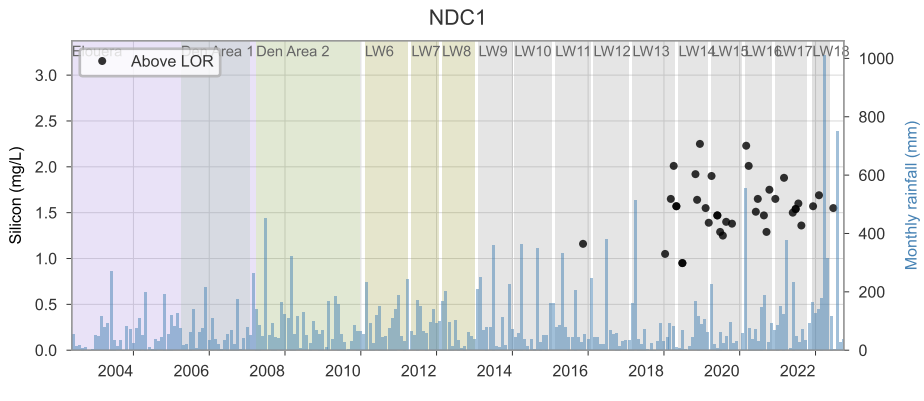
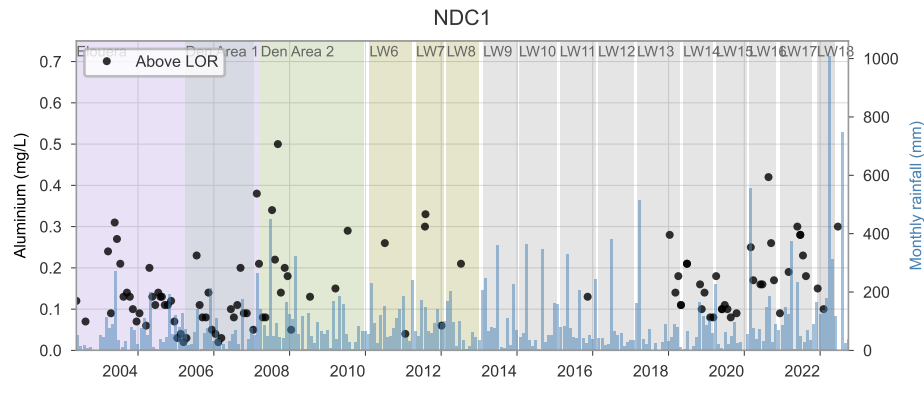
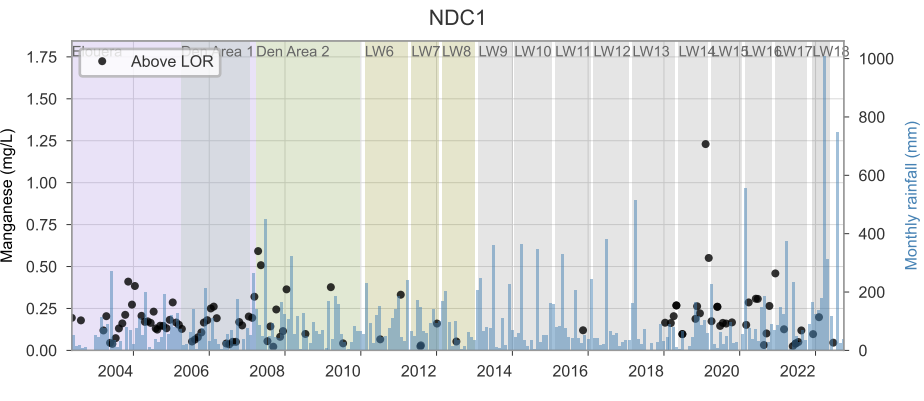
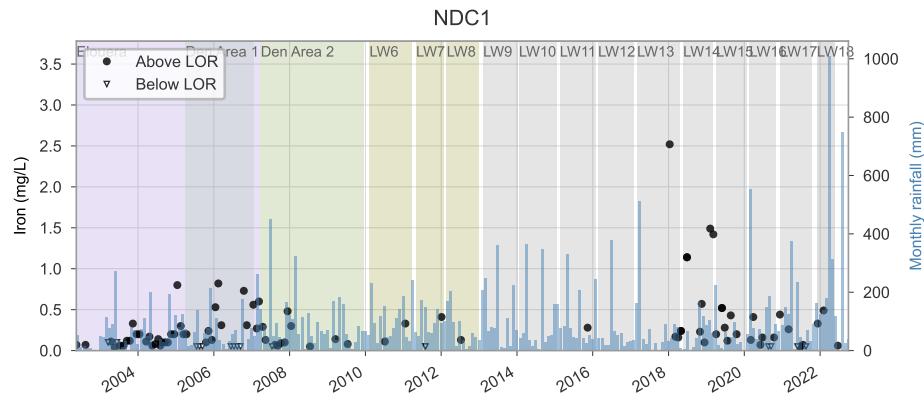
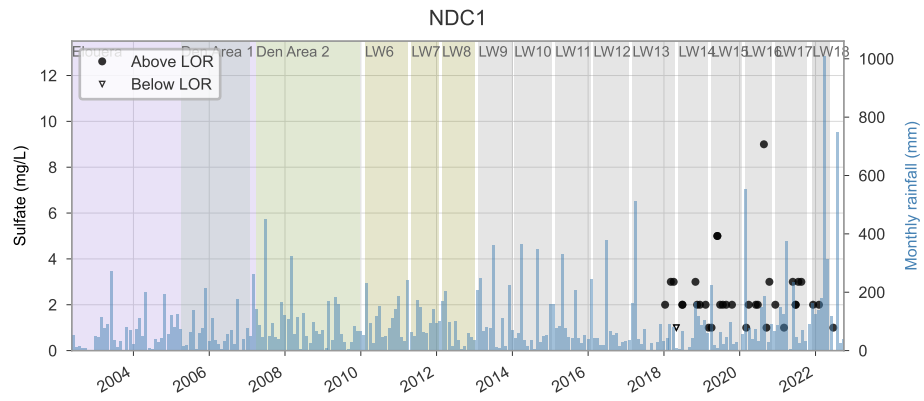


ND1_POOL2

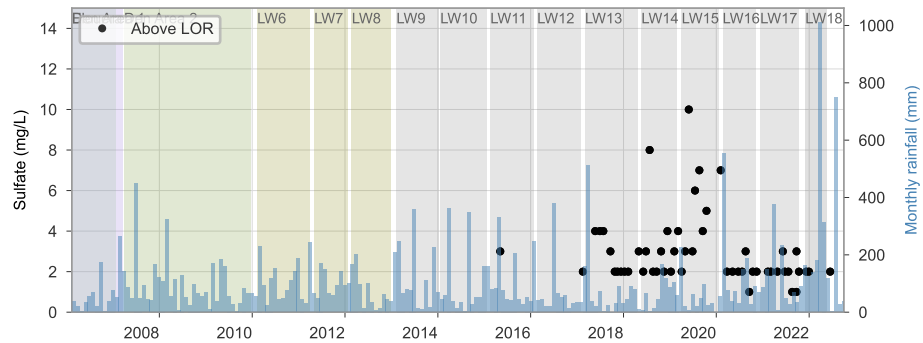


ND1_POOL2

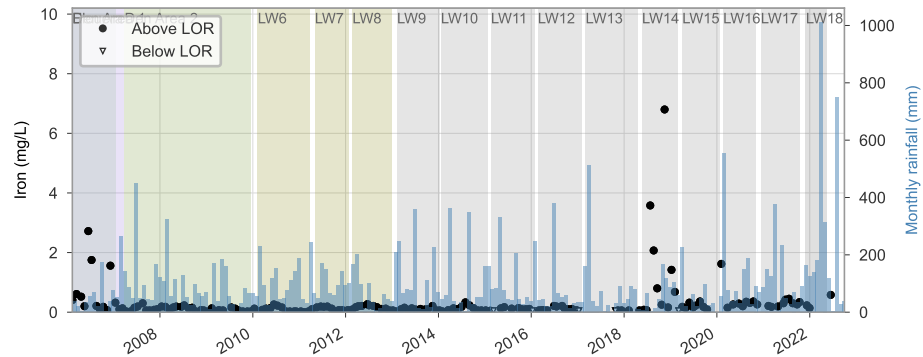




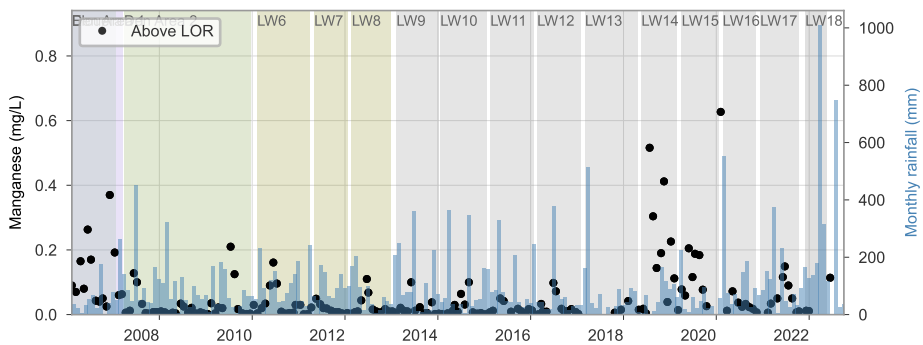
SANDY_CREEK_ARM



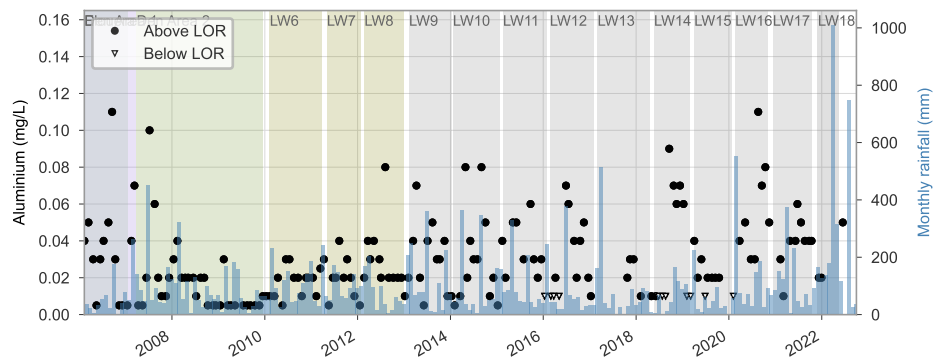
SANDY_CREEK_ARM



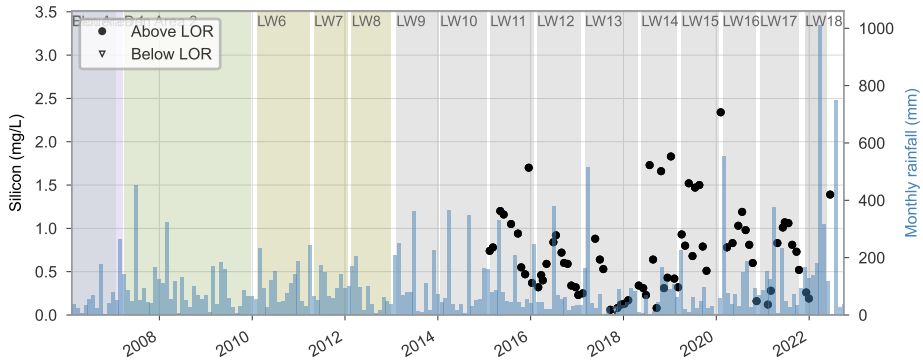
SANDY_CREEK_ARM



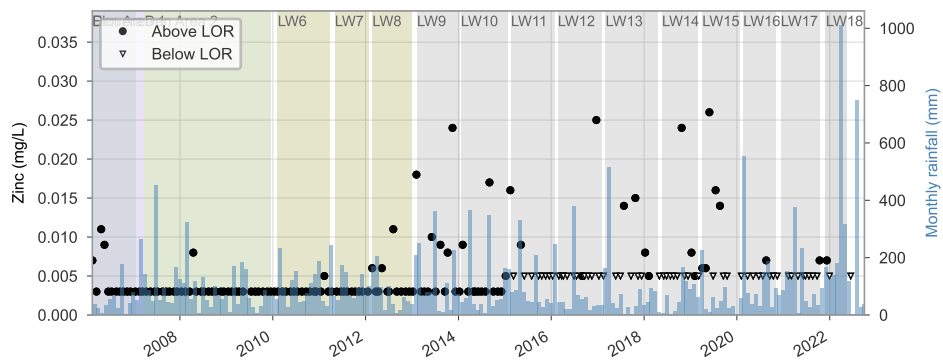
SANDY_CREEK_ARM



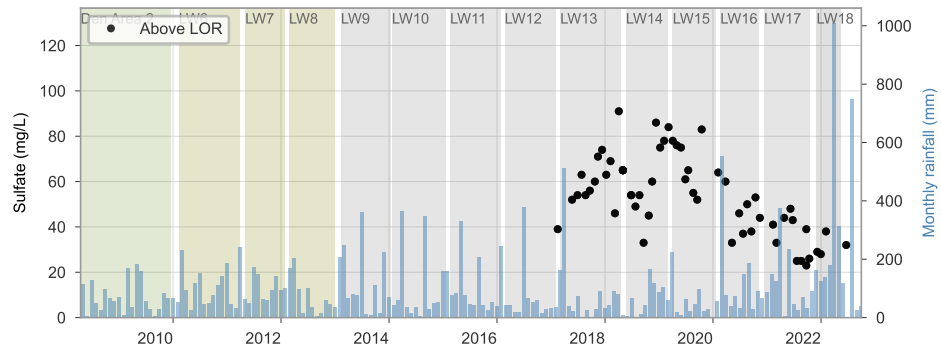
SANDY_CREEK_ARM



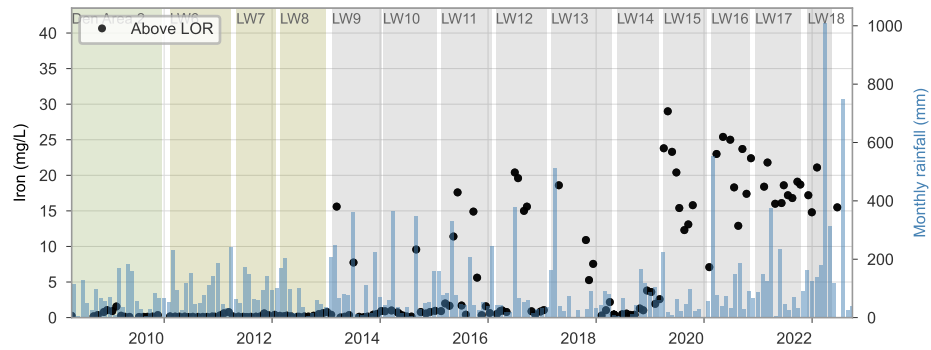
SANDY_CREEK_ARM



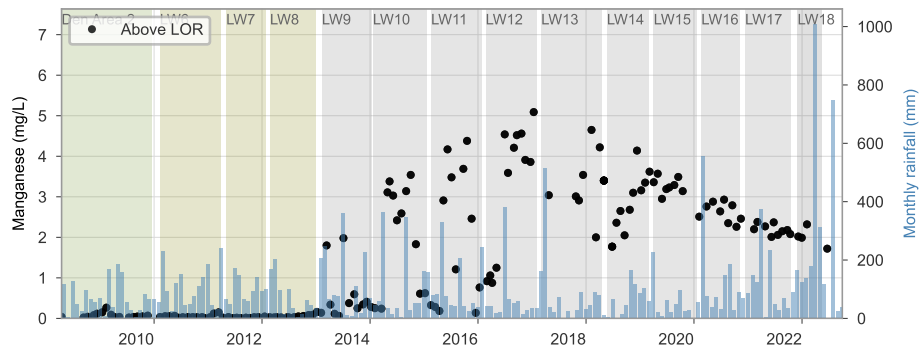
SC10C_POOL1



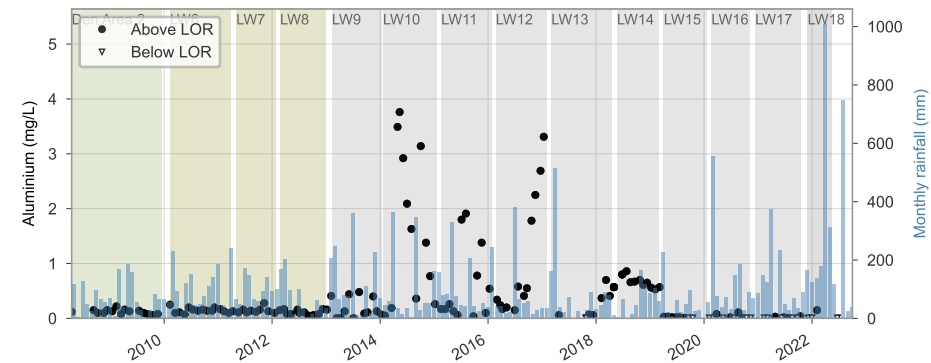
SC10C_POOL1



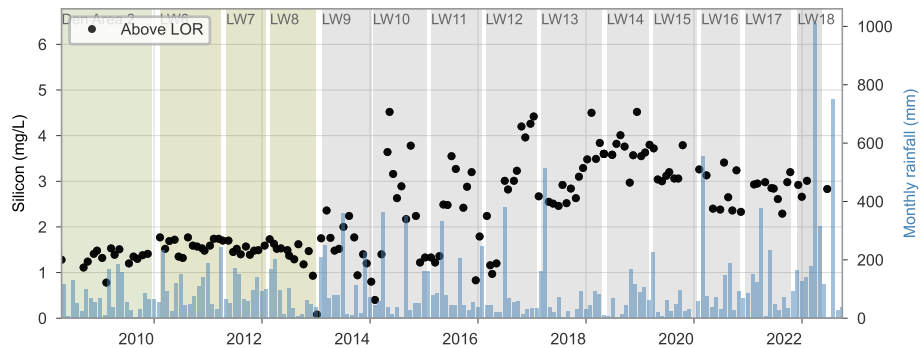
SC10C_POOL1



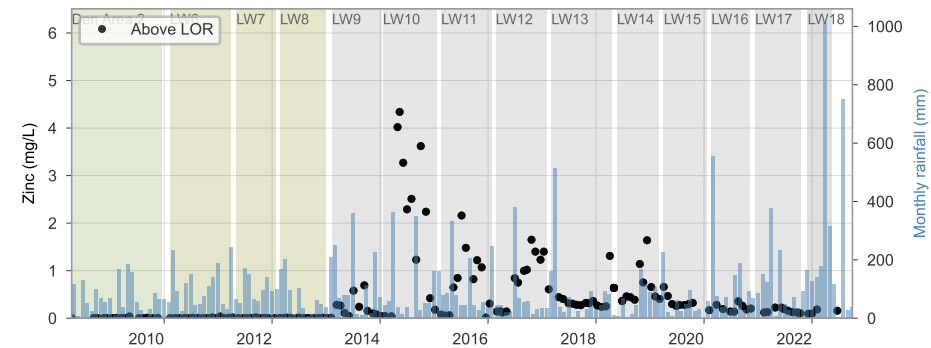
SC10C_POOL1

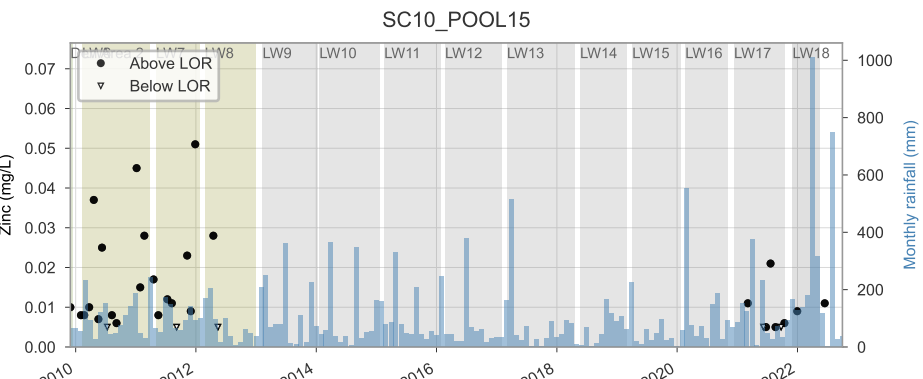
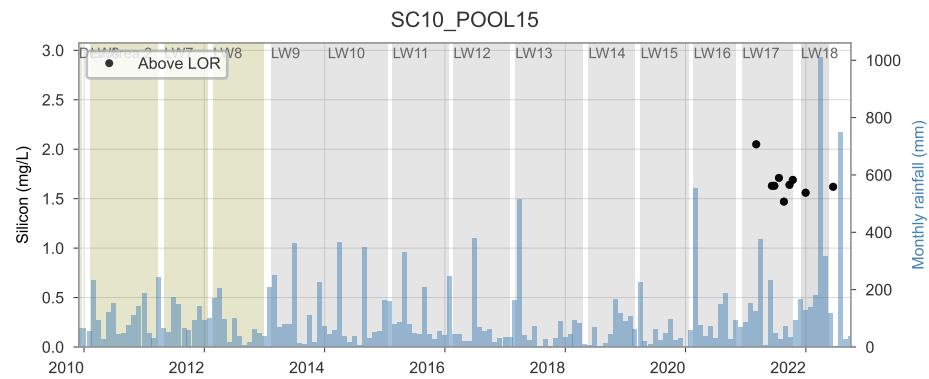
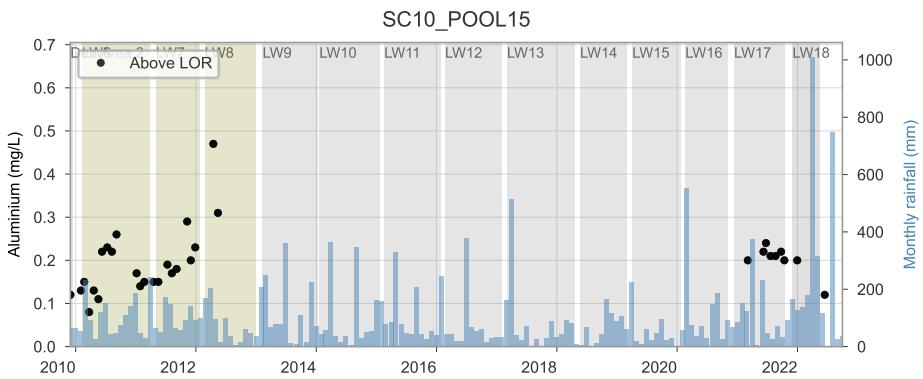
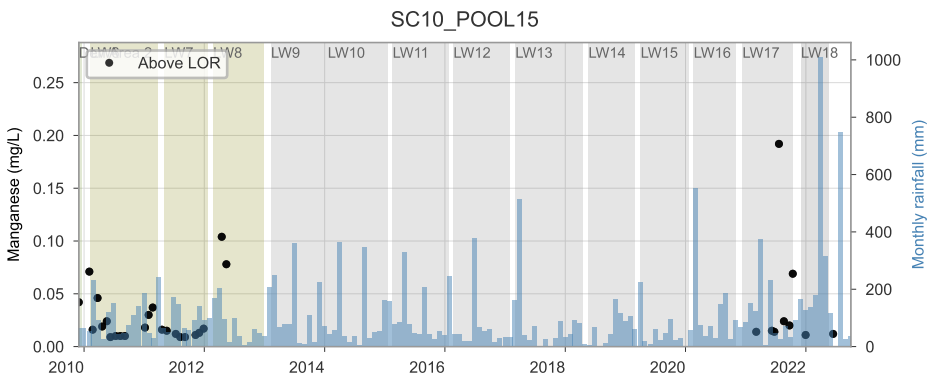
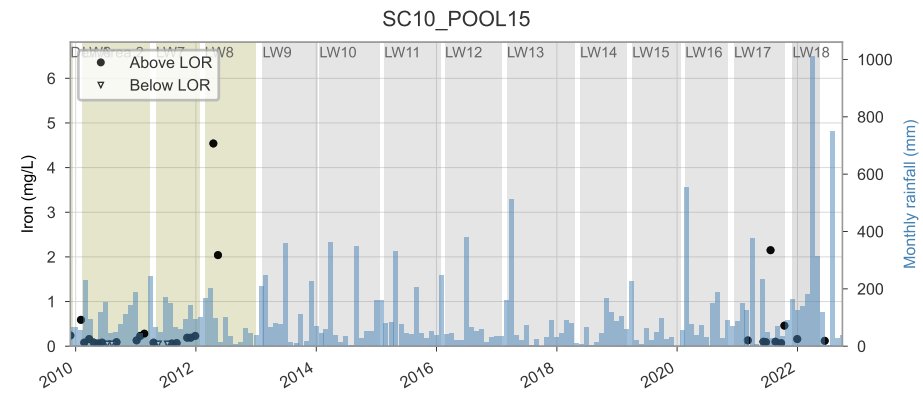
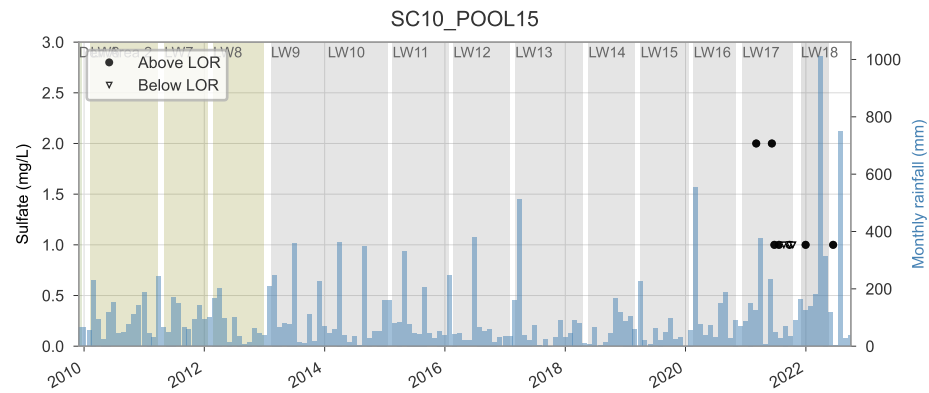


SC10C_POOL1

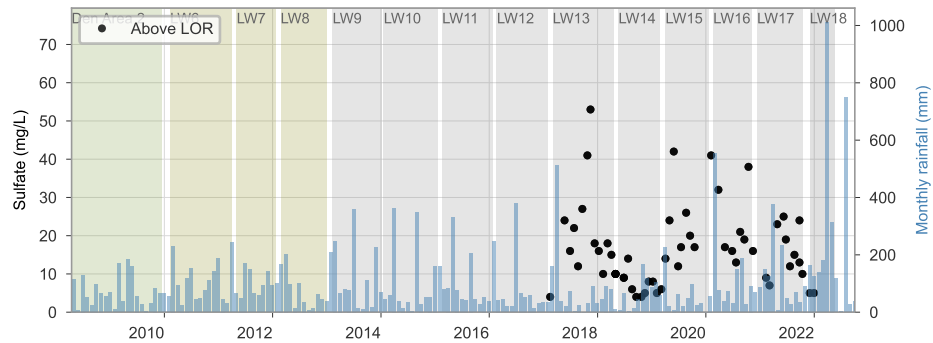


SC10C_POOL1

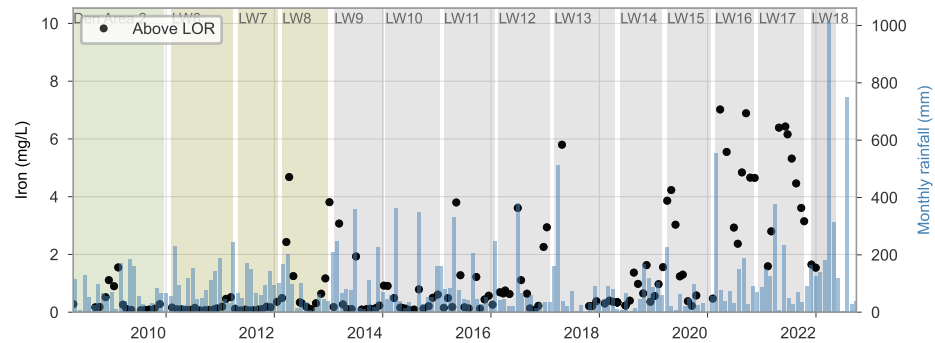




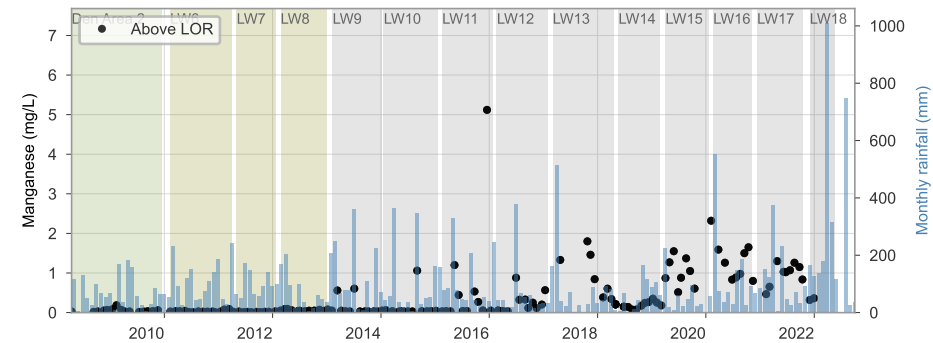
SC10_ROCKBAR3



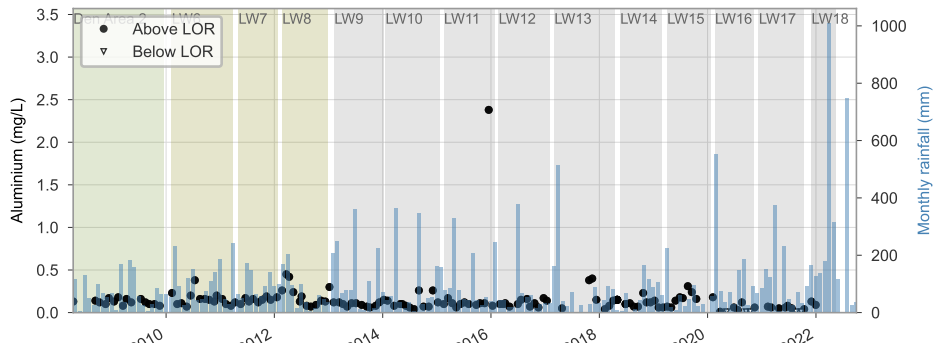
SC10_ROCKBAR3



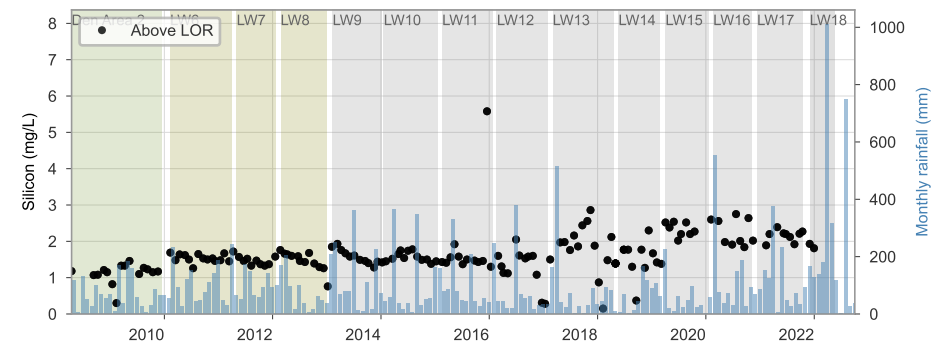
SC10_ROCKBAR3



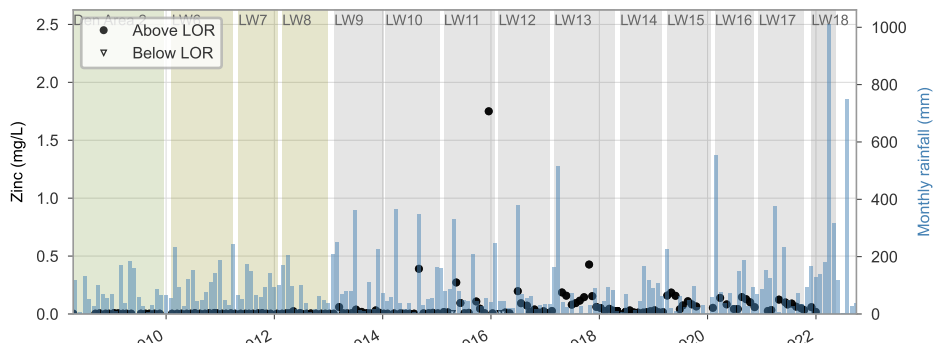
SC10_ROCKBAR3



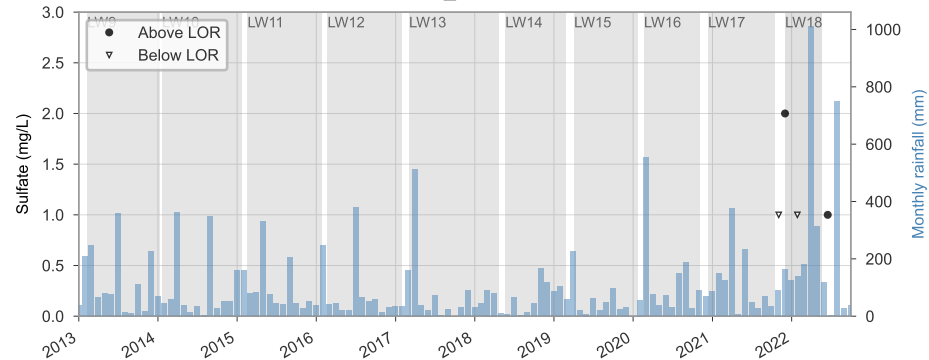
SC10_ROCKBAR3



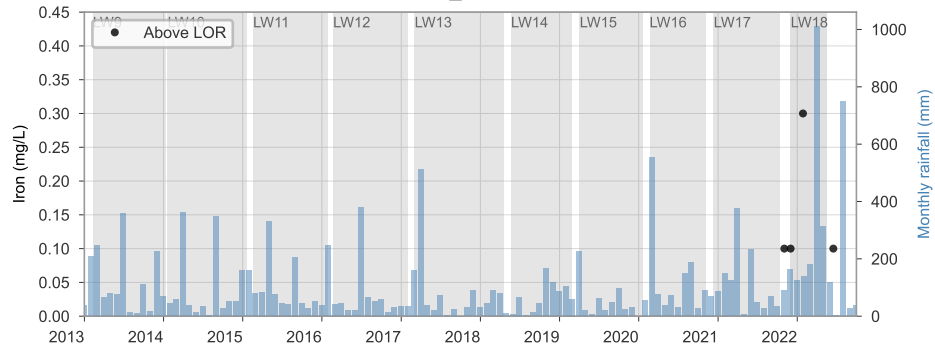
SC10_ROCKBAR3



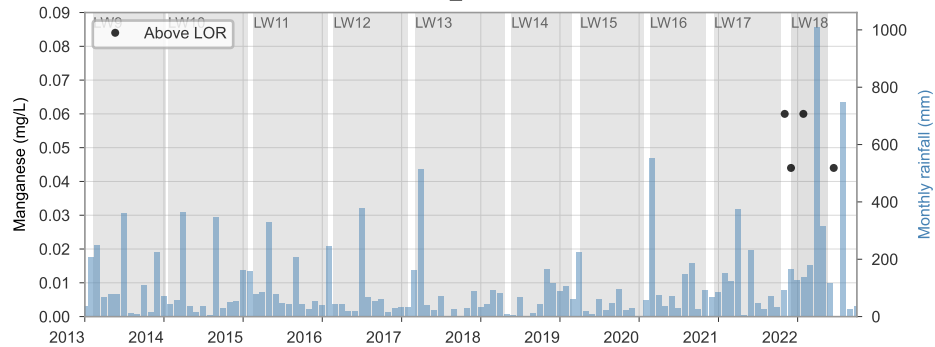
SC7_POOL1



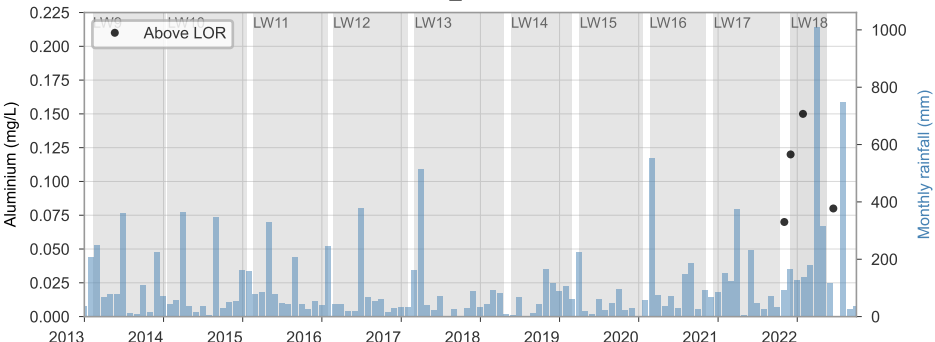
SC7_POOL1



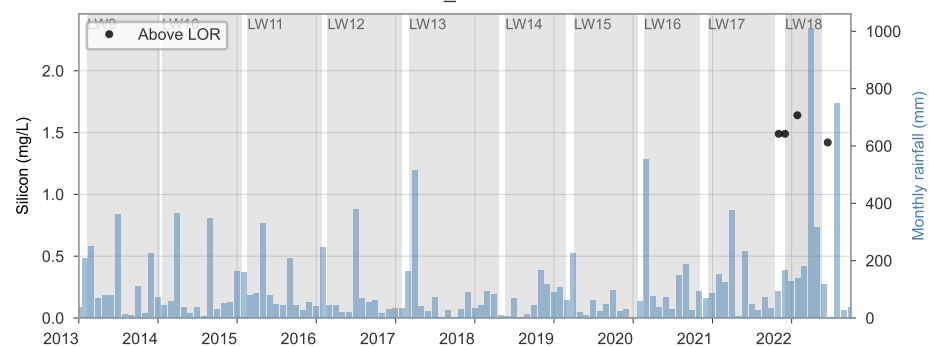
SC7_POOL1



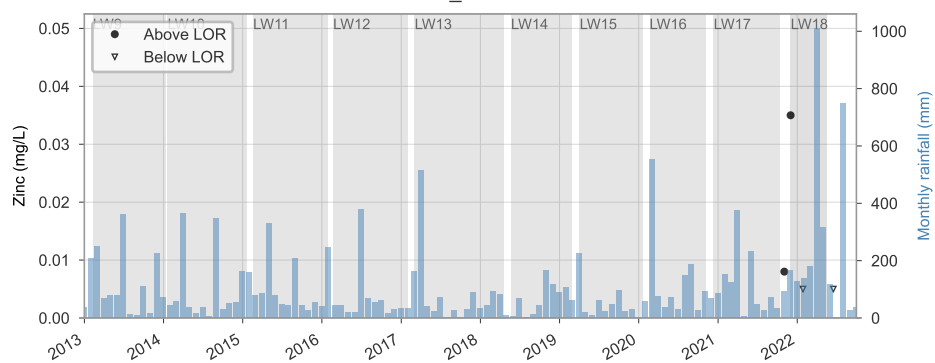
SC7_POOL1



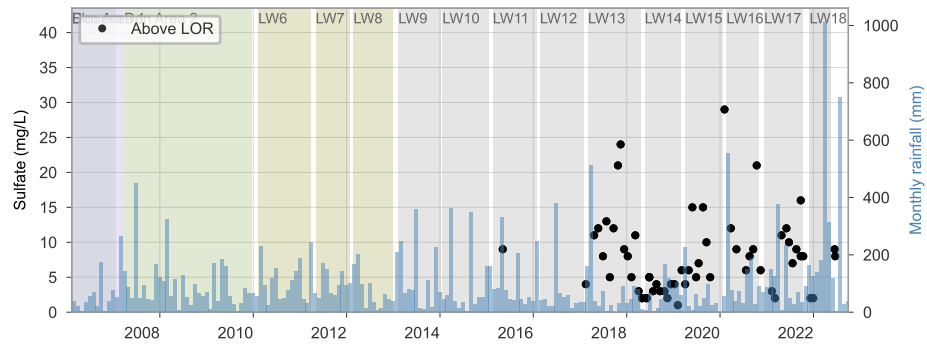
SC7_POOL1



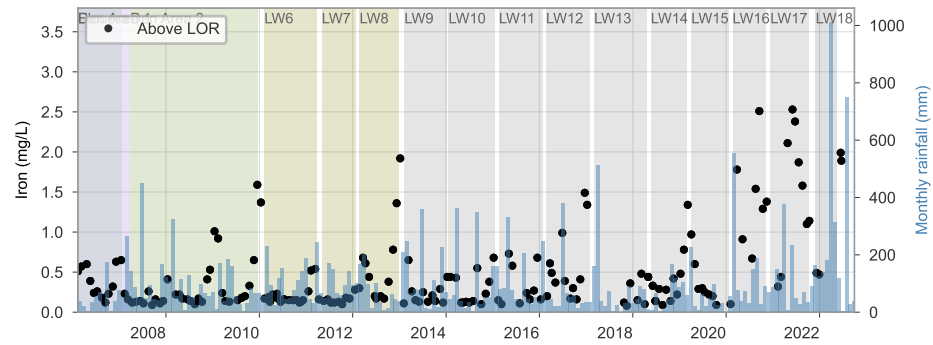
SC7_POOL1



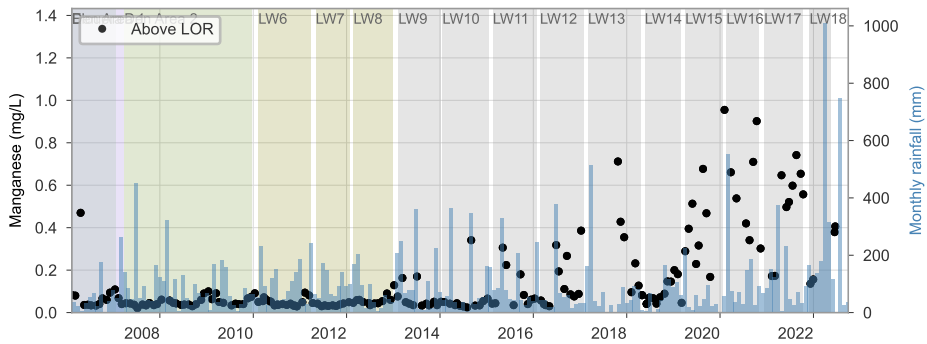
SCK_ROCKBAR5



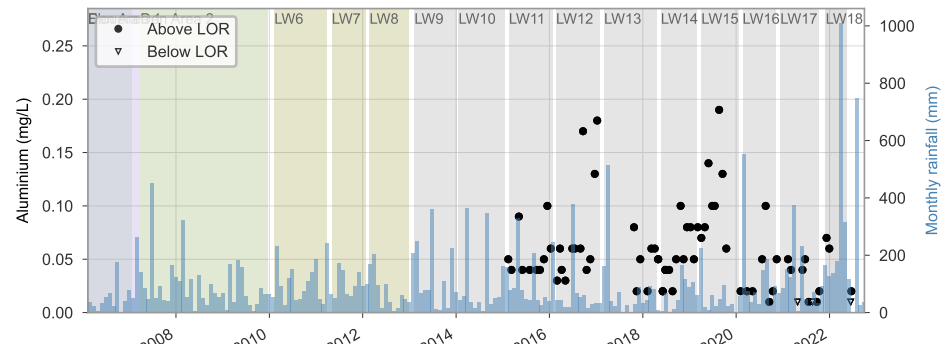
SCK_ROCKBAR5



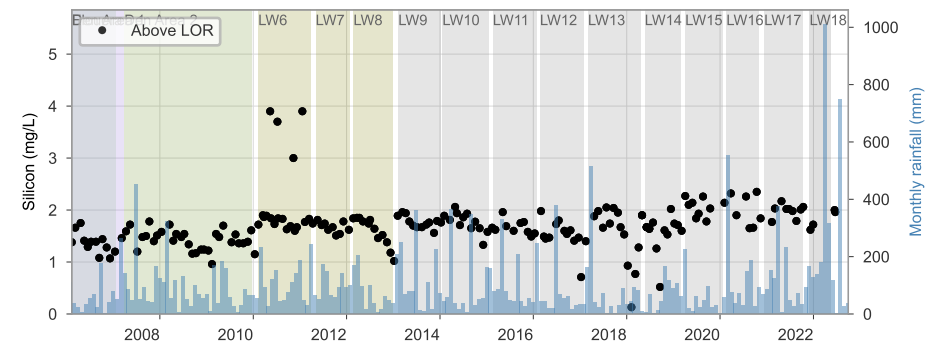
SCK_ROCKBAR5



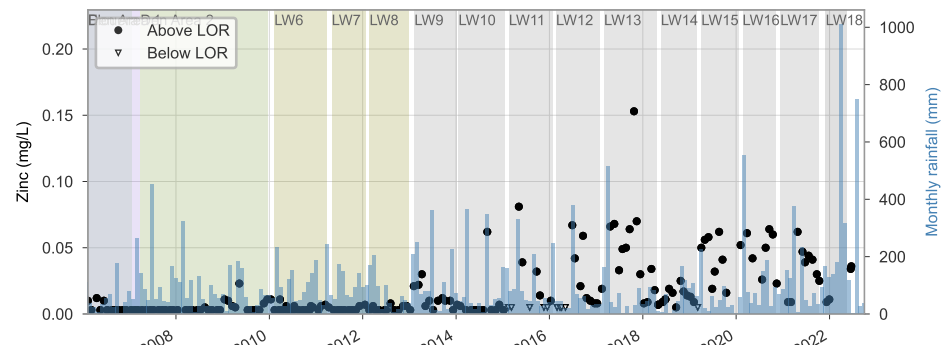
SCK_ROCKBAR5

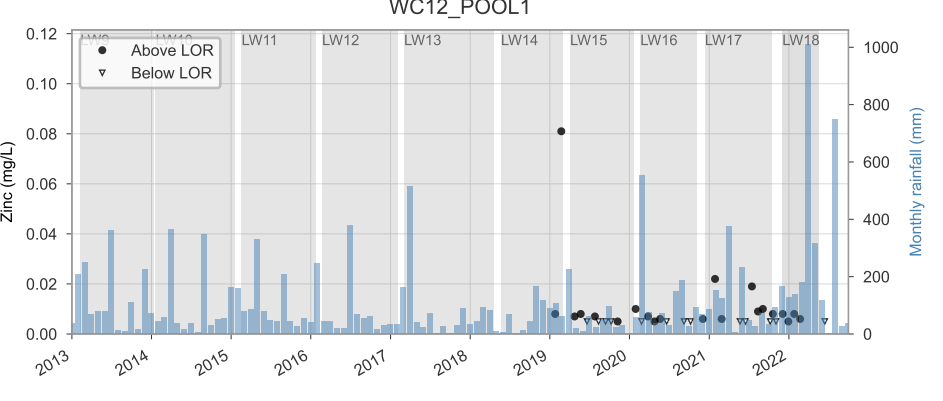
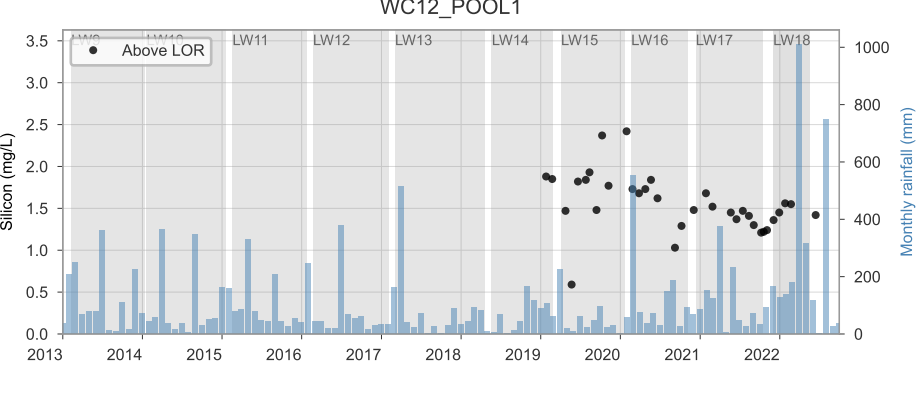
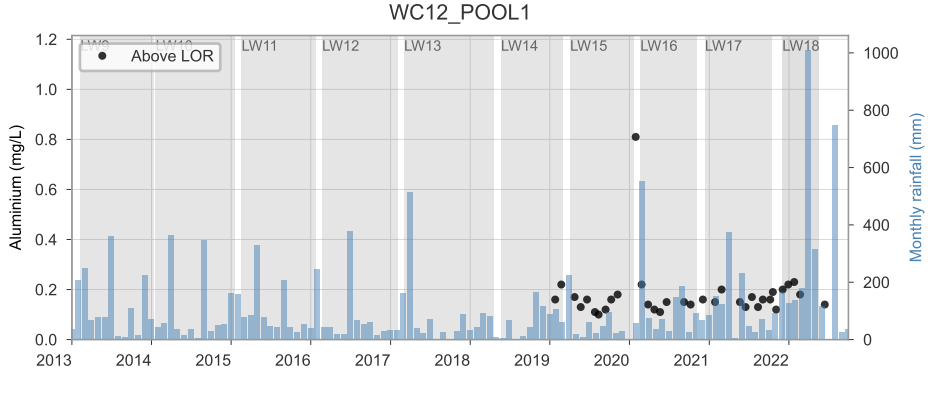
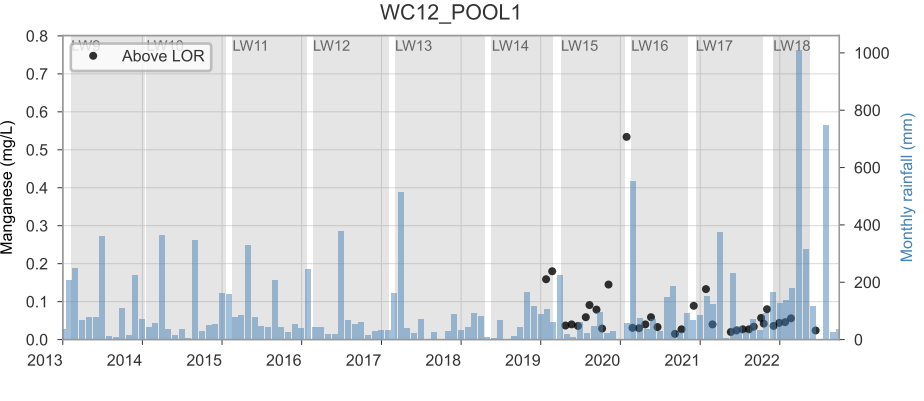
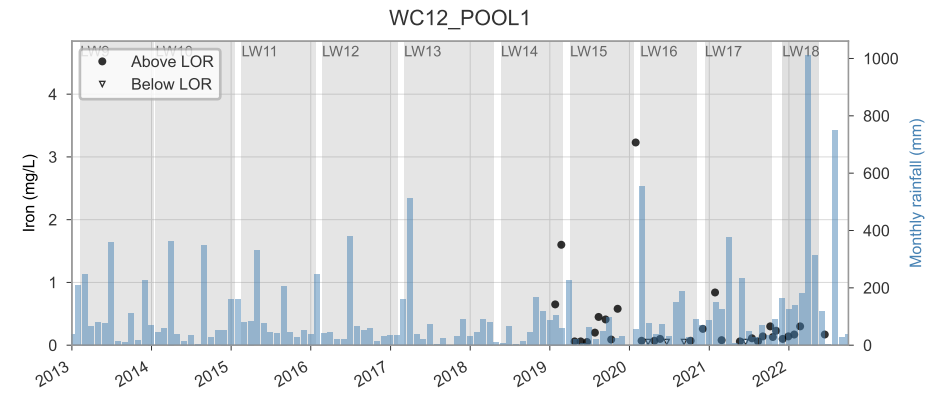
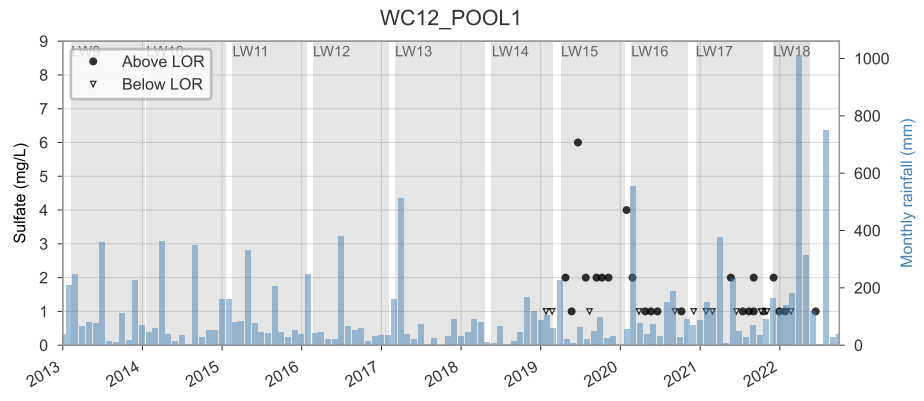


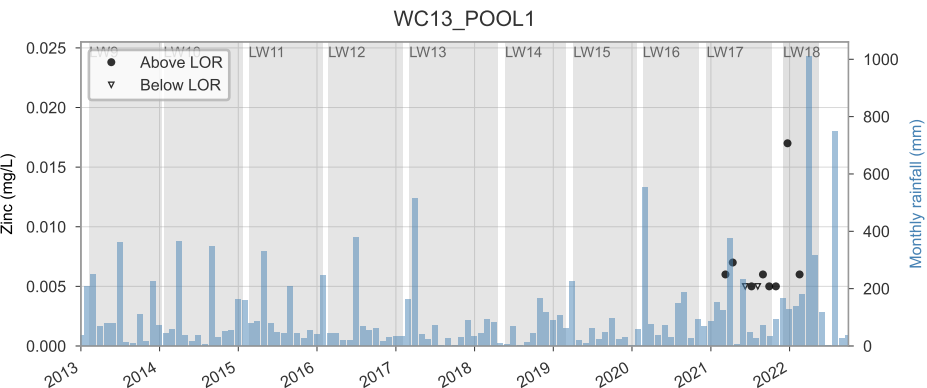
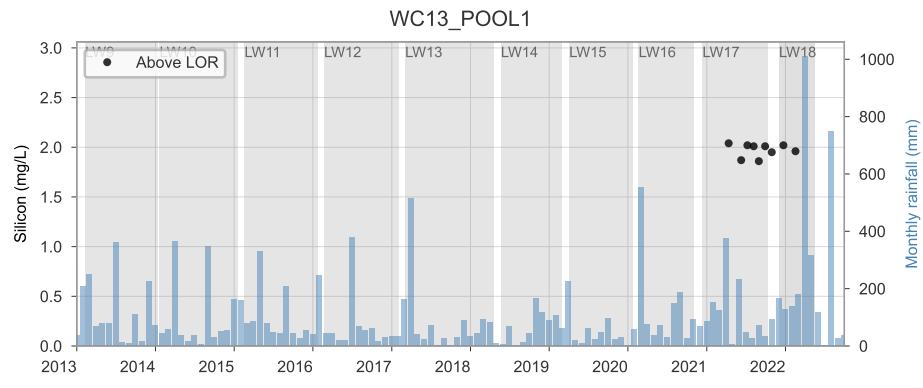
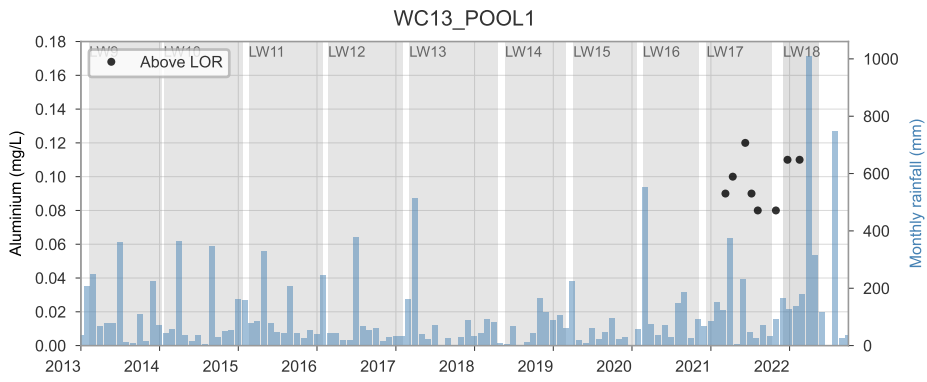
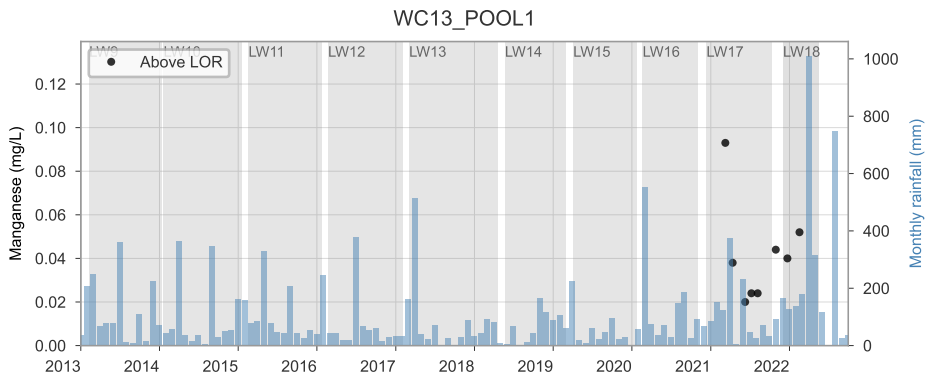
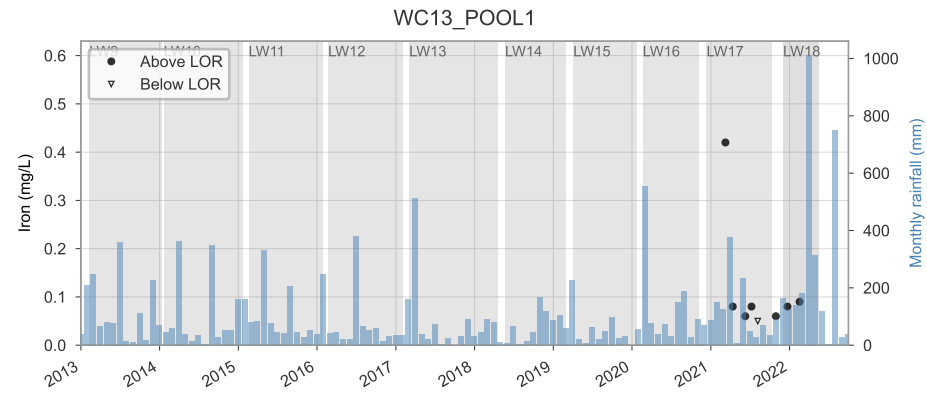
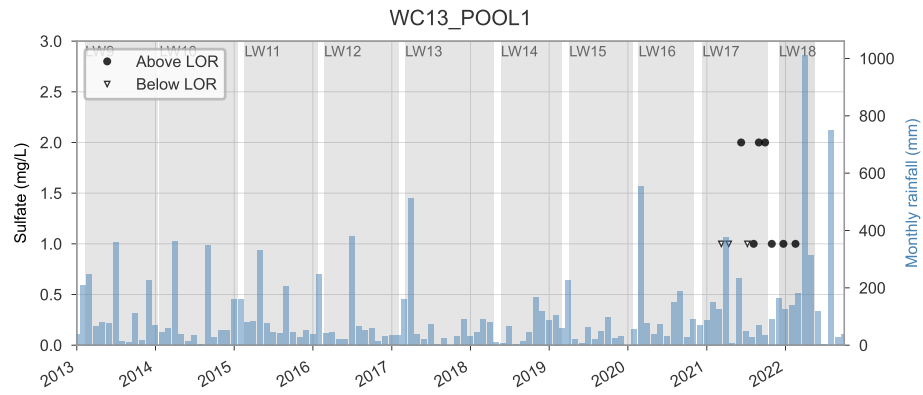
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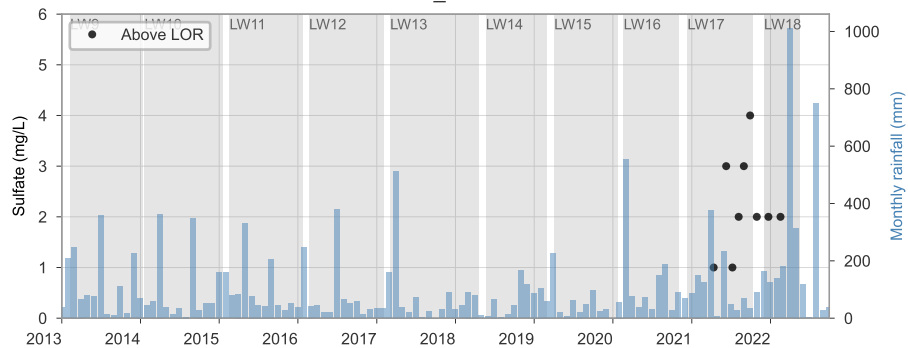
SCK_ROCKBAR5



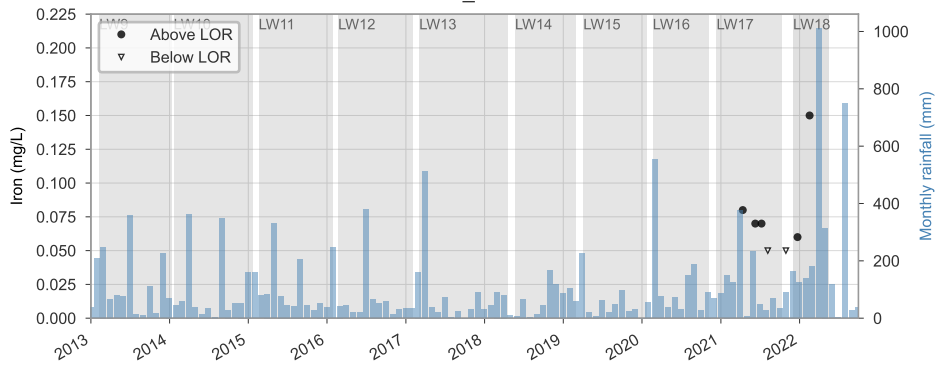




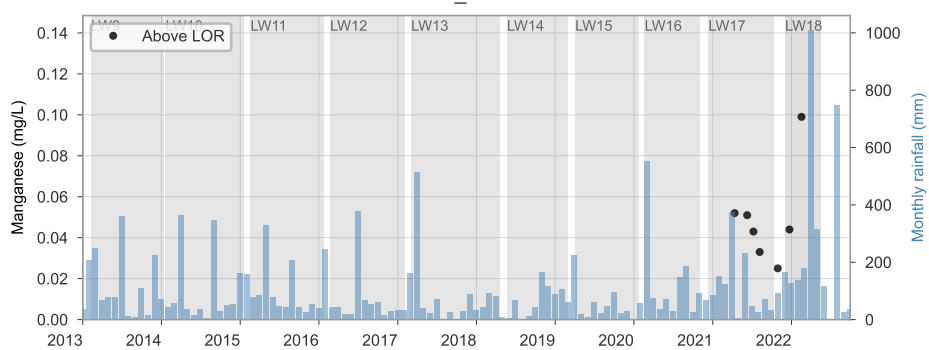
WC14_POOL3



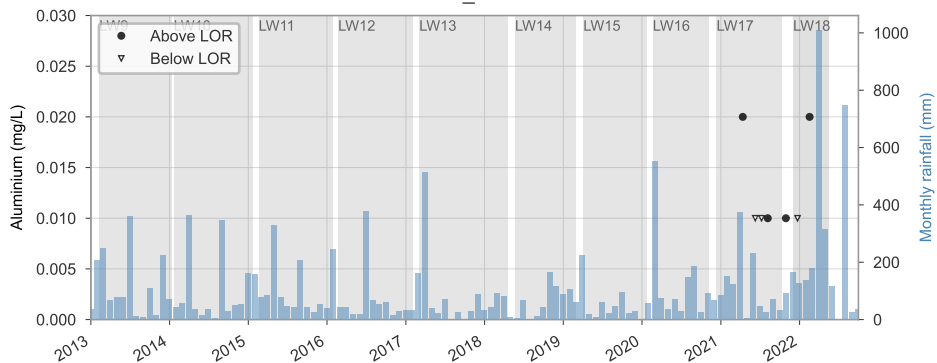
WC14_POOL3



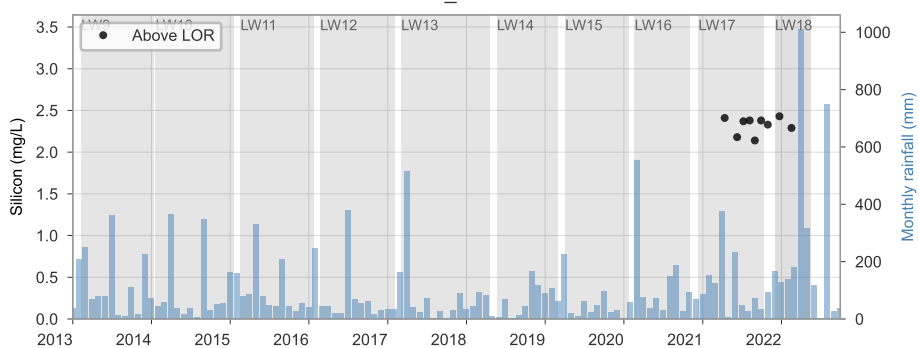
WC14_POOL3



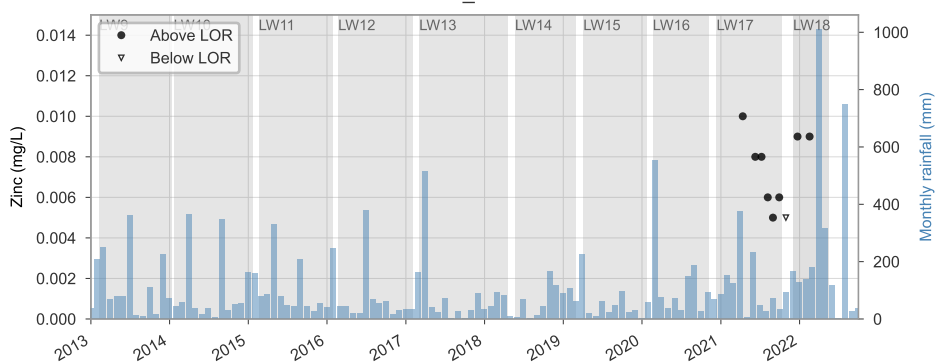
WC14_POOL3



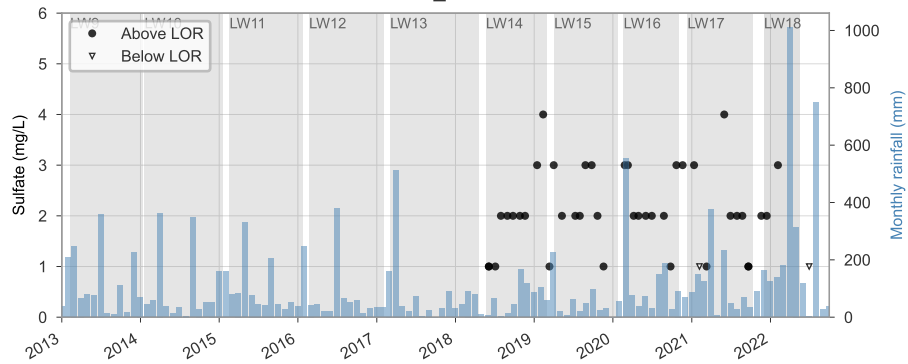
WC14_POOL3



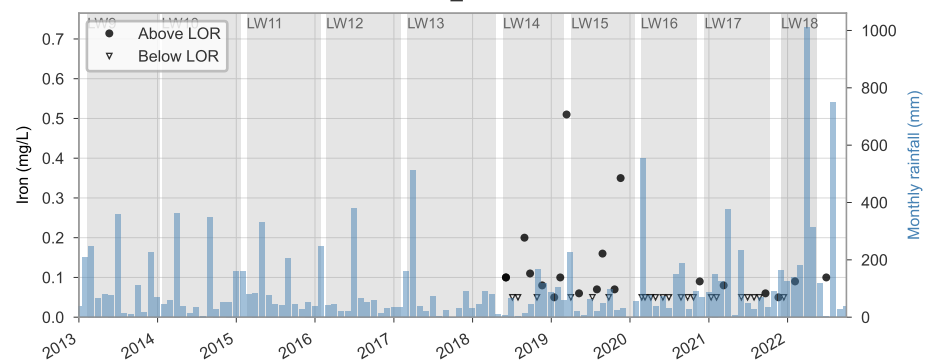
WC14_POOL3



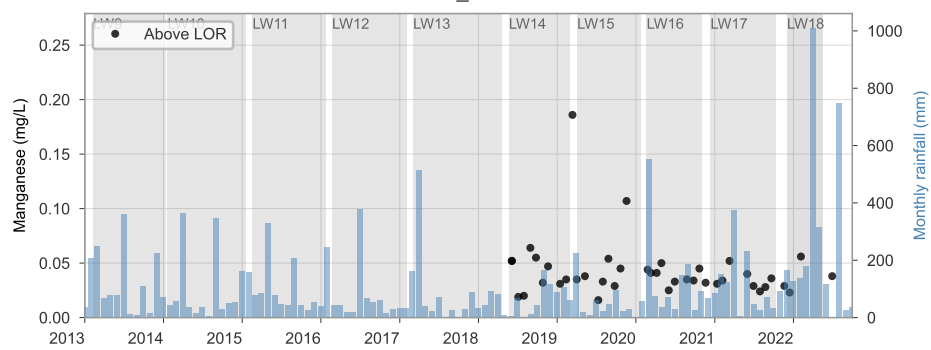
WC15_POOL2



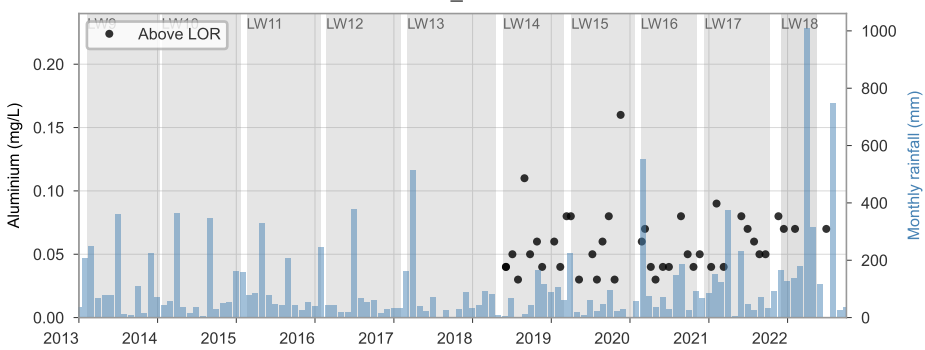
WC15_POOL2



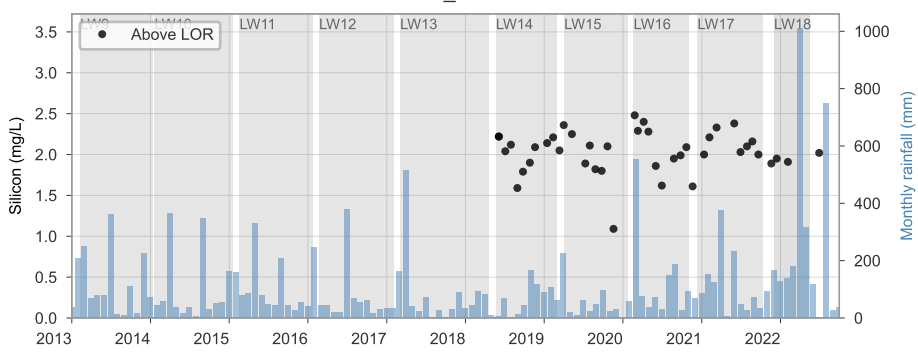
WC15_POOL2



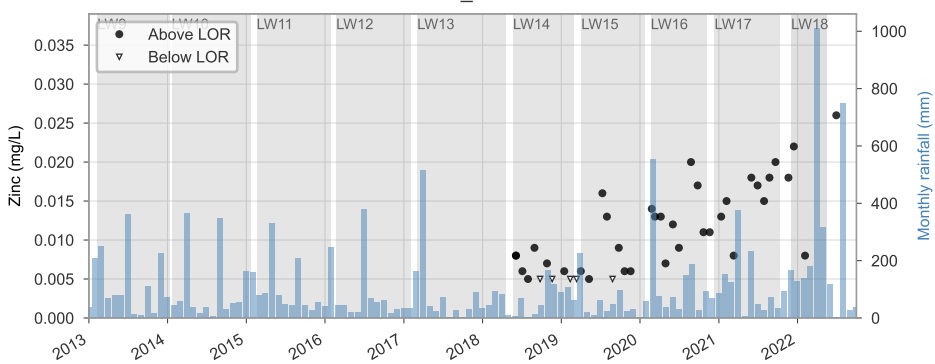
WC15_POOL2

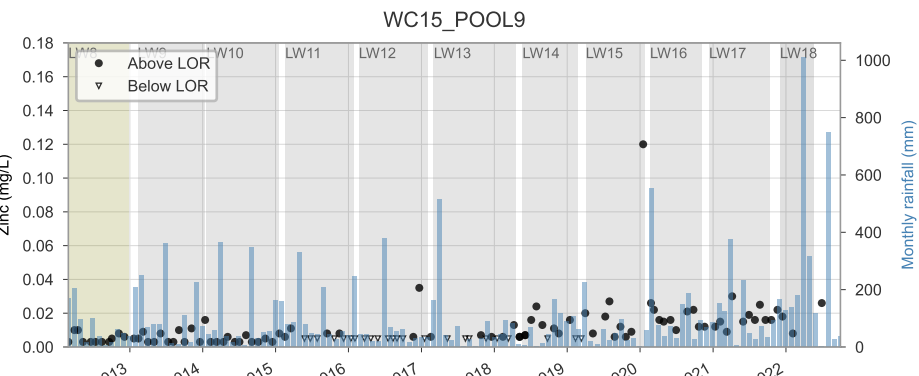
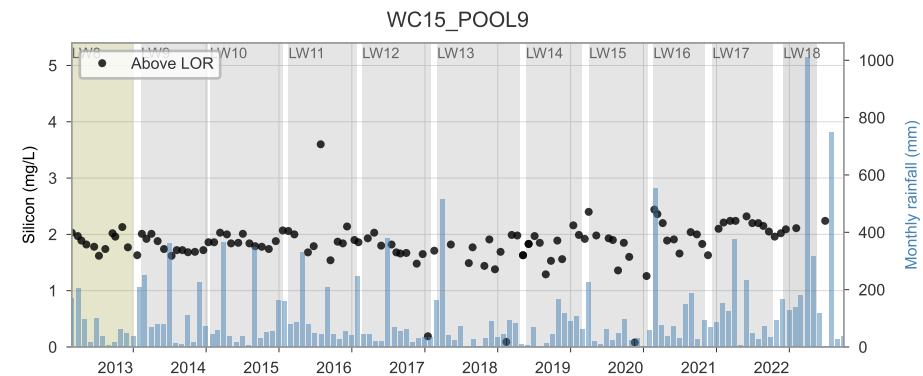
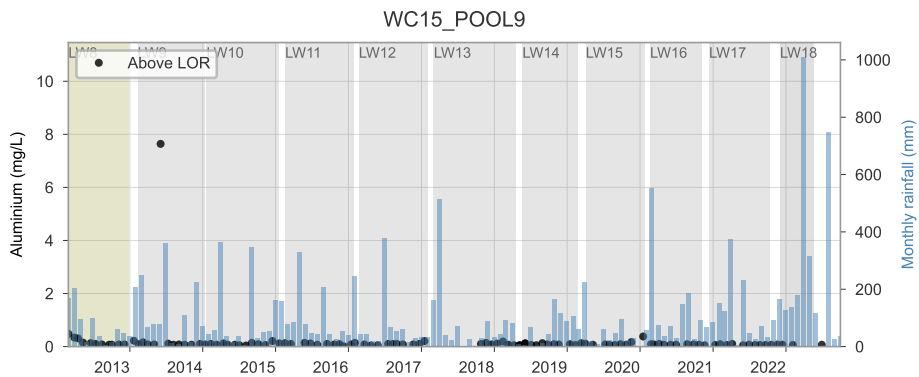
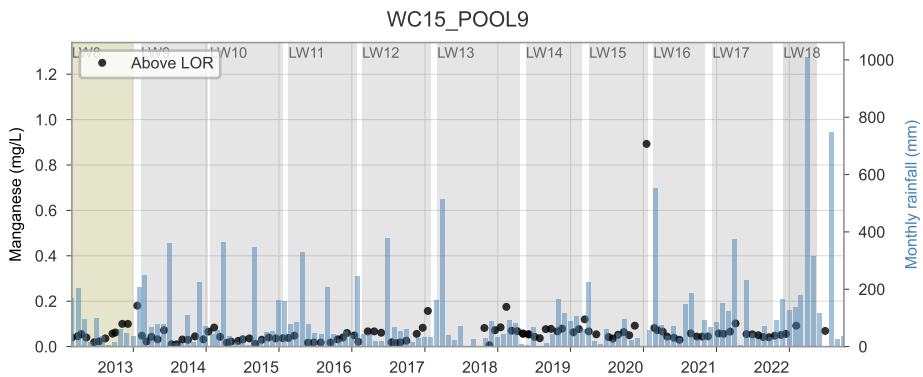
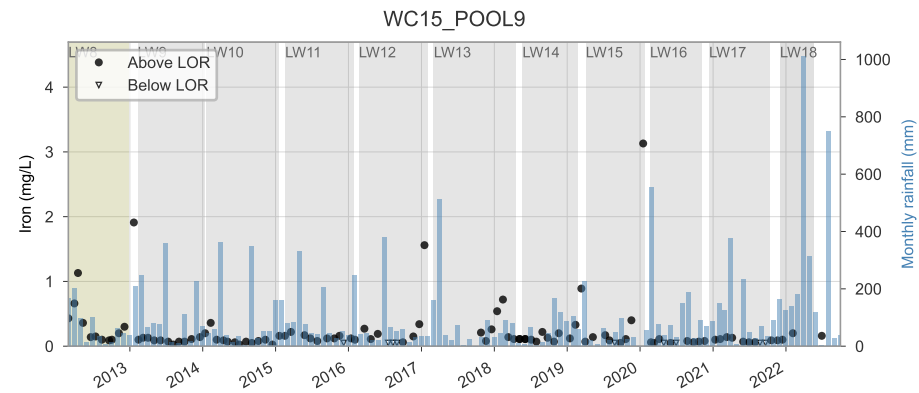
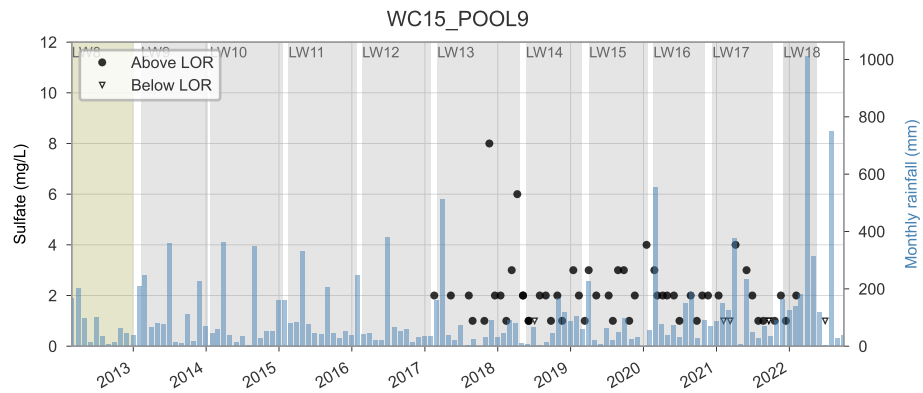


WC15_POOL2

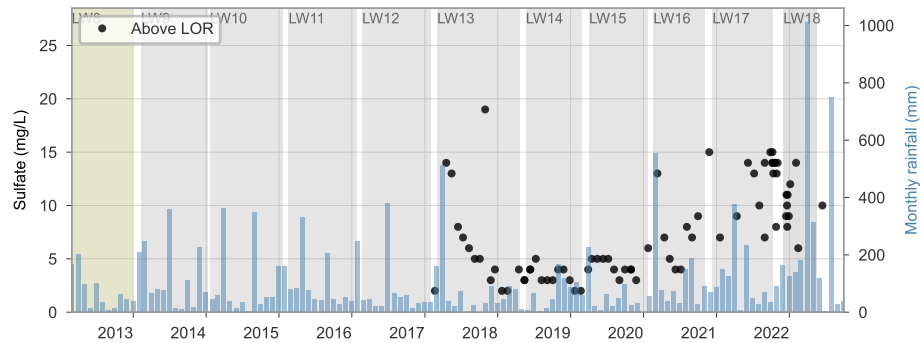


WC15_POOL2

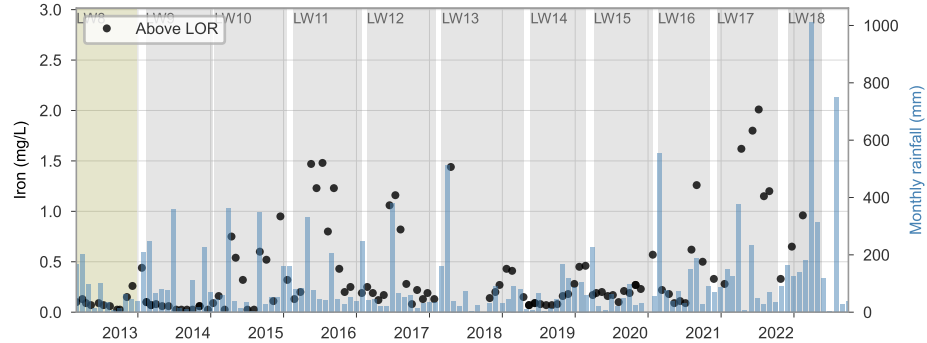




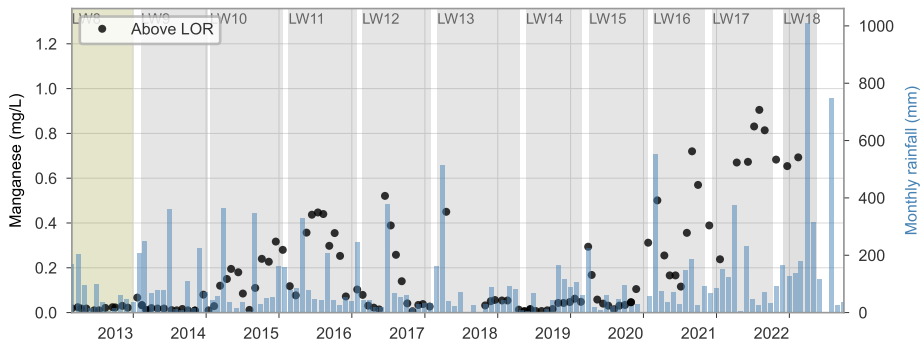
WC21_POOL5



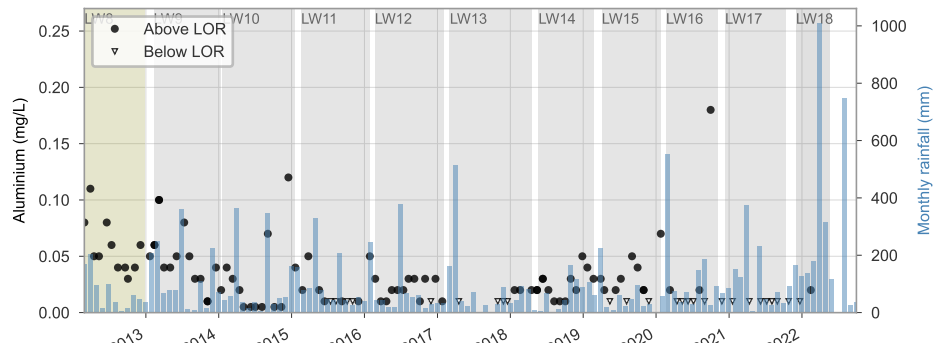
WC21_POOL5



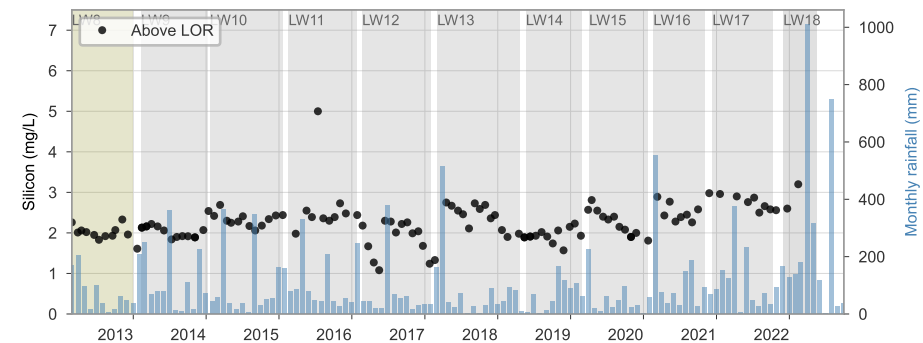
WC21_POOL5



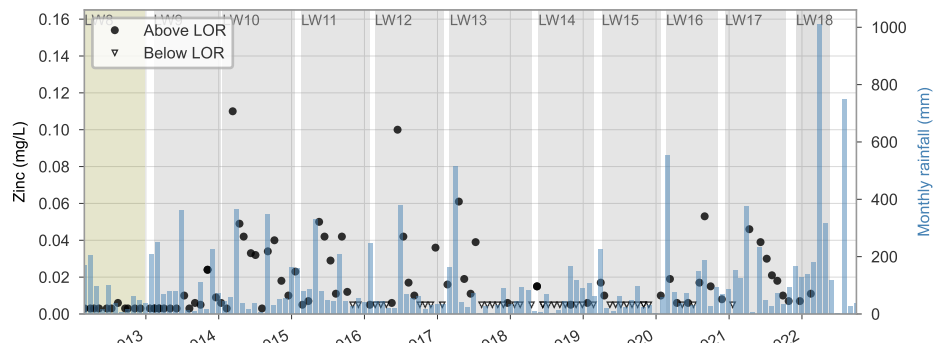
WC21_POOL5

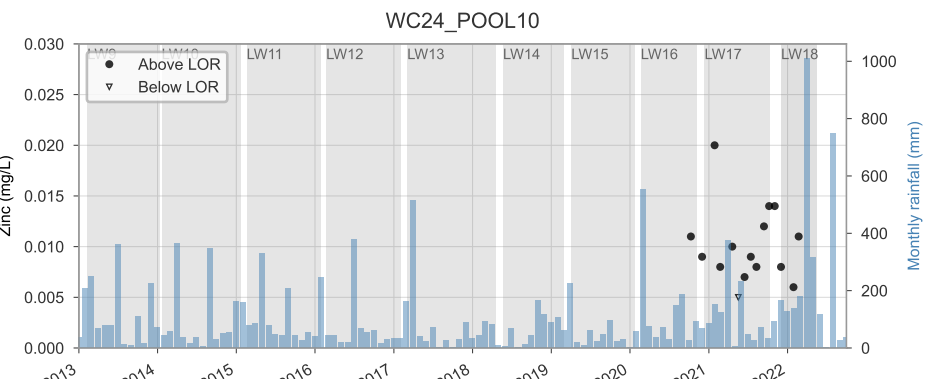
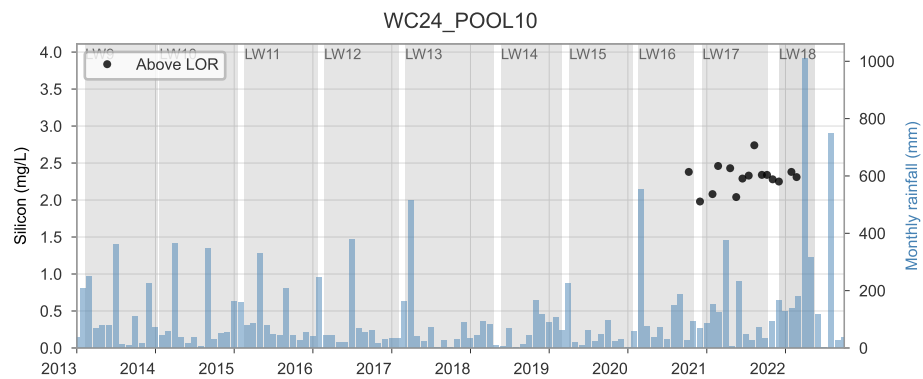
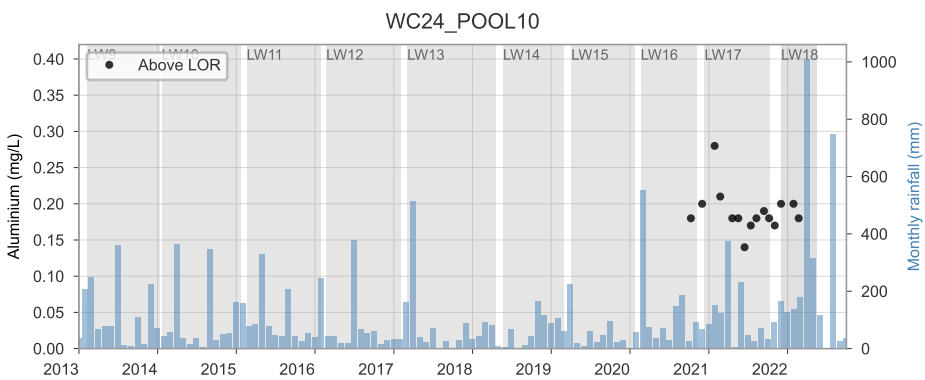
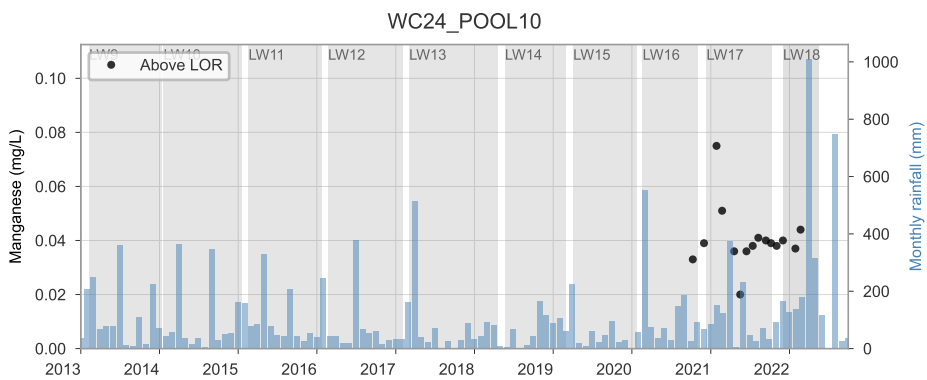
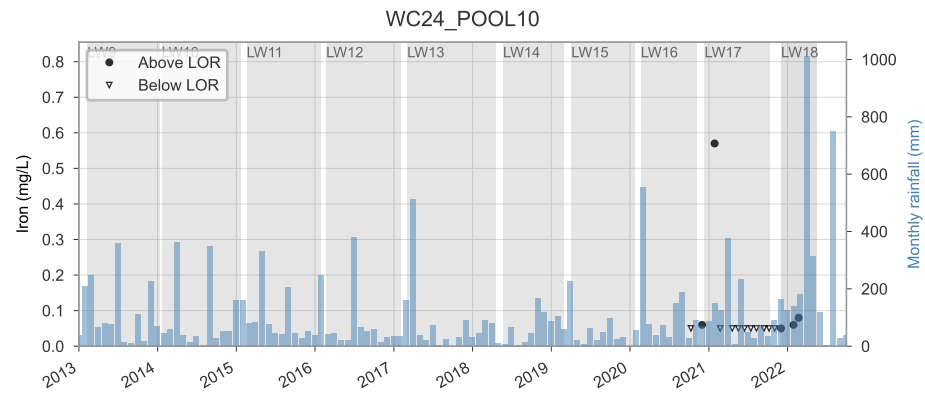
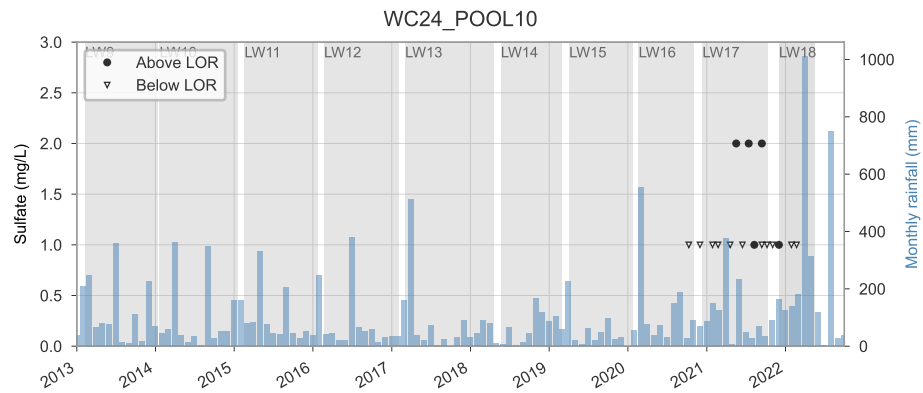


WC21_POOL5

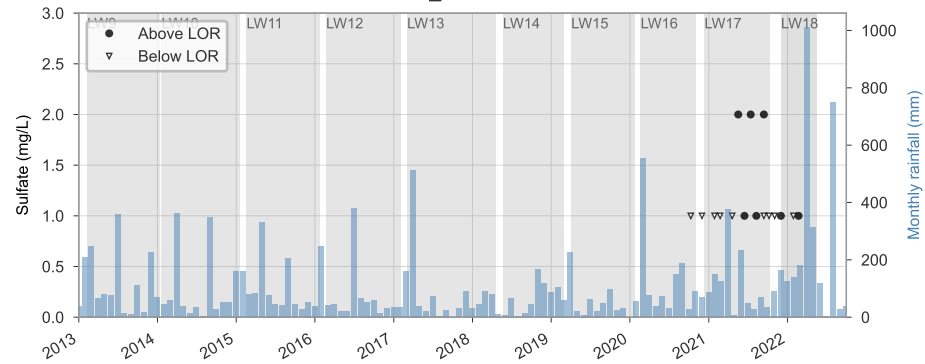


WC21_POOL5

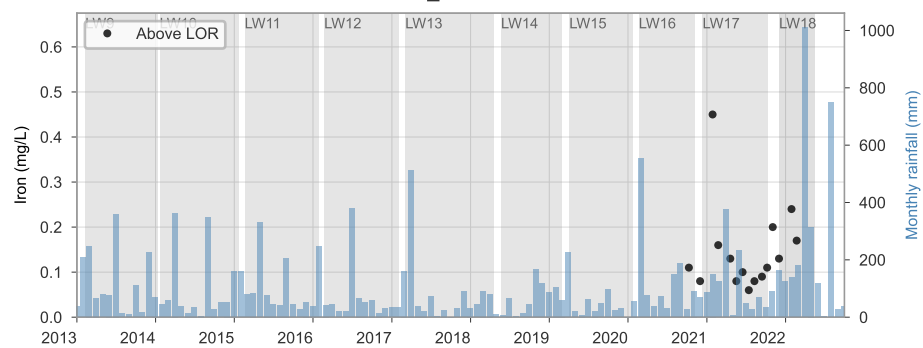




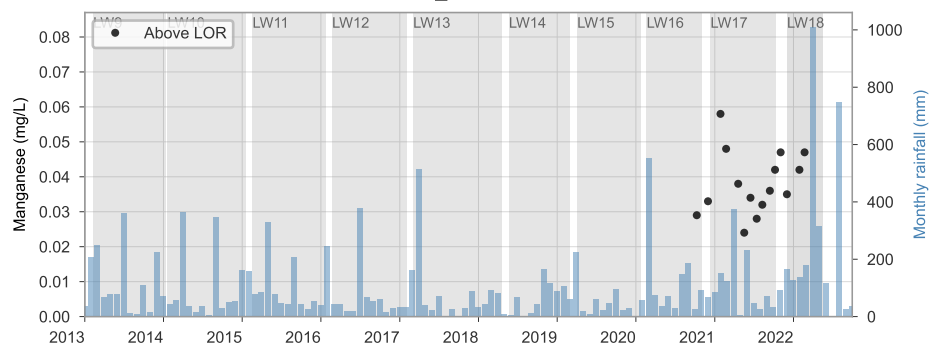
WC26_CHANNEL4



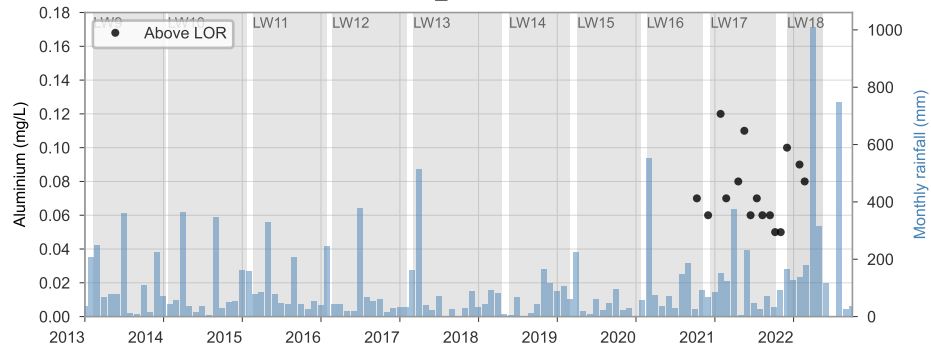
WC26_CHANNEL4



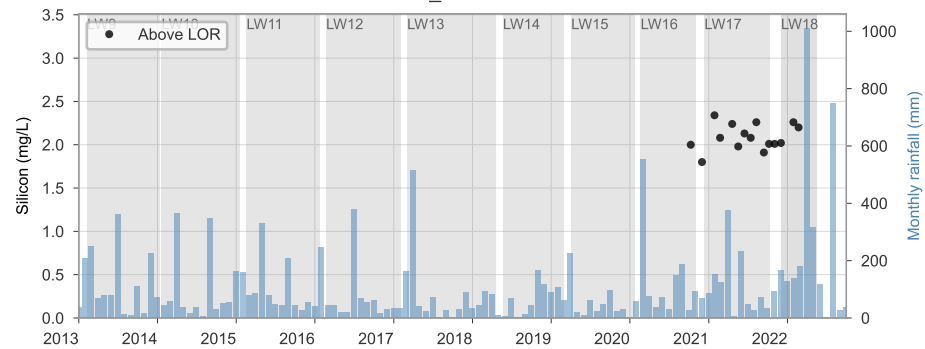
WC26_CHANNEL4



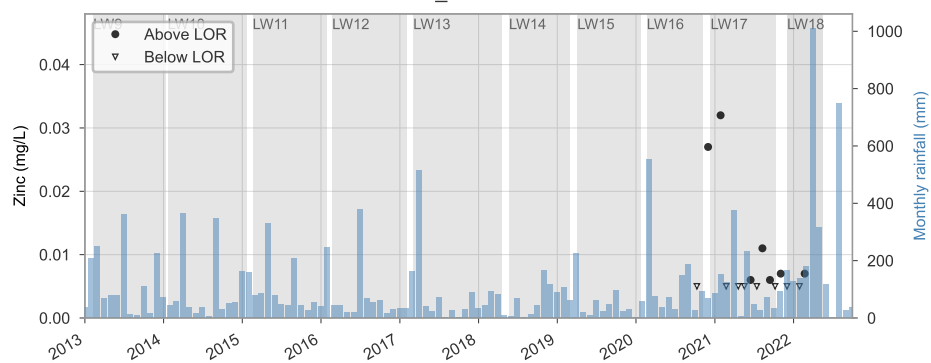
WC26_CHANNEL4



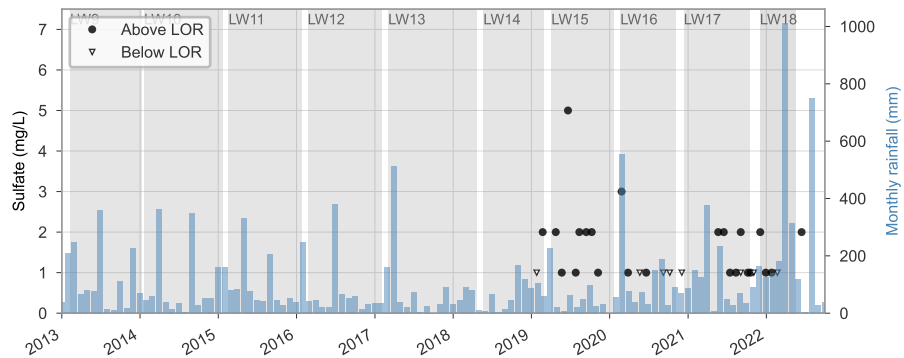
WC26_CHANNEL4



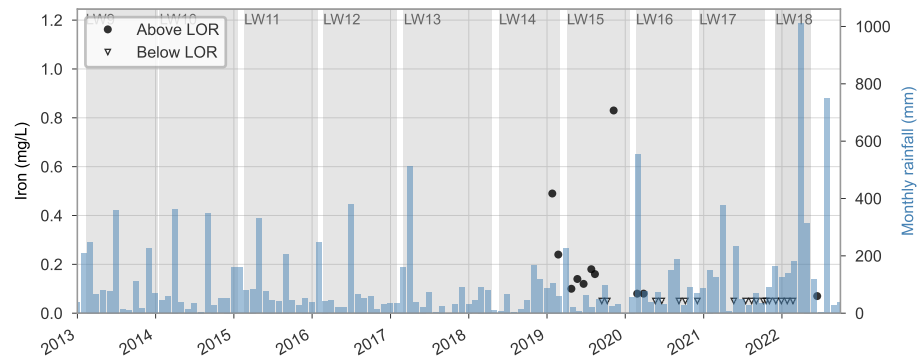
WC26_CHANNEL4



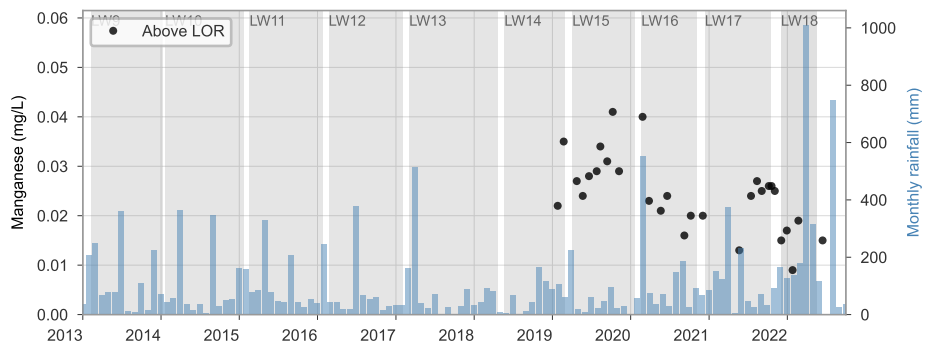
WC7_POOL1



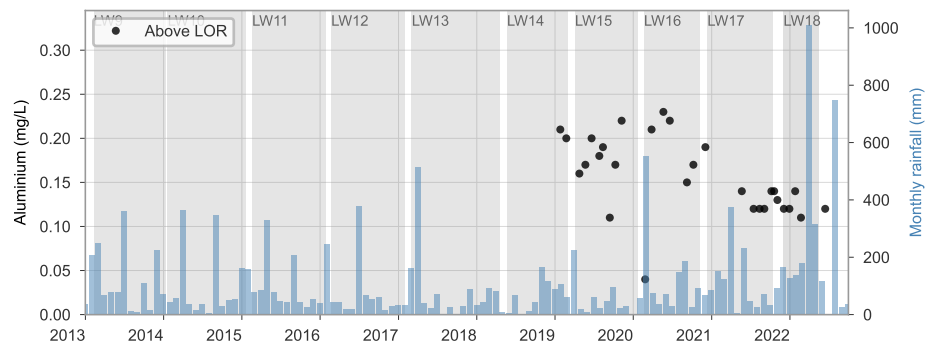
WC7_POOL1



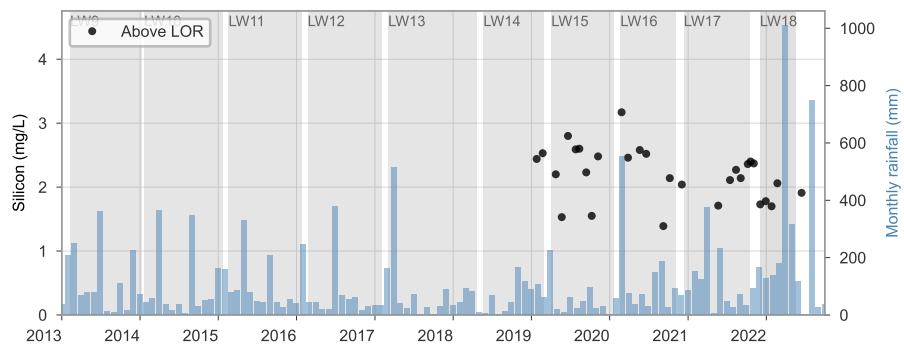
WC7_POOL1



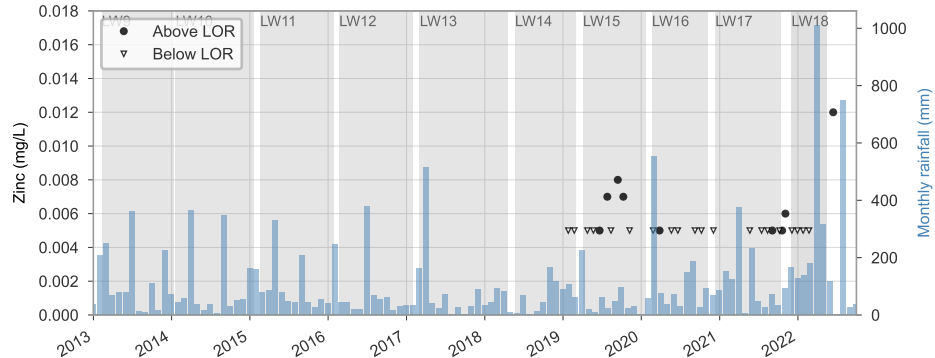
WC7_POOL1

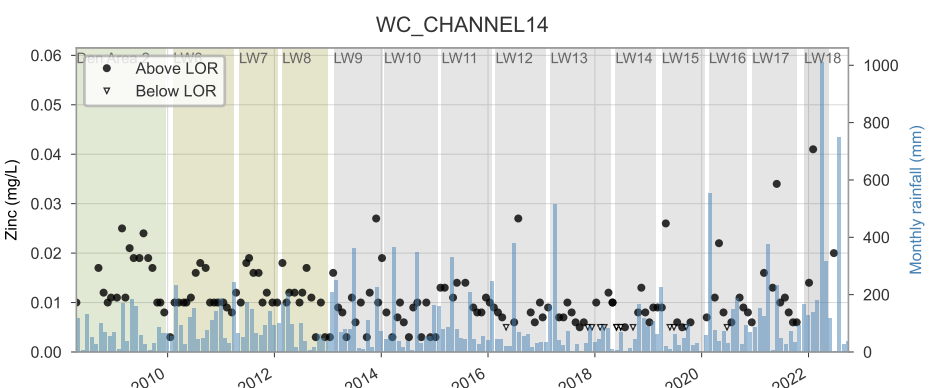
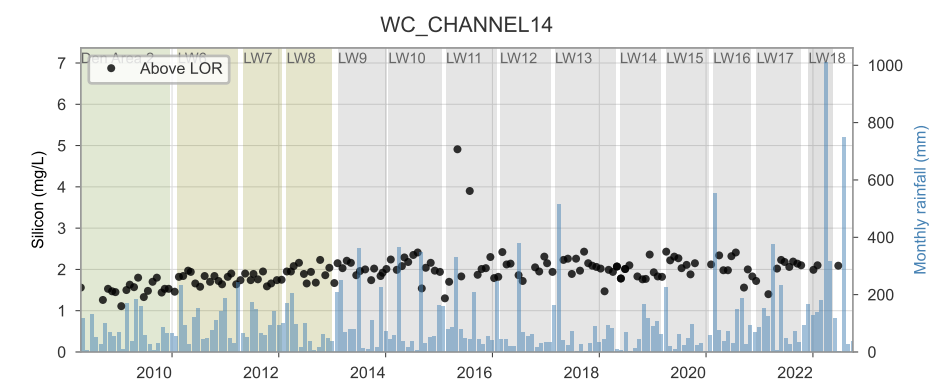
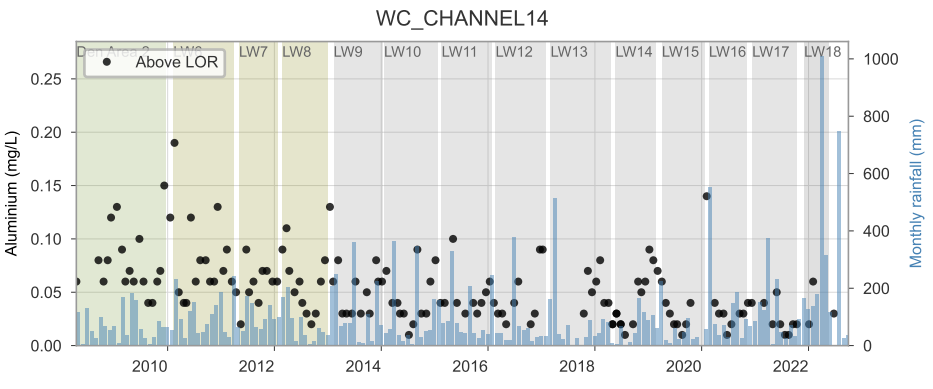
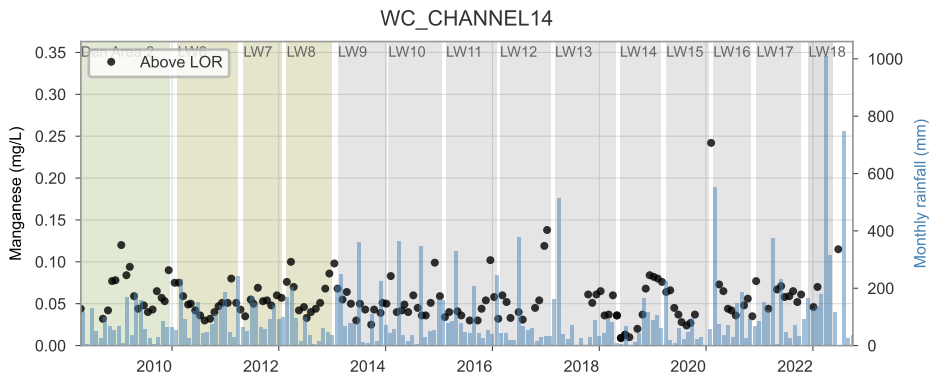
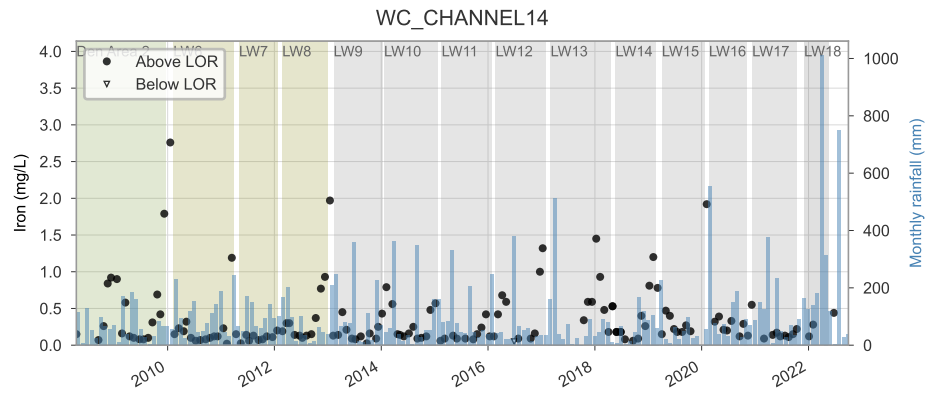
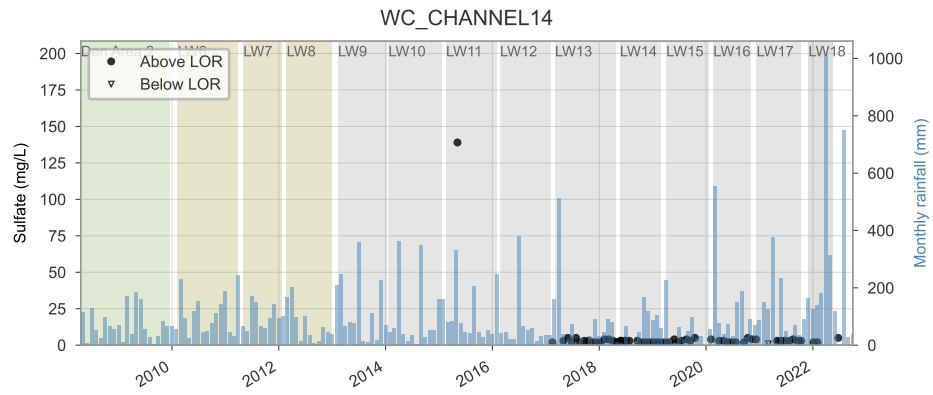


WC7_POOL1

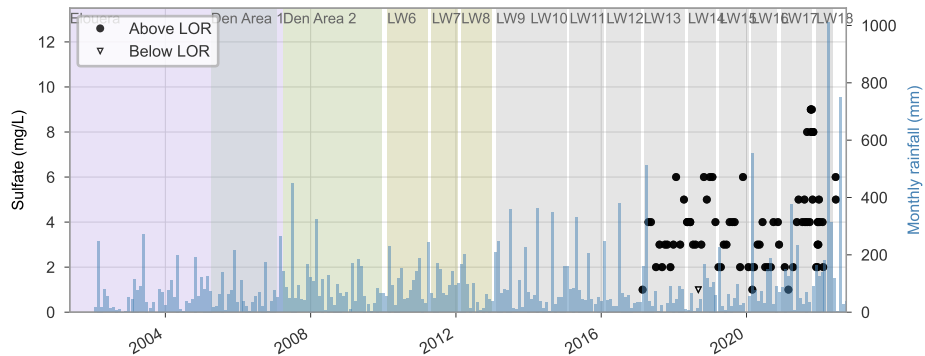


WC7_POOL1

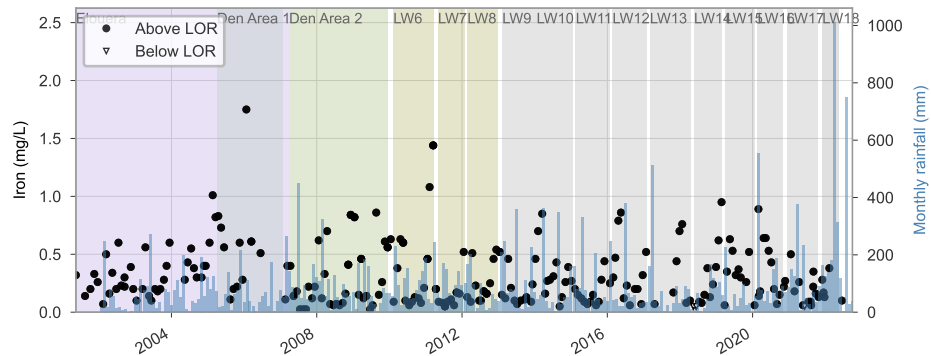




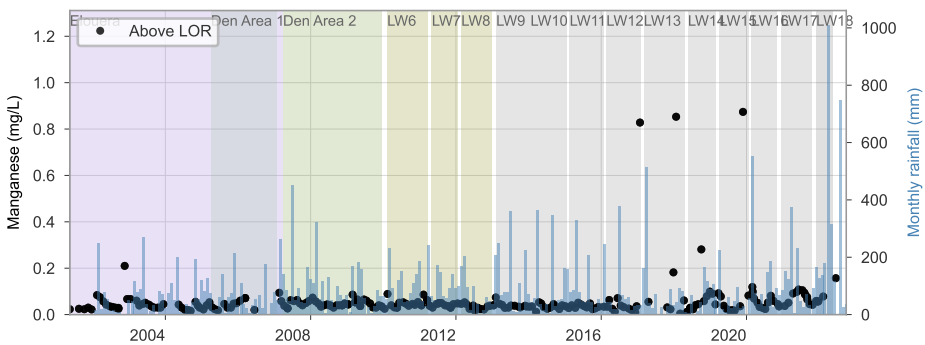
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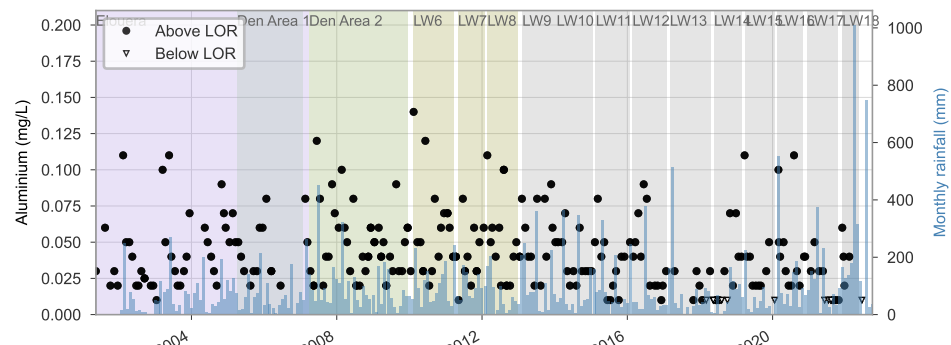
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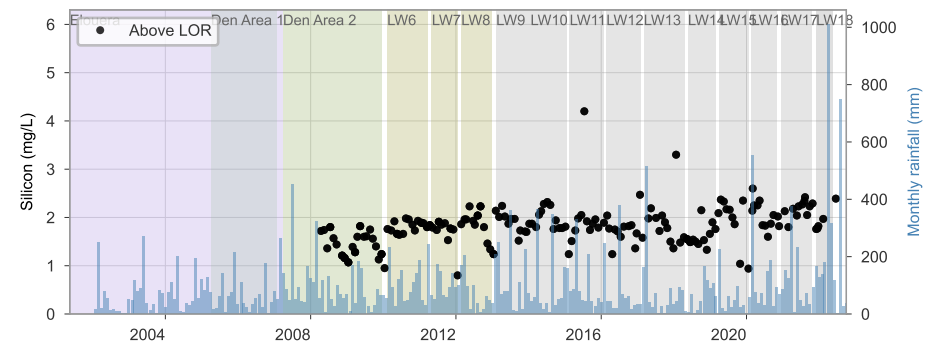
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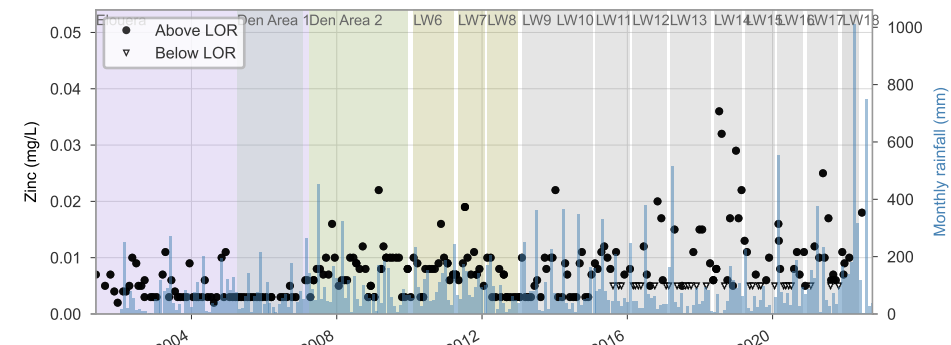
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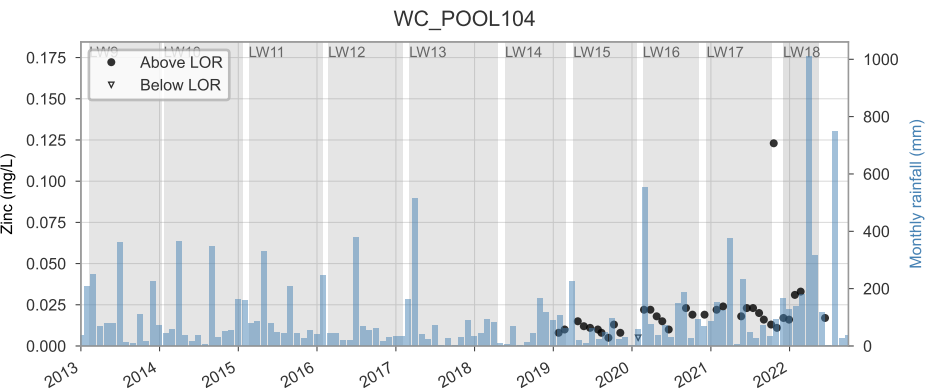
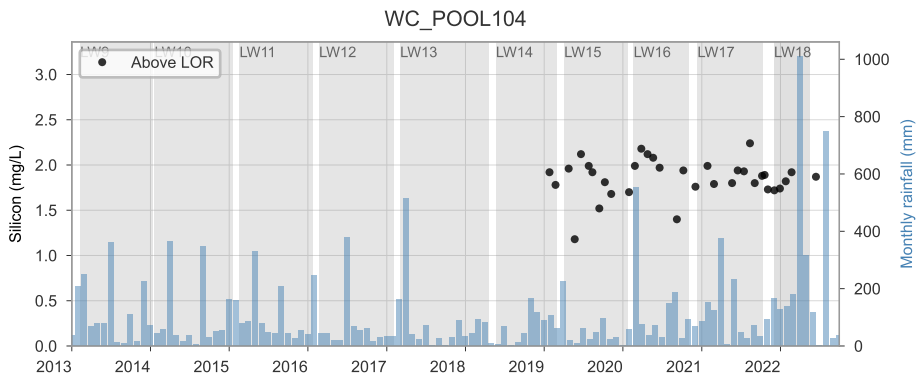
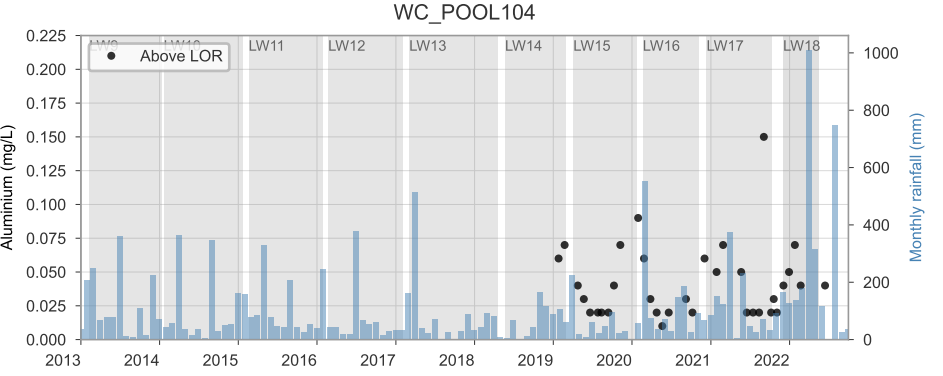
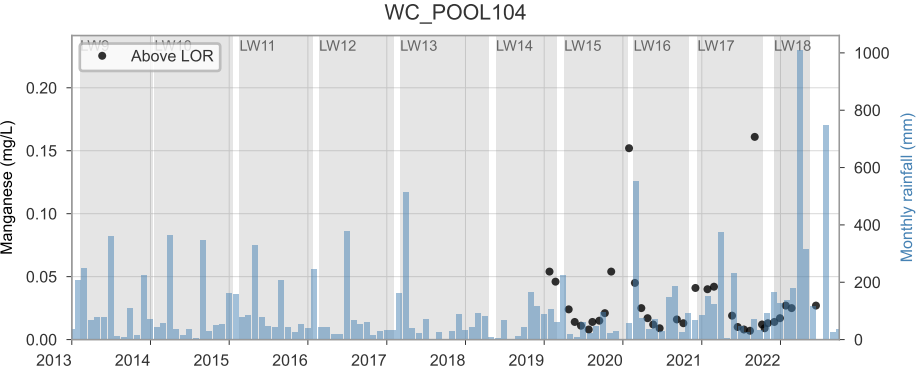
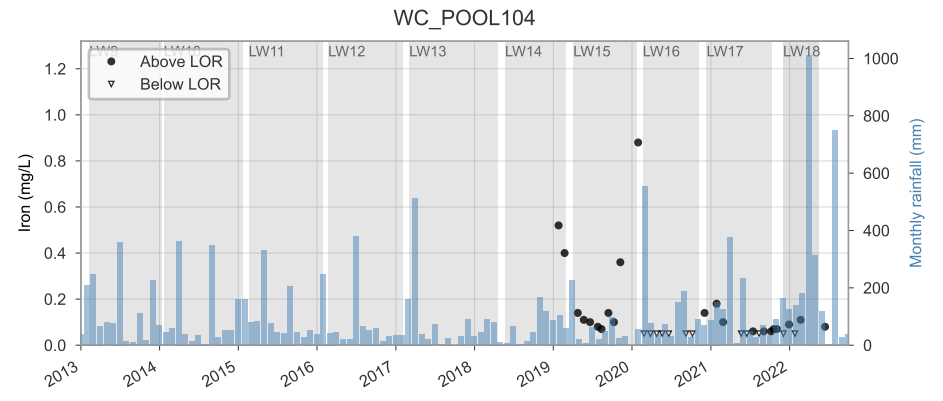
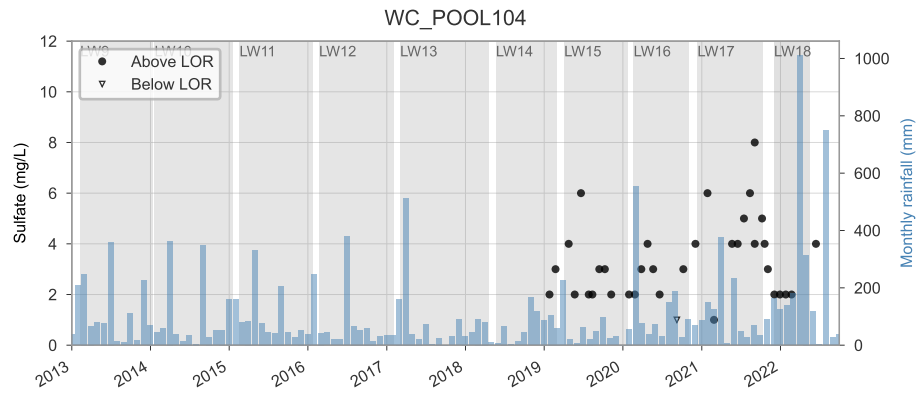


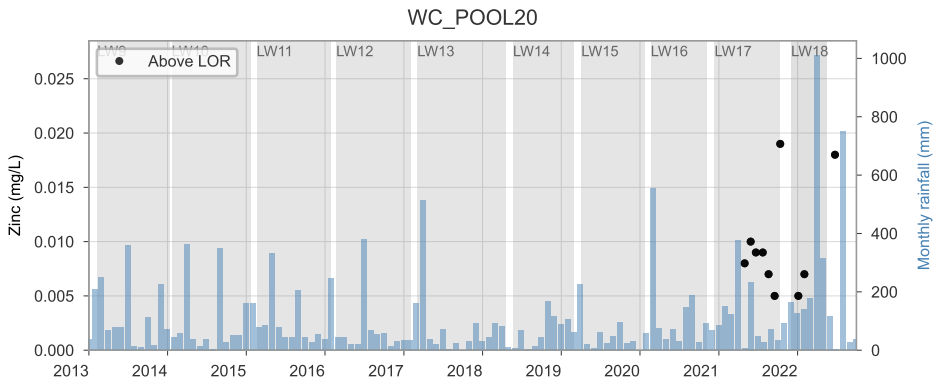
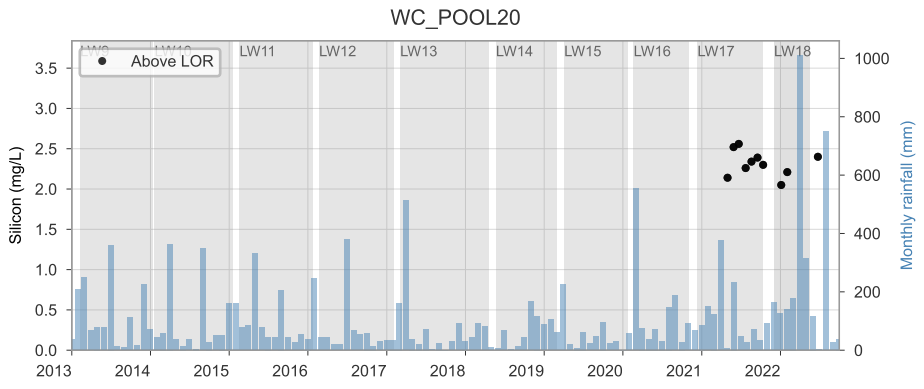
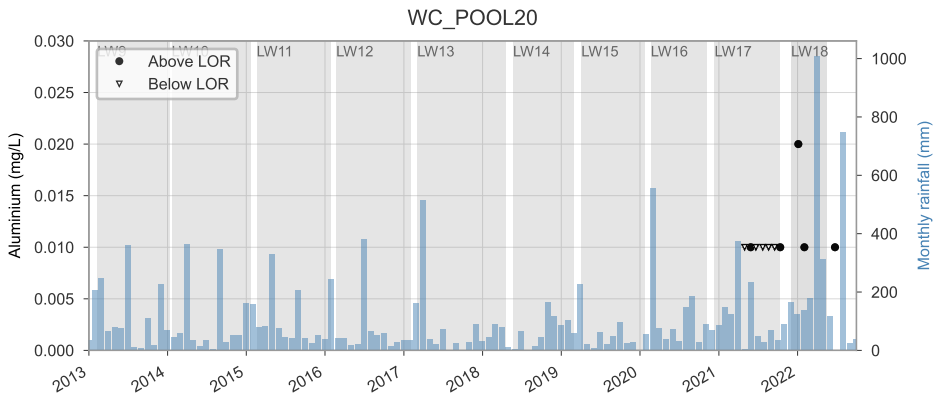
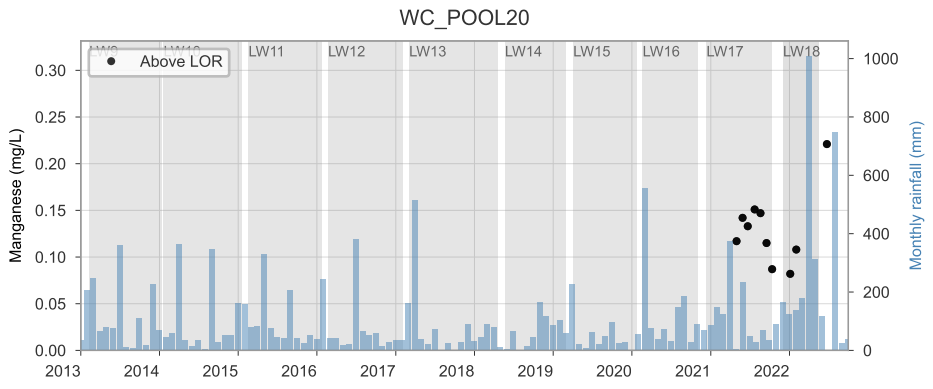
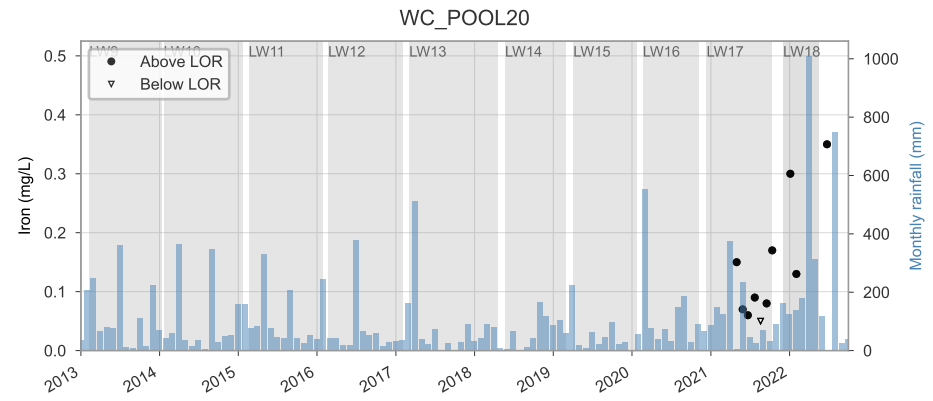
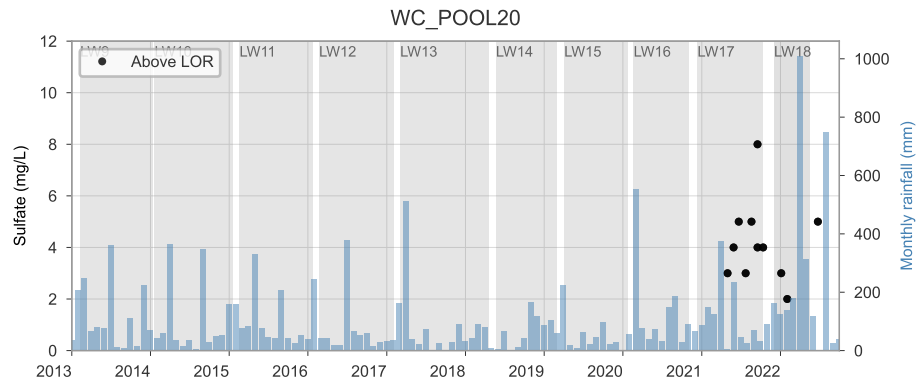
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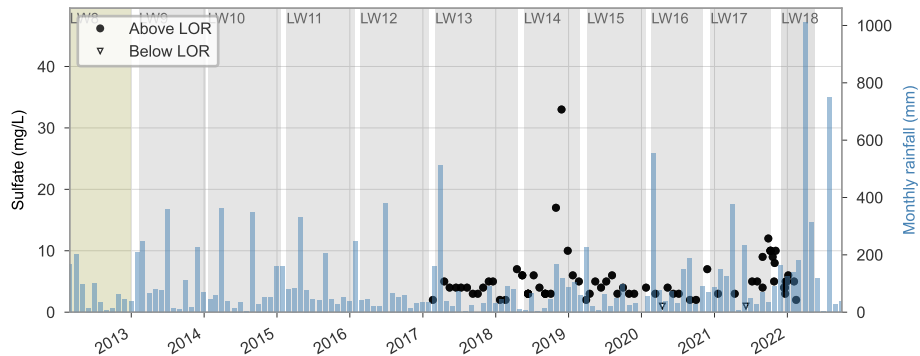
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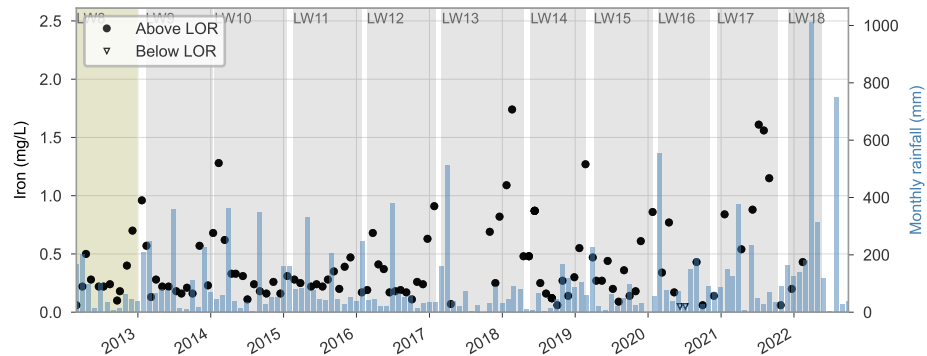




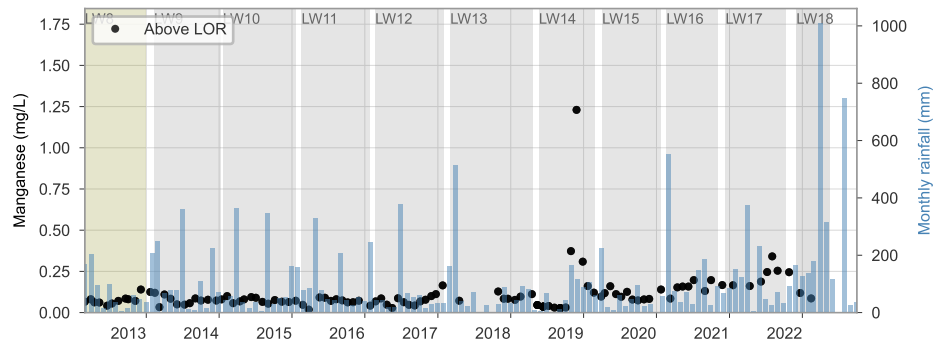
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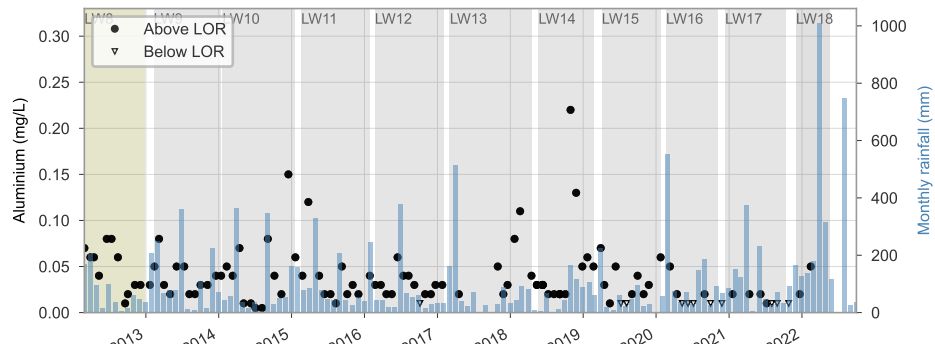
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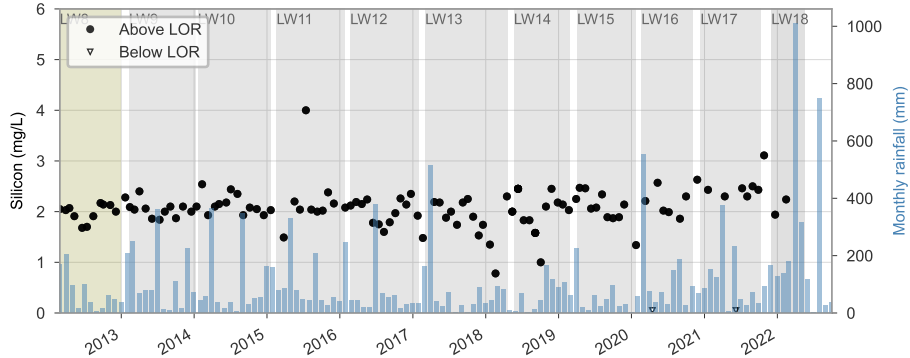
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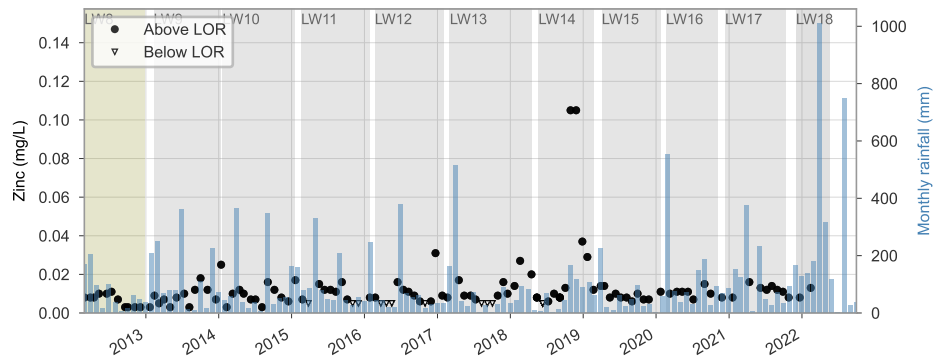
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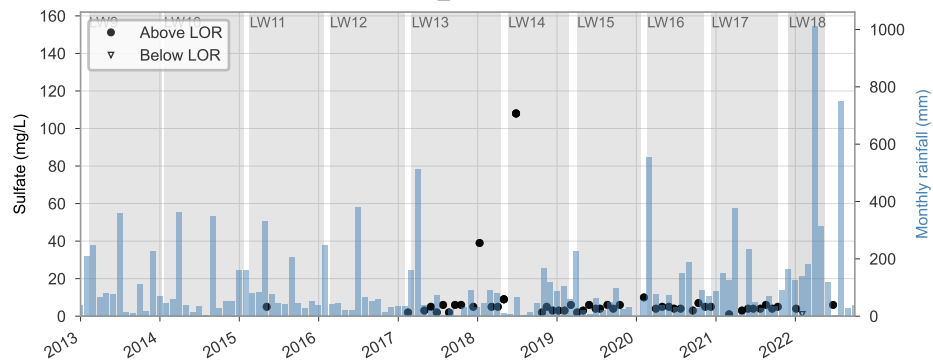
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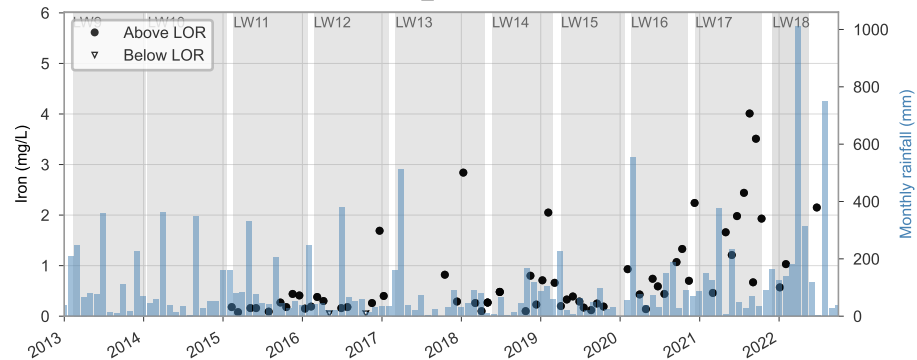
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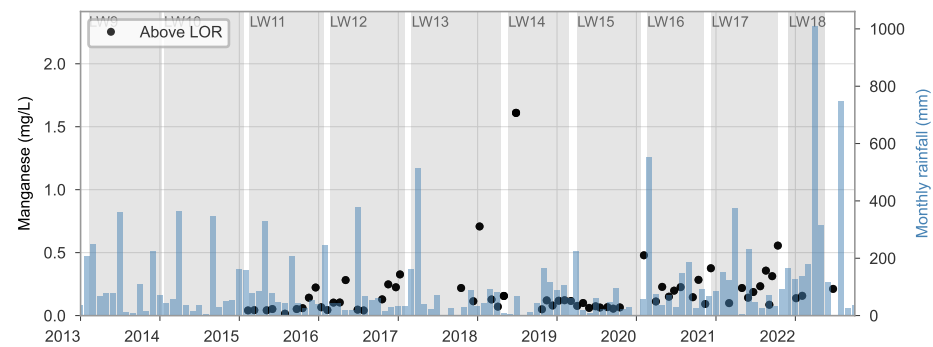
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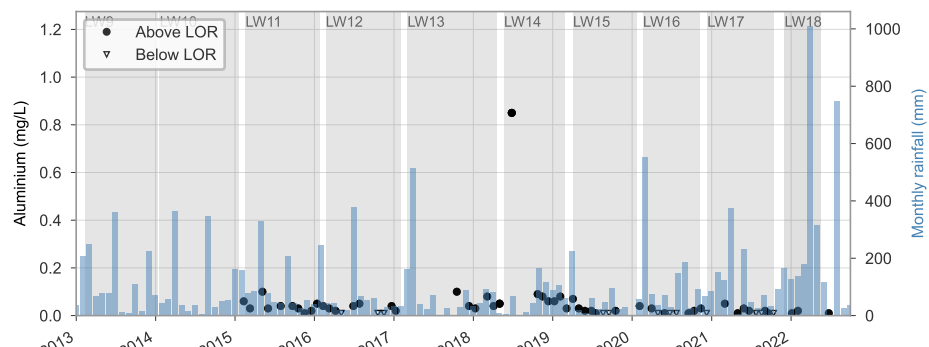
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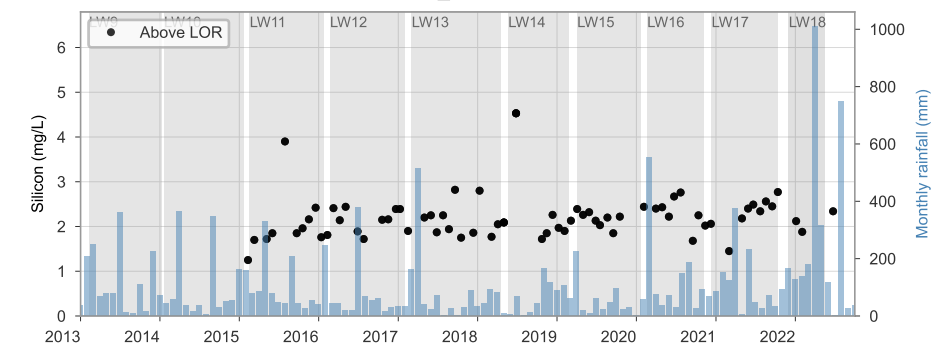
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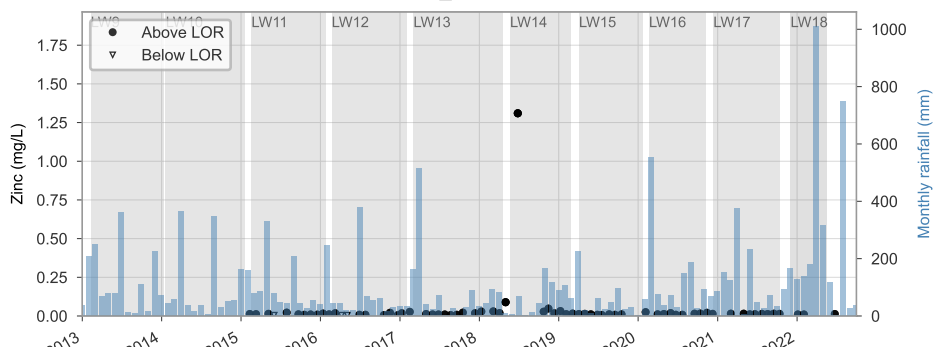
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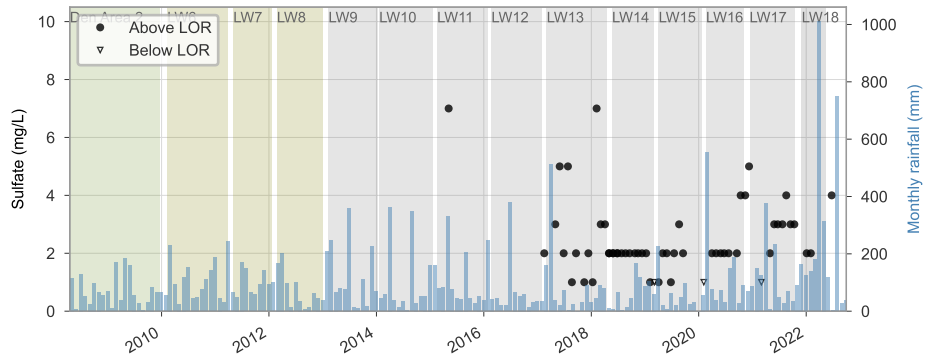
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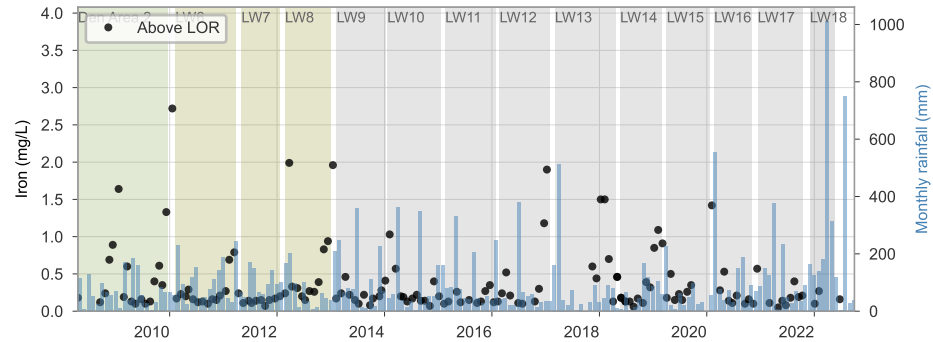
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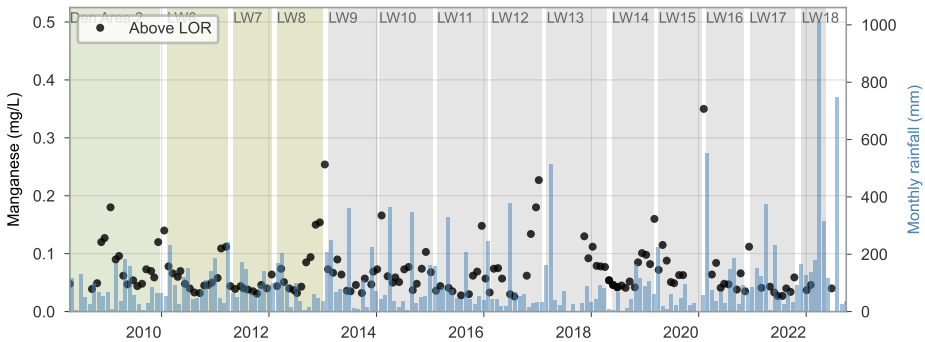
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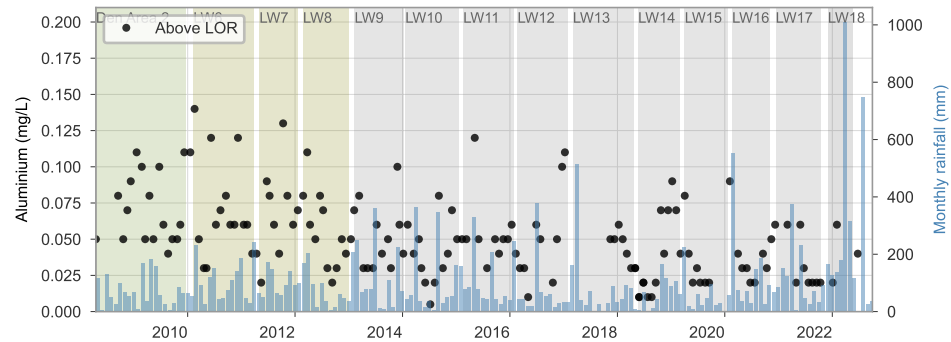
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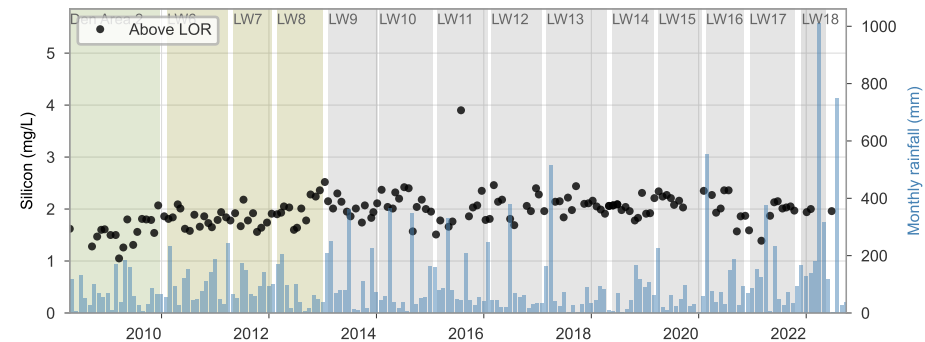
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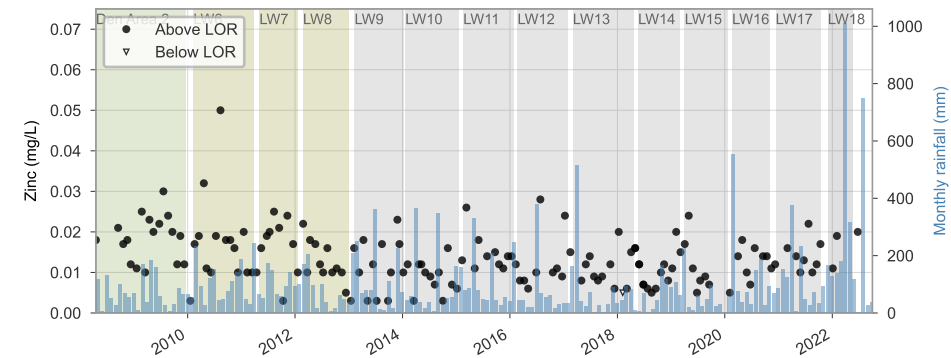
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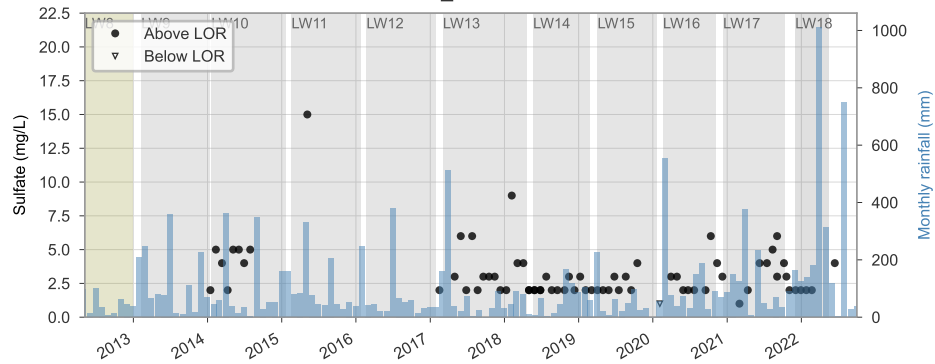
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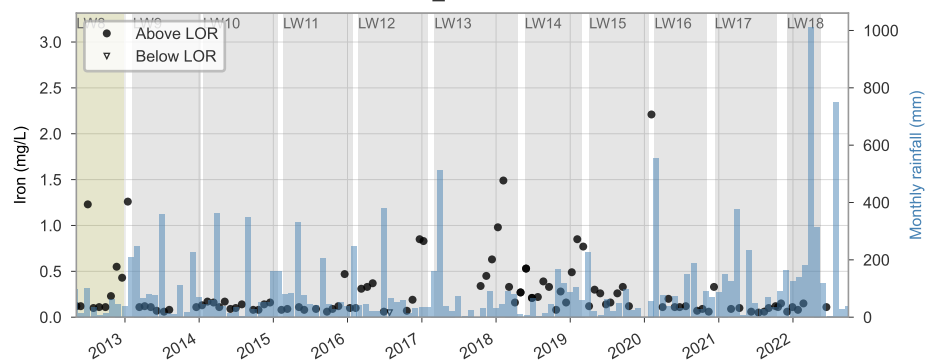
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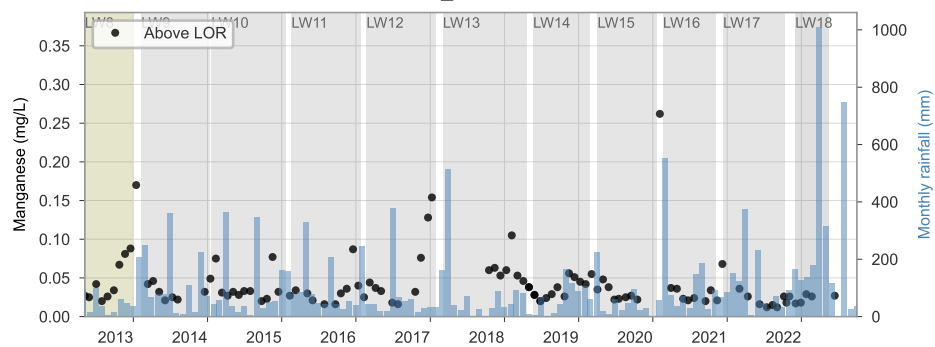
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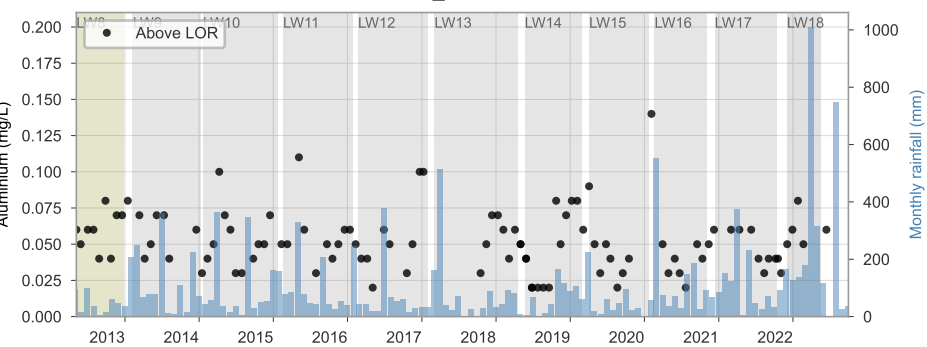
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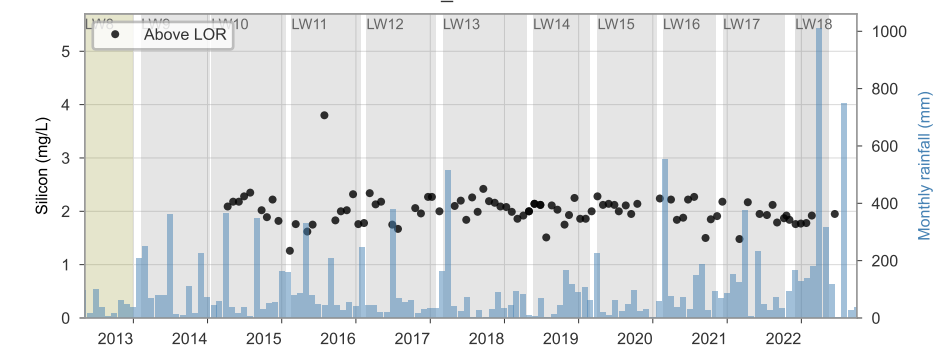
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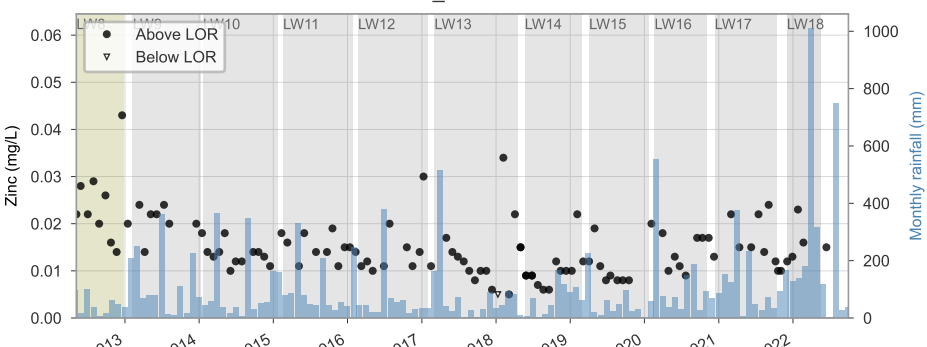
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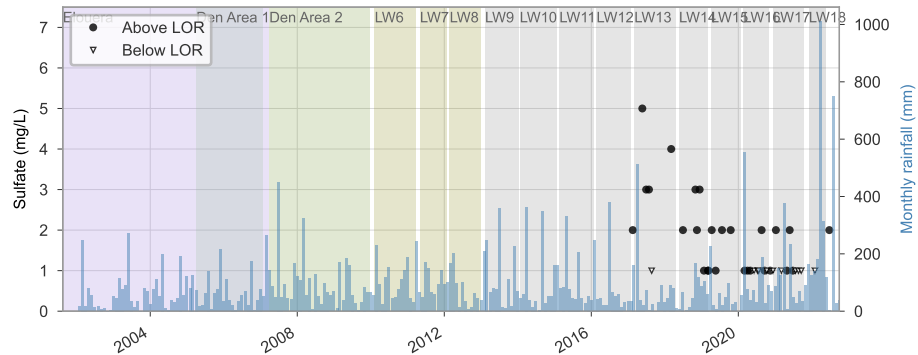
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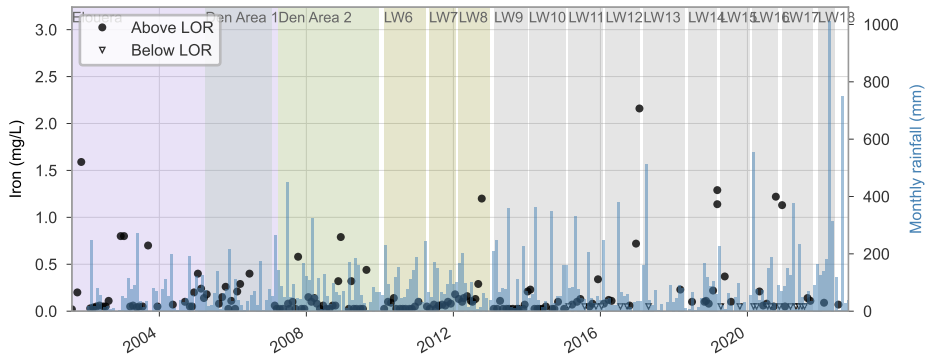
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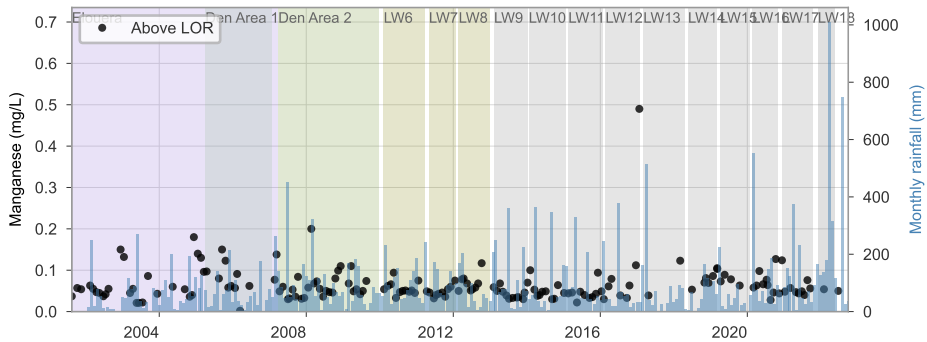
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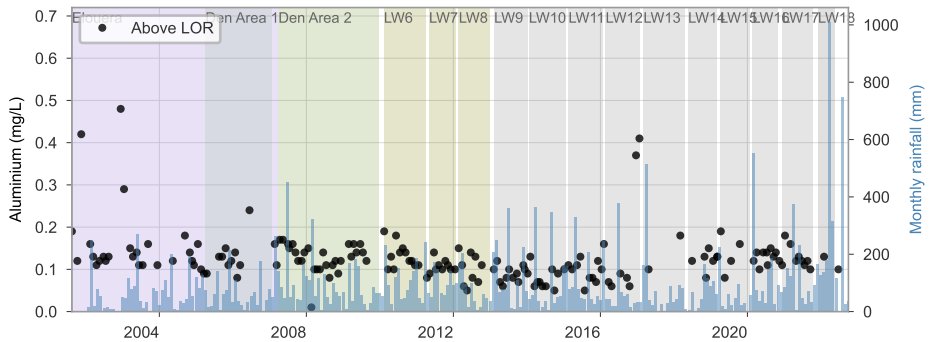
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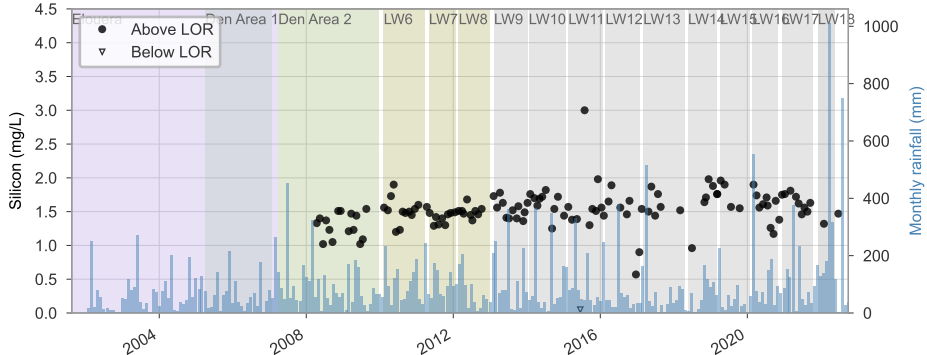
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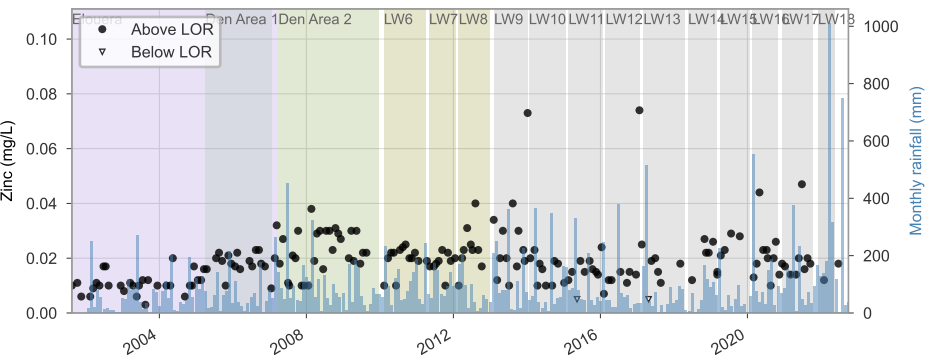
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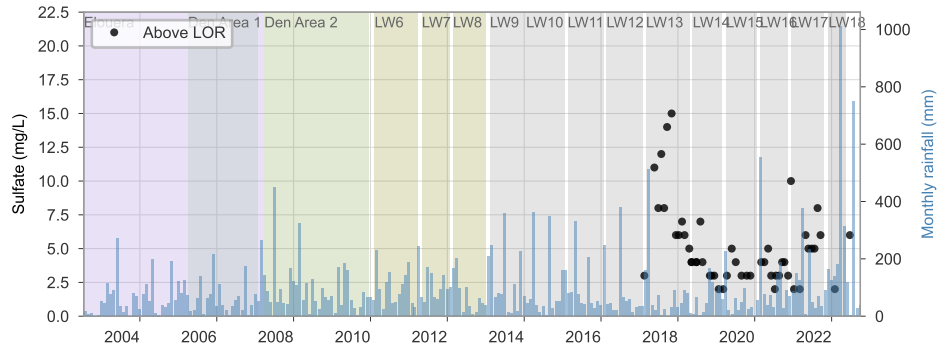
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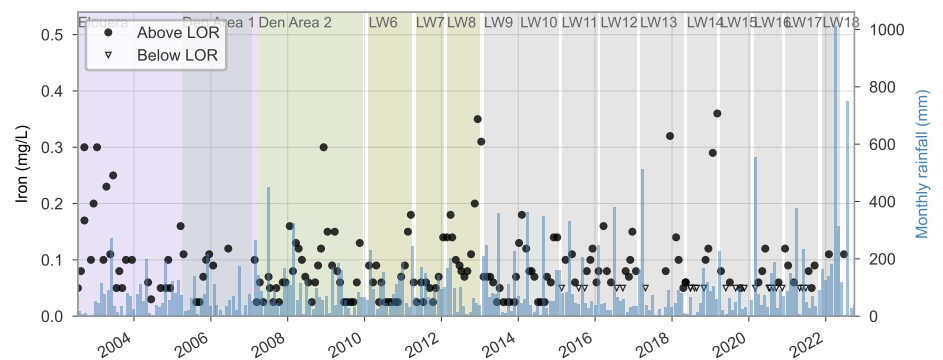
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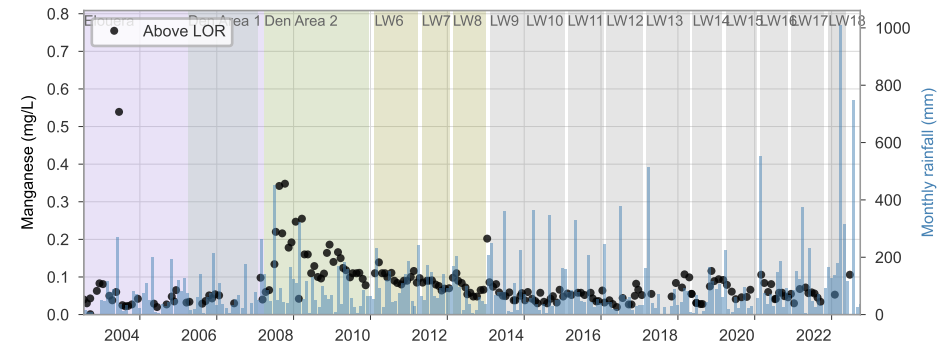
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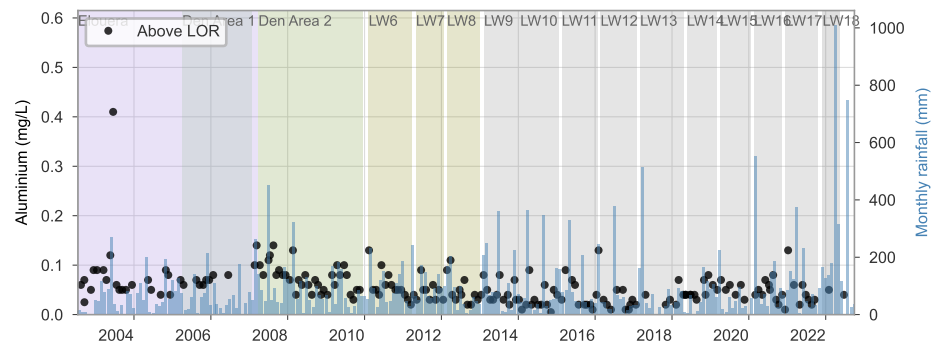
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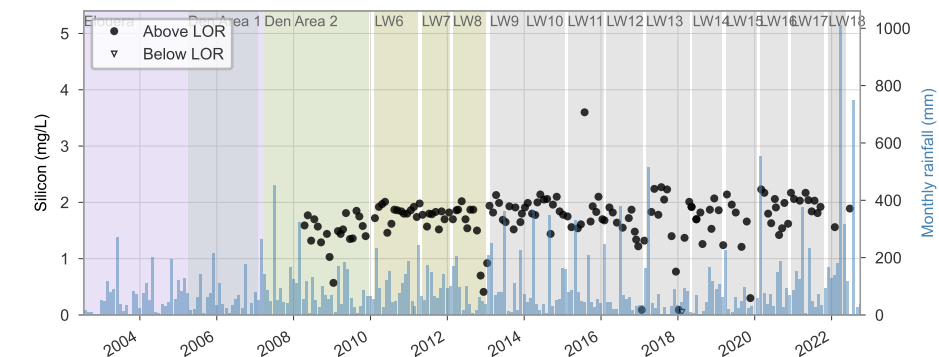
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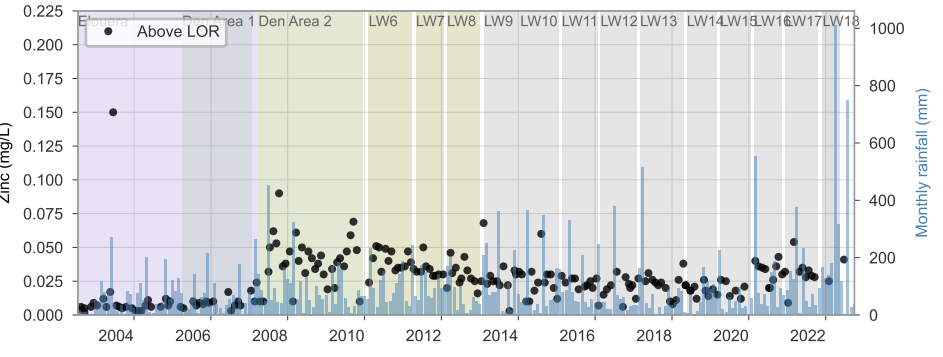
WWU4



WWU4



WWU4



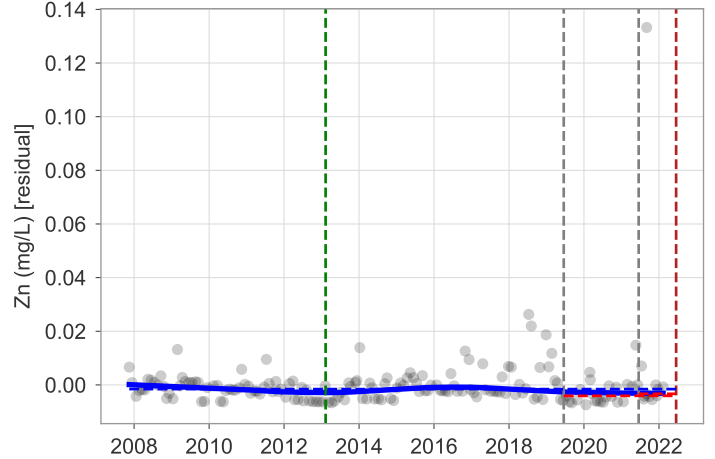
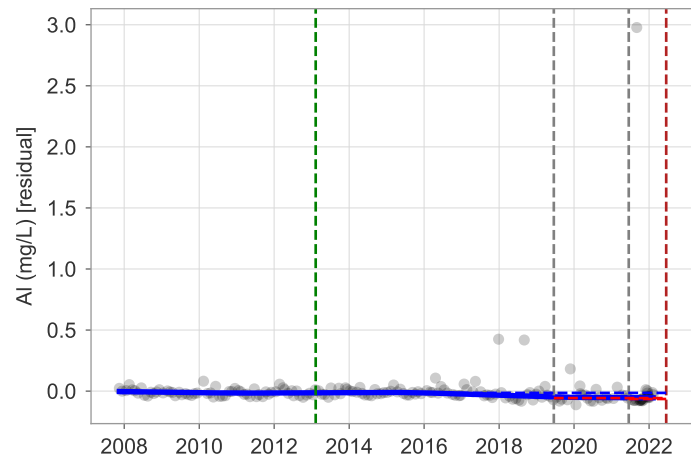
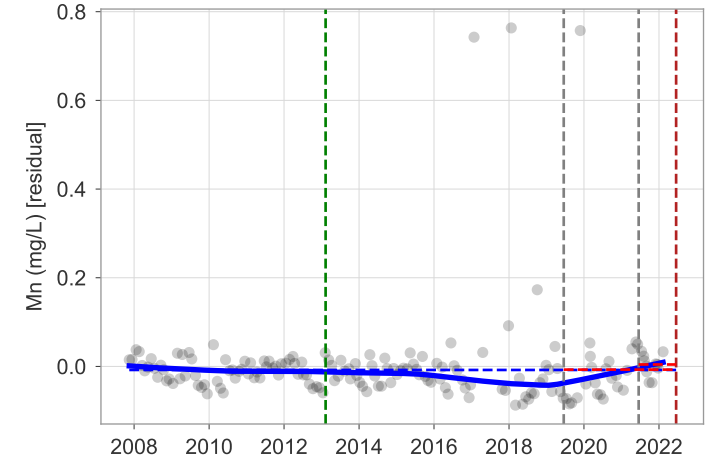
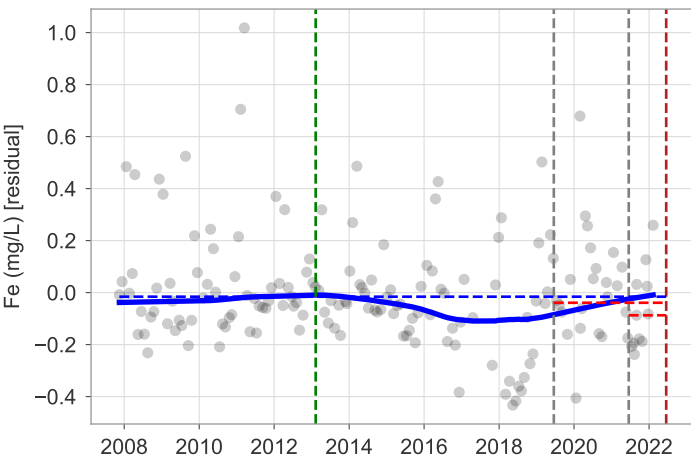
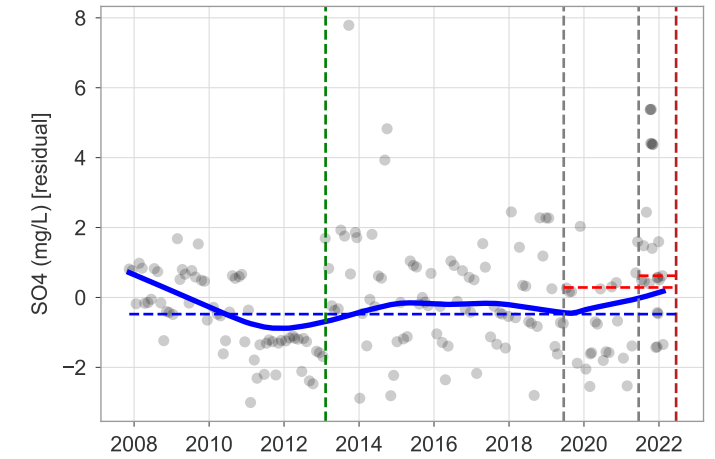
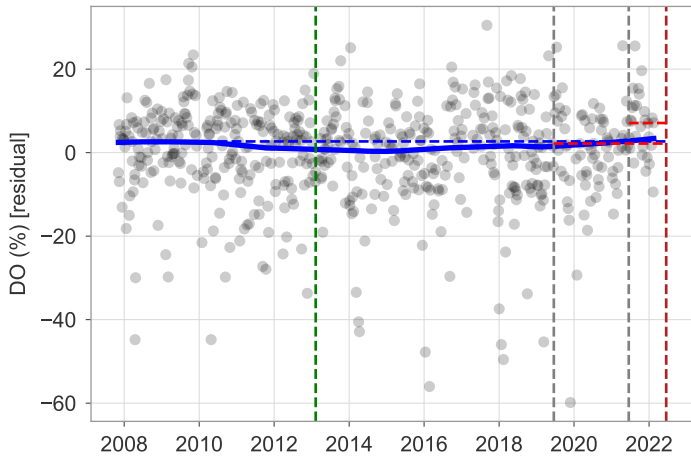
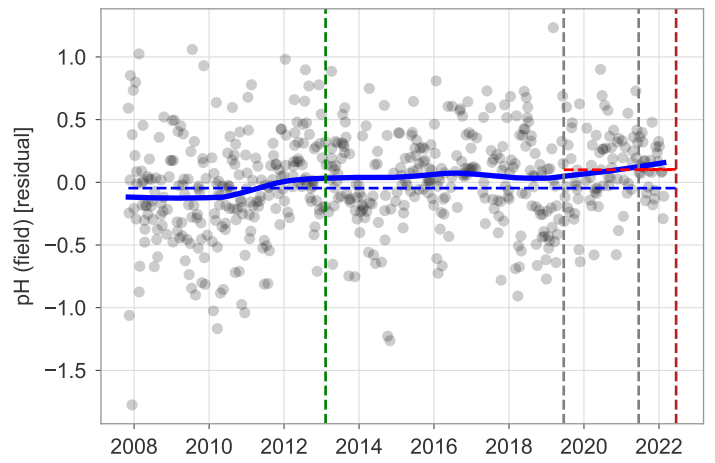
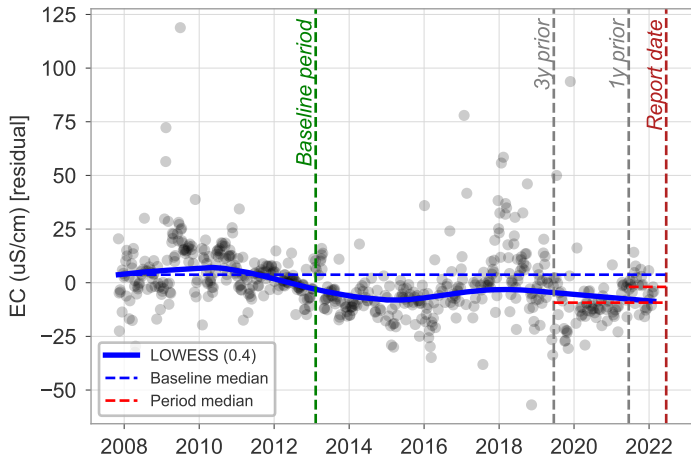
Appendix A2: Water quality trend analysis

Analysis of water quality trends in flow-corrected data

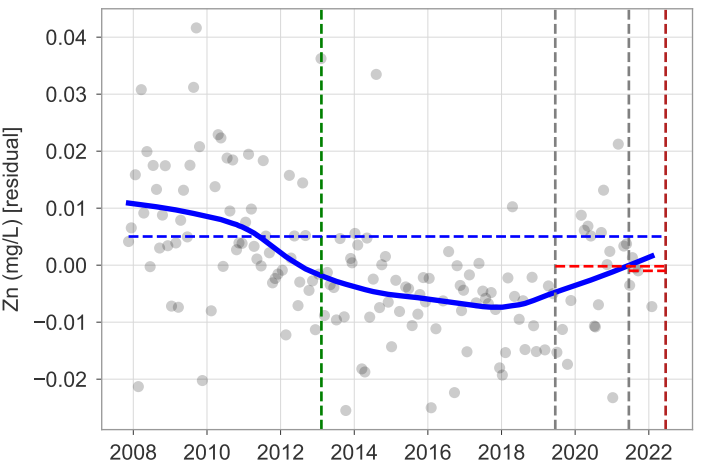
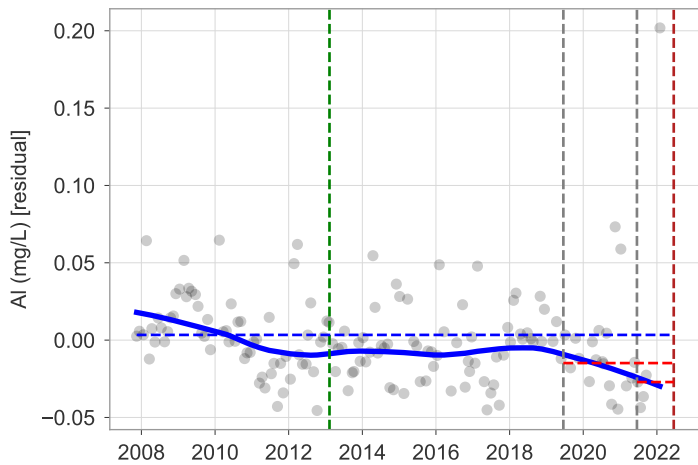
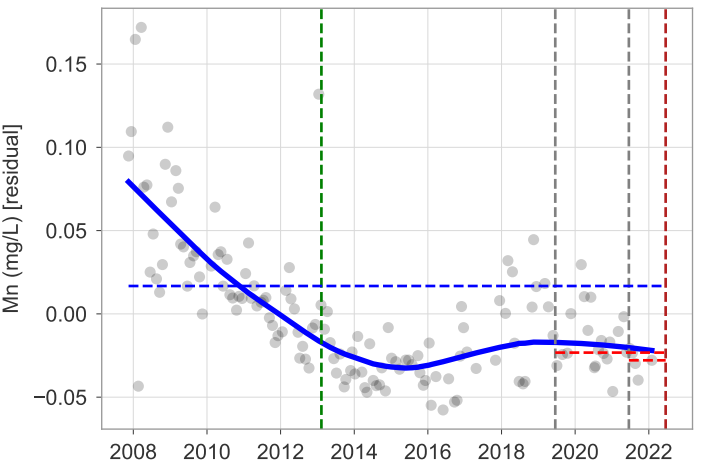
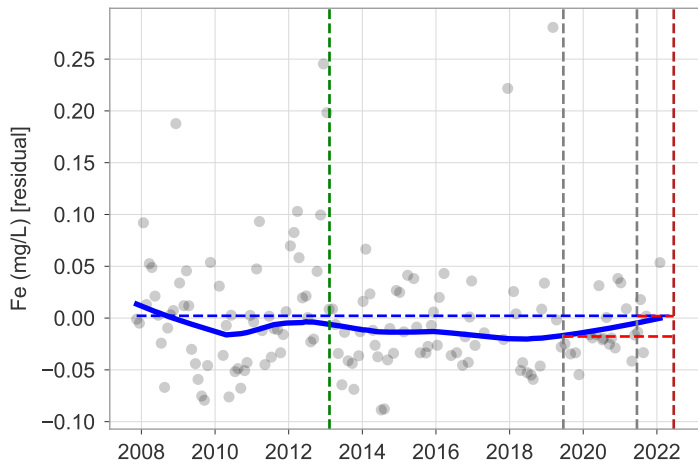
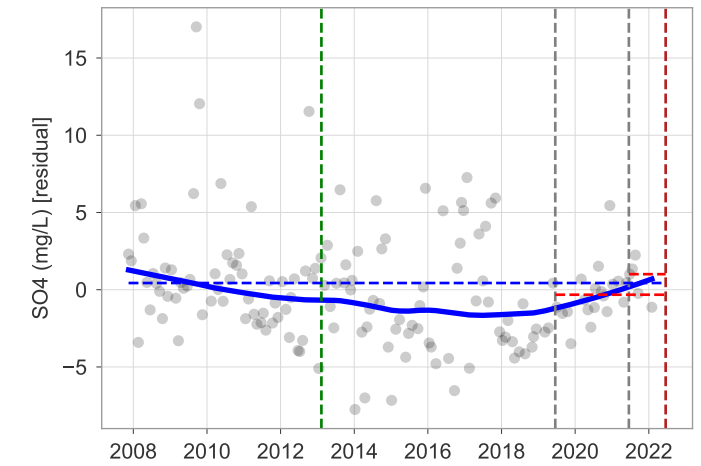
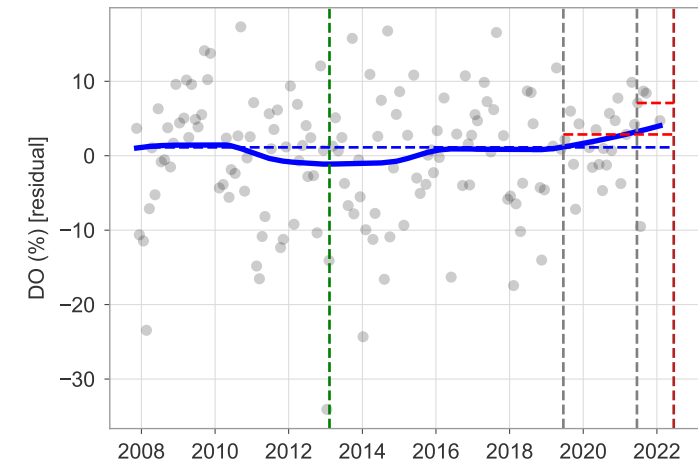
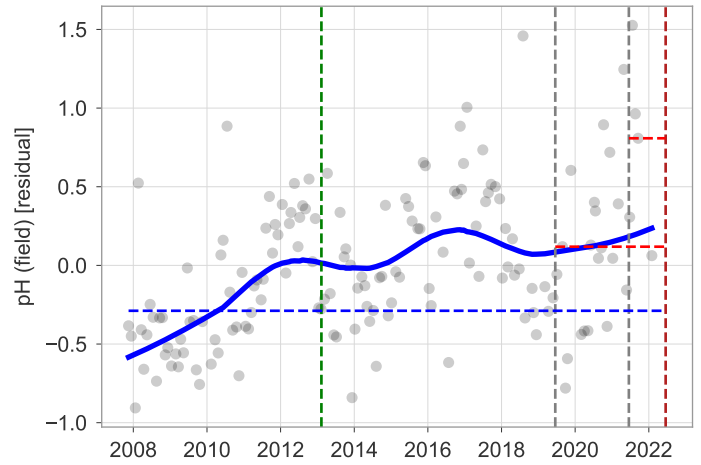
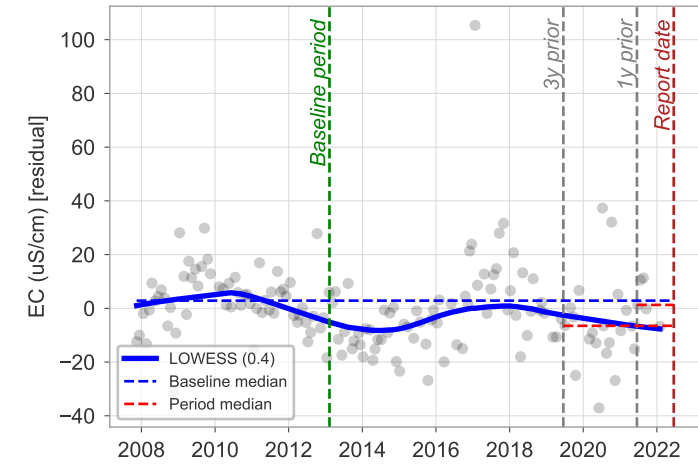
| WQ_site | Stream_gauge | Param | Theil-Sen slope | | | | Mann-Kendall serial correlation | | Raw data statistics | | | Flow-corrected statistics | | | Mann-Whitney U test | | 1-year mean test | 3-year mean test | | |
|--------------|--------------|----------|-----------------|-----------|-----------------|-----------------|---------------------------------|--------|---------------------|--------------|-----------|---------------------------|-----------|-----------|---------------------|--------------|------------------|------------------|----------------|---------|
| | | | 1-year slope | | 3-year slope | | 1-year | 3-year | 1-year_trend | 3-year_trend | median_BL | median_1y | median_3y | median_BL | median_1y | median_3y | | | 1-year | 3-years |
| | | | | | | | | | | | | | | | | | | | | |
| WC_FR6 | WWL | EC_uS/cm | -2.71E-02 | 1.17E-02 | 1.01E-01 | 1.11E-03 | | | Increasing | 96.35 | 96 | 97 | 3.741 | -1.937 | -9.279 | 0.000 | 0.000 | | | |
| WC_FR6 | WWL | pH_field | -5.10E-04 | 1.40E-04 | 3.77E-01 | 1.95E-01 | | | | 5.9 | 6.105 | 6.14 | -0.046 | 0.104 | 0.100 | 0.005 | 0.000 | | | |
| WC_FR6 | WWL | DO_% | -2.99E-02 | 5.02E-03 | 1.29E-01 | 1.13E-01 | | | | 93.4 | 97.7 | 91.4 | 2.687 | 7.101 | 2.174 | 0.005 | 0.357 | | | |
| WC_FR6 | WWL | SO4_mg/L | -1.28E-02 | 2.25E-03 | 1.27E-01 | 1.72E-03 | | | Increasing | 3 | 4 | 4 | -0.477 | 0.622 | 0.286 | 0.000 | 0.111 | Higher than BL | | |
| WC_FR6 | WWL | Fe_mg/L | 1.52E-03 | -6.56E-05 | 1.65E-02 | 3.21E-01 | | | Increasing | 0.16 | 0.16 | 0.22 | -0.016 | -0.087 | -0.039 | 0.078 | 0.328 | | | |
| WC_FR6 | WWL | Mn_mg/L | -1.43E-04 | 6.50E-05 | 2.83E-01 | 2.93E-02 | | | Increasing | 0.042 | 0.065 | 0.058 | -0.008 | 0.004 | -0.007 | 0.191 | 0.708 | | | |
| WC_FR6 | WWL | Zn_mg/L | 9.62E-06 | 3.40E-06 | 7.61E-01 | 4.86E-02 | | | Increasing | 0.008 | 0.007 | 0.007 | -0.002 | -0.003 | -0.004 | 0.693 | 0.062 | | | |
| WC_FR6 | WWL | Al_mg/L | 2.47E-04 | 9.73E-06 | 4.23E-03 | 5.27E-01 | | | Increasing | 0.065 | 0.03 | 0.045 | -0.015 | -0.065 | -0.057 | 0.000 | 0.000 | | | |
| DCC_FR6 | DCU | EC_uS/cm | 1.20E-01 | -8.48E-02 | 1.55E-05 | 2.35E-08 | | | Increasing | 120 | 112 | 140 | 6.265 | -25.713 | -8.878 | 0.000 | 0.002 | | | |
| DCC_FR6 | DCU | pH_field | 3.23E-03 | -3.14E-04 | 2.46E-04 | 1.23E-01 | | | | 5.22 | 4.76 | 4.85 | -0.167 | -0.627 | -0.333 | 0.000 | 0.002 | Lower than BL | Lower than BL | |
| DCC_FR6 | DCU | DO_% | -2.73E-02 | 1.41E-02 | 2.55E-01 | 5.89E-04 | | | | 89.1 | 95.2 | 89.05 | 1.780 | 9.390 | 3.435 | 0.000 | 0.077 | | | |
| DCC_FR6 | DCU | SO4_mg/L | 9.38E-03 | -1.48E-02 | 1.56E-01 | 1.41E-04 | | | | 2 | 5 | 9 | -2.831 | -5.685 | -3.925 | 0.004 | 0.984 | | | |
| DCC_FR6 | DCU | Fe_mg/L | 1.33E-03 | -5.48E-05 | 7.26E-02 | 3.75E-01 | | | | 0.11 | 0.065 | 0.12 | -0.114 | -0.196 | -0.078 | 0.413 | 0.824 | | | |
| DCC_FR6 | DCU | Mn_mg/L | 1.53E-04 | -4.76E-05 | 1.08E-01 | 2.39E-01 | | | | 0.043 | 0.087 | 0.113 | -0.014 | -0.021 | -0.016 | 0.386 | 0.768 | | | |
| DCC_FR6 | DCU | Zn_mg/L | 5.33E-05 | 9.23E-06 | 2.86E-02 | 1.57E-03 | | | Increasing | 0.005 | 0.0195 | 0.018 | -0.003 | 0.007 | 0.003 | 0.000 | 0.000 | Higher than BL | Higher than BL | |
| DCC_FR6 | DCU | Al_mg/L | 2.22E-03 | 1.83E-04 | 1.08E-01 | 2.94E-01 | | | | 0.15 | 0.345 | 0.35 | -0.071 | -0.177 | -0.168 | 0.941 | 0.389 | | | |
| SCK_ROCKBAR5 | SC10S1 | EC_uS/cm | -2.52E-01 | -2.55E-03 | 1.36E-01 | 8.96E-01 | | | | 96.4 | 114 | 122 | 1.433 | 14.120 | 9.290 | 0.242 | 0.155 | | | |
| SCK_ROCKBAR5 | SC10S1 | pH_field | -2.72E-03 | 1.03E-03 | 2.39E-01 | 1.25E-02 | | | | 5.42 | 6.04 | 5.91 | -0.215 | 0.328 | 0.180 | 0.002 | 0.005 | | | |
| SCK_ROCKBAR5 | SC10S1 | DO_% | -1.95E-02 | -4.58E-03 | 3.81E-01 | 3.36E-01 | | | | 78.8 | 90.1 | 90 | -5.893 | 1.629 | 1.671 | 0.015 | 0.001 | | | |
| SCK_ROCKBAR5 | SC10S1 | SO4_mg/L | -3.67E-02 | -2.55E-03 | 2.39E-01 | 4.76E-01 | | | | 2 | 8 | 8.5 | -1.660 | 2.185 | 2.338 | 0.075 | 0.000 | | Higher than BL | |
| SCK_ROCKBAR5 | SC10S1 | Fe_mg/L | -1.00E-02 | 1.09E-03 | 2.78E-03 | 6.98E-02 | | | | 0.2 | 1.14 | 1.005 | -0.154 | 0.494 | 0.327 | 0.007 | 0.312 | Higher than BL | | |
| SCK_ROCKBAR5 | SC10S1 | Mn_mg/L | -2.51E-03 | 9.32E-05 | 2.39E-01 | 5.72E-01 | | | | 0.046 | 0.557 | 0.509 | -0.078 | 0.292 | 0.242 | 0.061 | 0.003 | | Higher than BL | |
| SCK_ROCKBAR5 | SC10S1 | Zn_mg/L | -2.01E-04 | -1.82E-05 | 3.02E-02 | 3.37E-01 | | | | 0.003 | 0.03 | 0.04 | -0.013 | 0.008 | 0.018 | 0.001 | 0.000 | Higher than BL | Higher than BL | |
| SCK_ROCKBAR5 | SC10S1 | Al_mg/L | 1.45E-04 | -6.60E-05 | 3.81E-01 | 3.39E-02 | | | | 0 | 0.05 | 0.0625 | 0.000 | -0.028 | -0.014 | 0.000 | 0.000 | | | |
| WWU4 | WWU | EC_uS/cm | -4.23E-02 | 1.03E-02 | 4.83E-01 | 4.18E-01 | | | | 85.1 | 93 | 93 | 2.866 | 1.278 | -6.504 | 0.917 | 0.008 | | | |
| WWU4 | WWU | pH_field | -5.49E-03 | 1.35E-03 | 4.83E-01 | 3.94E-03 | | | | 5.28 | 6.47 | 5.77 | -0.288 | 0.808 | 0.119 | 0.001 | 0.005 | | | |
| WWU4 | WWU | DO_% | 3.20E-03 | 7.67E-03 | 1.00E+00 | 8.83E-02 | | | | 93.1 | 100.4 | 95.3 | 1.124 | 7.085 | 2.859 | 0.168 | 0.247 | | | |
| WWU4 | WWU | SO4_mg/L | -1.13E-02 | 3.29E-03 | 4.83E-01 | 9.53E-03 | | | Increasing | 6.5 | 5 | 4 | 0.434 | 1.006 | -0.324 | 0.602 | 0.373 | | | |
| WWU4 | WWU | Fe_mg/L | 2.45E-04 | 3.71E-05 | 4.83E-01 | 3.39E-02 | | | Increasing | 0.07 | 0.08 | 0.05 | 0.002 | 0.002 | -0.018 | 0.917 | 0.182 | | | |
| WWU4 | WWU | Mn_mg/L | -1.31E-04 | -2.09E-05 | 2.33E-01 | 1.74E-01 | | | | 0.0985 | 0.053 | 0.0555 | 0.017 | -0.028 | -0.023 | 0.000 | 0.000 | | | |
| WWU4 | WWU | Zn_mg/L | -2.94E-05 | 6.56E-06 | 4.83E-01 | 7.13E-01 | | | | 0.0355 | 0.028 | 0.0295 | 0.005 | -0.001 | 0.000 | 0.034 | 0.007 | | | |
| WWU4 | WWU | Al_mg/L | 4.31E-04 | -2.60E-05 | 2.33E-01 | 1.59E-01 | | | | 0.065 | 0.035 | 0.055 | 0.003 | -0.027 | -0.015 | 0.063 | 0.004 | | | |

- Notes:
- Theil-Sen slope is the median of the slopes between all pairs of x-y points in the data. It is a non parametric estimator of median slope
 - Mann-Kendal test for serial correlation (p-value): the probability of obtaining a correlation result at least as extreme due to chance. Results significant at the 95% level are indicated in **bold** (values <0.05)
 - Flow-corrected statistics are the median values for the periods indicated, based on the flow-corrected residuals (not actual measurement values)
 - Mann-Whitney U test: A non parametric rank-sum test for the difference in means between two samples (in this case, observations in different time intervals). Expressed as a p-value with significance at the 95% level indicated in **bold**.

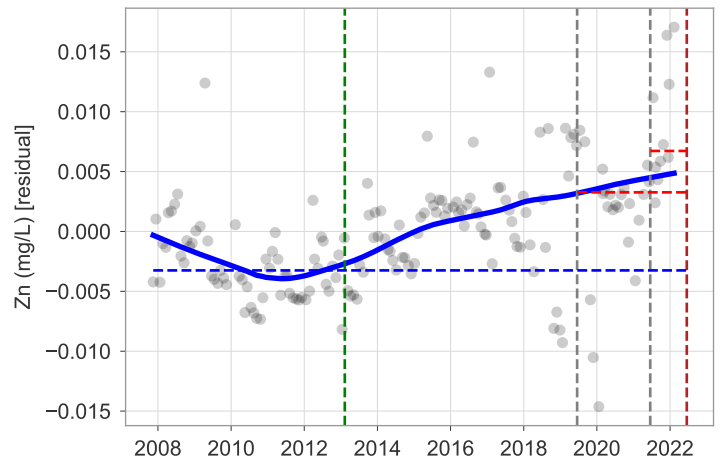
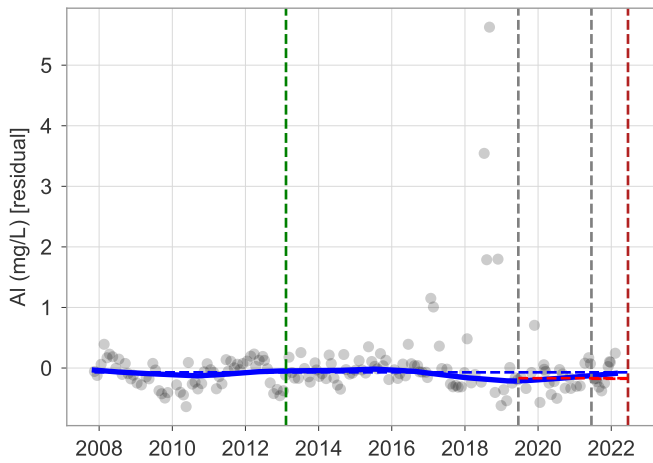
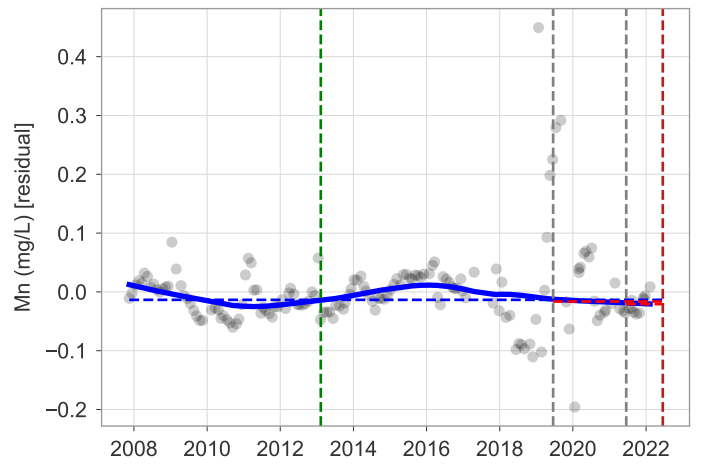
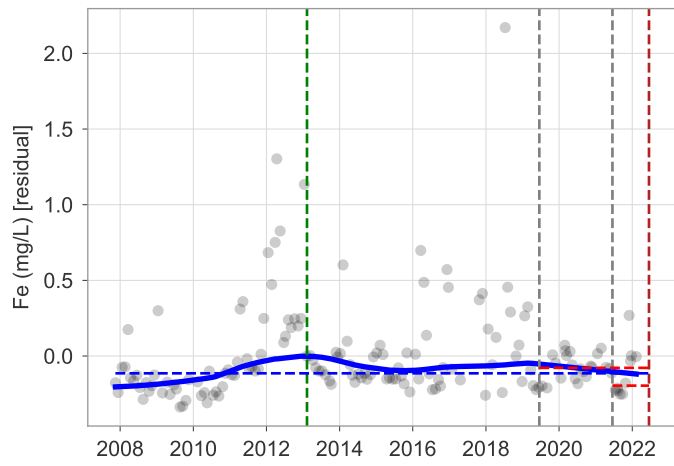
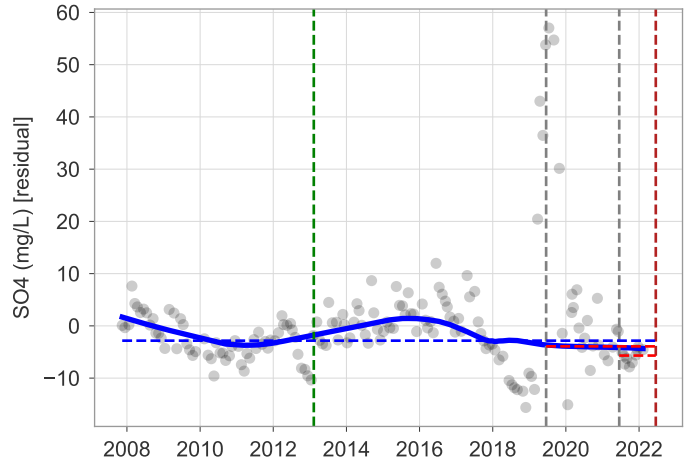
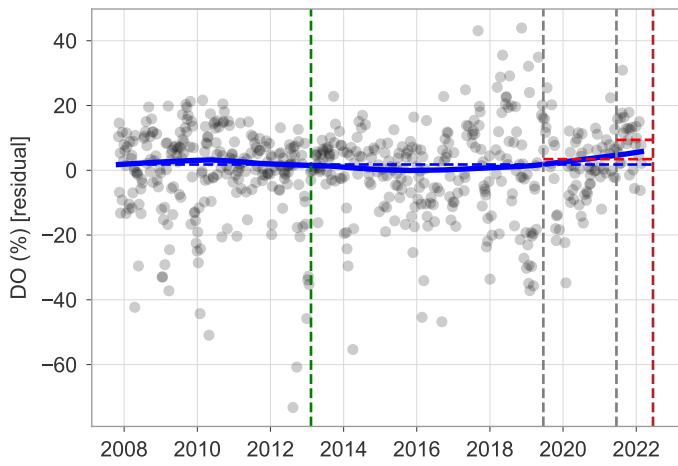
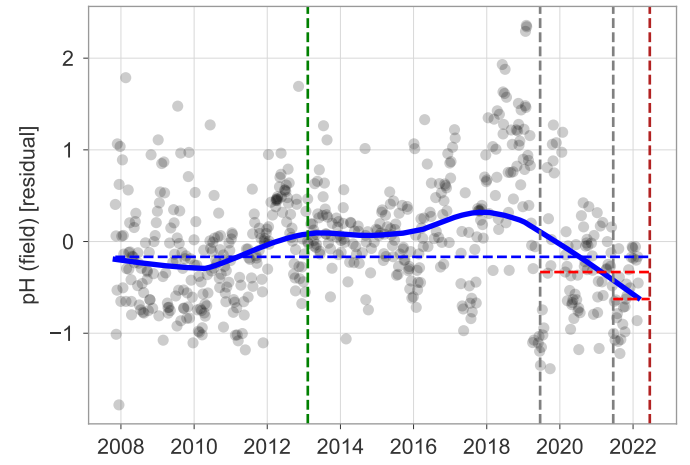
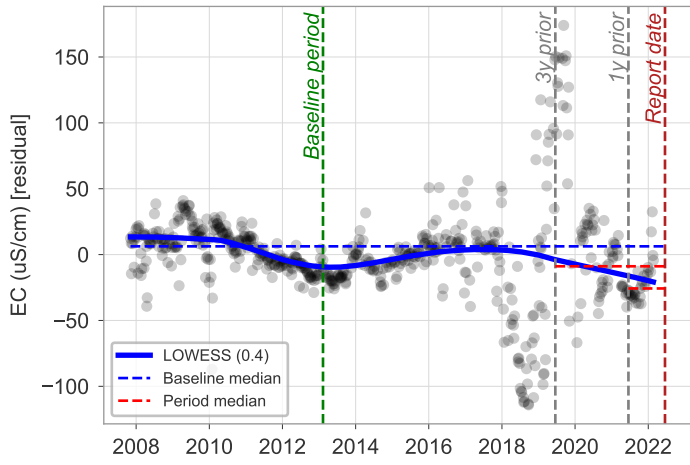
WC_FR6 Flow-corrected time series



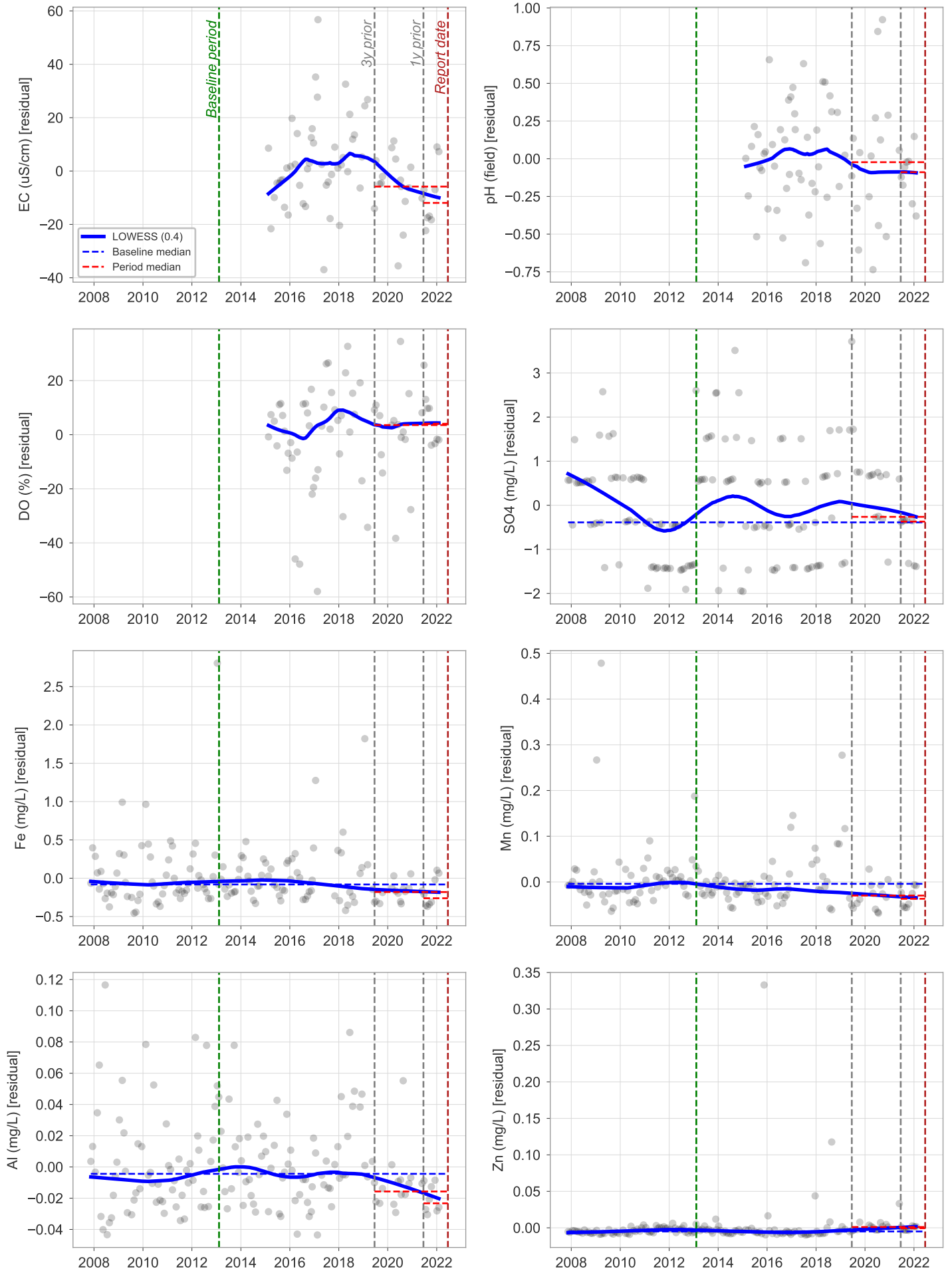
WWU4 Flow-corrected time series



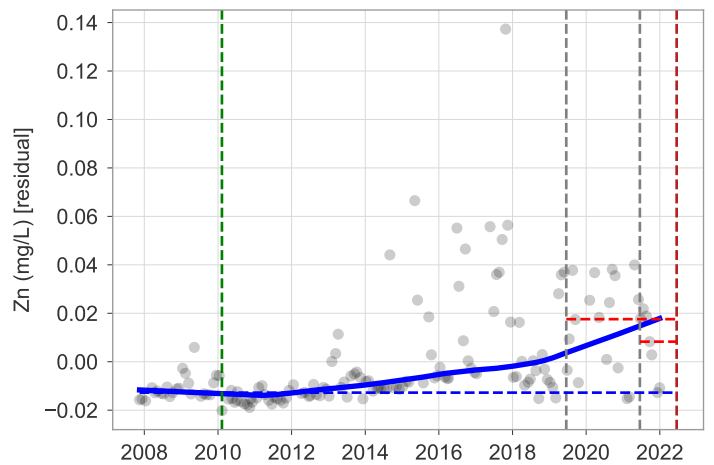
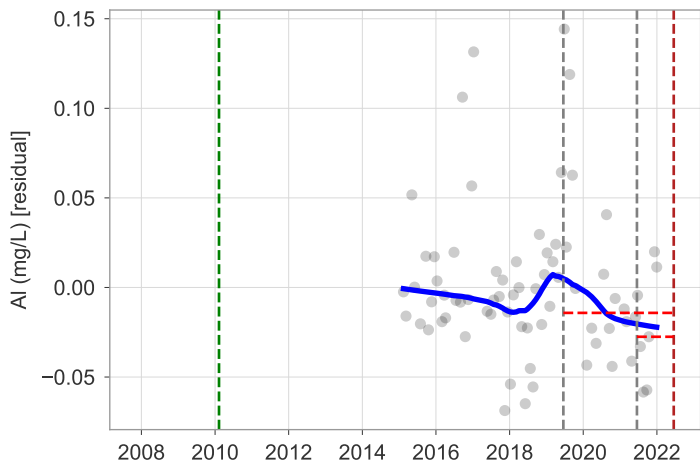
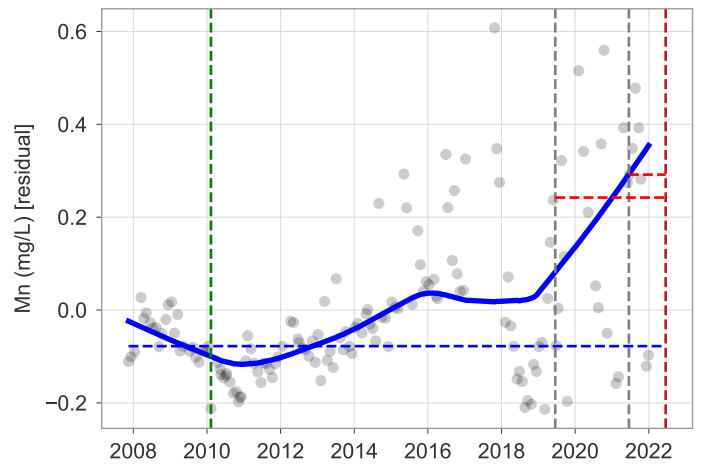
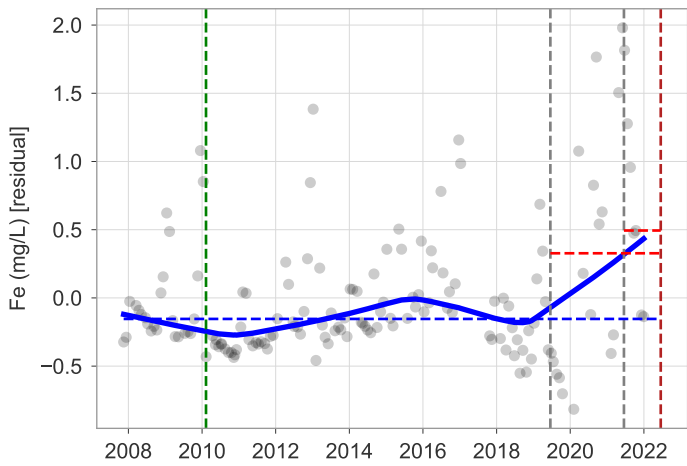
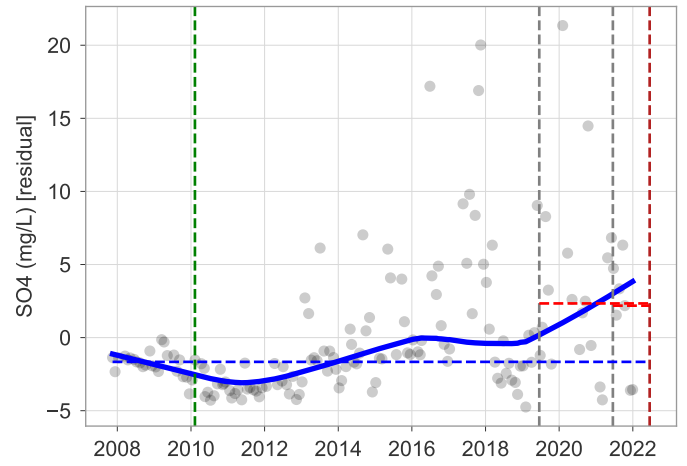
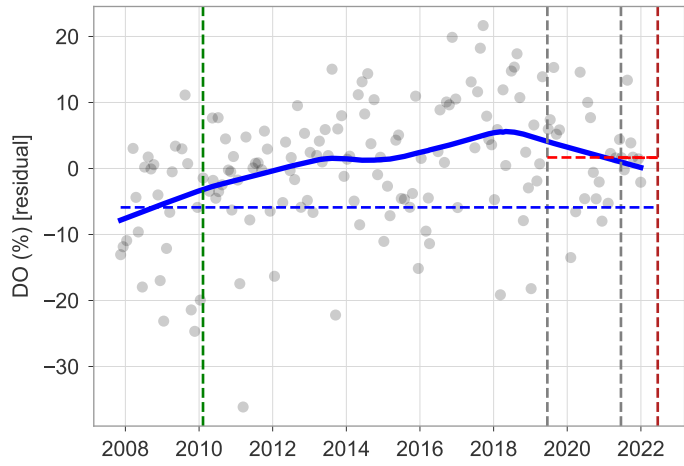
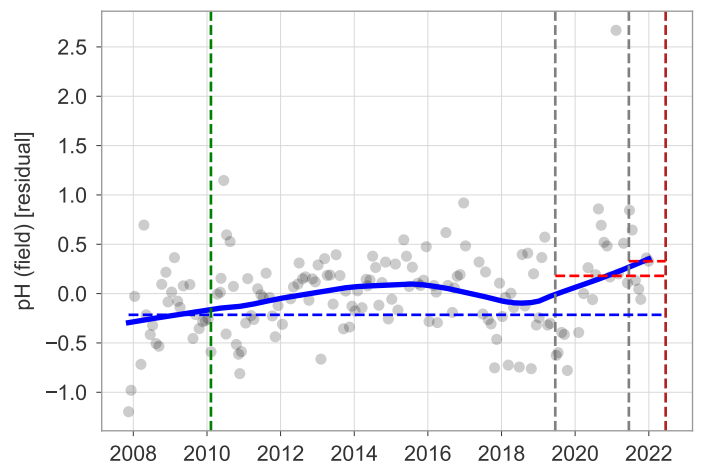
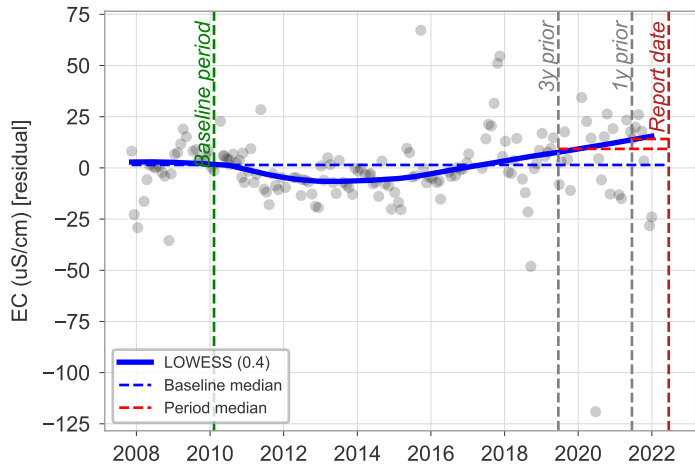
DCC_FR6 Flow-corrected time series



DCL3 Flow-corrected time series



SCK_ROCKBAR5 Flow-corrected time series



Appendix B: Rainfall data

Monitoring Data

The geographic distribution of the various data sources at which rainfall and stream flow is measured around Dendrobium are shown on Figure B1.

Rainfall data

Rainfall data for the Dendrobium area is available from three primary sources (Figure B1):

- A series of rainfall gauges owned by IMC (“site data”) and currently operated by the hydrographic consultants ALS;
- A rainfall gauge (“Browns Road”) operated by WaterNSW (located within Area 3B); and
- Series of “infilled” data available from the SILO service, a cooperative initiative of the Queensland Government’s Department of Environment and Science (DES) and the Australian Bureau of Meteorology (BoM).

The details of the various data sources is summarised in Table B1.

Measurement uncertainty

Based on manufacturers specifications¹, “the Hyquest Solutions TB3 Model Tipping Bucket Rain Gauge is recognised as the world standard for measuring rainfall and precipitation in remote and unattended locations. The TB3 is the rain gauge of choice to the Australian Bureau of Meteorology and other organisations world wide.”

ALS have instructed that the manufacturer’s stated accuracy is:

| Rainfall intensity | Accuracy |
|---------------------|-----------------|
| 0-250 mm per hour | +/-2% |
| 250-500 mm per hour | +/-3% |
| Measurement range | 700 mm per hour |

These accuracies are independent of the siting of the gauge itself. ALS have stated that the siting of the gauges has been carried out consistent with RMS, 2016², which is itself compiled from Australian Standards. ALS can provide further information on request.

¹ <https://www.hyquestsolutions.com/products-services/products-hardware/meteorology/model-tb3-tipping-bucket-rain-gauge/>

² NSW RMS, 2016. Automatic weather stations: QA specification R272. Edition 1, rev 1, May 2016.

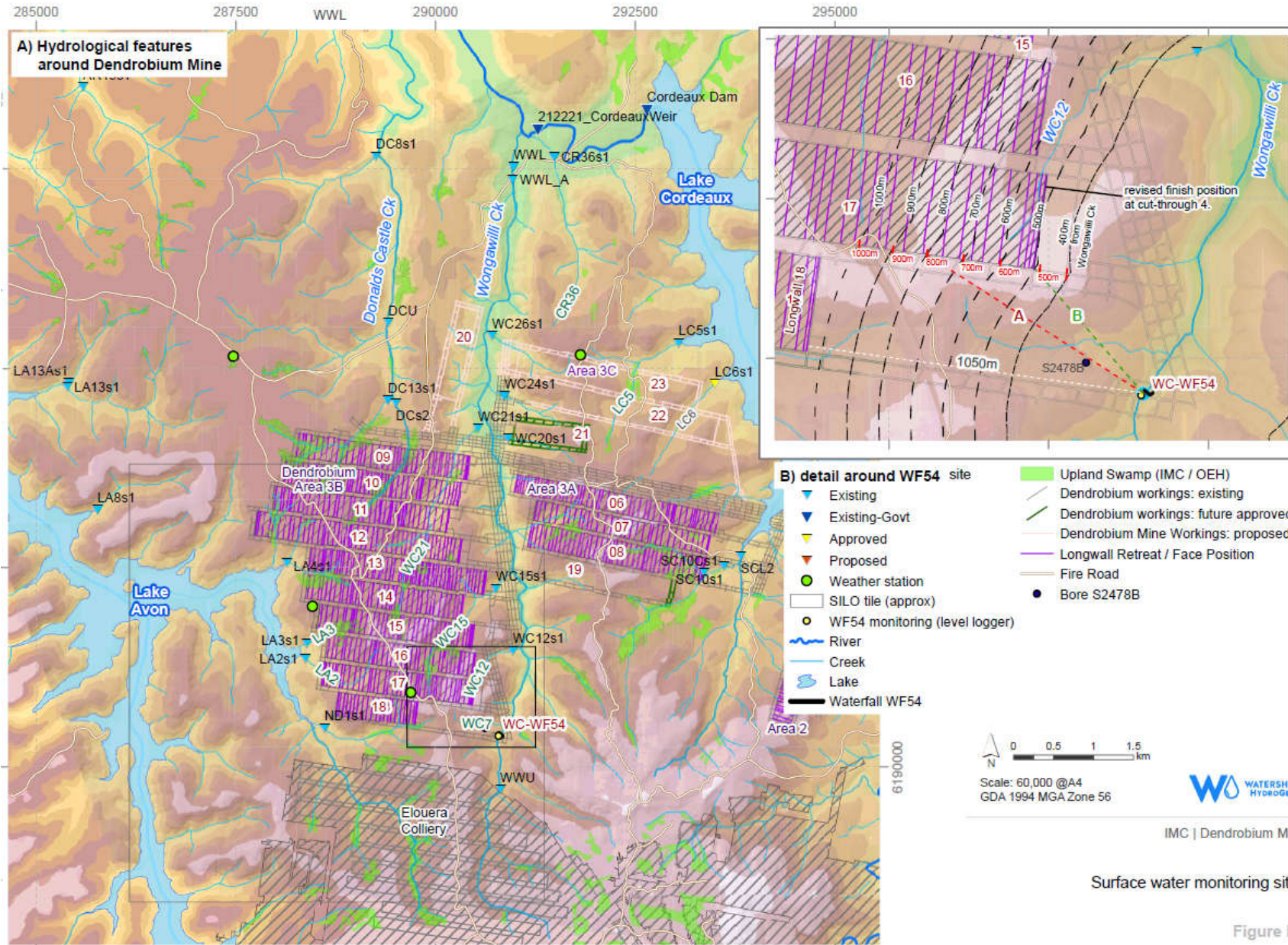


Figure B1 Location of rainfall and flow data sources

Table B1 Rainfall data sources

| SITE_NAME | ID | EASTING (Z56) | NORTHING (Z56) | Z_ELEVATION [MAHD] | OPERATOR | START DATE | MEASUREMENT FREQUENCY | GAUGE TYPE | LOGGER |
|------------------------------|--------|------------------|-------------------|-----------------------|----------|---------------|--|---|--|
| Dendrobium "Centroid" A3A | | 291815 | 6195170 | 403.1 | IMC | 28/10/2007 | event based and midnight timestamp | Hyquest Tipping Bucket Rain Gauge model TB3 / 0.5mm | Hyquest Minilog |
| DA3B Weather Station | | 288458 | 6192012 | 413.4 | IMC | 1/06/2012 | event based and midnight timestamp, 15 minute on CR800 | Hyquest Tipping Bucket Rain Gauge model TB3 / 0.5mm | Hyquest Minilog and Campbells Scientific CR800 |
| DA5 Rainfall Gauge | | 287468 | 6195153 | 401.6 | IMC | 19/07/2017 | event based and midnight timestamp | Hyquest Tipping Bucket Rain Gauge model TB3 / 0.5mm | Hyquest Minilog |
| DA6 Rainfall Gauge | | 291749 | 6200383 | 352.4 | IMC | 17/06/2017 | event based and midnight timestamp | Hyquest Tipping Bucket Rain Gauge model TB3 / 0.5mm | Hyquest Minilog |
| Cordeaux Site Rain Gauge | | 294658 | 6199531 | 373.5 | IMC | 1/01/2002 | data recorded @ 00:00 for previous 24 hrs | | |
| Browns Road | 568061 | 289690 | 6190930 | 442.0 | WaterNSW | 31/03/1983 | data recorded @ 00:00 for previous 24 hrs | | |
| SILO Data drill - "A3B" | | Long: 150.70 | Lat: -34.40 | | SILO | 1/01/1900 | 24hr total to 9am, interpolated and averaged for 0.05x0.05 degree tile | | |

Notes:

IMC sites maintained by ALS.

Browns Road data obtained from WaterNSW.

SILO data from <https://www.longpaddock.qld.gov.au/silo/datadrill/>

A comparison of the recent data from these sources is presented below.

Figure B2 shows recent annual totals (since the commencement of Dendrobium Mine). For these, the SILO and Dendrobium Area 3B record is 100% complete, Area 3A is 100% complete, while the Browns Road record is 94.5% complete.

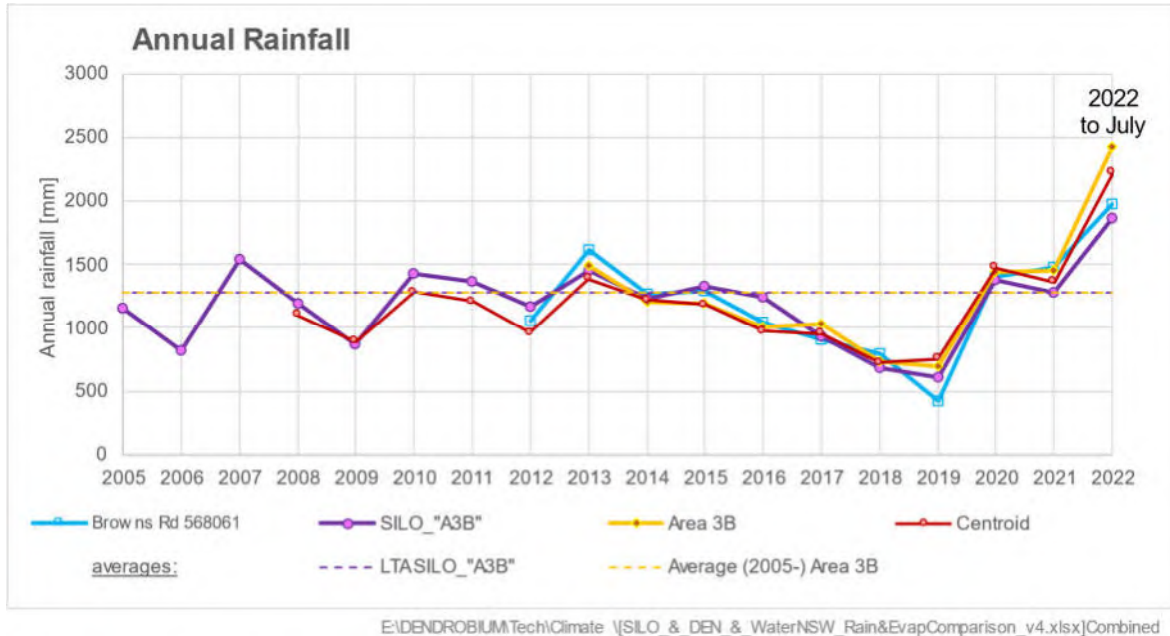


Figure B2 Annual rainfall totals

Figure B2 shows that the different records show some variability in annual totals. 2013, 2020 and 2021 were the wettest years of the recent period, until 2022 – the 7 months of 2022 have seen record rainfall (across much of eastern NSW). In the wettest years, the variability between the different monitoring records shown above, from minimum to maximum, was 14%, 7%, 23% and 23% in 2013, 2020, 2021 and 2022 (to date) respectively. In 2019 (the driest of the selected years), the variability was 44%. All stations show the same broad trends across those selected years, including the severe and persistent rainfall deficit in 2017-2019, and the return to higher rainfall across 2020-21 and then the record rainfall of 2022.

There was variation in which station or data source was wettest or driest in each year but in terms of cumulative rainfall 2013-2022 to date (excluding 2012 because the Area 3B gauge did not commence until May that year), there is 10% difference between the four monitoring records shown on Figure B2.

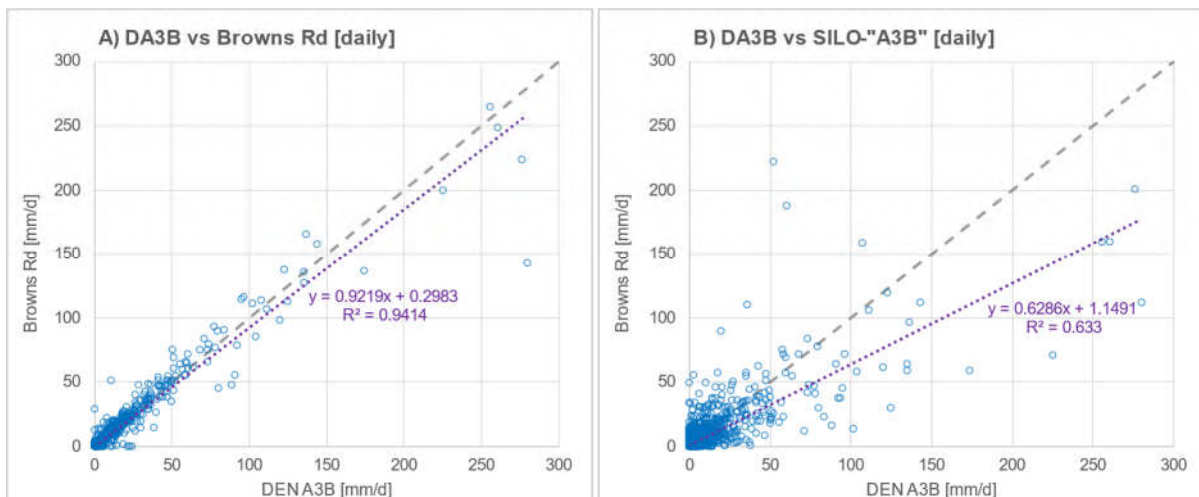


Figure B3 Correlation of daily rainfall around Area 3B

Figure B3 shows a comparison of daily totals for the three stations most relevant to Area 3B.

Figure B3a shows that the A3B station and the Browns Road station, which are 1.6 km apart, are highly correlated ($R^2 = 0.94$). Figure B3b shows that there is a substantially weaker correlation between the local SILO record and the Browns Road station (and therefore, also with the A3B record) ($R^2 = 0.63$).

To assess this further, the accumulated rainfall for the available record during the period 2005-2020 is plotted on Figure B4. The A3B rainfall is not included here due to that record beginning in 2012.

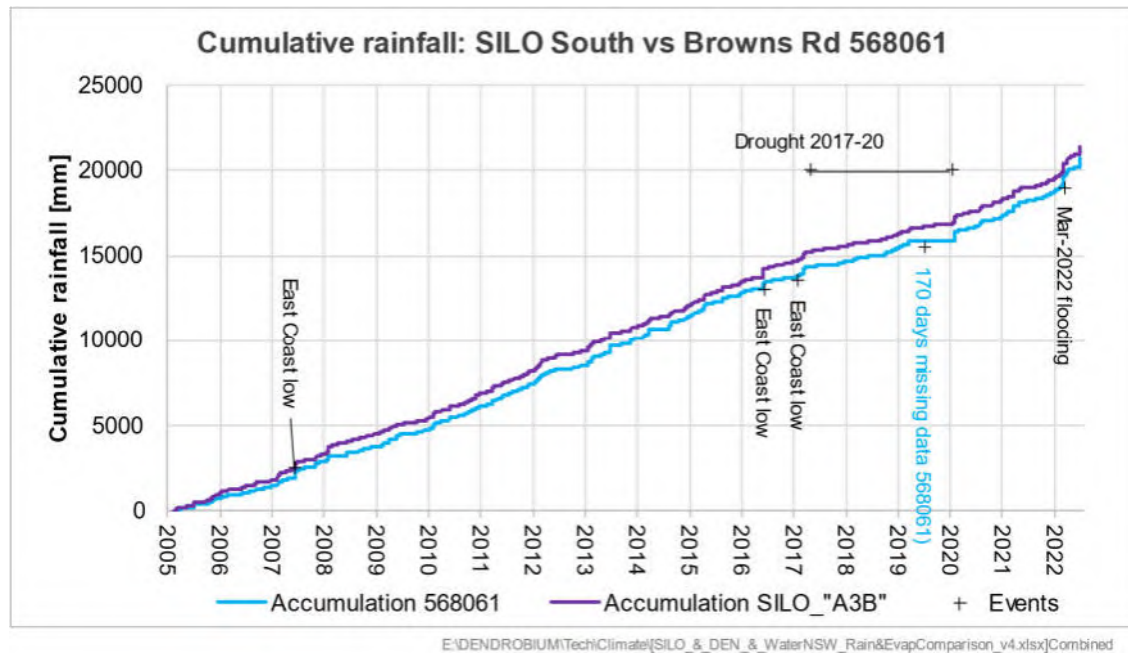


Figure B4 Rainfall accumulation near to the Area 3B domain

Figure B4 indicates that while there is a weaker correlation on a daily basis, and some significant variances in total rainfall on specific days (Figure B3b), the correlation is strong overall. That is, the accumulation shows very similar trends through the 17 year record, and the overall total accumulated rainfall varies by 3% to July-2022.

This analysis of the data indicates that broadly there is general agreement in the data on a monthly or annual timeframe, but that short-term differences exist. These differences may be due to:

- Measurement error (in the case of the site data, Browns Road and the raw BoM data behind the SILO “Data Drill” record);
- Mismatch in timing between totals to midnight (site and Browns Road data) versus totals to 9am (BoM / SILO records);
- Interpolation or infilling errors due to temporal or spatial infilling or averaging (in the SILO data);
- True variations in rainfall in space and time due to weather patterns, e.g. topographic effects and/or the effects of localised weather systems.

These measurement errors or true variations will propagate into any rainfall-runoff modelling and simulation of daily flow around Area 3B. Smaller catchments will be less prone to spatial variation in rainfall across the catchment, but maybe affected by not having a truly local rainfall record. Therefore, while the rainfall gauging network at Dendrobium (including SILO and WaterNSW data) is considered adequate, the potential localisation of rainfall patterns will result in some unavoidable discrepancies between modelled and observed flows.

Appendix C: Flow gauge data

Flow data for the Dendrobium area is available from a series of flow gauges owned by IMC and currently operated by the hydrographic consultants ALS. The details of these are summarised in Table C1.

These gauging stations provide estimates of stream flow via:

- (A) A structure behind which water pools and flows over. Structures can be:
 - a. natural, e.g. a rock bar, or
 - b. man-made, e.g. a half-pipe flume.
- (B) A sensor and logger that measure and record the water level (“stage”) in the pool at 15-minute intervals
- (C) A “rating curve” which is a chart or graph of discharge (flow) versus stage for each gauging station. The rating curve is developed via periodic measurements of flow in the channel at a known water level.
- (D) Estimates of mean daily flow are then provided.

IMC commissioned an independent hydrologist (Enviromon) to systematically identify and quantify the accuracy or error involved in each part of the process (Enviromon, 2019 and Enviromon, 2020). This process has been carried out for a selection of gauging sites (provided in C5, below), and is currently being applied to the remaining in gauging sites, with some fine-tuning of the method for gauging sites where ‘underflow’ occurs (i.e. flow beneath the monitoring structure). The objective is to re-assess each gauging station for each future End of Panel report, and to use the results to reduce uncertainty where practicable (i.e. additional data-gathering and improved measurement methods).

Table C1 Flow gauge information

| Watercourse | Site | Area | Easting (z56) | Northing (z56) | Z_Elevation [mAHD] | Catchment area [km ²] | Installation | Structure type | | Logger |
|----------------|---------|-----------|---------------|----------------|--------------------|-----------------------------------|-------------------------|--------------------------|--------|------------------------------------|
| Wongawilli Ck | WWU | u/s A3B | 290808 | 6189716 | 352.94 | 3.211 | Stainless Steel housing | Natural control | | Orpheus |
| Wongawilli Ck | WWL | d/s A3A,B | 290975 | 6197526 | 261.86 | 20.079 | Stainless Steel housing | Natural control | | Diver |
| Wongawilli Ck | WWL_A | d/s A3A,B | 290962 | 6197370 | 263.22 | 19.602 | PVC housing | Half pipe | 225 mm | Orpheus |
| WC21 | WC21S1 | A3B | 290529 | 6194255 | 283.07 | 2.434 | Stainless Steel housing | Natural control | | Diver |
| WC15 | WC15S1 | A3B | 290754 | 6192239 | 324.71 | 1.192 | PVC housing | Natural control | | Diver |
| WC12 | WC12S1 | A3B | 290964 | 6191459 | 322.34 | 0.380 | Polypipe housing | Weir and half pipe flume | 150 mm | Orpheus |
| LA2 | LA2S1 | A3B | 288364 | 6191364 | 324.65 | 0.824 | Polypipe housing | Weir and half pipe flume | 150 mm | Orpheus |
| LA3 | LA3S1 | A3B | 288385 | 6191548 | 323.82 | 0.375 | Polypipe housing | Weir and half pipe flume | 150 mm | Orpheus |
| LA4 | LA4S1 | A3B | 288134 | 6192565 | 322.98 | 0.817 | Stainless Steel housing | Modified control | 150 mm | Diver |
| ND1 | ND1S1 | A3B | 288607 | 6190491 | 325.11 | 1.130 | Polypipe housing | Weir and half pipe flume | 150 mm | Orpheus |
| DC13 | DC13S1 | A3B | 289401 | 6194605 | 339.50 | 1.638 | PVC housing | Natural control | | Diver |
| Donalds Castle | DCS2 | A3B | 289502 | 6194572 | 341.27 | 1.084 | PVC housing | Natural control | | Diver |
| Donalds Castle | DCU | A3B | 289407 | 6195577 | 322.42 | 6.219 | Stainless Steel housing | Natural control | | Diver |
| SC10 | SC10S1 | A3A | 293608 | 6192516 | 333.03 | 2.771 | Stainless Steel housing | Natural control | | Diver, recently updated to Orpheus |
| SC10C | SC10CS1 | A3A | 293358 | 6192433 | 340.78 | 0.817 | Stainless Steel housing | Natural control | | Diver |
| Sandy Ck | SCL2 | A3A | 293819 | 6192648 | 328.61 | 7.029 | Stainless Steel housing | Modified control (leaky) | | Diver |
| | 2022205 | A3A | 293819 | 6192648 | 328.61 | 7.029 | WaterNSW site | Modified control (leaky) | | Unknown |
| LC5 | LC5S1 | A3A,C | 293043 | 6195327 | 318.10 | 1.861 | Polypipe housing | Weir and half pipe flume | 225 mm | Orpheus |
| LA8 | LA8S1 | A5 | 285764 | 6193225 | 331.56 | 0.93 | Polypipe housing | Weir and half pipe flume | 150 mm | Orpheus |
| LA13A | LA13AS1 | A5 | 285401 | 6194826 | 319.93 | 1.04 | Polypipe housing | Weir and half pipe flume | 150 mm | Orpheus level, Diver EC |
| LA13 | LA13S1 | A5 | 285384 | 6194777 | 320.35 | 2.79 | PVC housing | Weir and half pipe flume | 225 mm | Orpheus |

| Watercourse | Site | Area | Easting (z56) | Northing (z56) | Z_Elevation [mAHD] | Catchment area [km ²] | Installation | Structure type | | Logger |
|-------------|--------|------|---------------|----------------|--------------------|-----------------------------------|-------------------------|--------------------------|--------|-------------------------|
| AR31 | AR31S1 | A5 | 283999 | 6197770 | 270.70 | 2.96 | Polypipe housing | Weir and half pipe flume | 225 mm | Orpheus level, Diver EC |
| AR32 | AR32S1 | A5 | 283945 | 6197576 | 266.10 | 1.5 | PVC housing | Weir and half pipe flume | 150 mm | Orpheus |
| AR19 | AR19S1 | A5 | 285584 | 6198528 | 382.91 | 3.53 | Polypipe housing | Weir and half pipe flume | 225 mm | Orpheus level, Diver EC |
| DC8 | DC8S1 | A5 | 289249 | 6197663 | 301.20 | 2.61 | Polypipe housing | Weir and half pipe flume | 225 mm | Orpheus level, Diver EC |
| CR29 | CR29S1 | A6 | 289969 | 6201109 | 257.13 | 2.33 | Polypipe housing | Weir and half pipe flume | 225 mm | Orpheus level, Diver EC |
| CR31 | CR31S1 | A6 | 290062 | 6200056 | 248.33 | 2.55 | Polypipe housing | Weir and half pipe flume | 225 mm | Orpheus level, Diver EC |
| CR36 | CR36S1 | A3C | 291487 | 6197650 | 272.82 | 1.7 | PVC housing | Weir and half pipe flume | 225 mm | Orpheus |
| WC20 | WC20s1 | A3C | 290906 | 6194133 | 303.86 | 0.44 | Approved, not installed | Weir and half pipe flume | 150 mm | Orpheus |
| WC24 | WC24s1 | A3C | 290863 | 6194658 | 286.90 | 0.5 | <i>PVC housing</i> | Weir and half pipe flume | 150 mm | Orpheus |
| WC26 | WC26s1 | A3C | 290703 | 6195411 | 277.85 | 0.55 | <i>PVC housing</i> | Weir and half pipe flume | 150 mm | Orpheus |
| LC6 | LC6s1 | A3C | 293495 | 6194813 | 328 | 1.16 | Awaiting approval | | | |

Reference Sites

| | | | | | | | | | | |
|------------|--------|------------|--------|---------|--------|------|--|---------|--|--|
| O'Hares Ck | 213200 | Wedderburn | 300657 | 6217589 | 166.87 | 73.0 | | V-notch | | |
|------------|--------|------------|--------|---------|--------|------|--|---------|--|--|

alternative Reference Sites

| | | | | | | | | | | |
|--------------|---------|-------------------|--------|---------|--------|------|-------------------------|-------------------------------|--|--------------------------|
| Bomaderry Ck | 215016 | Bomaderry | 279354 | 6142065 | 25.822 | 31.0 | Vandalised, stolen 2019 | Natural rock bar and boulders | | Vandalised, stolen 2019* |
| Cordeaux R | 2122204 | Cordeaux Dam No.1 | 295413 | 6188702 | | 9.3^ | | | | |

Notes:

IMC gauging stations operated by ALS. *Italicised text* = to be confirmed.

Reference gauge data from WaterNSW.

* WaterNSW advise that Bomaderry Creek gauging site repeatedly vandalised, equipment stolen (2019-20) and not yet replaced.

^ not from WaterNSW - estimated in GIS

Table C2 ALS data quality codes

| DATA QUALITY CODE | DESCRIPTION | |
|-------------------|---|--|
| 1 | Good continuous records | Notes: Negative values of these codes may be shown on the data quality charts shown later in this Appendix. These indicate where a record has been processed or infilled. |
| 2 | Reliable Edited Data | |
| 3 | Unreliable Edited Data | |
| 5 | Non-Continuous Data | |
| 55 | Fair quality data | |
| 69 | Fair Quality Rating Extrapolated | |
| 104 | Records estimated | |
| 109 | Poor quality data | |
| 140 | Level below CTF (cease-to-flow) | |
| 145 | Discharge not reliable Rating under review | |
| 150 | Rating table extrapolated due to inadequate gauging information | |
| 151 | Data not yet available | |
| 160 | Water level below sensor | |
| 161 | Poor quality data from debris affecting sensor | |
| 205 | Data lost | |
| 255 | No data exists | |

Table C3 Data quality assessment for Reference Gauges

| Watercourse | Gauge Id | Gauge Name | Start Date | End Date | No. of Records | % available | % suspect | % infilled | Status |
|-----------------|----------|---------------------------------------|------------|-------------|----------------|-------------|-----------|------------|---|
| Wongawilli Ck | 300024 | WWU | 2/11/2007 | 18/11/2021 | 5071 | 99.8% | 2.1% | 0.9% | Primary Reference |
| O'Hares Creek | 213200 | Wedderburn (213200) | 2/02/1978 | 21/12/2021 | 16029 | 99.9% | 0.1% | 0.1% | Primary Reference |
| Bomaderry Creek | 215016 | Bomaderry Creek at Bomaderry (215016) | 4/07/2003 | 10/05/2020* | 6156 | 99.0% | 1.0% | | Alternative Reference – equipment vandalised |
| Woronora River | 2132101 | Woronora River Fire Rd (2132101) | 21/02/2007 | 27/03/2020 | 4784 | 81.3% | 35.9% | | Alternative Reference (low quality) |
| LC5 | 300094 | LC5S1 | 4/04/2019 | 11/01/2021 | 649 | 100.0% | 2.6% | 0.0% | Future Reference site |

Additional Reference Sites might be adopted for future analysis – a subset of Area 5 sites (LA8S1, DC8S1, AR19S1) will continue to be monitored and be used as Reference Sites.

* WaterNSW advise that Bomaderry Creek gauging site repeatedly vandalised, equipment stolen (2019-20) and after May-2020, not yet replaced.

Table C4 Data quality assessment for Area 3B and relevant Assessment Sites

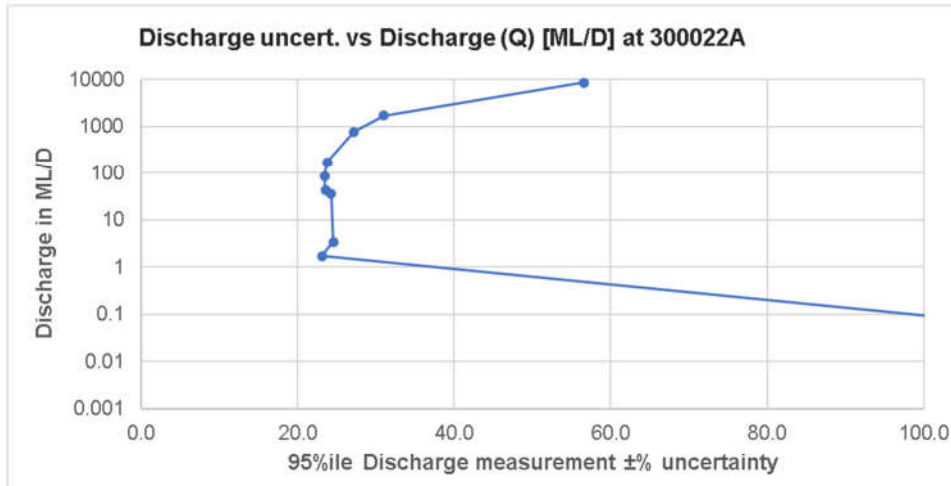
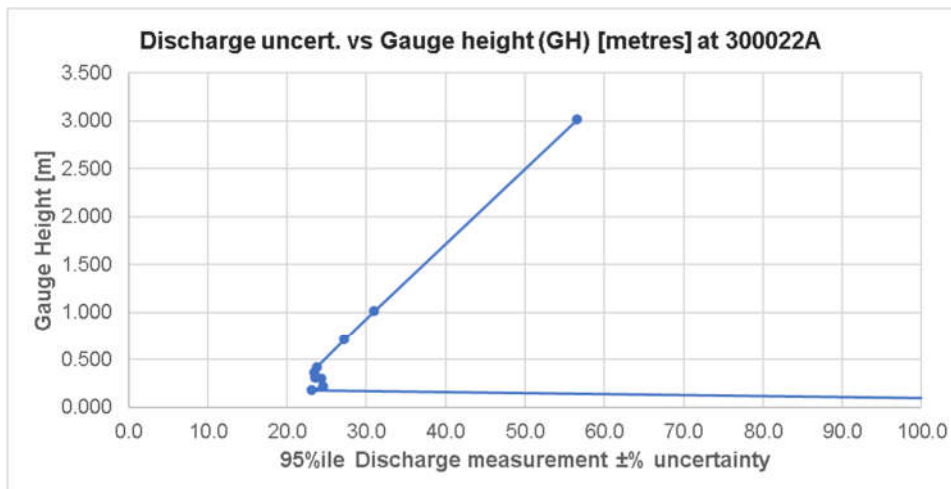
| Assessment Sites | | | Summary - Pre-Mining | | | | | | Summary - Post-Mining | | | | | |
|----------------------|-------------|------------|----------------------|------------|----------------|-------------|-----------|------------|-----------------------|------------|----------------|-------------|-----------|------------|
| Watercourse | Hydstra No. | Gauge Name | Start Date | End Date | No. of Records | % available | % suspect | % infilled | Start Date | End Date | No. of Records | % available | % suspect | % infilled |
| Donalds Castle Creek | 300023 | DCU | 2/11/2007 | 8/02/2013 | 1926 | 100.0% | 0.0% | 0.0% | 9/02/2013 | 19/11/2021 | 3206 | 99.8% | 14.6% | 0.2% |
| DC13 | 300067 | DC13S1 | 27/06/2012 | 8/02/2013 | 227 | 100.0% | 0.0% | 0.0% | 9/02/2013 | 19/11/2021 | 3206 | 97.8% | 4.3% | 2.2% |
| Donalds Castle Creek | 300068 | DCS2 | 27/06/2012 | 9/07/2013 | 378 | 100.0% | 0.0% | 0.0% | 10/07/2013 | 19/11/2021 | 3055 | 98.0% | 9.9% | 2.0% |
| Wongawilli Creek | 300022 | WWL | 2/11/2007 | 8/02/2010 | 830 | 96.4% | 3.8% | 3.6% | 9/02/2010 | 29/11/2021 | 4312 | 100.0% | 1.0% | 0.1% |
| Wongawilli Creek | 300022A | WWLA | | | | | | | 23/08/2018 | 29/11/2021 | 1195 | 100.0% | 5.8% | 1.0% |
| WC21 | 300069 | WC21S1 | 20/06/2012 | 4/10/2013 | 472 | 99.8% | 0.2% | 0.2% | 5/10/2013 | 18/11/2021 | 2967 | 100.0% | 0.1% | 0.0% |
| WC15 | 300071 | WC15S1 | 20/06/2012 | 27/01/2017 | 1683 | 91.2% | 15.9% | 9.0% | 28/01/2017 | 18/11/2021 | 1756 | 100.0% | 17.4% | 0.0% |
| WC12 | 300092 | WC12S1 | 5/04/2019 | 17/10/2020 | 562 | 100.0% | 5.0% | 0.0% | 18/10/2020 | 12/11/2021 | 391 | 100.0% | 4.9% | 0.0% |
| LA4 | 300070 | LA4S1 | 24/09/2012 | 31/03/2015 | 919 | 100.0% | 0.0% | 0.0% | 1/04/2015 | 19/11/2021 | 2425 | 77.3% | 35.1% | 0.3% |
| LA3 | 300091 | LA3S1 | 3/02/2019 | 27/04/2019 | 84 | 100.0% | 2.4% | 0.0% | 28/04/2019 | 12/11/2021 | 930 | 100.0% | 7.3% | 0.2% |
| LA2 | 300090 | LA2S1 | 4/02/2019 | 29/02/2020 | 391 | 100.0% | 14.6% | 0.0% | 1/03/2020 | 12/11/2021 | 622 | 100.0% | 1.61% | 0.0% |
| ND1 | 300093 | NDS1 | 3/03/2019 | 17/04/2021 | 777 | 100.0% | 8.8% | 0.0% | 18/04/2021 | 27/11/2021 | 224 | 100.0% | 1.79% | 0.0% |

E:\DENDROBIUM\Tech\SurfaceWater\SWFlowData_Compiled_Wshed_20211217_v4.xlsx

“Suspect” data based on raw data assigned ALS quality codes >145.

C5) Flow gauge uncertainty (assessed by Enviromon)

Gauge 300022A – WWL_A



The use of an artificial control (halfpipe) leads to generally lower uncertainty than for the natural control sites (e.g. DCS2, WC21 or WWL). High flow uncertainty related to flow above the structure and few high flow gaugings. High flow gaugings, possibly via alternative methods, are recommended.

Comments from Enviromon:

1) There are 13 rating table points, but only 2 of which are within the half pipe depth, to explain and define its low flow rating, which is insufficient to cover this range of most interest;

For example, after CTF level of 0.012m, there are 0.022m (10mm above CTF), then 0.024m (12mm above CTF), then 0.178m (54mm above the top of the half pipe, and spread across the wide bund).

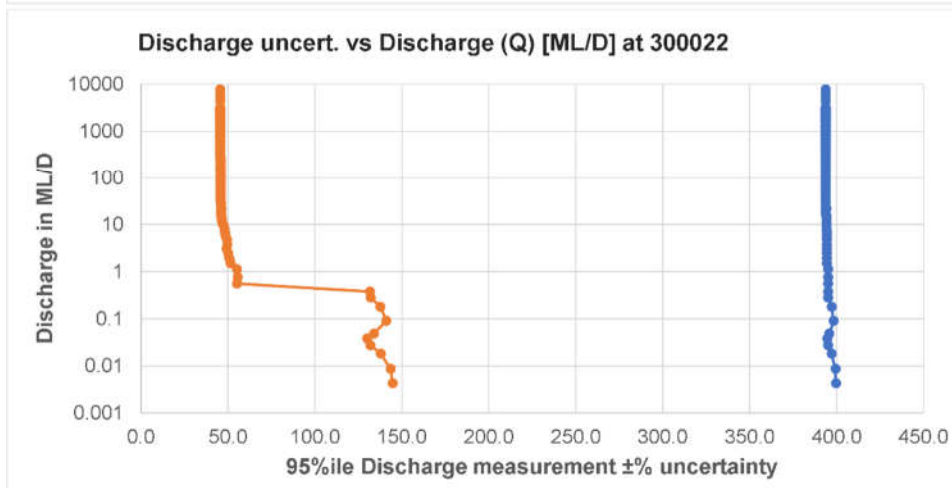
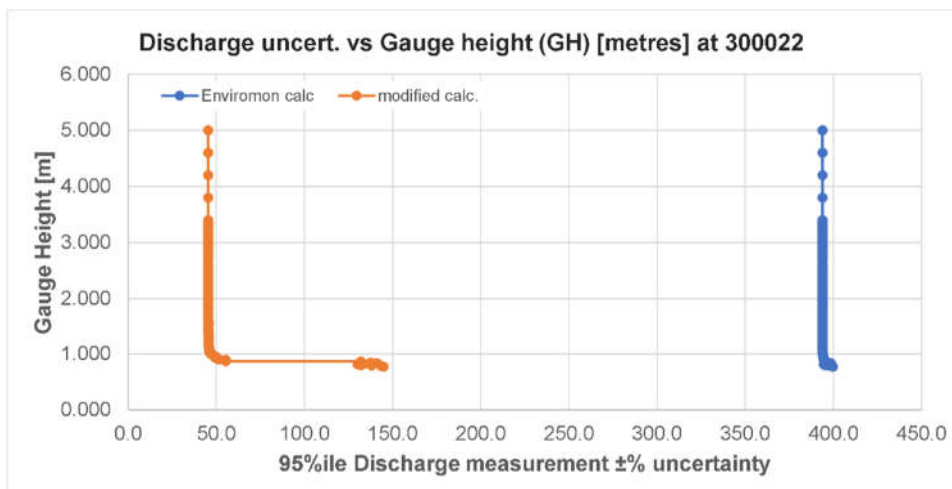
2) ALS could be asked to consider using the generic 225mm half pipe rating done by Enviromon for South32, which does have far more rating points;

3) Most of the low flow gaugings have been done with a pygmy meter held in the centre line of the half pipe. Although practical, this method is unconventional.

ALS could be asked to prove this method against, say, an experimental half pipe setup in their office, versus a V-notch measured flow to the half pipe.

4) In very low flows (e.g. at gauge height 0.022m), discharge uncertainty is dominated by the effects of level measurement uncertainty, whereas for depths above the half pipe, uncertainty is dominated by the uncertainty attributable to the scatter of field gaugings about the rating.

Gauge 300022 – WWL

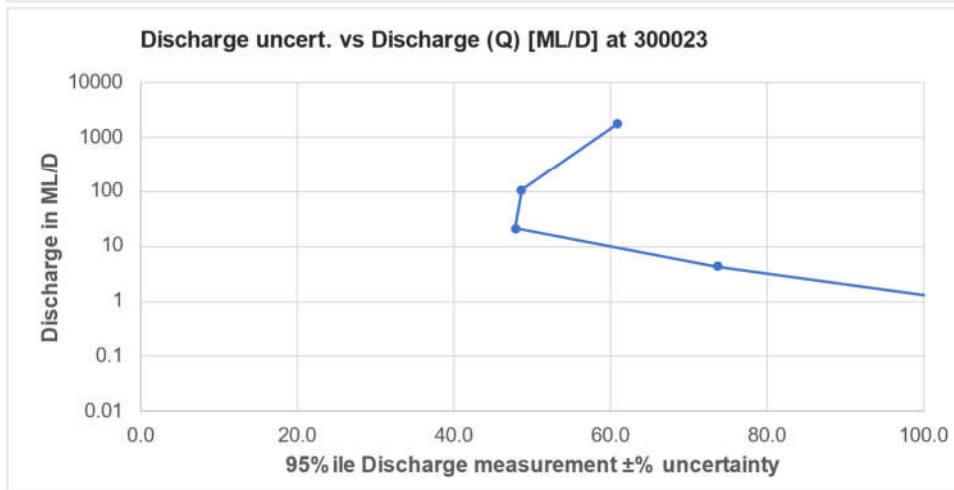
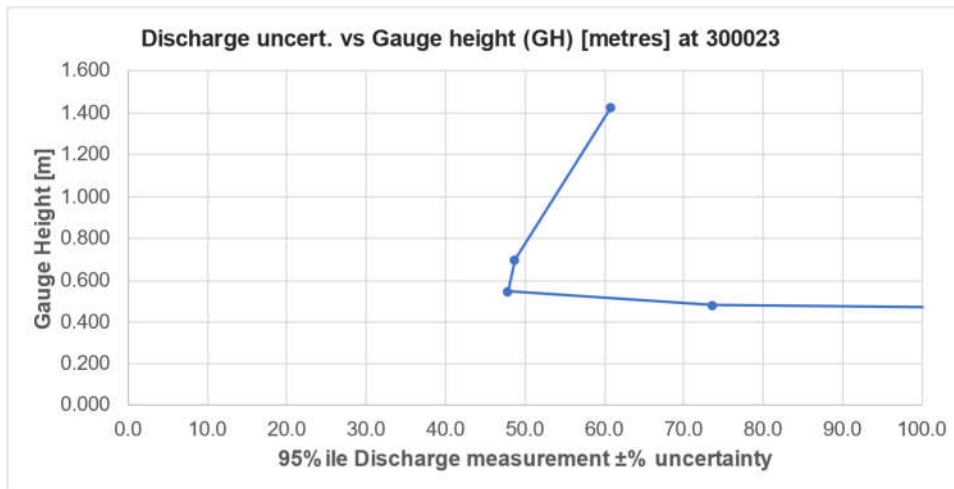


Enviromon developed the blue uncertainty curve, and a modified version of that has been calculated by Watershed HydroGeo (orange series). The high uncertainty at low flow relates to underflow at the (permeable) control. High flow gaugings, possibly via alternative methods, are recommended.

Comments from Enviromon:

- 1) There are 263 rating table change points- which is really excessive for such a site;
It is expected that 10 to 20 would be sufficient, like for site DCU (300023).
- 2) Note the comment in sheet 1 concerning the mismatch between precise CTF level and rounded off CTF level (0.700 vs 0.7085) which needs confirmation/correction.
- 3) Out of the 32 gaugings taken only 18 were suitable for uncertainty analysis, as the other 14 were either no flow or flow so low that a valid gauging could not be taken, OR LEAKAGE PAST THE CONTROL WAS OBSERVED
- 4) Level uncertainty used is the generic +/-8.2mm for a Diver sensor, as there were no comparative field difference readings to analyse
- 5) The very high uncertainty due to gauging deviations is mostly due to "leakage" effects- for example the three gaugings taken on 24/10/2019 which each showed more than 600% deviation mismatch. Using these to quantify low-flow uncertainty alone (and not moderate/high flow uncertainty) is the basis for the orange chart series.

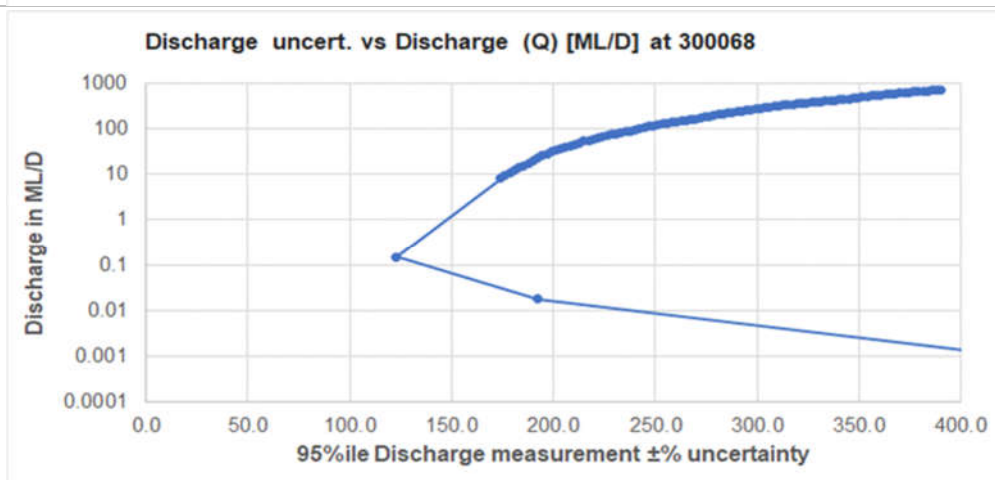
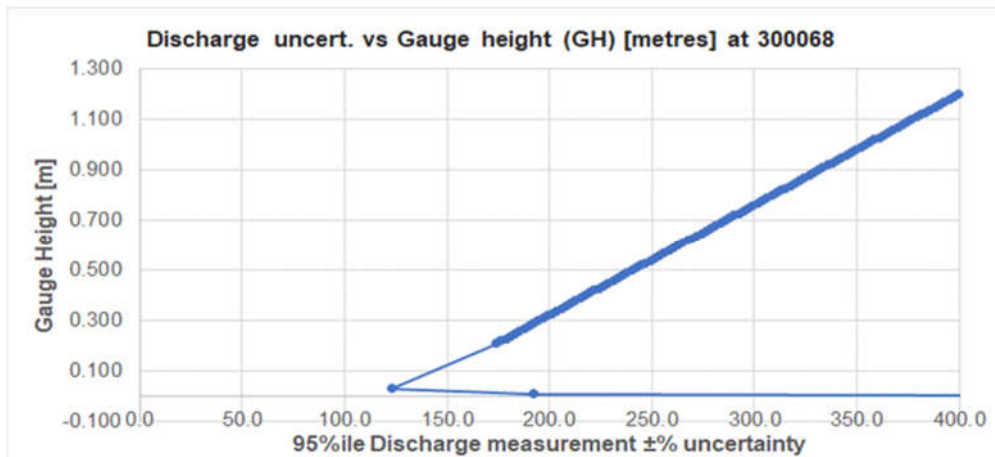
Gauge 300023 – DCU



Comments from Enviromon:

- 1) Out of the 29 gauging records, only 9 were valid for this analysis, due to 20 being either zero flow or so low that no valid measurement could be taken
- 2) Gauging 28 on 4/3/21 had the highest deviation- South 32 noted leakage:- Serviced by S32. Pygmy av. = 0.05844 MLD and comment of 'approx. 10% leakage'
 Could this be a general issue with this site?
- 3) Level uncertainty used is the generic +/-8.2mm for a Diver sensor, as there were no comparative field difference readings to analyse

Gauge 300098 – DCS2



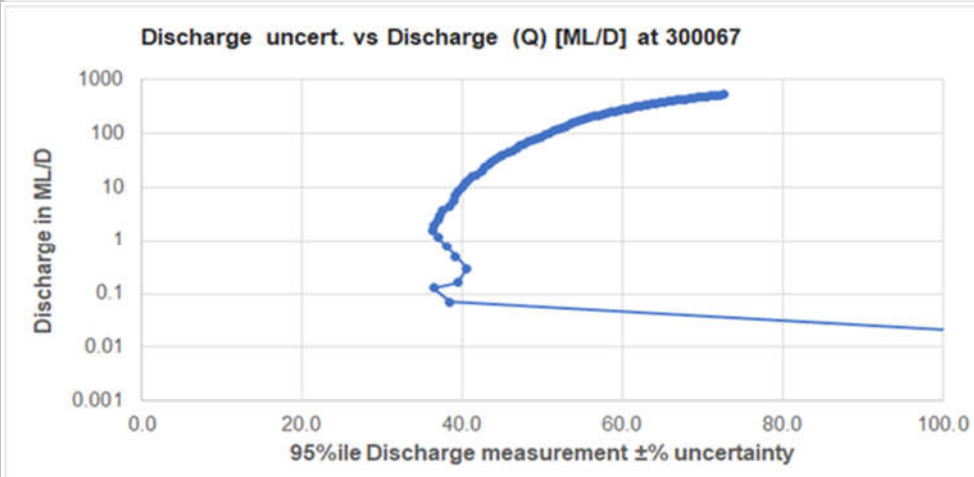
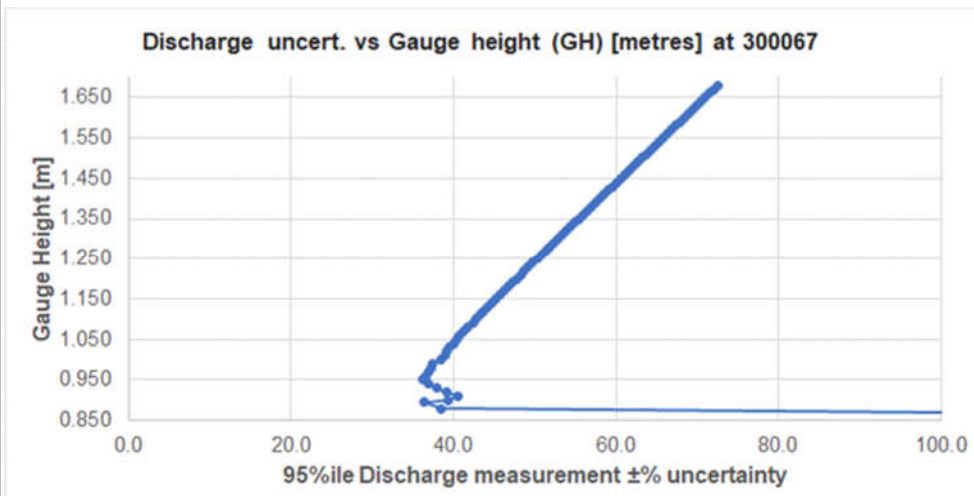
Enviromon (2022) noted that there is not a tight reliable and unique rating relationship between water level and discharge, ranging from -36% to +175% different from the rating. Based on this analysis, Enviromon has sent questions to ALS relating to the quality of gaugings used to develop the rating curve and the rock bar control.

High flow gaugings, possibly via alternative methods, are required to improve high flow.

Comments from Enviromon:

This site is affected by intermittent leakage past the control, which causes gauging deviations of greater than 100% in at least 2 of the 7 valid gaugings, and hence very high “scatter of gaugings” rating uncertainty. If this leakage, which could be related to mining subsidence and fracturing, cannot be located and sealed, then relocation should be considered (or grouting beneath the control?).

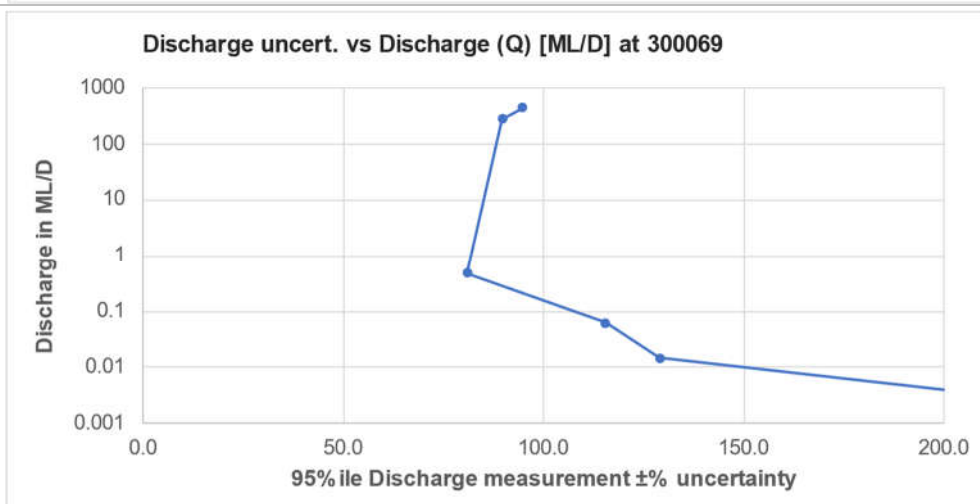
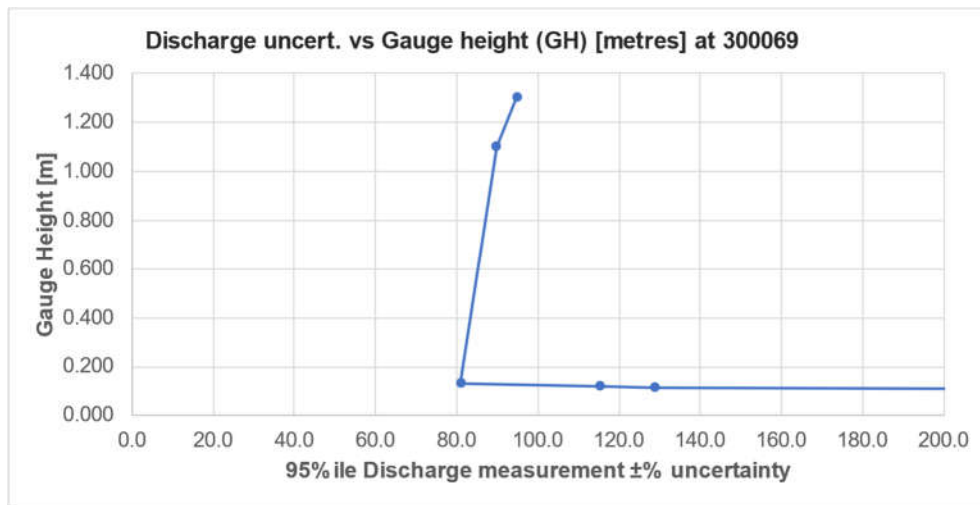
Gauge 300067 – DC13S1



Comments from Enviromon:

Enviromon (2022) noted that there is not a tight reliable and unique rating relationship between water level and discharge, ranging from -29% to +44% different from the rating. The rock bar control makes low flow measurement very sensitive to level measurement uncertainty. Enviromon has sent questions to ALS relating to the quality of gaugings used for the rating curve and other issues.

Gauge 300069 – WC21S1



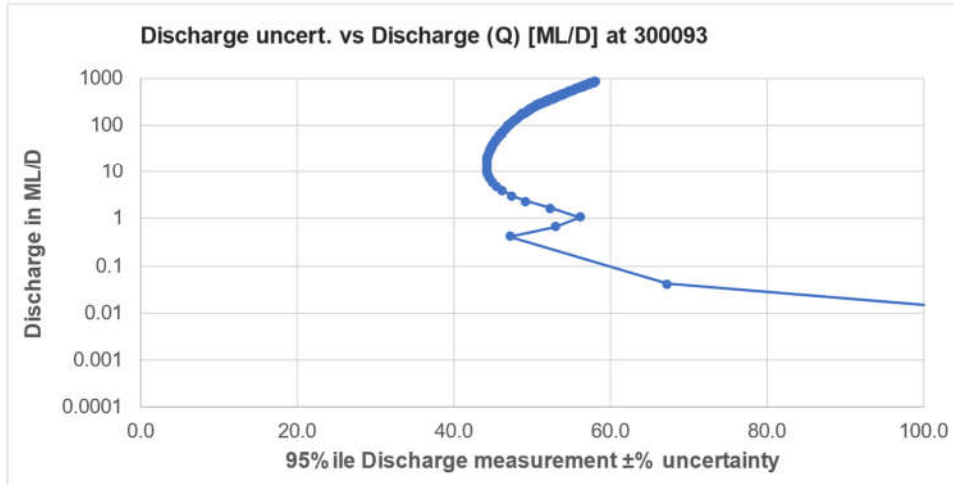
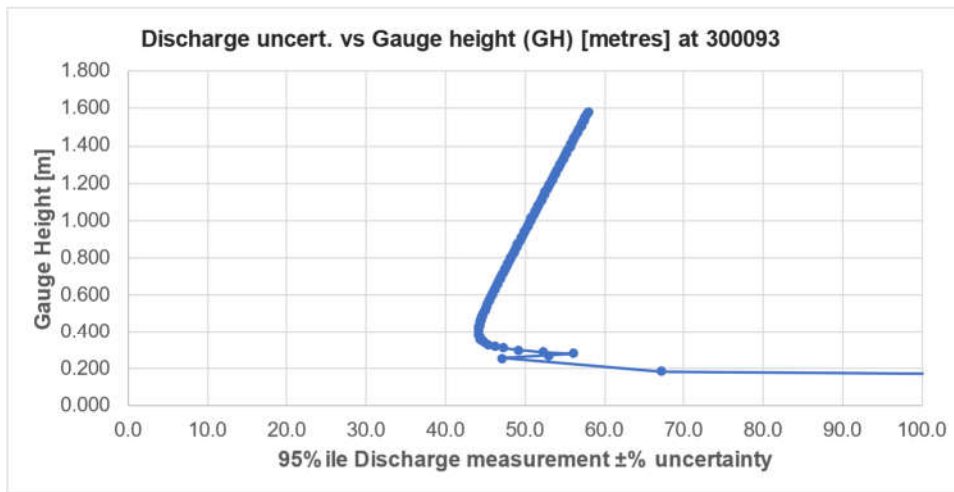
Comments from Enviromon:

There is not a tight reliable and unique rating relationship between water level and discharge, ranging from - 76% to +121% different from the rating. Intermittent leakage under the control is suspected as the main cause of this scatter of gaugings (which could be mining-related?).

Based on this analysis, Enviromon has sent questions to ALS relating to the quality of gaugings used to develop the rating curve and the rock bar control. If the location of the low flow leakage cannot be found then site relocation should be considered (or grouting beneath the control?).

High flow gaugings are required to improve high flow rating.

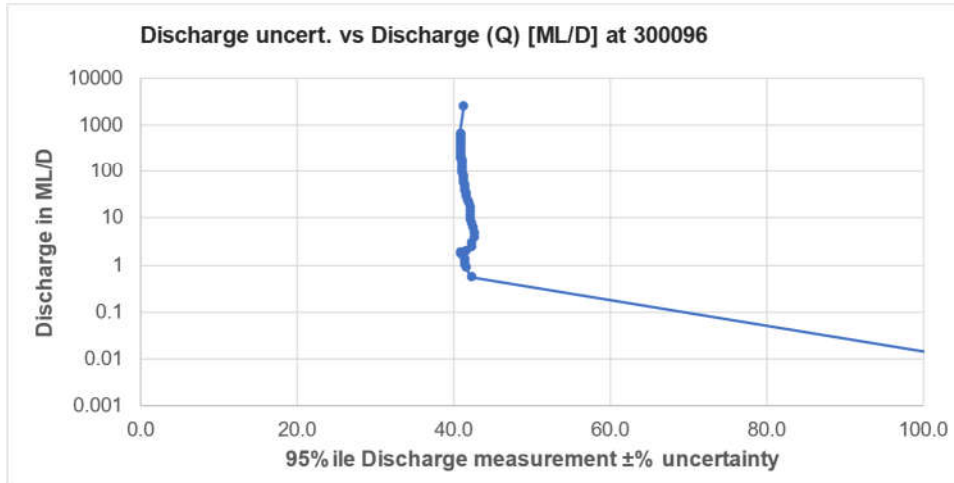
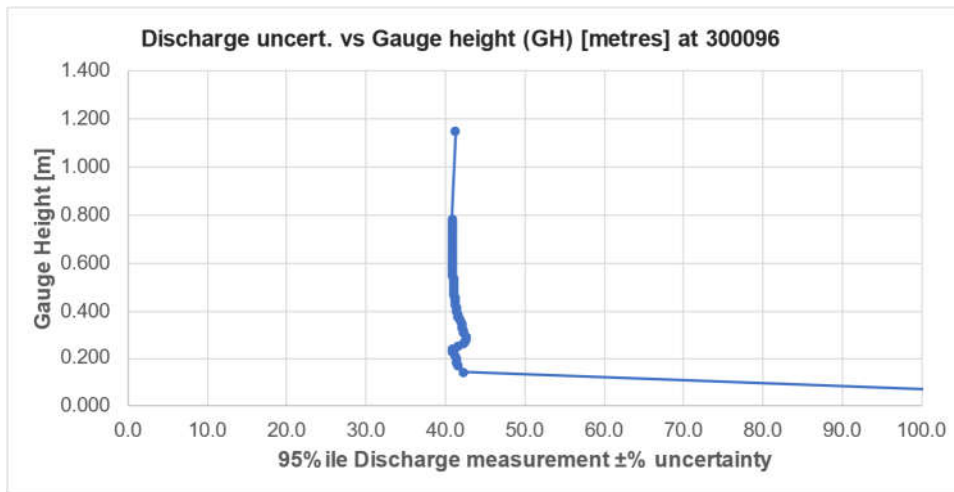
Gauge 30022A – ND1



Comments from Enviromon:

- 1) There are 139 rating table change points- which seems excessive for such a site;
- 2) The last 2 gaugings are done with pygmy meter in half pipe, and show similar rating differences (e.g. circa 30%) as volumetrics
- 3) On sheet 5 "Gauge height uncertainty"- the calculated (from field differences) uncertainty of 0.007238 is greater than the generic Orpheus value of 0.0066, then raise this as an issue with ALS
- 4) The sudden zig-zag in uncertainty at 0.255m is because this is when flow starts to break out of the half pipe and spread wide across the bund, with rapid increase in discharge for small increase in level.
This slowly recovers as level rises further and discharge does not increase so much per mm of level rise.

Gauge 300096 – CR36



The use of an artificial control (halfpipe) leads to generally lower uncertainty than for the natural control sites (e.g. DCS2, WC21).

Comments from Enviromon:

- 1) There are 68 rating table change points- which seems excessive for such a site;
- 2) Most of the gaugings are at very small flowrates and depths, making the uncertainty of individual gaugings (and hence the rating curve plotted through them) to also be higher;
- 3) Note the existence of the one high gauging at 0.91m, quite a rarity- this stops the extra uncertainty of rating curve extension for most of the range of interest.
- 4) The % deviation varies of similar magnitude whether gauging is volumetric or pygmy meter in half pipe

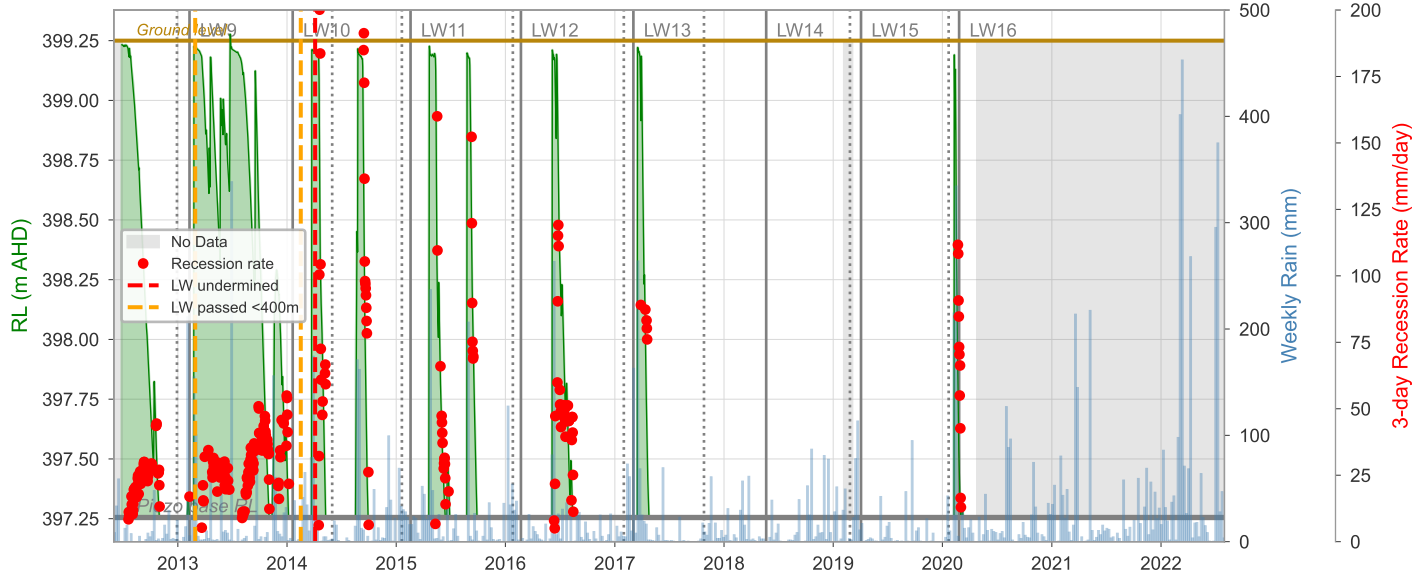
C6) Summary charts illustrating 'raw' flow data and processed flow data used for TARP Assessments

Insert:

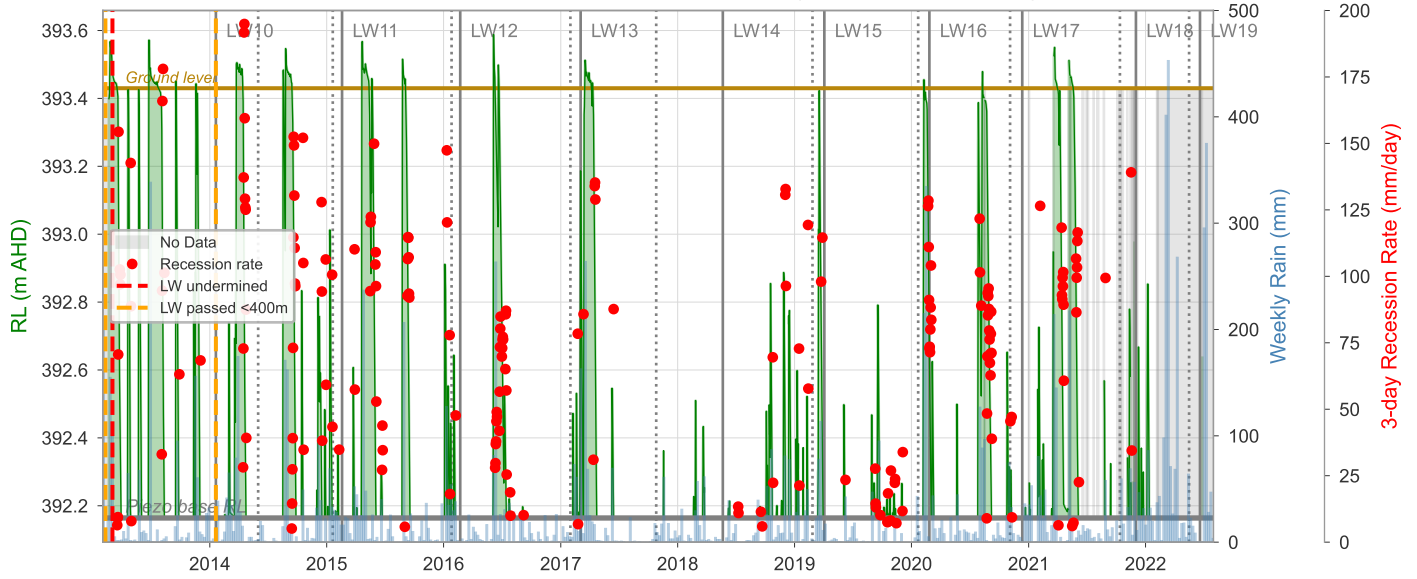
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Appendix D: Shallow groundwater hydrographs

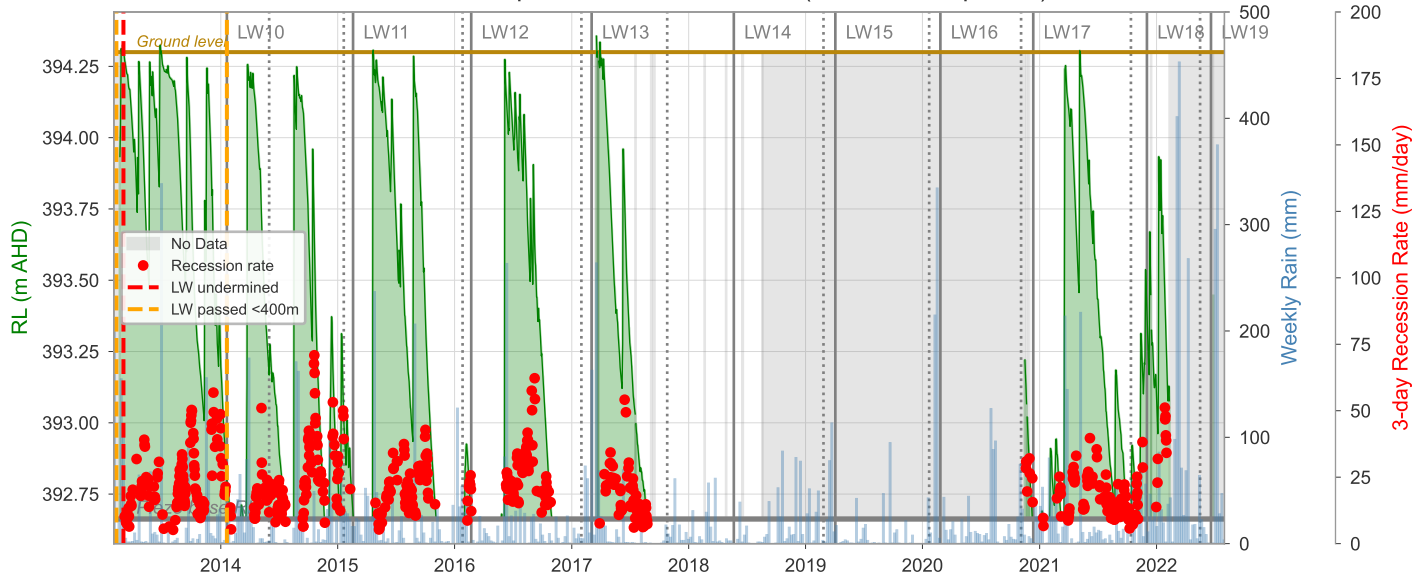
Dendrobium Swamp 01A: Piezometer 01 (Within swamp EEC)



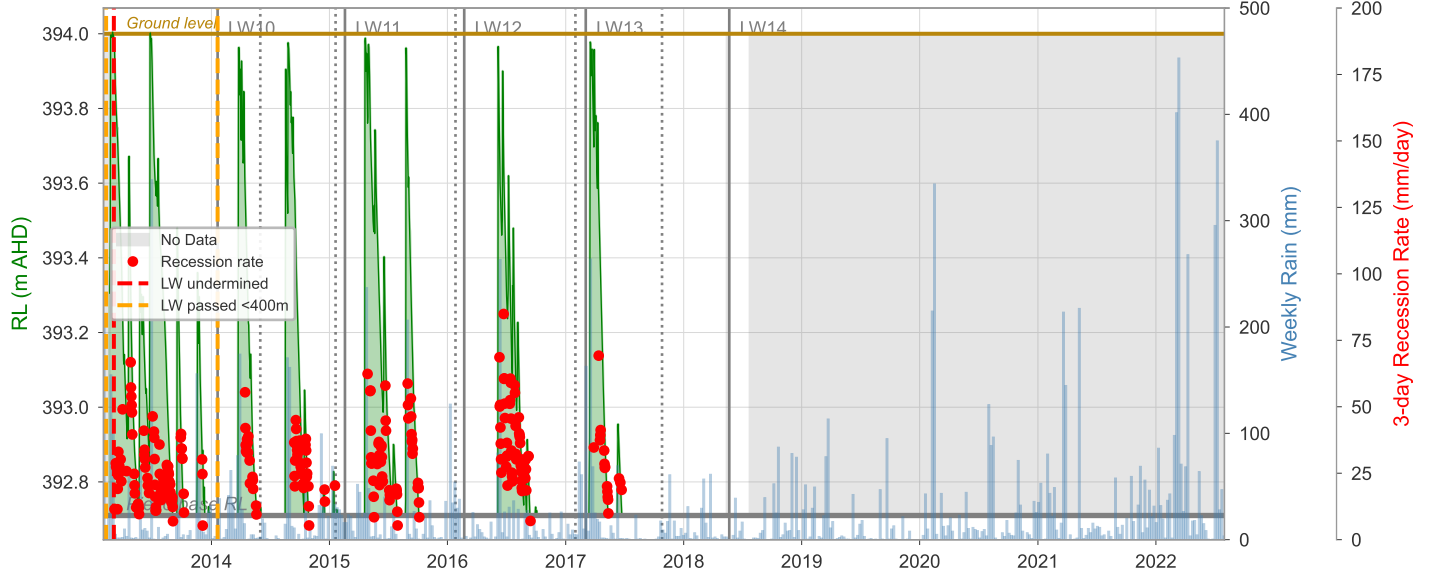
Dendrobium Swamp 01B: Piezometer 02IV (Within swamp EEC)



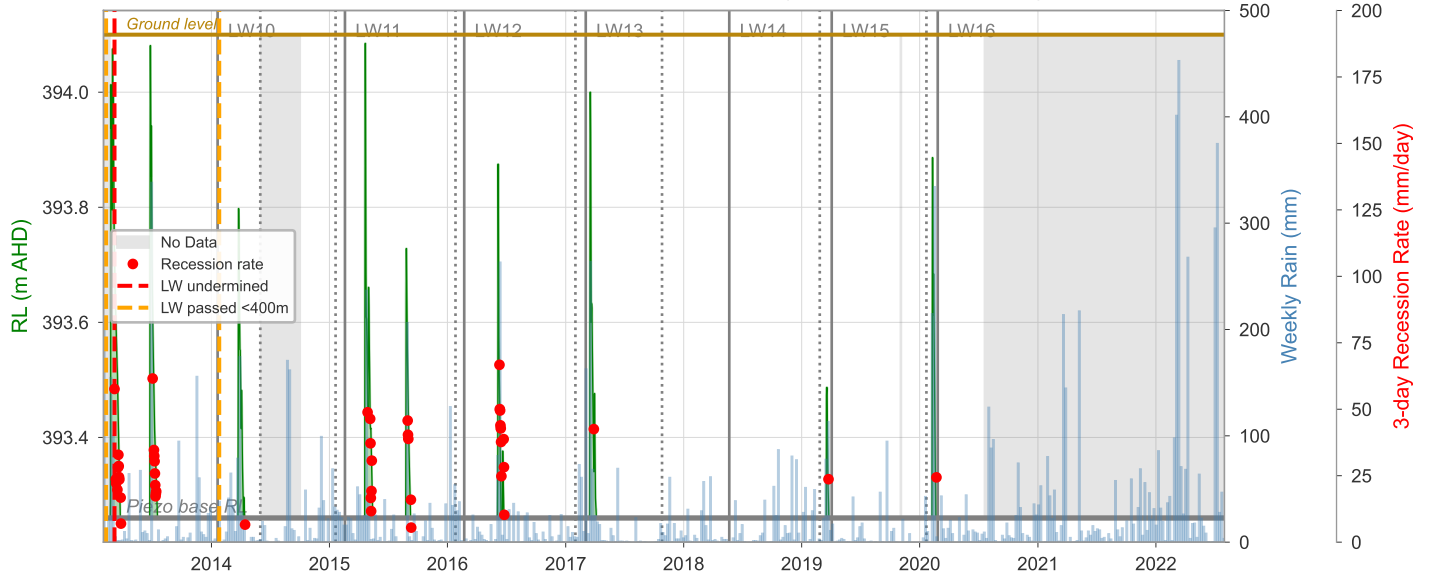
Dendrobium Swamp 01B: Piezometer 02III (Within swamp EEC)



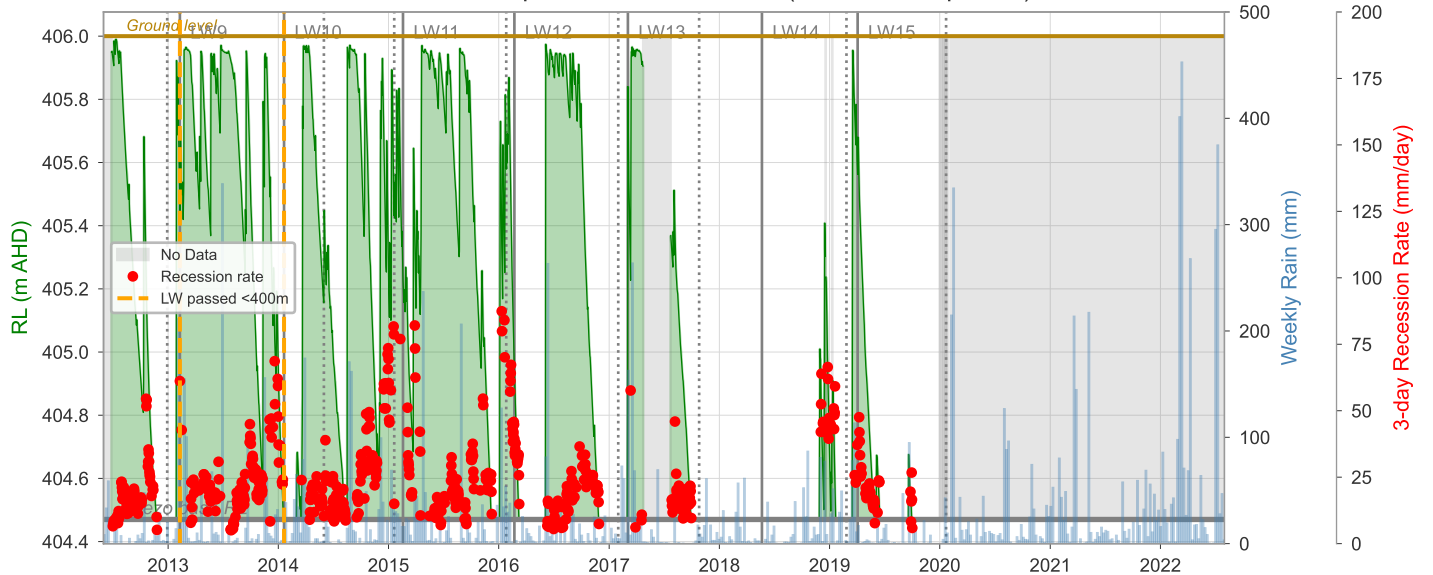
Dendrobium Swamp 01B: Piezometer 02II (Within swamp EEC)



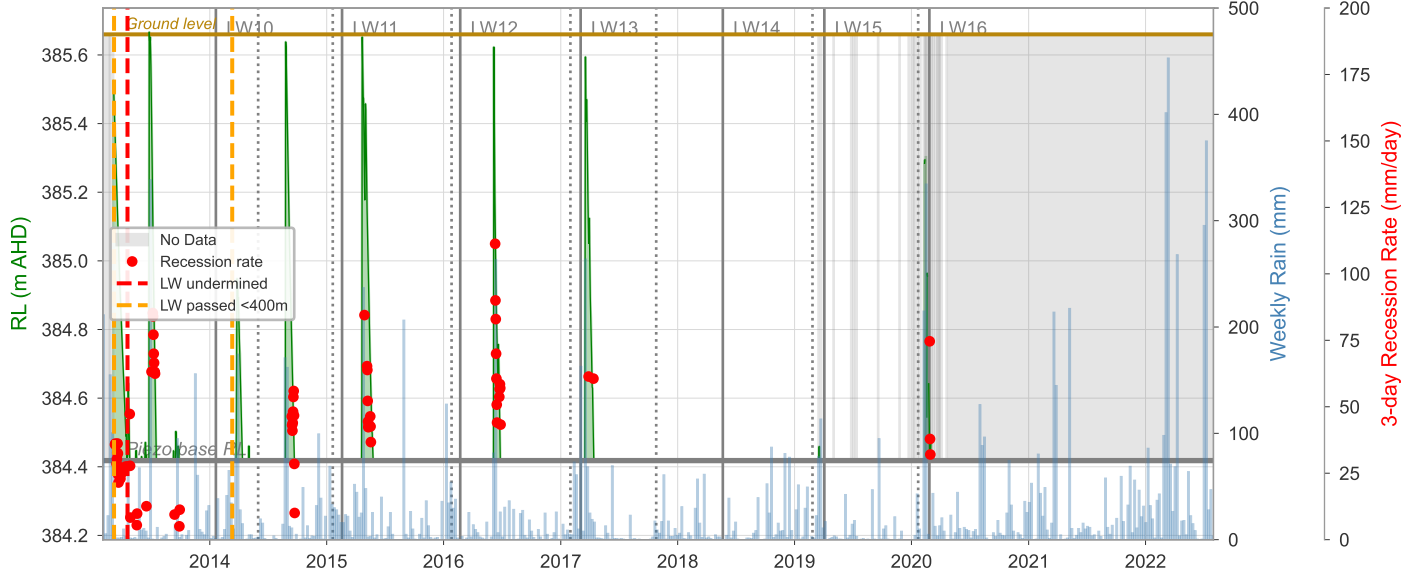
Dendrobium Swamp 01B: Piezometer 02I (Outside swamp EEC)



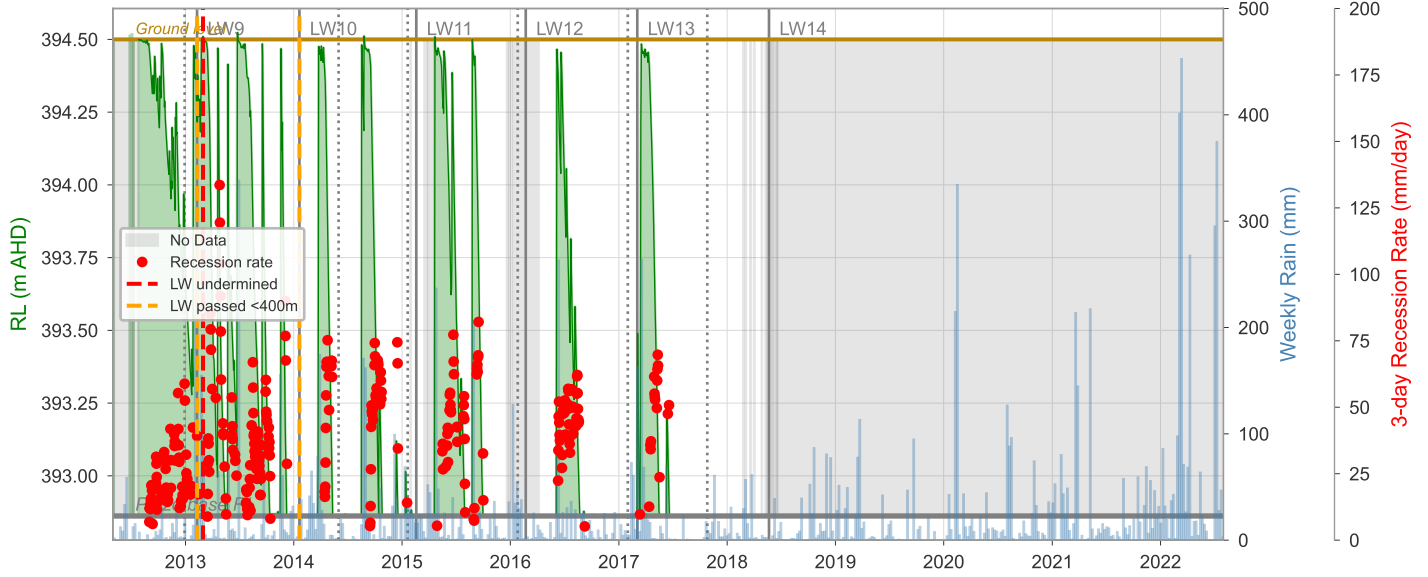
Dendrobium Swamp 01B: Piezometer 01 (Within swamp EEC)



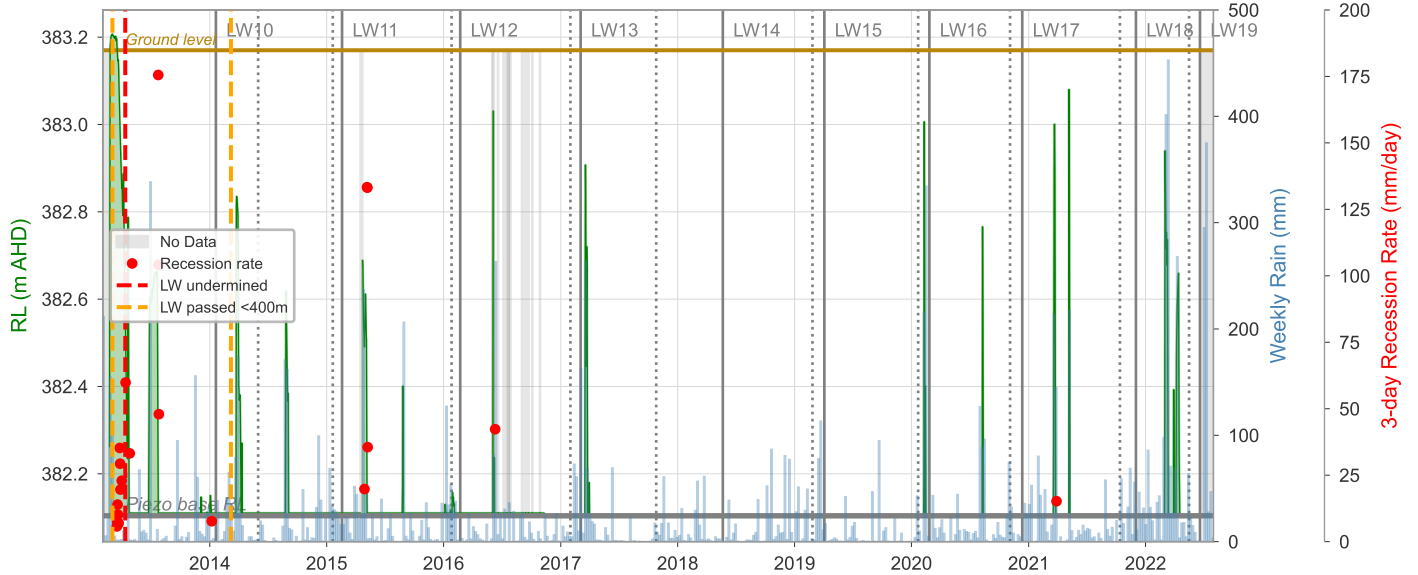
Dendrobium Swamp 01A: Piezometer 04V (Outside swamp EEC)



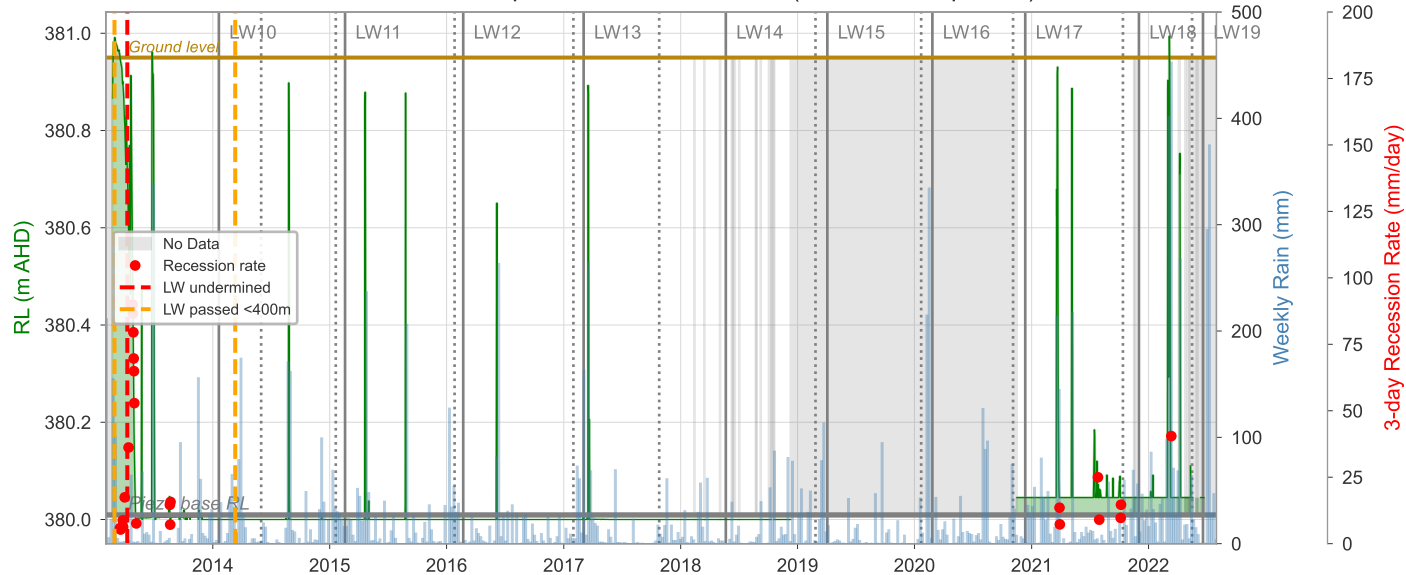
Dendrobium Swamp 01B: Piezometer 02 (Within swamp EEC)



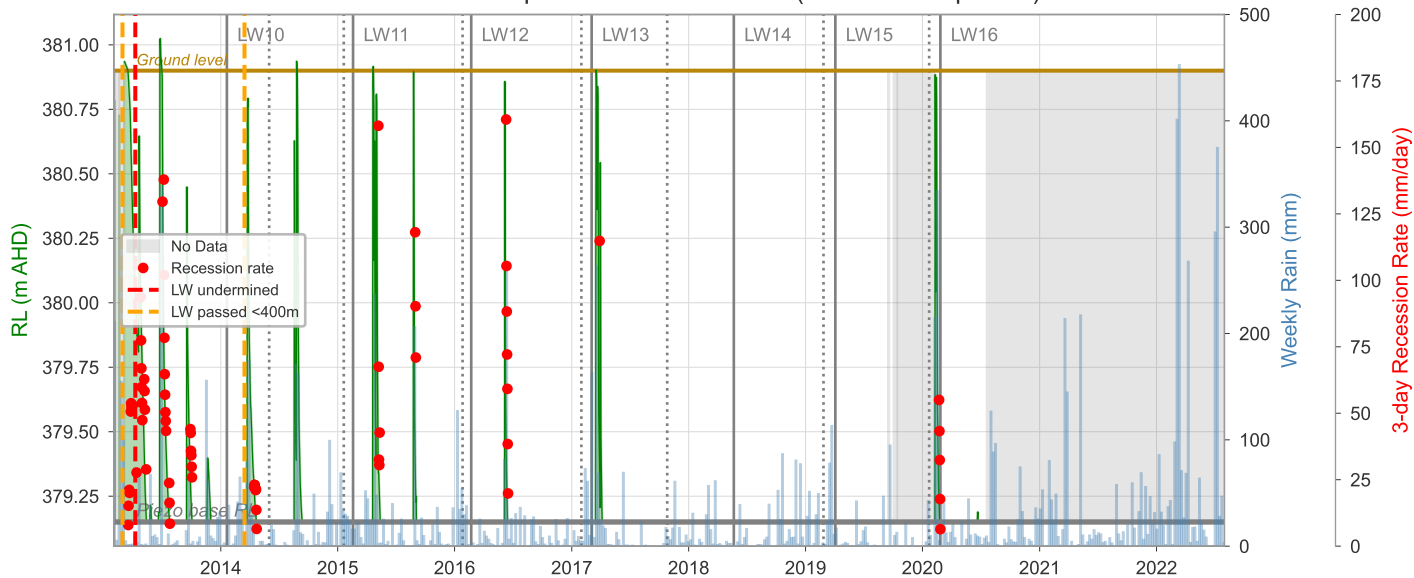
Dendrobium Swamp 01A: Piezometer 04III (Within swamp EEC)



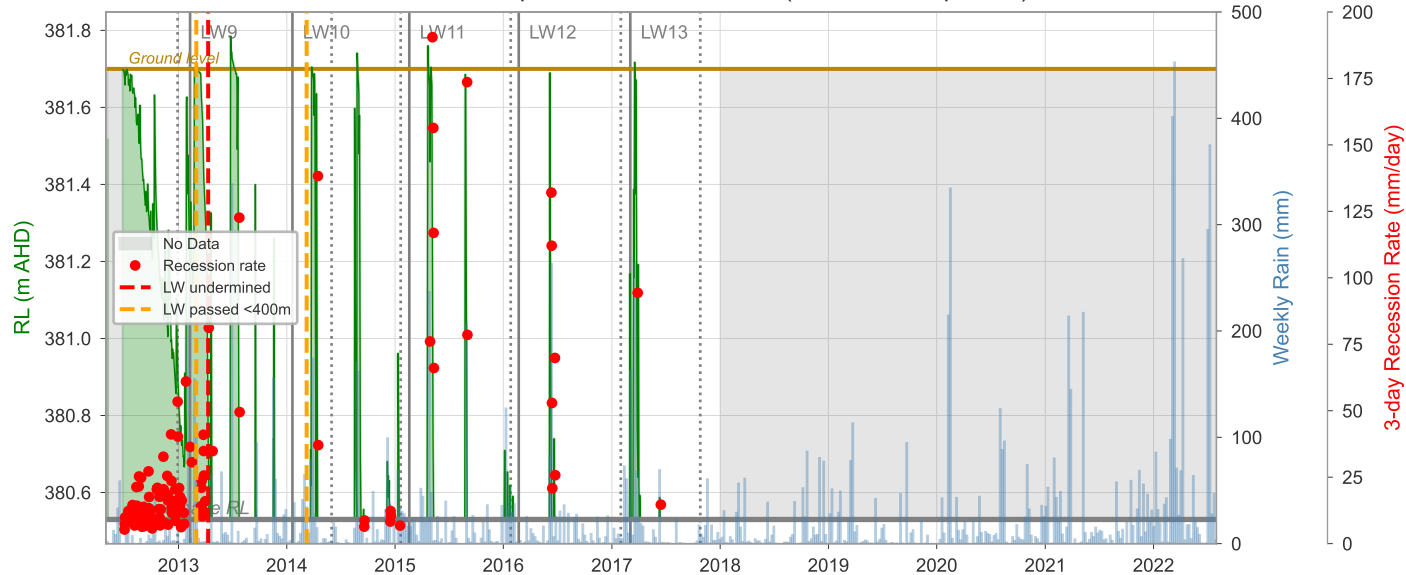
Dendrobium Swamp 01A: Piezometer 04II (Within swamp EEC)



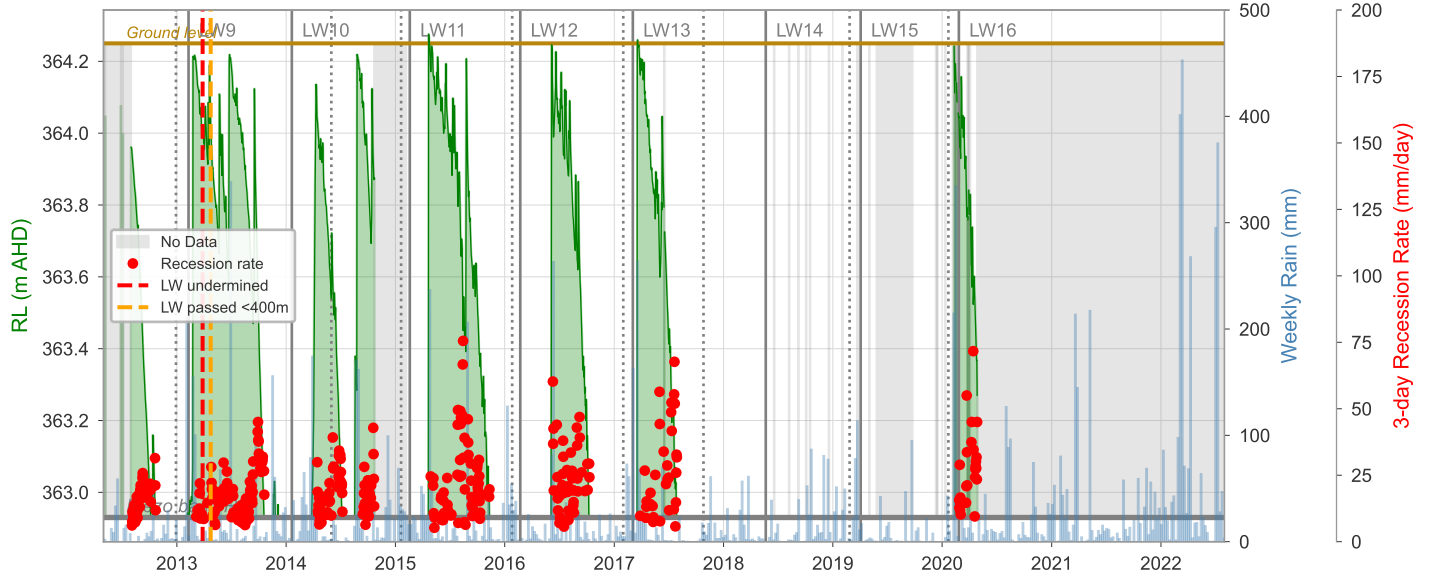
Dendrobium Swamp 01A: Piezometer 04I (Within swamp EEC)



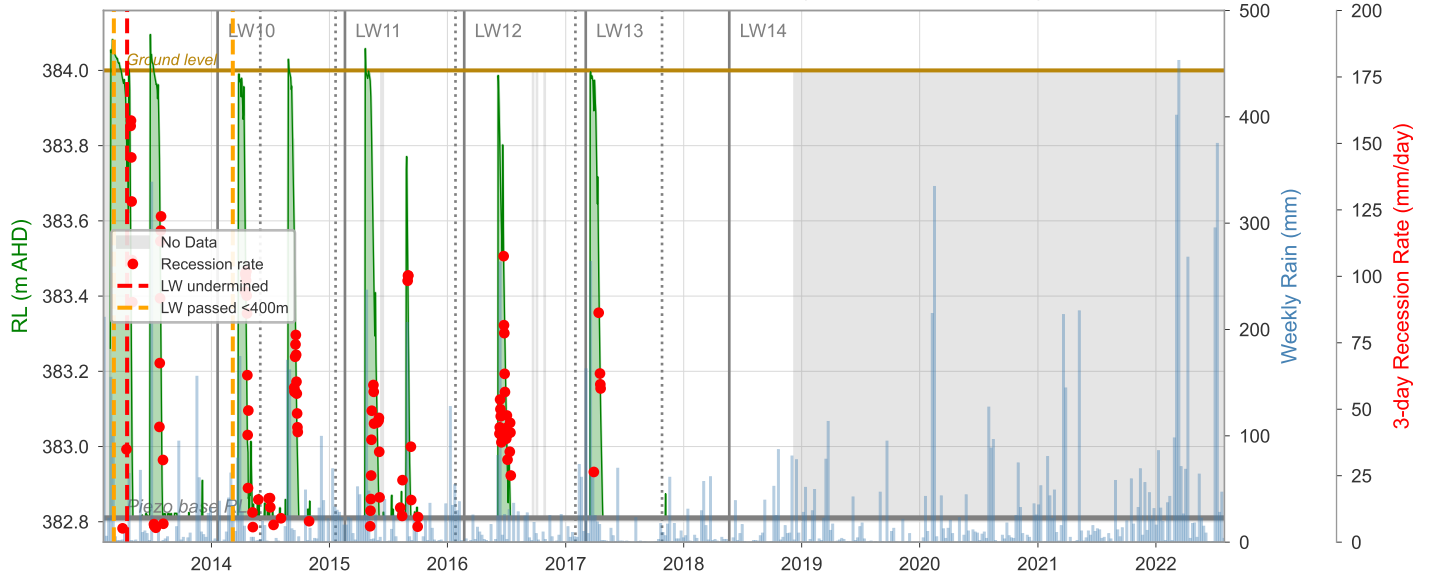
Dendrobium Swamp 01A: Piezometer 04 (Within swamp EEC)



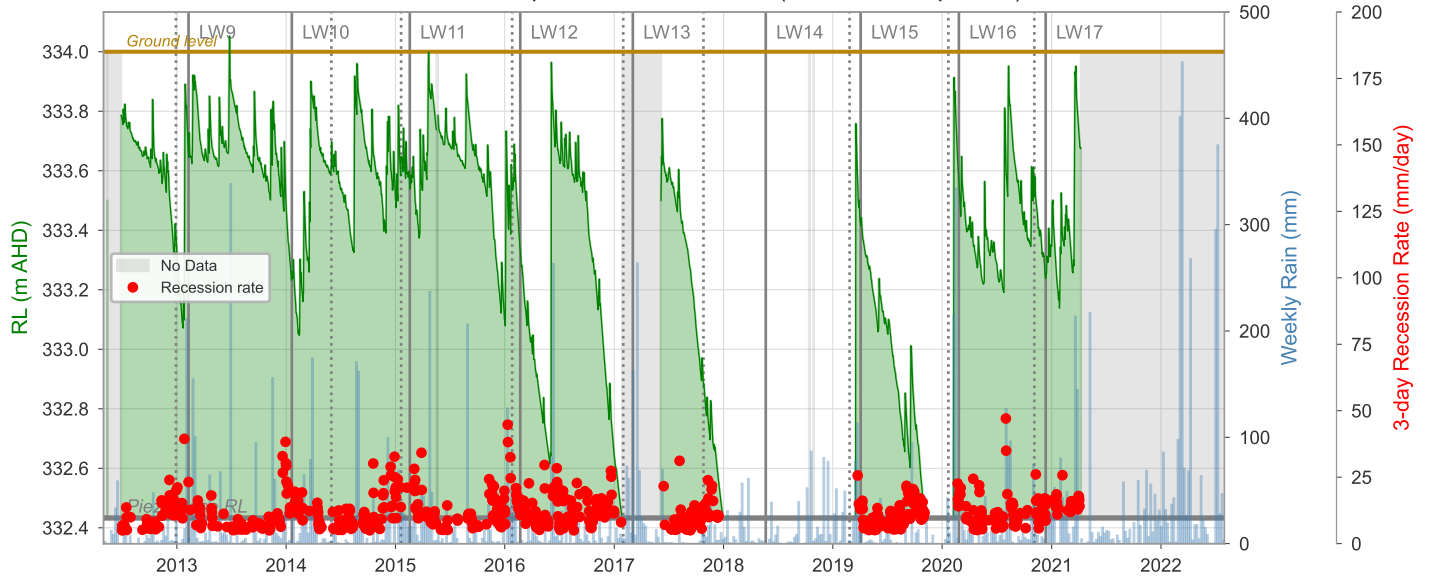
Dendrobium Swamp 01A: Piezometer 02 (Within swamp EEC)



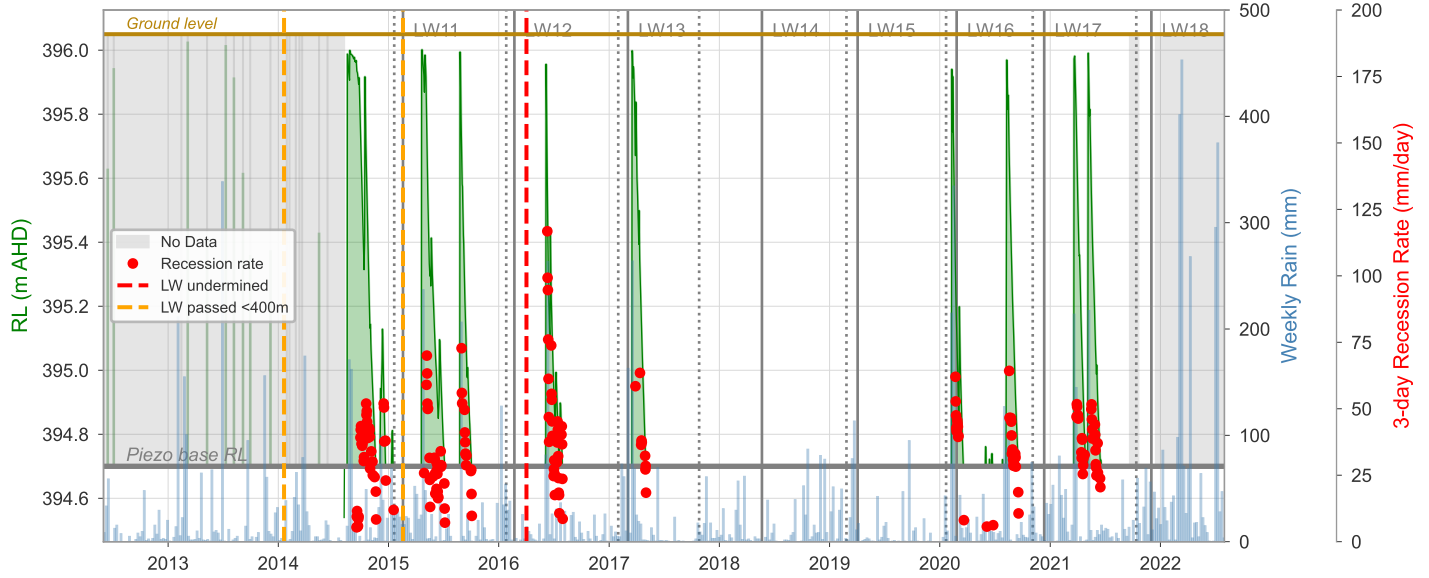
Dendrobium Swamp 01A: Piezometer 04IV (Within swamp EEC)



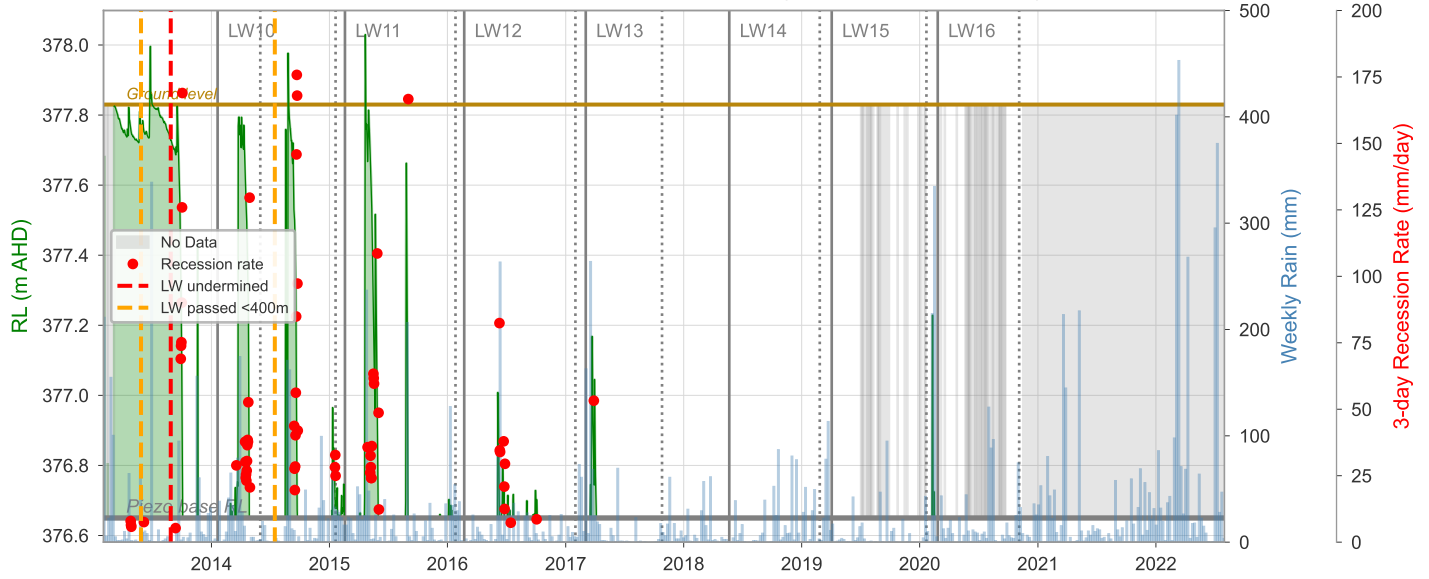
Dendrobium Swamp 02: Piezometer 01 (Within swamp EEC)



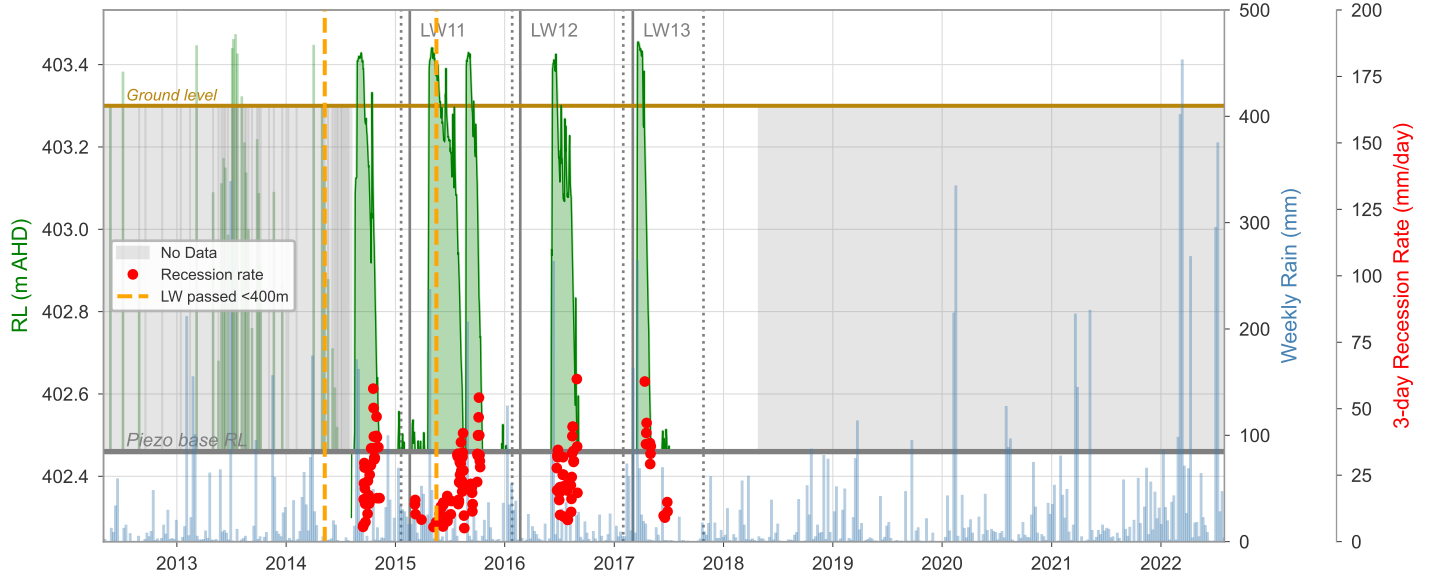
Dendrobium Swamp 03: Piezometer 01 (Within swamp EEC)



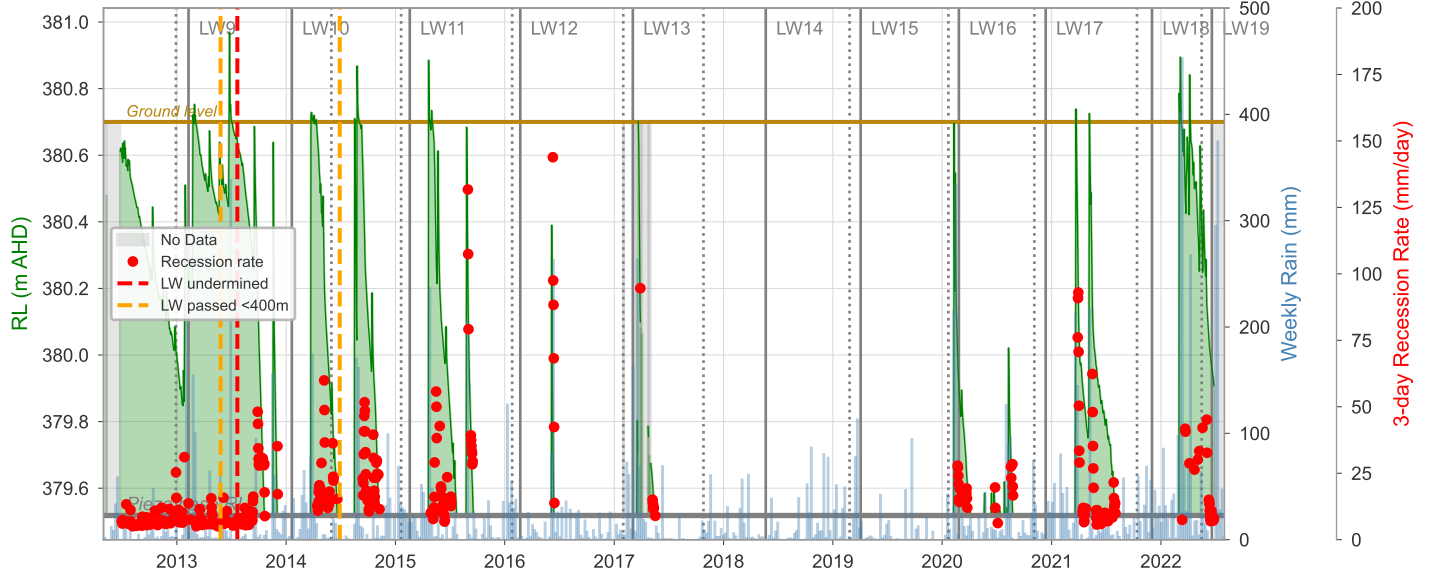
Dendrobium Swamp 05: Piezometer 03III (Outside swamp EEC)



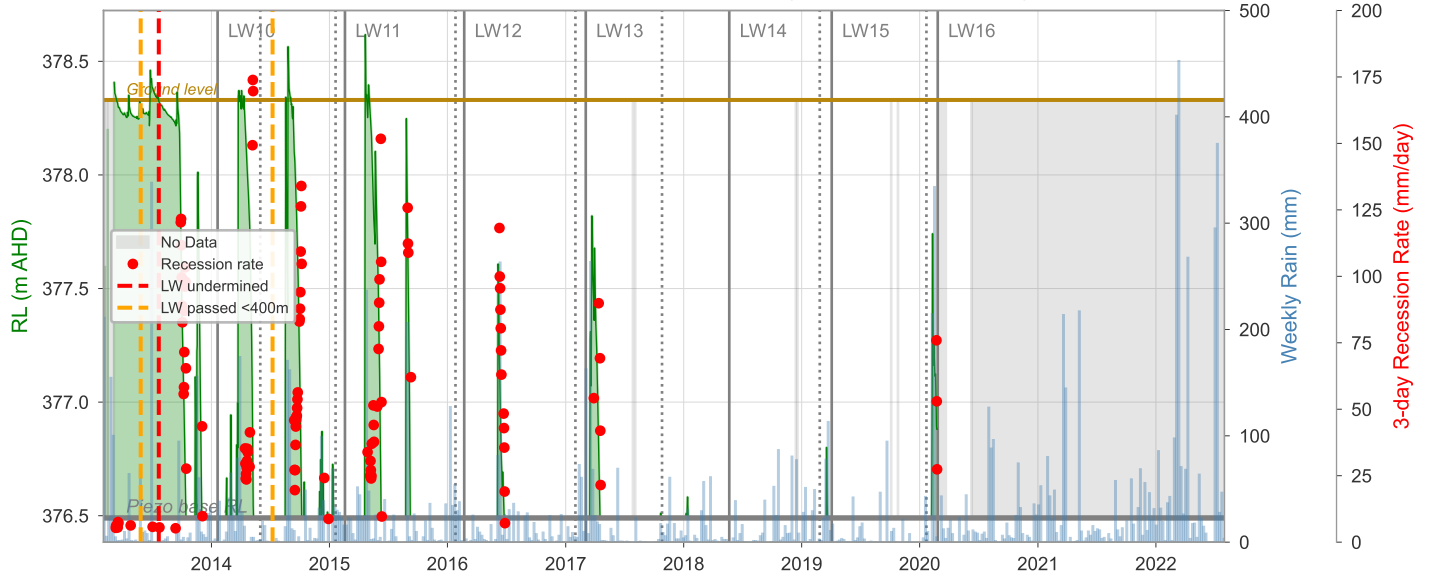
Dendrobium Swamp 05: Piezometer 05 (Within swamp EEC)



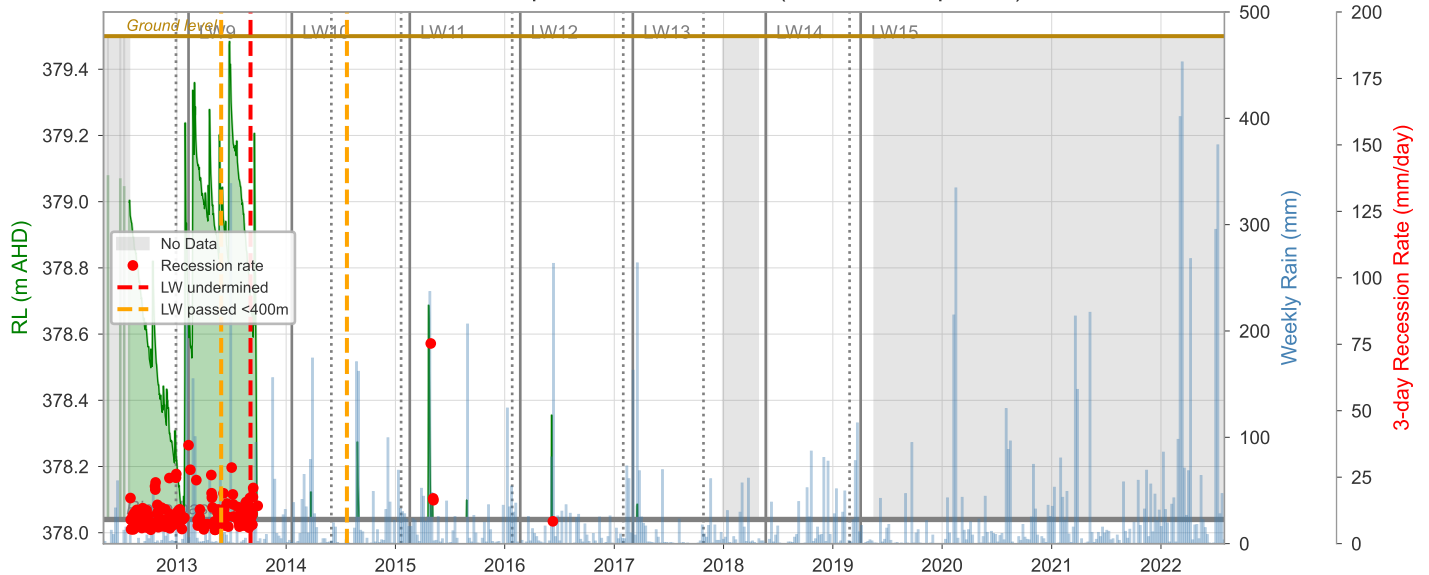
Dendrobium Swamp 05: Piezometer 04 (Within swamp EEC)



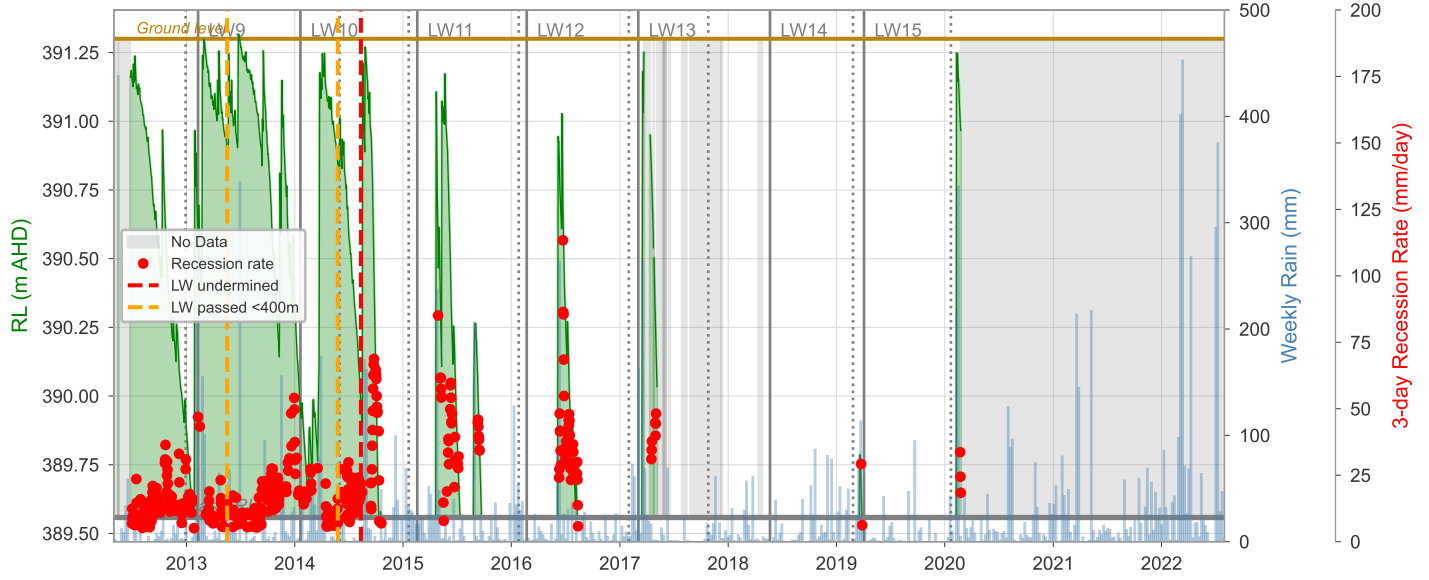
Dendrobium Swamp 05: Piezometer 03II (Within swamp EEC)



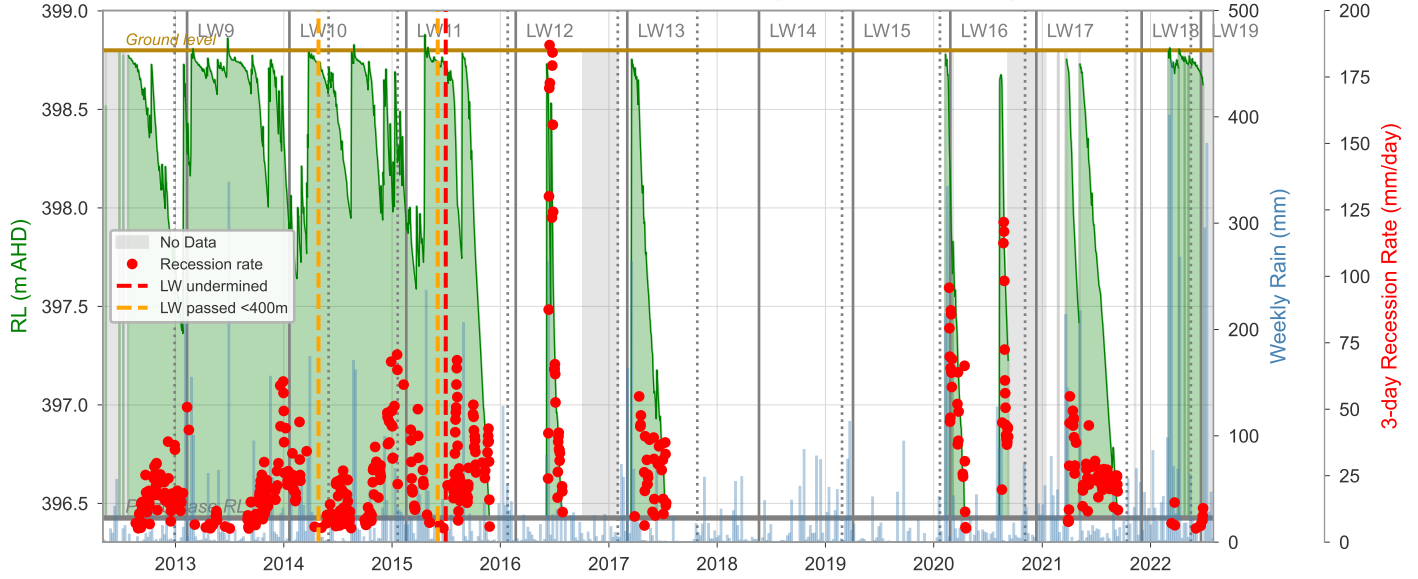
Dendrobium Swamp 05: Piezometer 03 (Within swamp EEC)



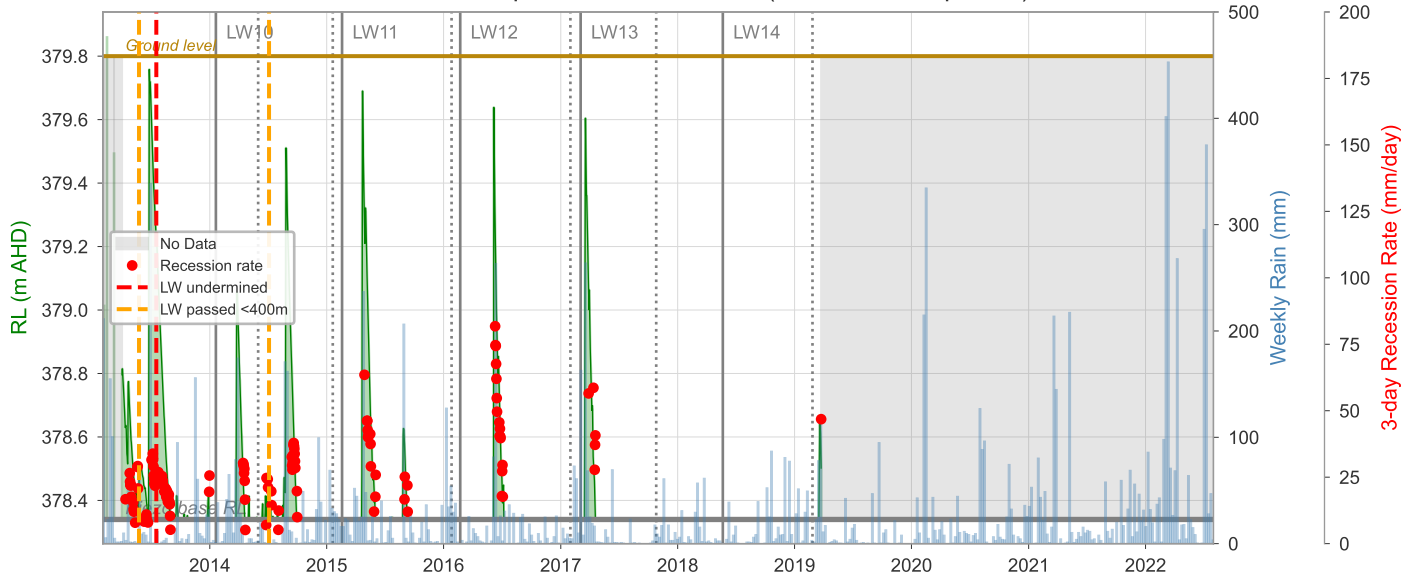
Dendrobium Swamp 05: Piezometer 02 (Within swamp EEC)



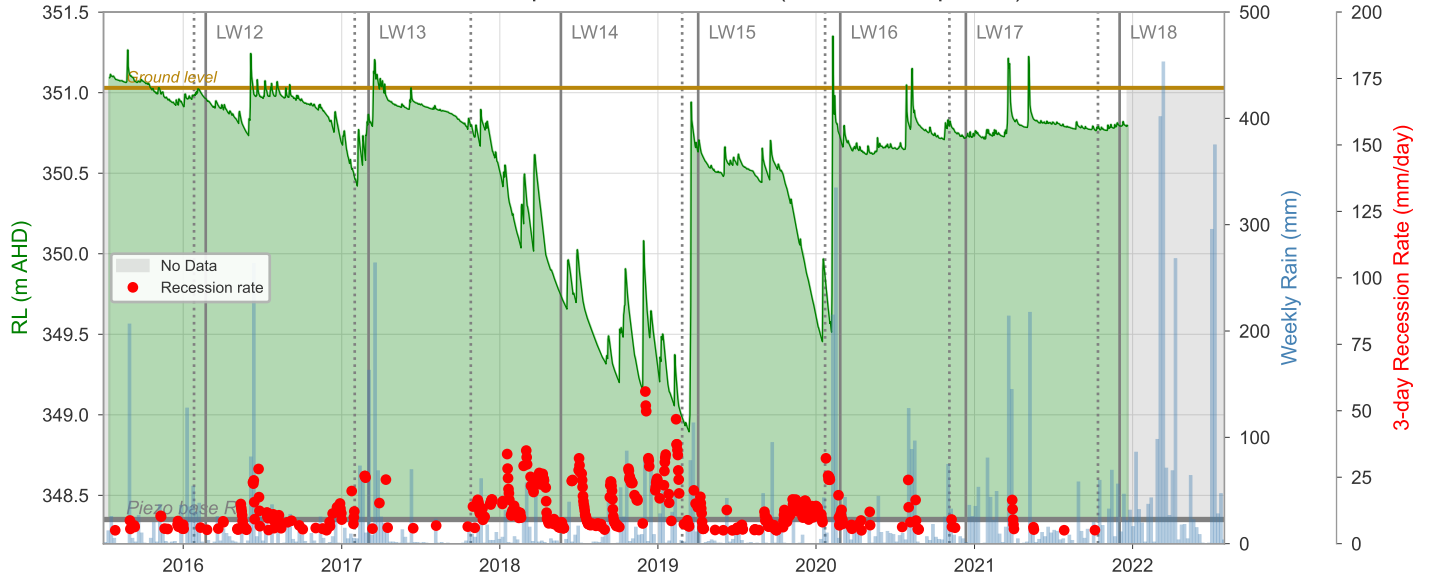
Dendrobium Swamp 05: Piezometer 01 (Within swamp EEC)



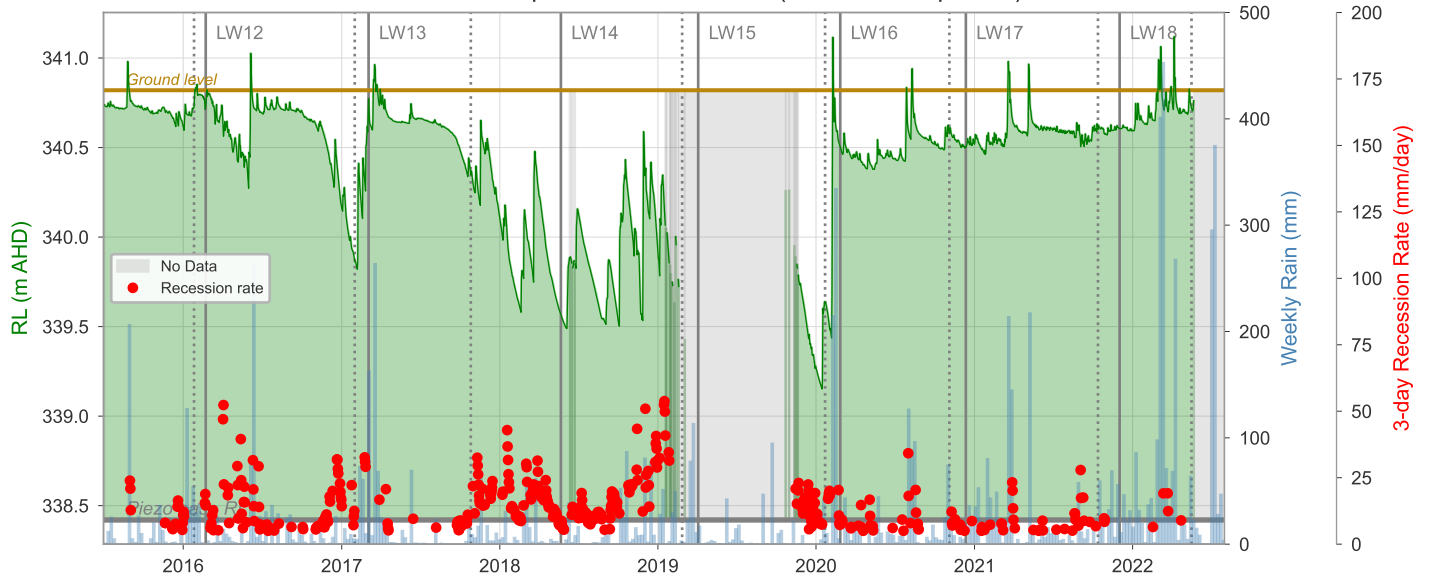
Dendrobium Swamp 05: Piezometer 03I (Outside swamp EEC)



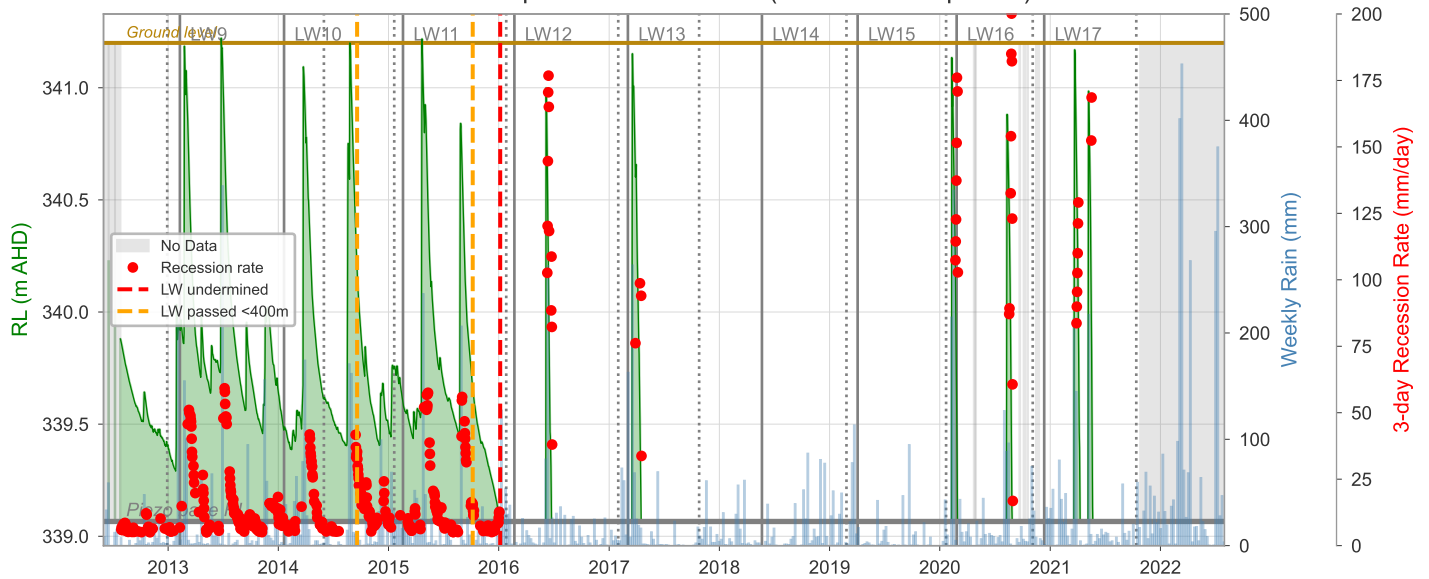
Dendrobium Swamp 07: Piezometer 05 (Within swamp EEC)



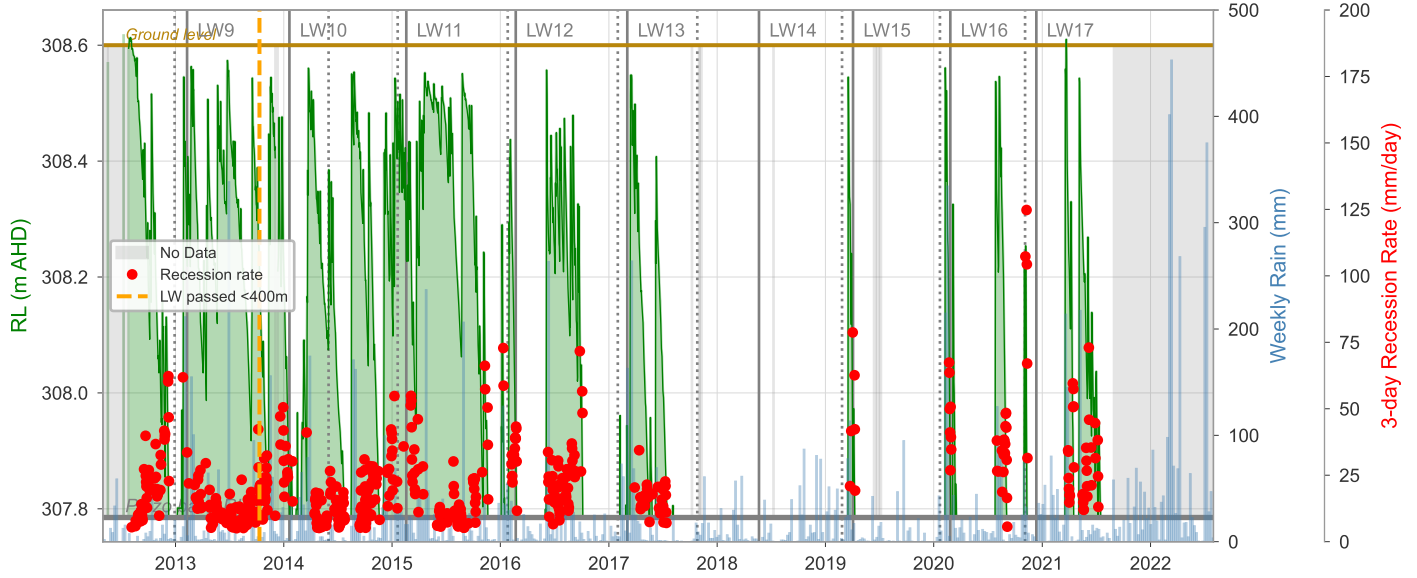
Dendrobium Swamp 07: Piezometer 06 (Within swamp EEC)



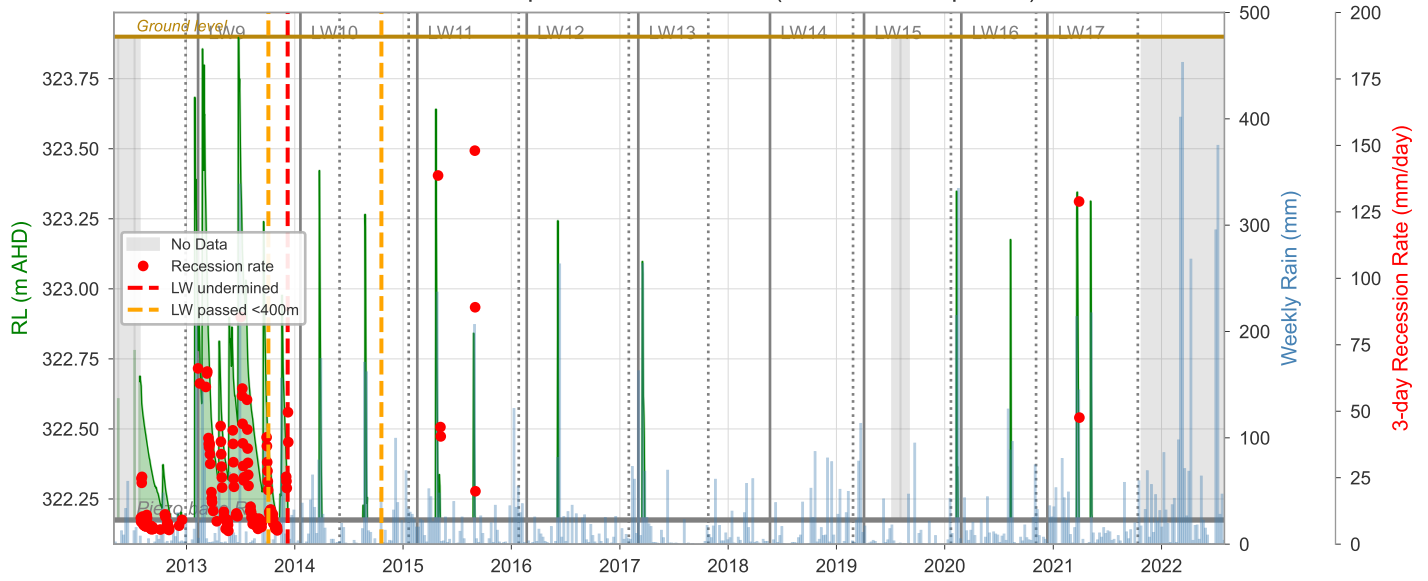
Dendrobium Swamp 08: Piezometer 01 (Outside swamp EEC)



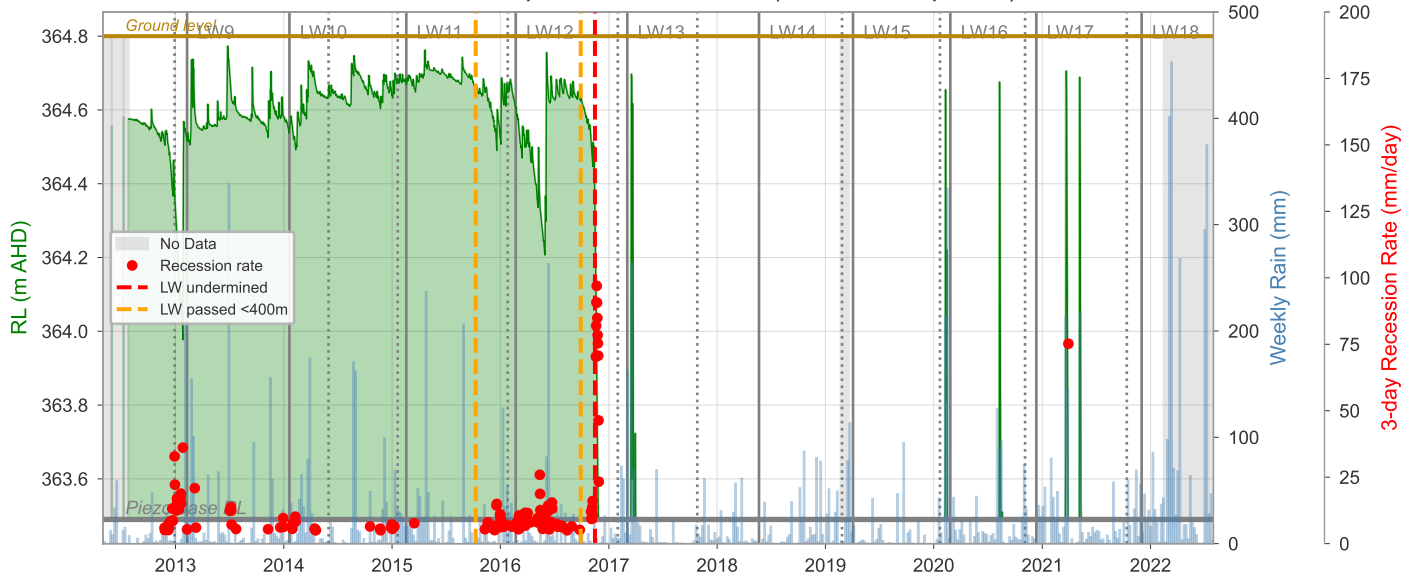
Dendrobium Swamp 08: Piezometer 02 (Outside swamp EEC)



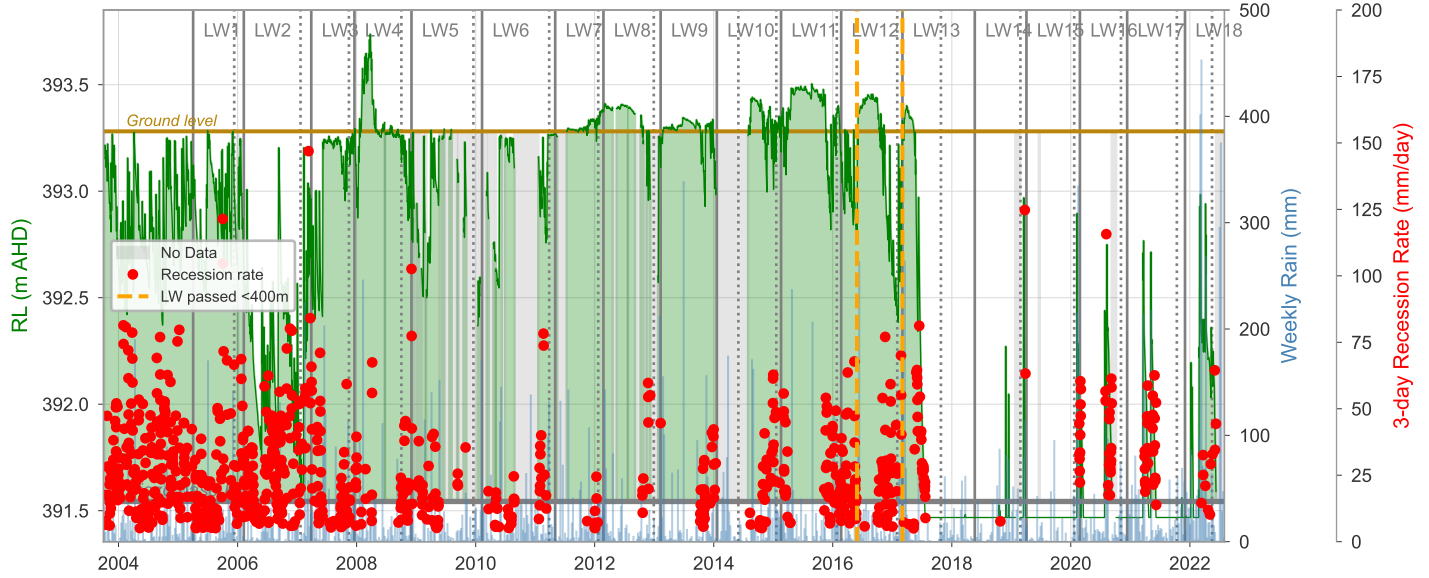
Dendrobium Swamp 08: Piezometer 04 (Outside swamp EEC)



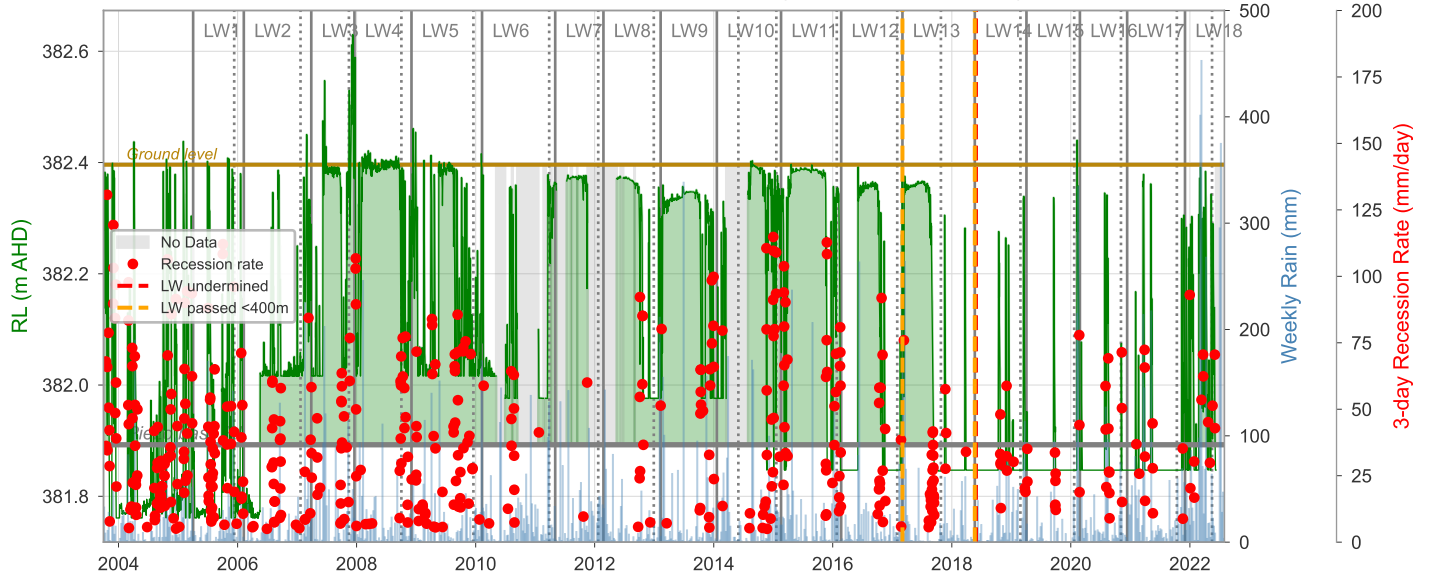
Dendrobium Swamp 10: Piezometer 01 (Within swamp EEC)



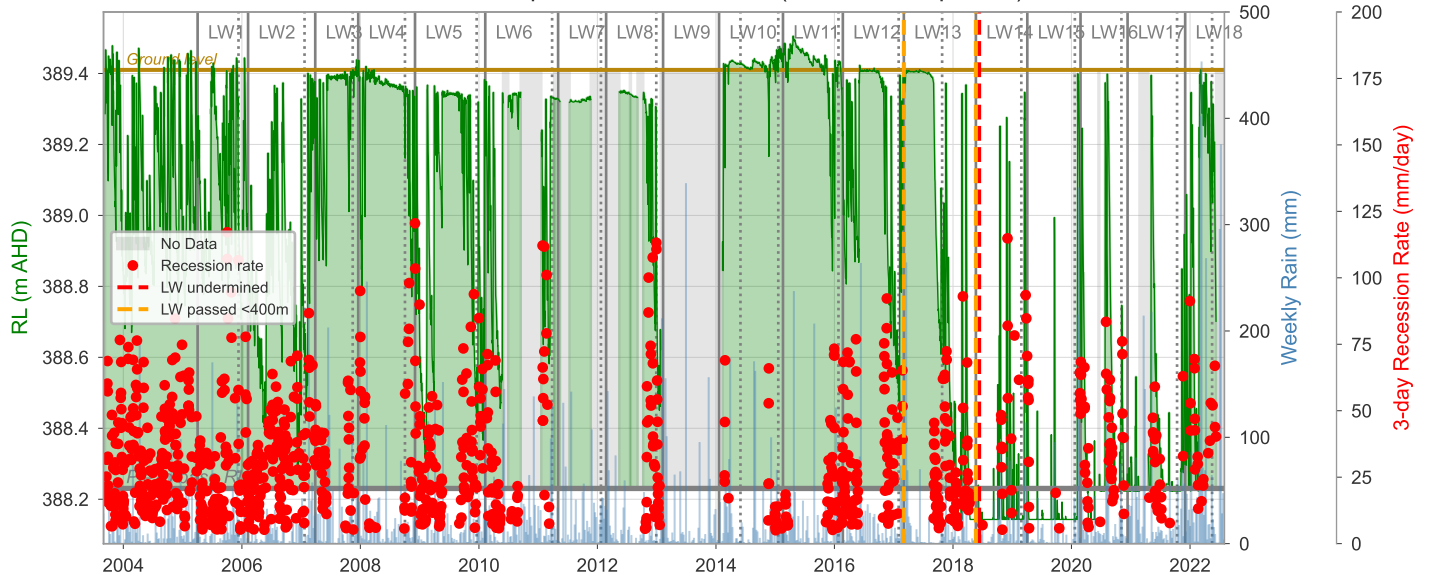
Dendrobium Swamp 11: Piezometer H1 (Within swamp EEC)



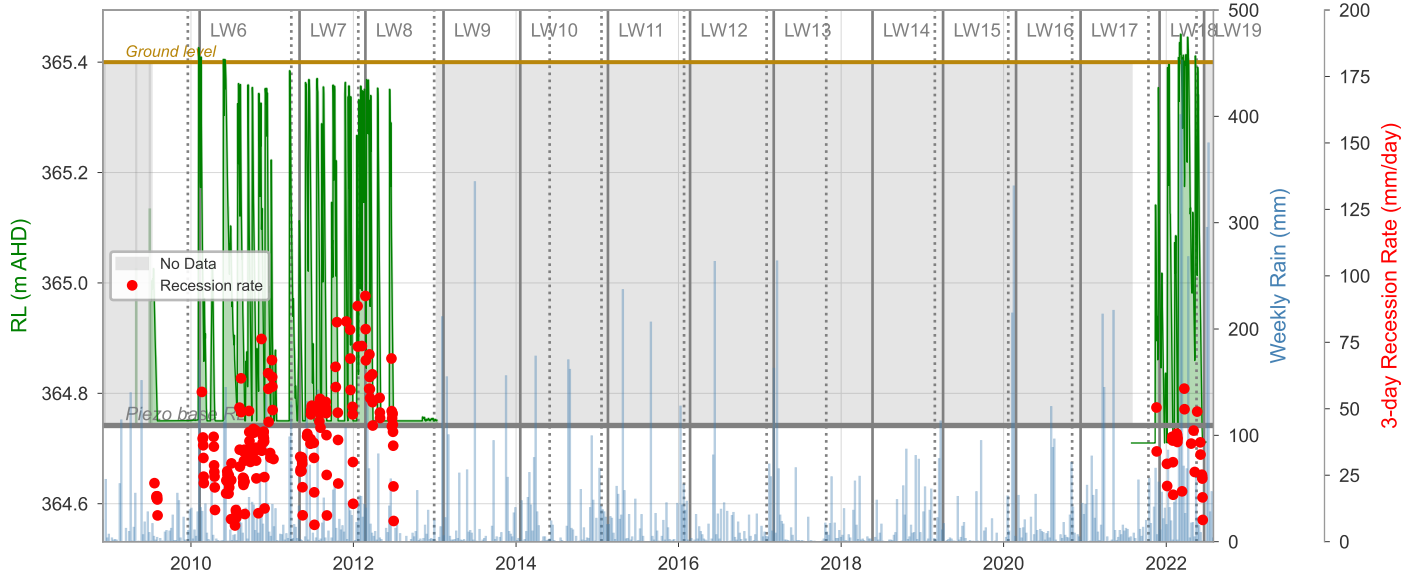
Dendrobium Swamp 11: Piezometer H3 (Within swamp EEC)



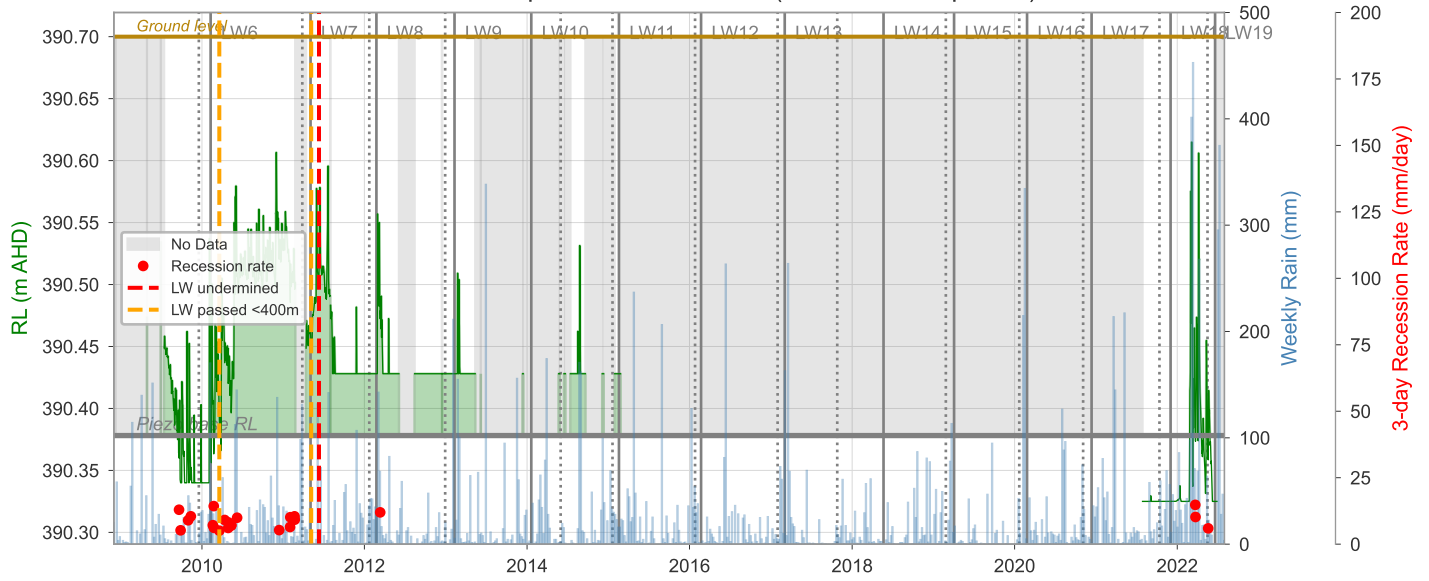
Dendrobium Swamp 11: Piezometer H2 (Within swamp EEC)



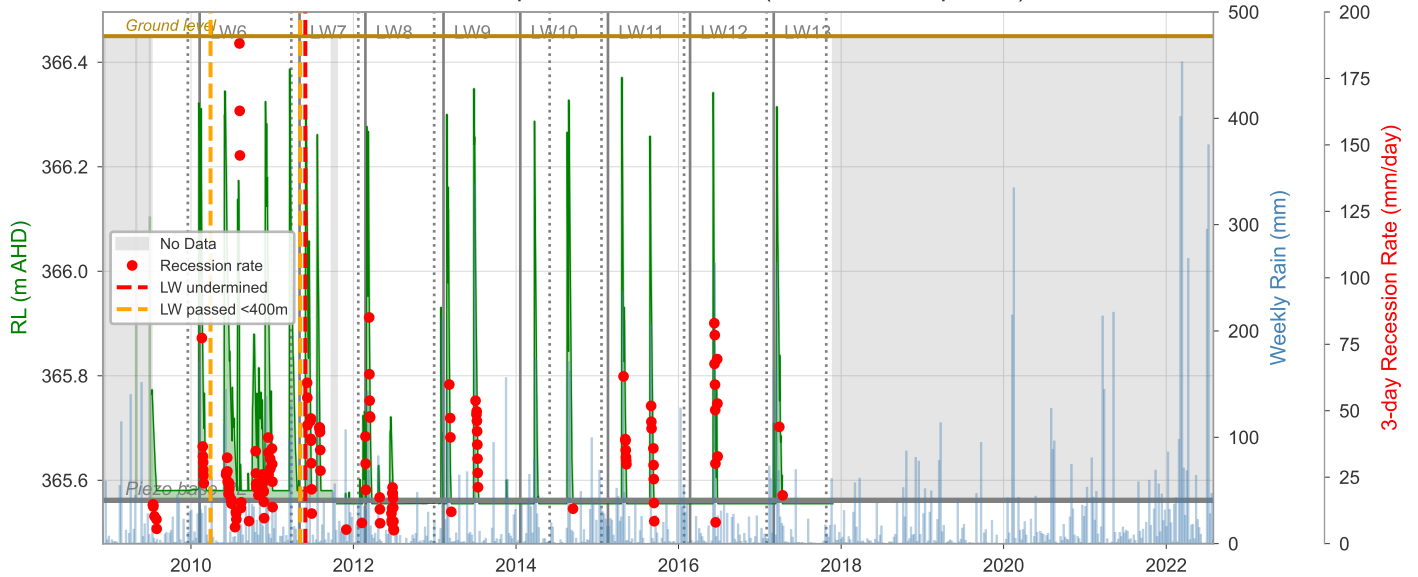
Dendrobium Swamp 12: Piezometer 04 (Outside swamp EEC)



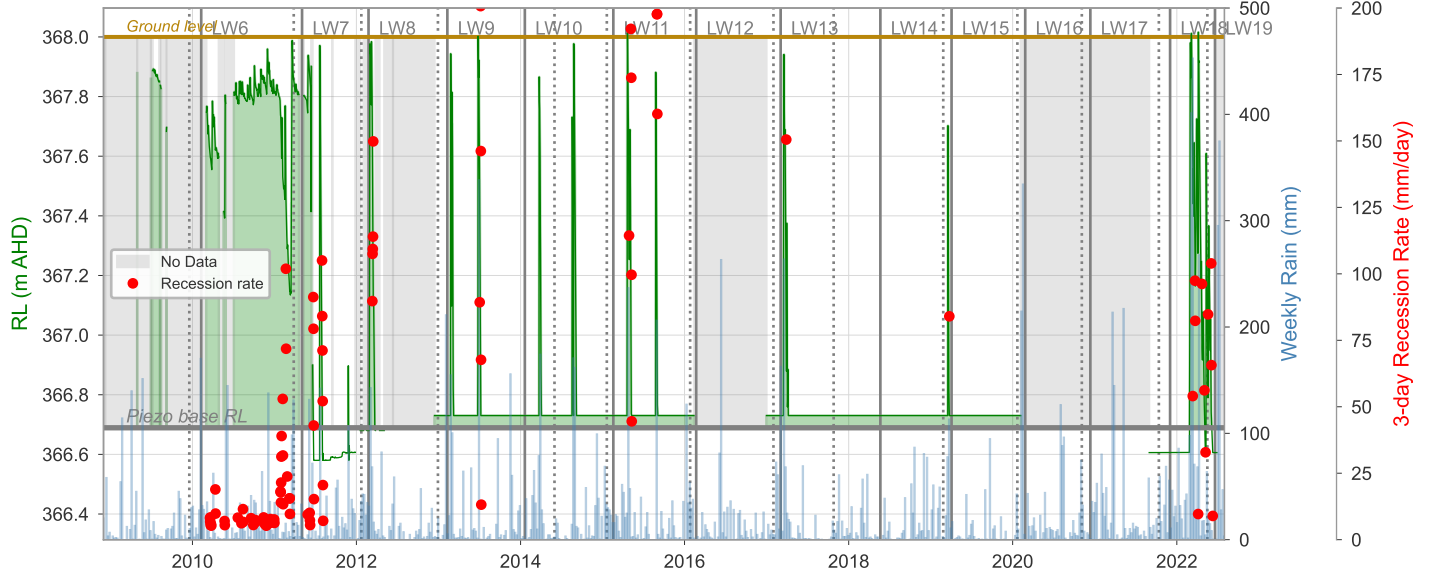
Dendrobium Swamp 12: Piezometer 03 (Outside swamp EEC)



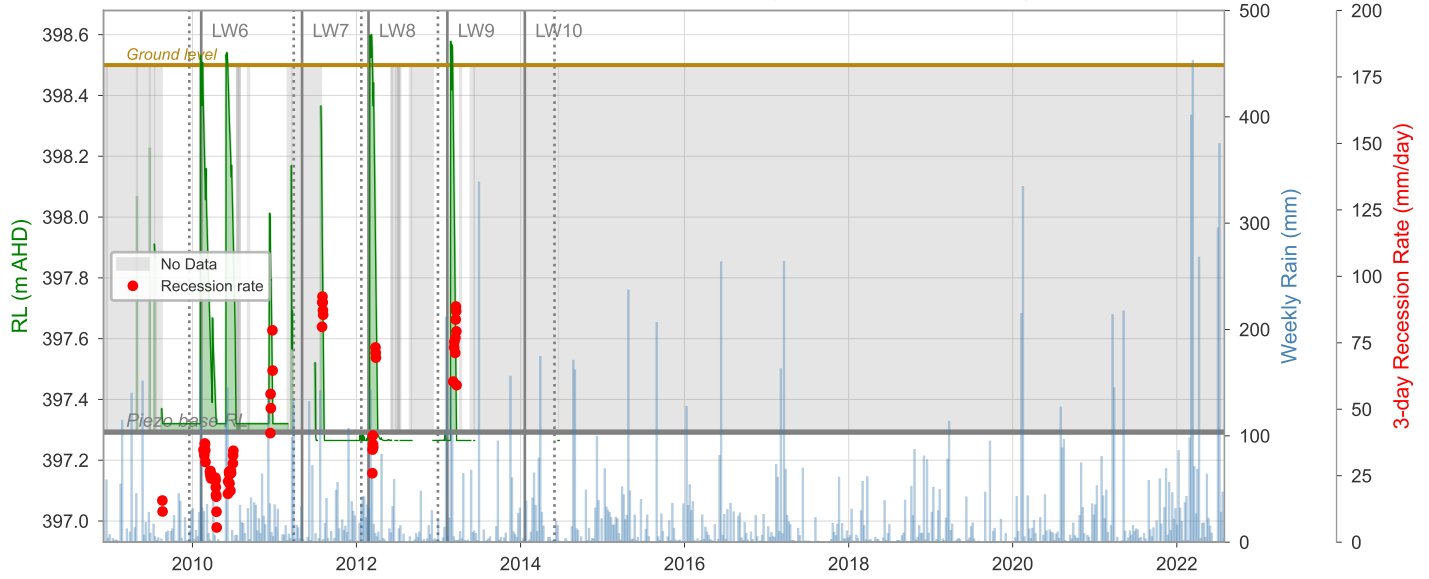
Dendrobium Swamp 12: Piezometer 02 (Outside swamp EEC)



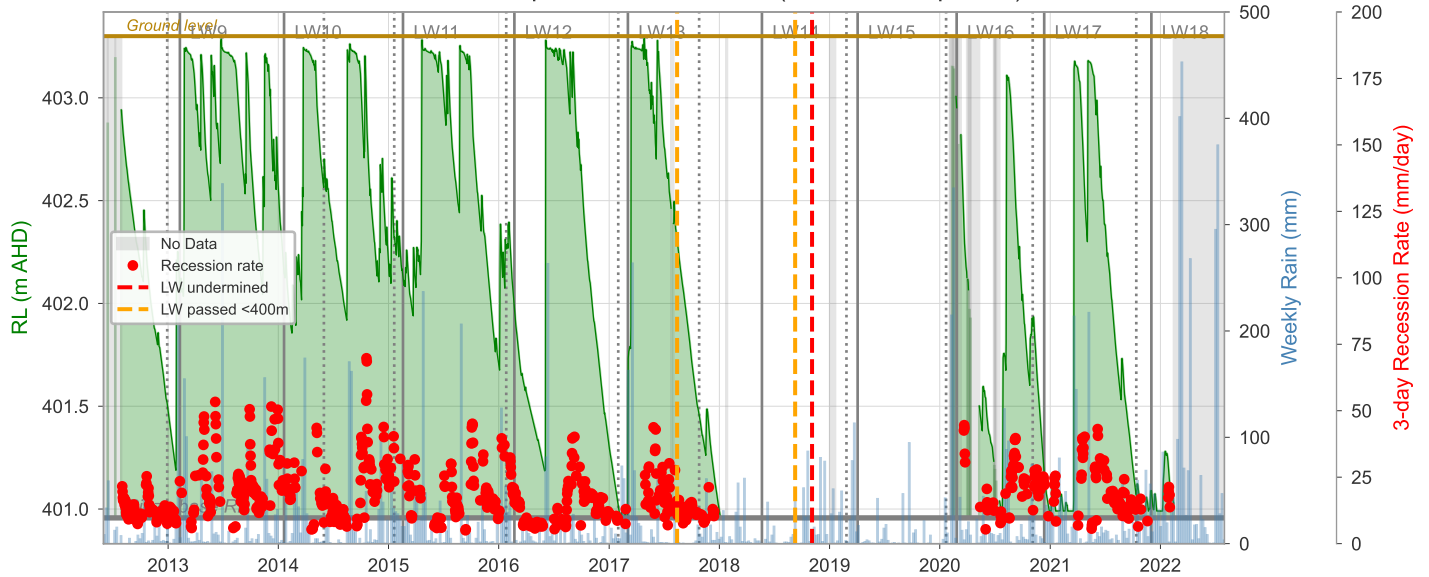
Dendrobium Swamp 12: Piezometer 01 (Within swamp EEC)



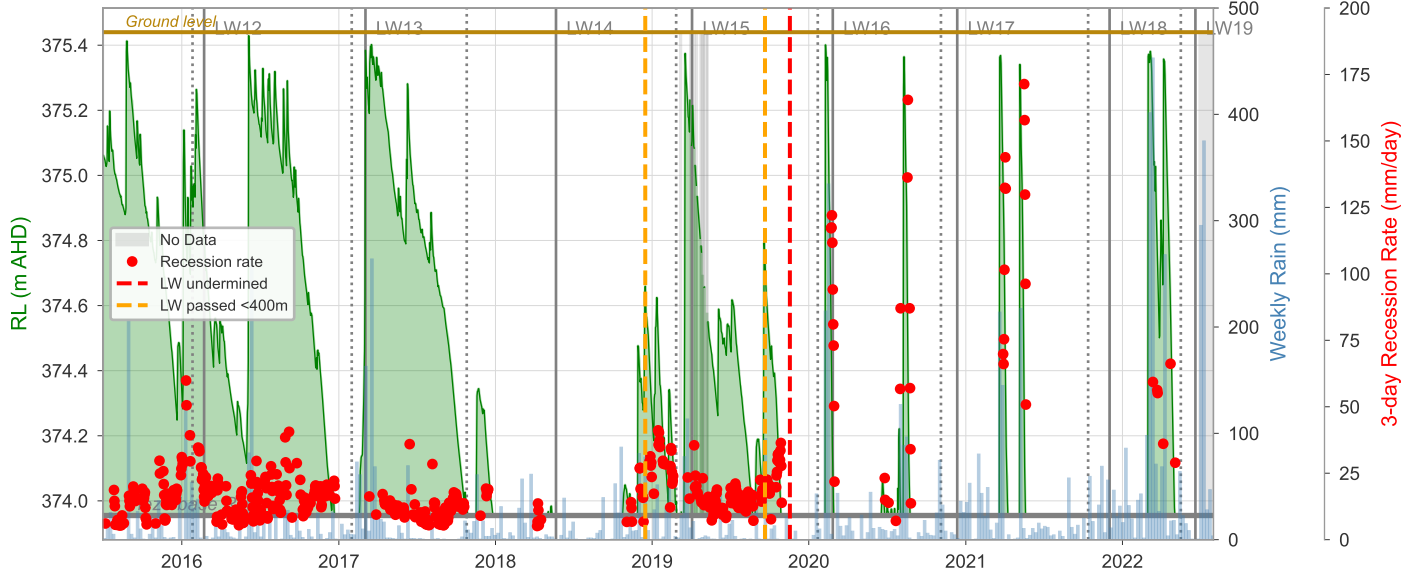
Dendrobium Swamp 12: Piezometer 05 (Outside swamp EEC)



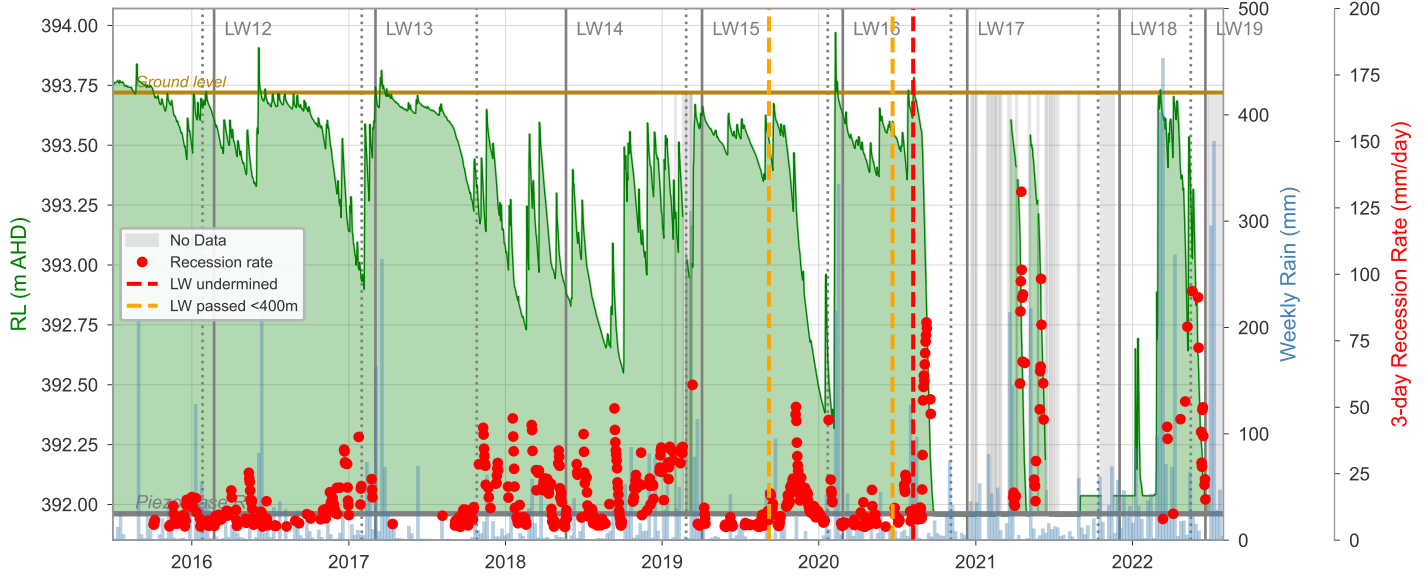
Dendrobium Swamp 13: Piezometer 01 (Within swamp EEC)



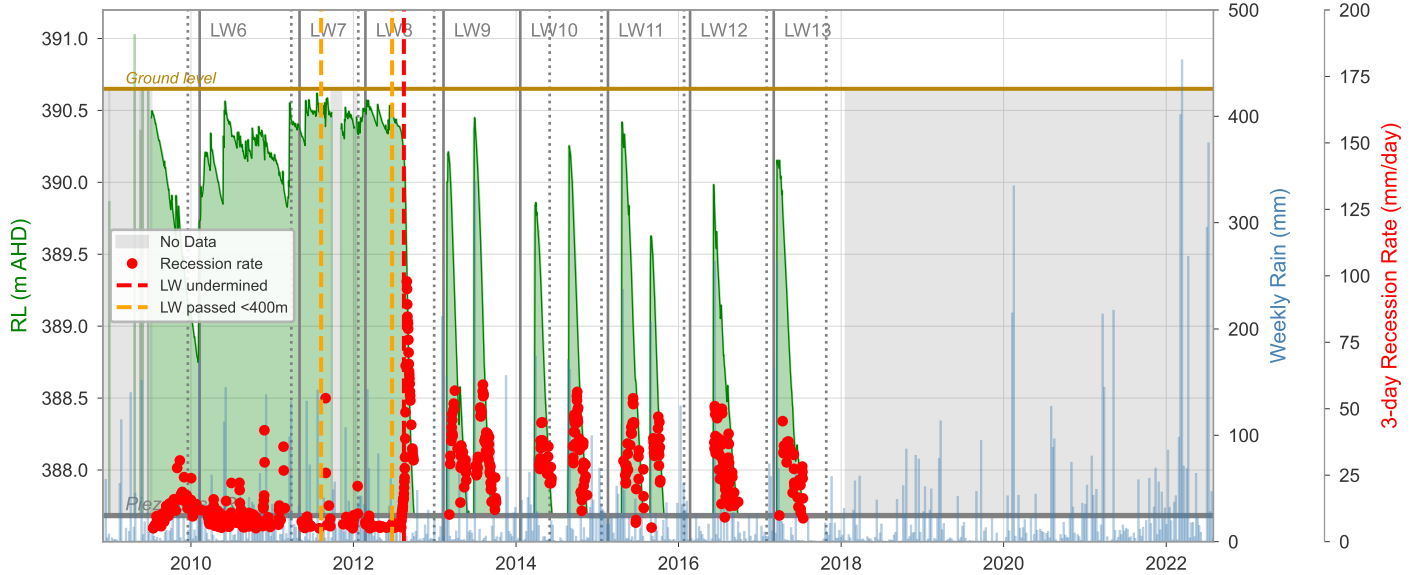
Dendrobium Swamp 14: Piezometer 02 (Within swamp EEC)



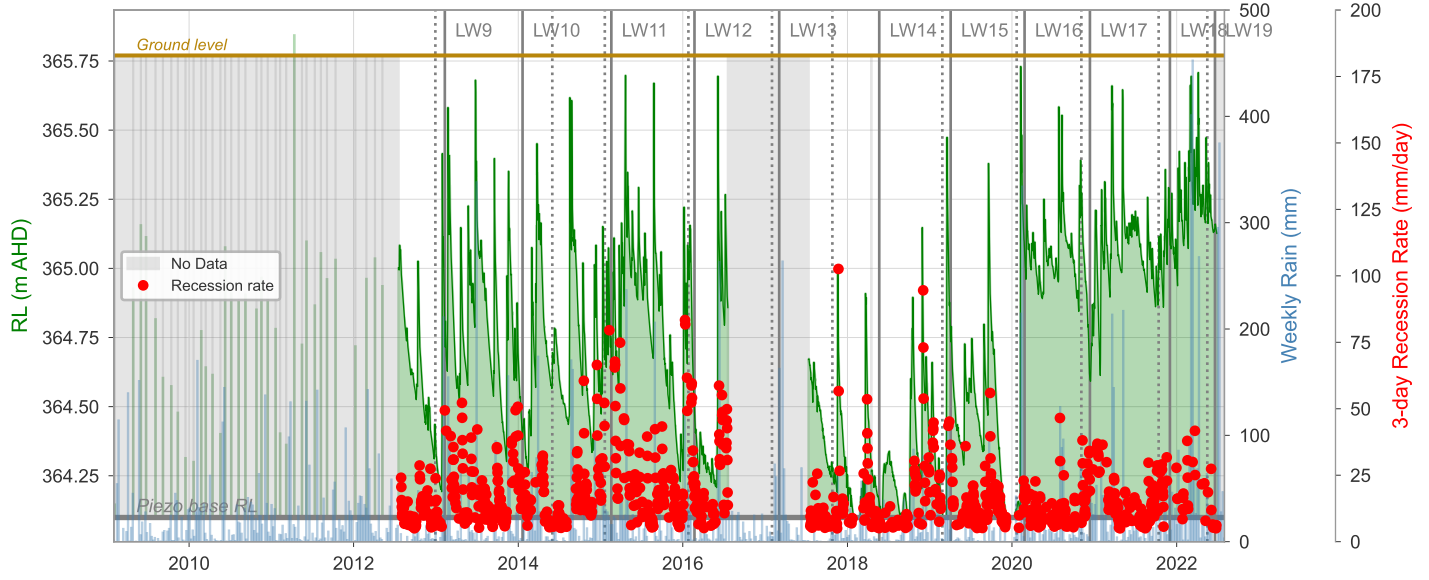
Dendrobium Swamp 14: Piezometer 01 (Within swamp EEC)



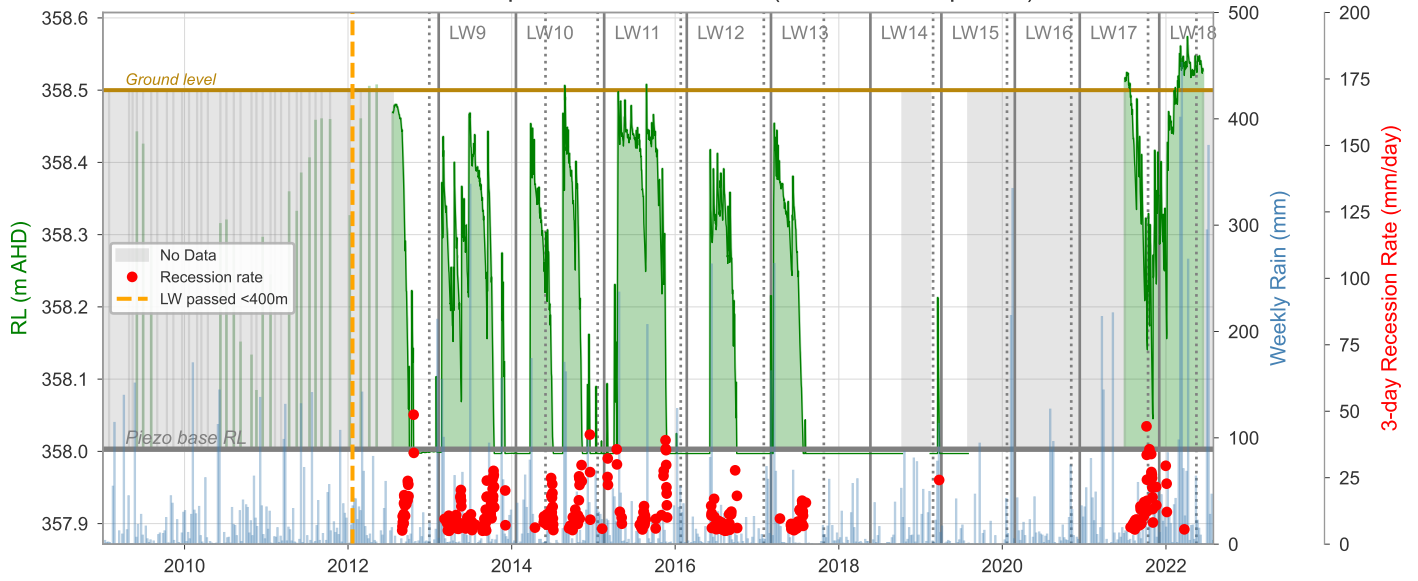
Dendrobium Swamp 15B: Piezometer 27 (Within swamp EEC)



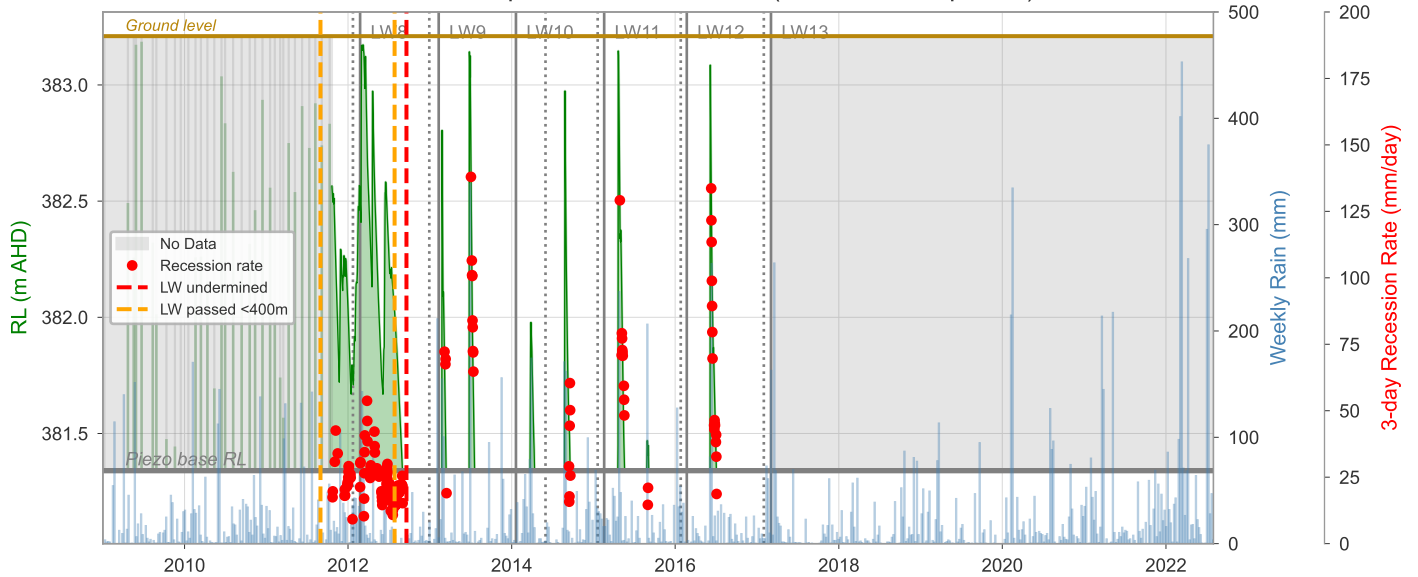
Dendrobium Swamp 15A: Piezometer 07 (Outside swamp EEC)



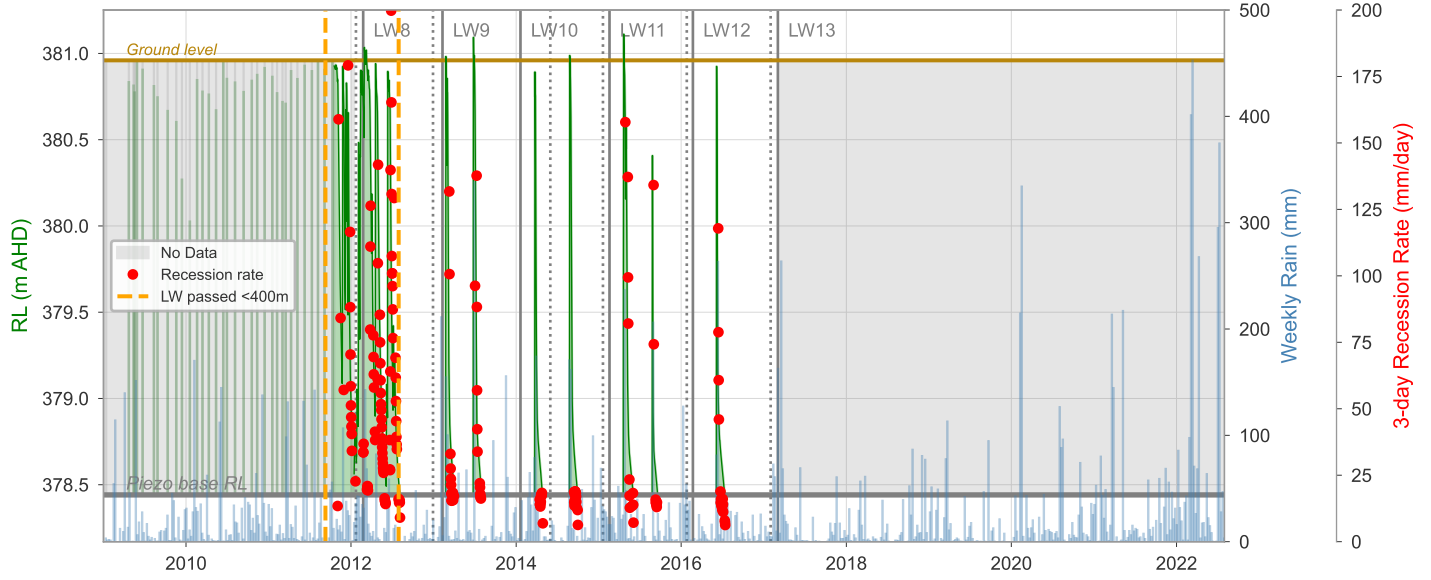
Dendrobium Swamp 15A: Piezometer 18 (Outside swamp EEC)



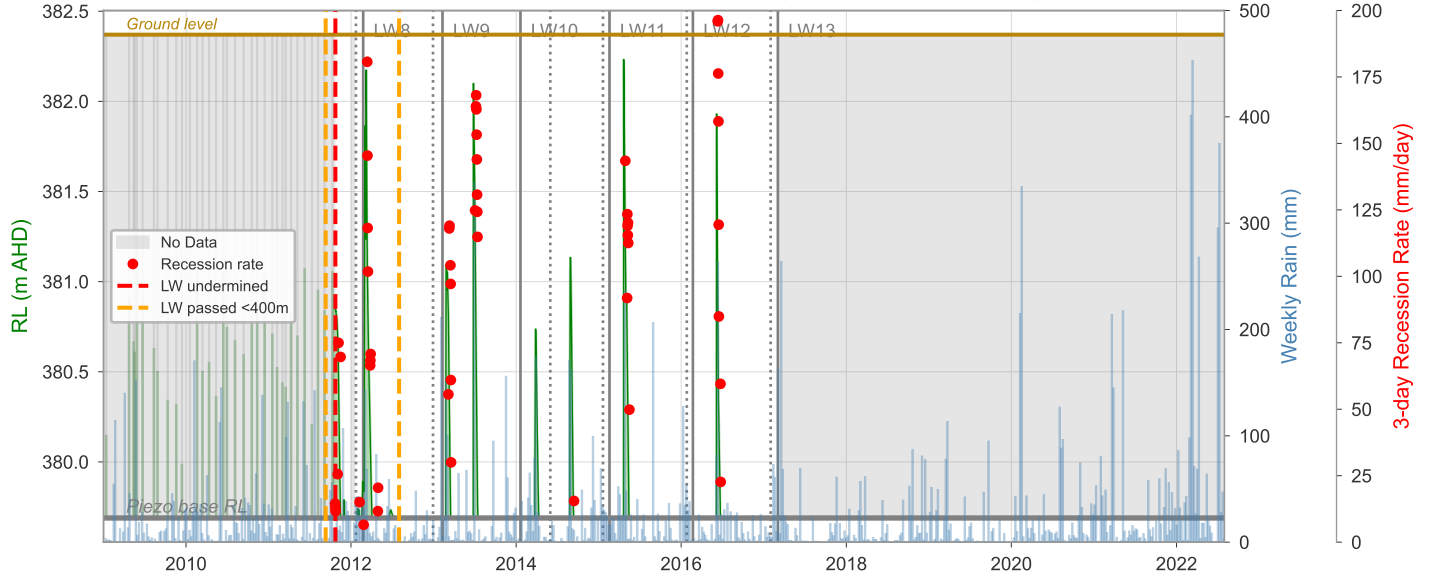
Dendrobium Swamp 15B: Piezometer 22 (Outside swamp EEC)



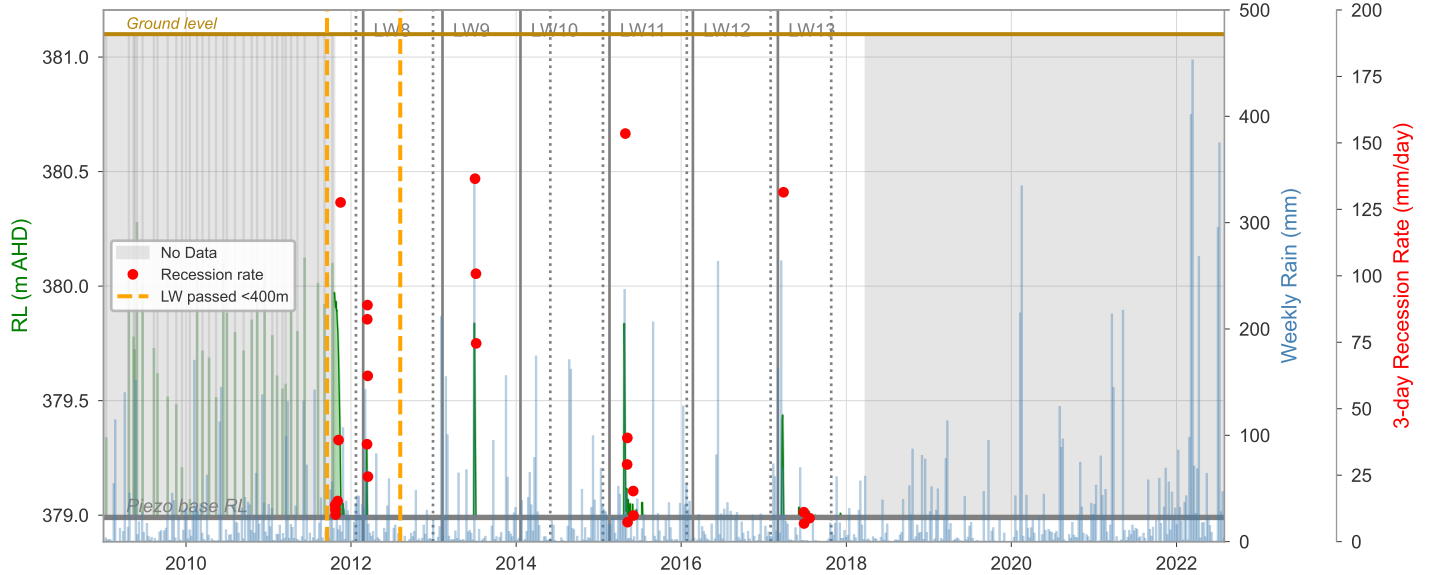
Dendrobium Swamp 15B: Piezometer 23 (Within swamp EEC)



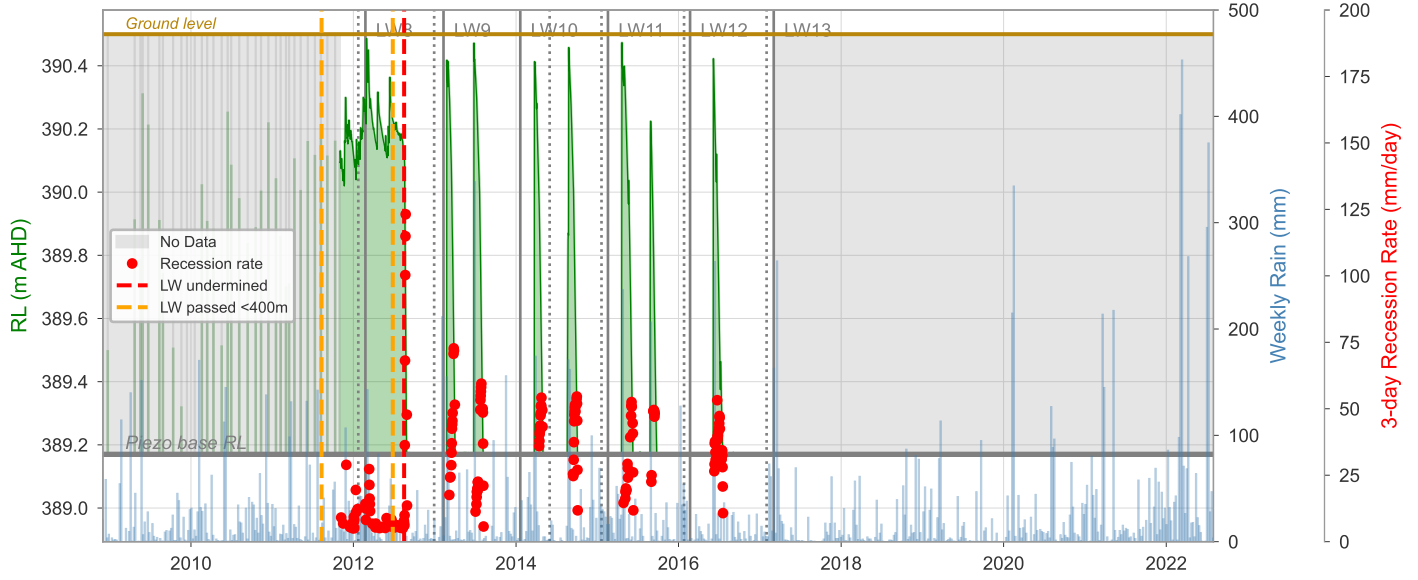
Dendrobium Swamp 15B: Piezometer 24 (Outside swamp EEC)



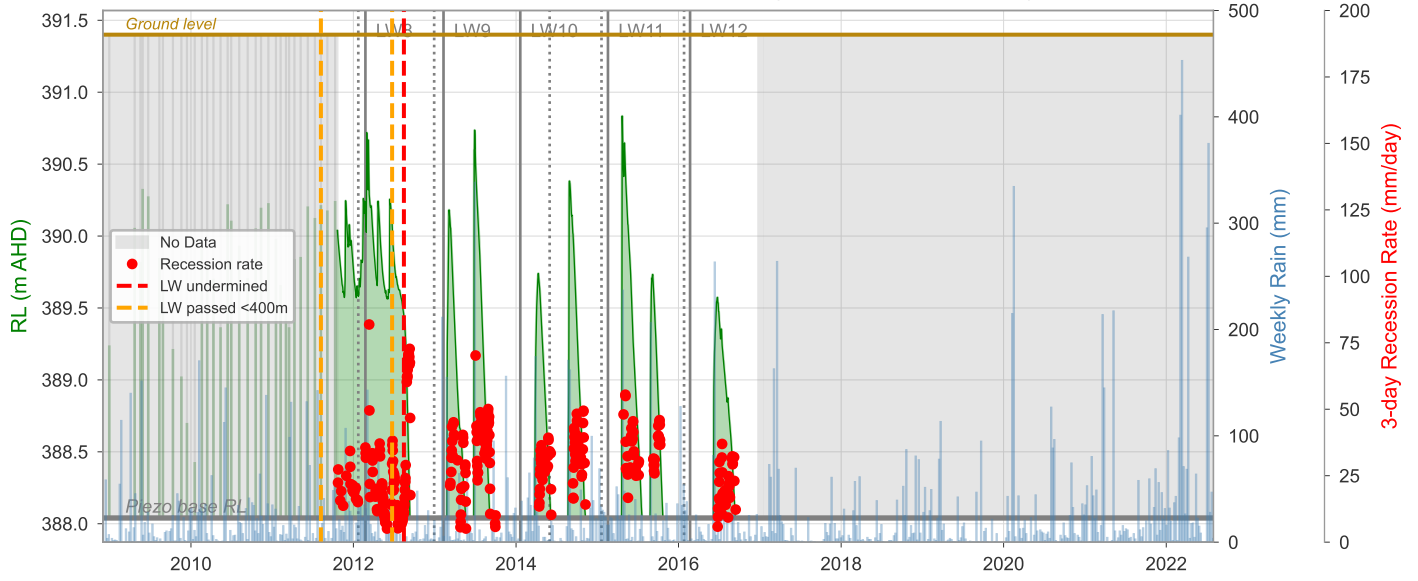
Dendrobium Swamp 15B: Piezometer 25 (Outside swamp EEC)



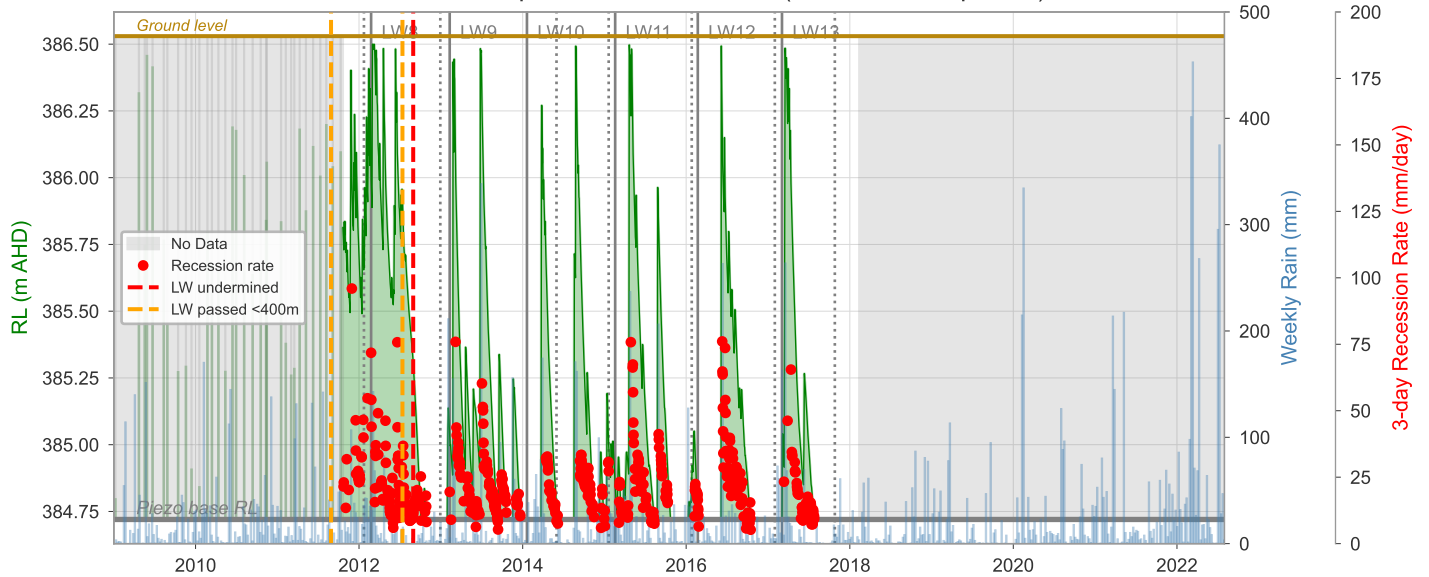
Dendrobium Swamp 15B: Piezometer 26 (Within swamp EEC)



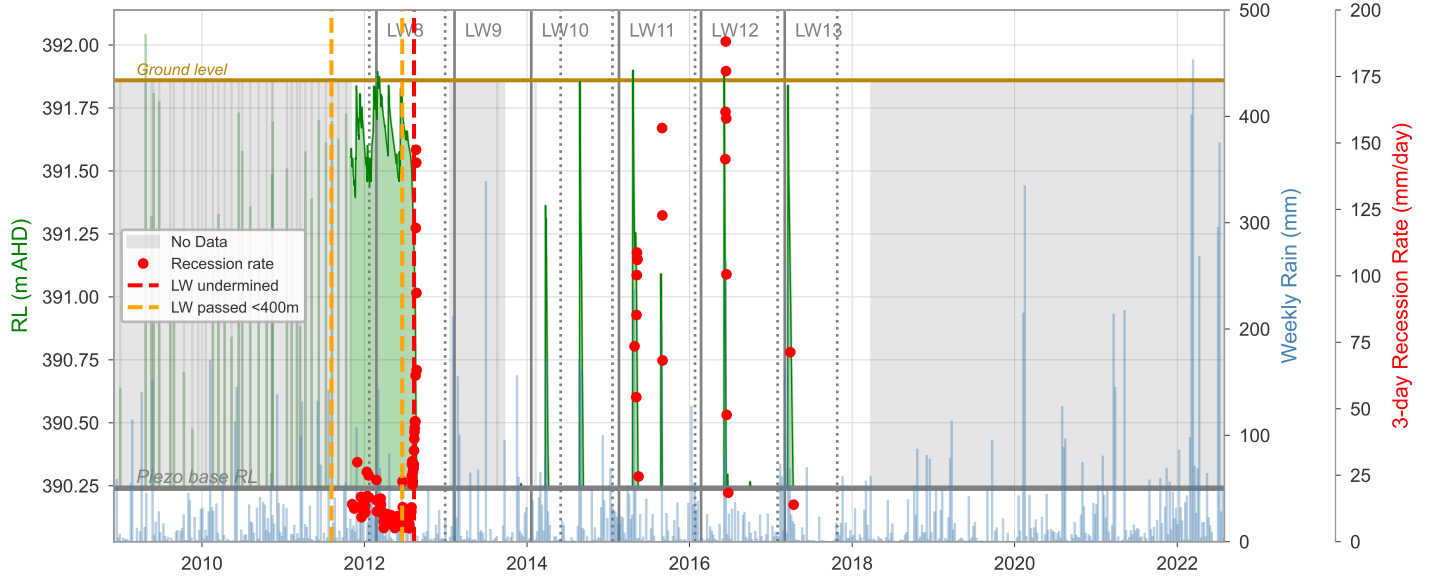
Dendrobium Swamp 15B: Piezometer 28 (Outside swamp EEC)



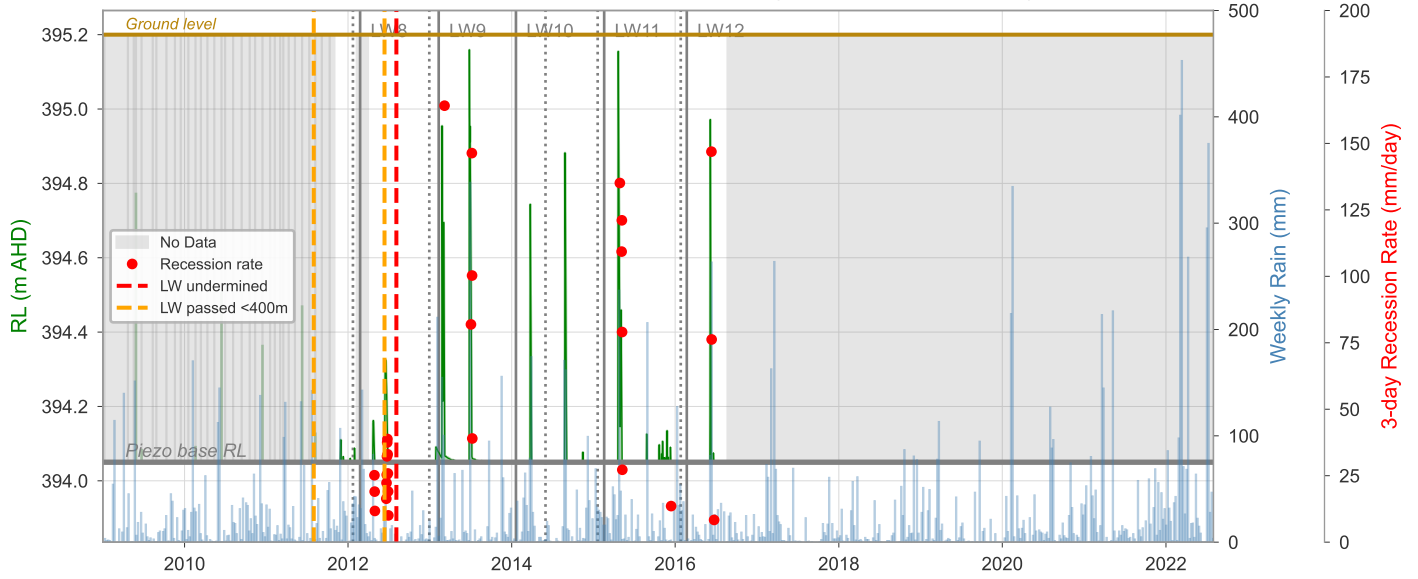
Dendrobium Swamp 15B: Piezometer 29 (Outside swamp EEC)



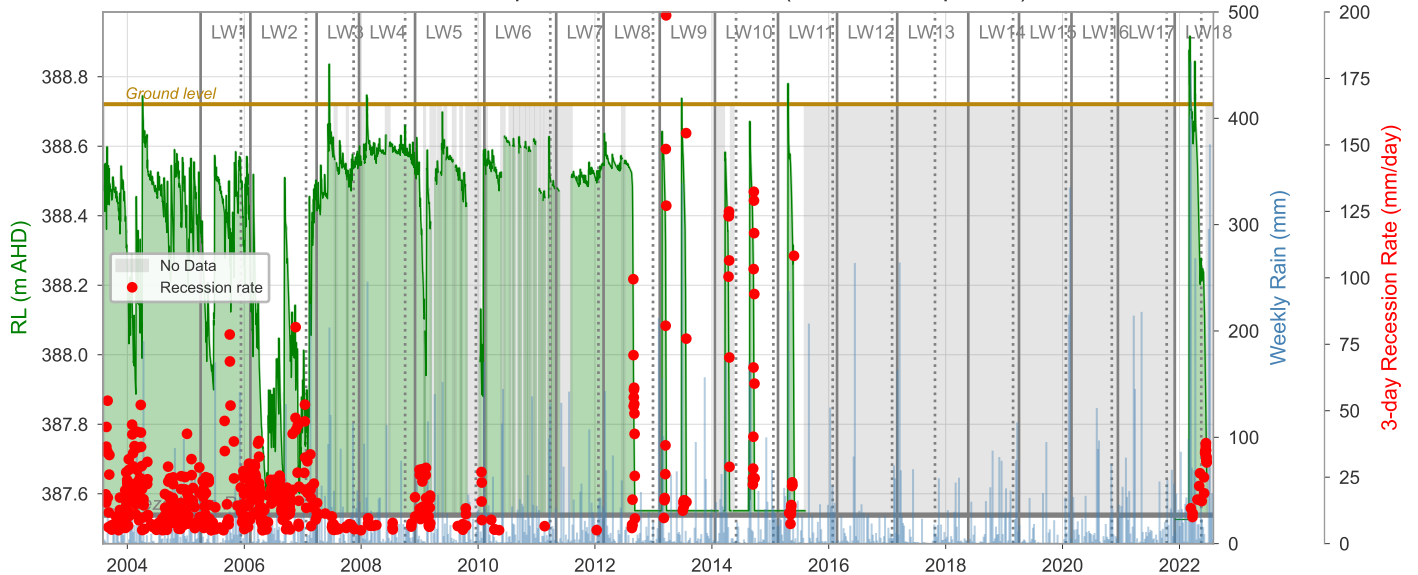
Dendrobium Swamp 15B: Piezometer 30 (Within swamp EEC)



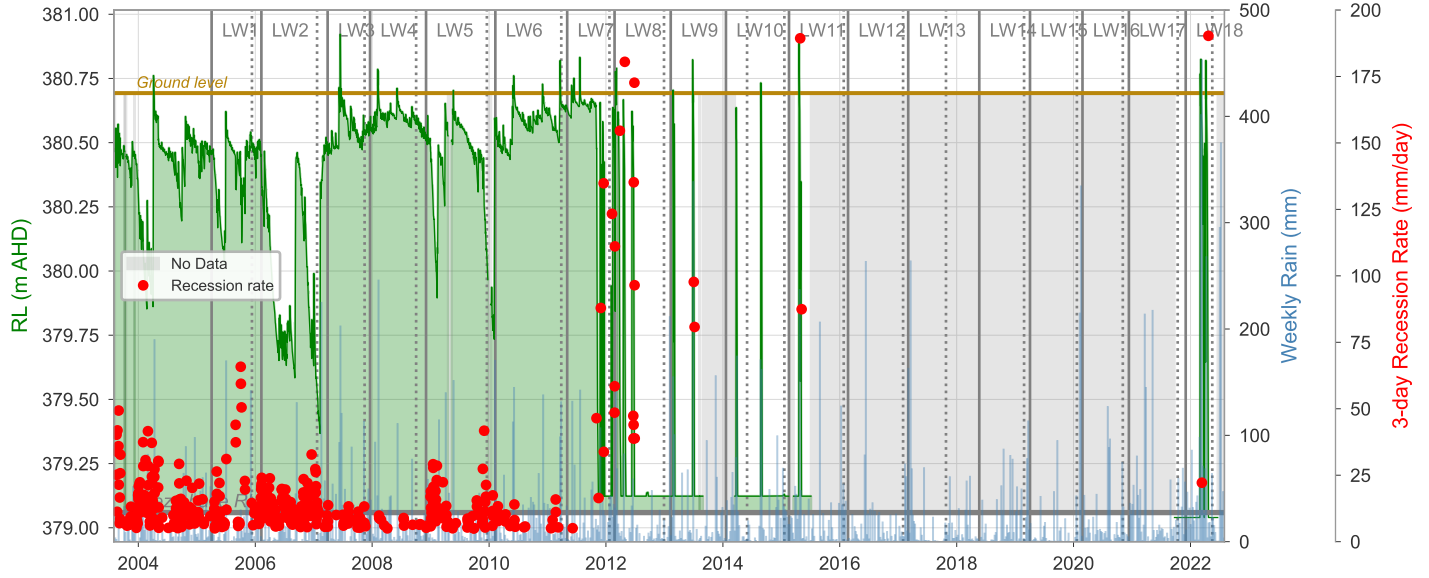
Dendrobium Swamp 15B: Piezometer 31 (Outside swamp EEC)



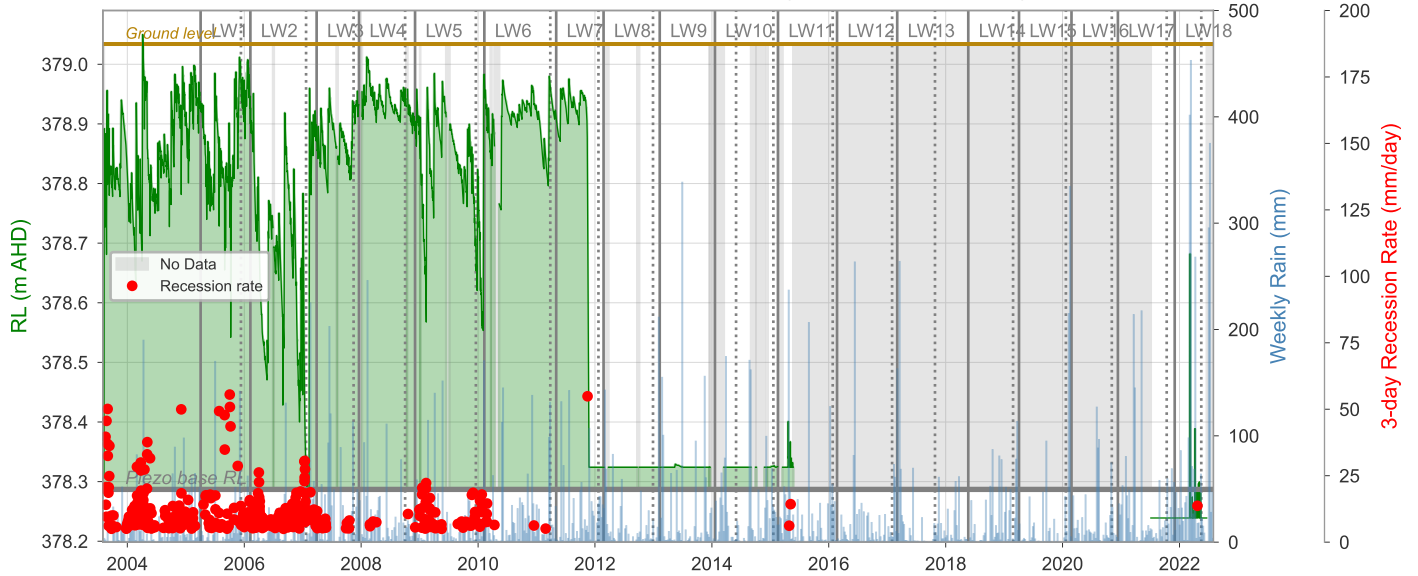
Dendrobium Swamp 15B: Piezometer H1 (Within swamp EEC)



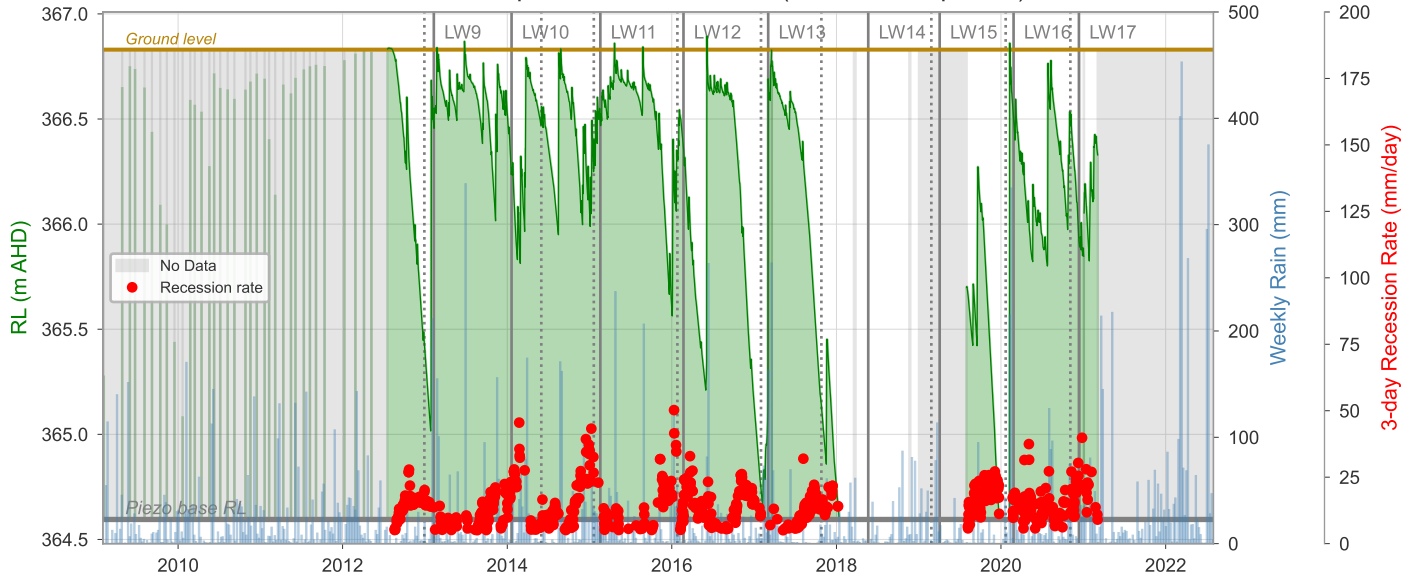
Dendrobium Swamp 15B: Piezometer H2 (Within swamp EEC)



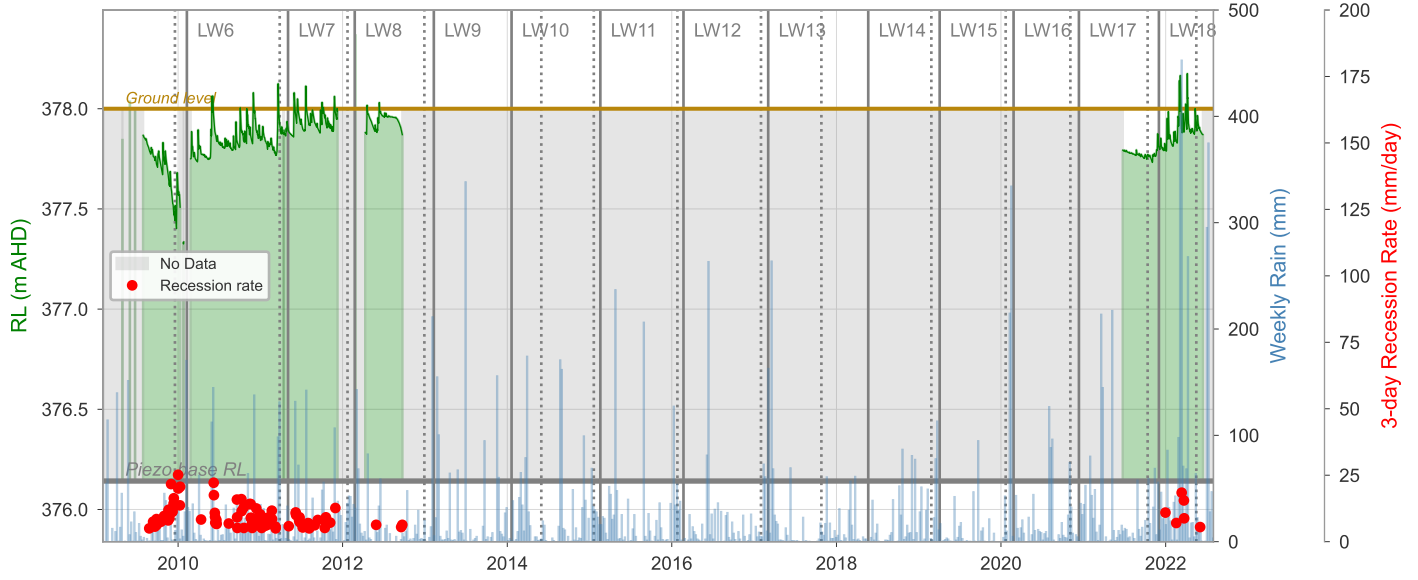
Dendrobium Swamp 15B: Piezometer H3 (Within swamp EEC)



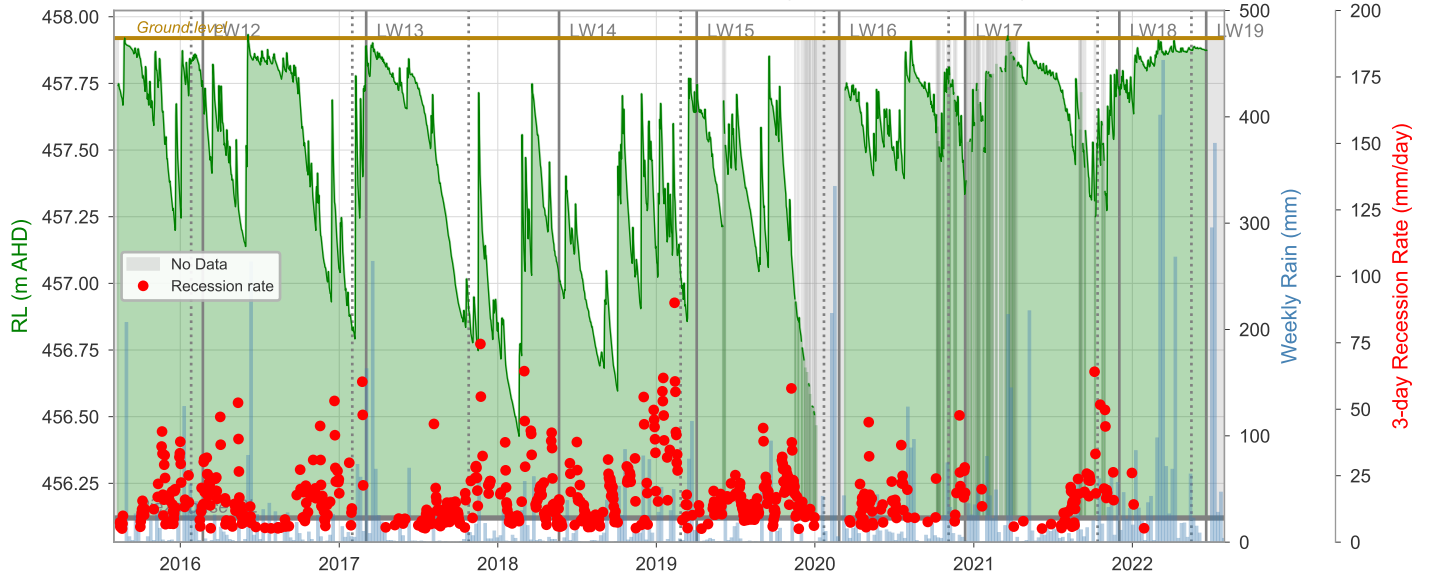
Dendrobium Swamp 15A: Piezometer 06 (Within swamp EEC)



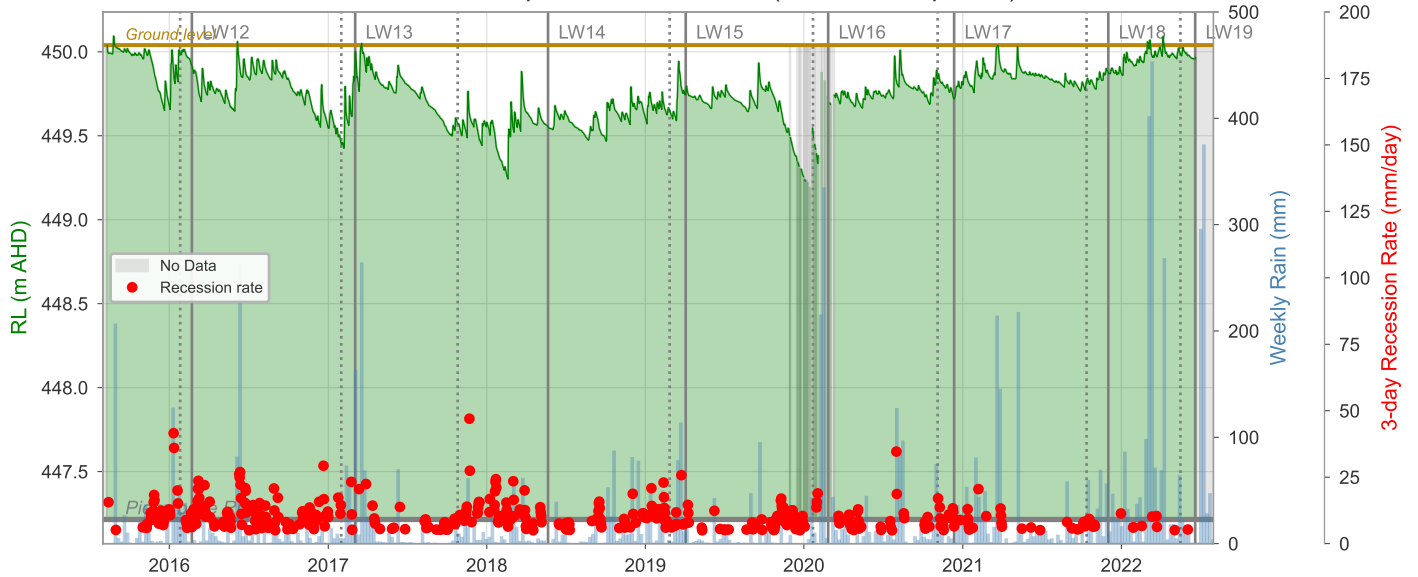
Dendrobium Swamp 15A: Piezometer 03 (Within swamp EEC)



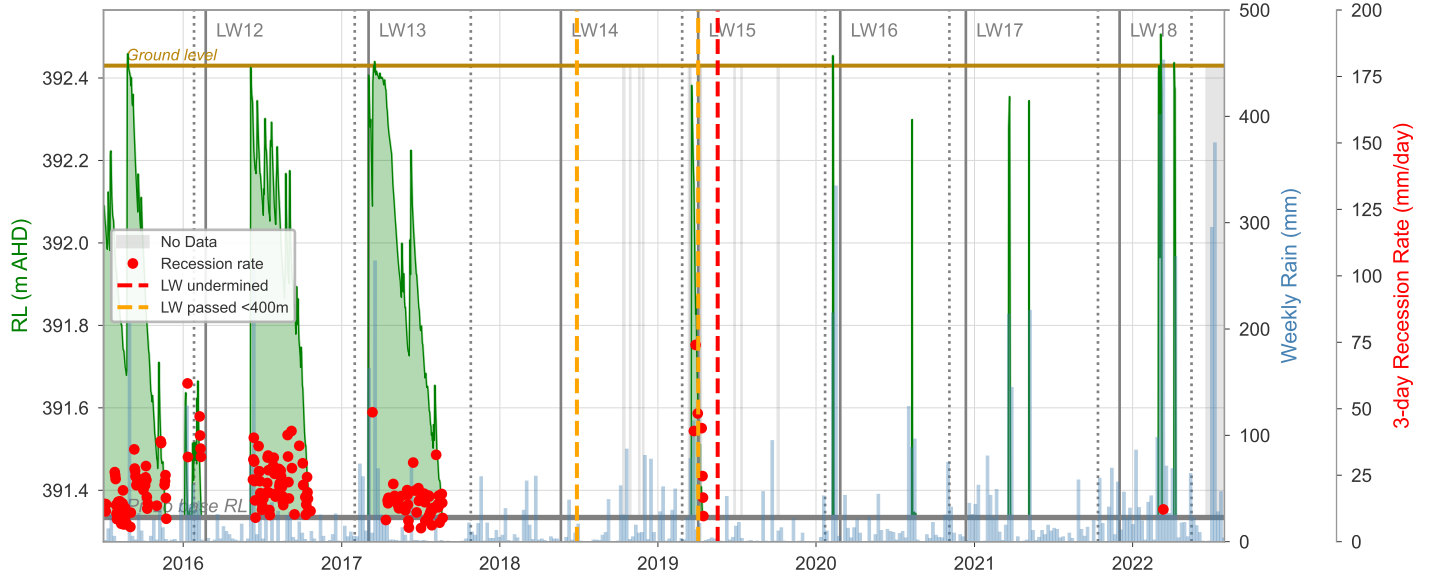
Dendrobium Swamp 22: Piezometer 01 (Within swamp EEC)



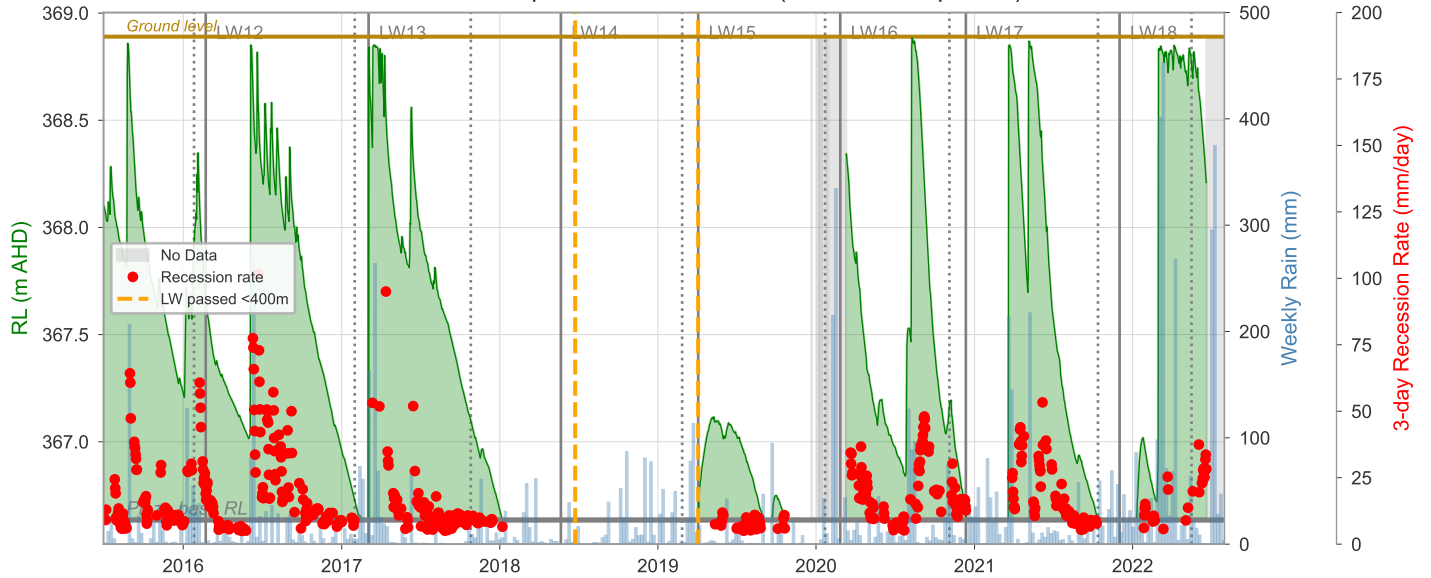
Dendrobium Swamp 22: Piezometer 02 (Within swamp EEC)



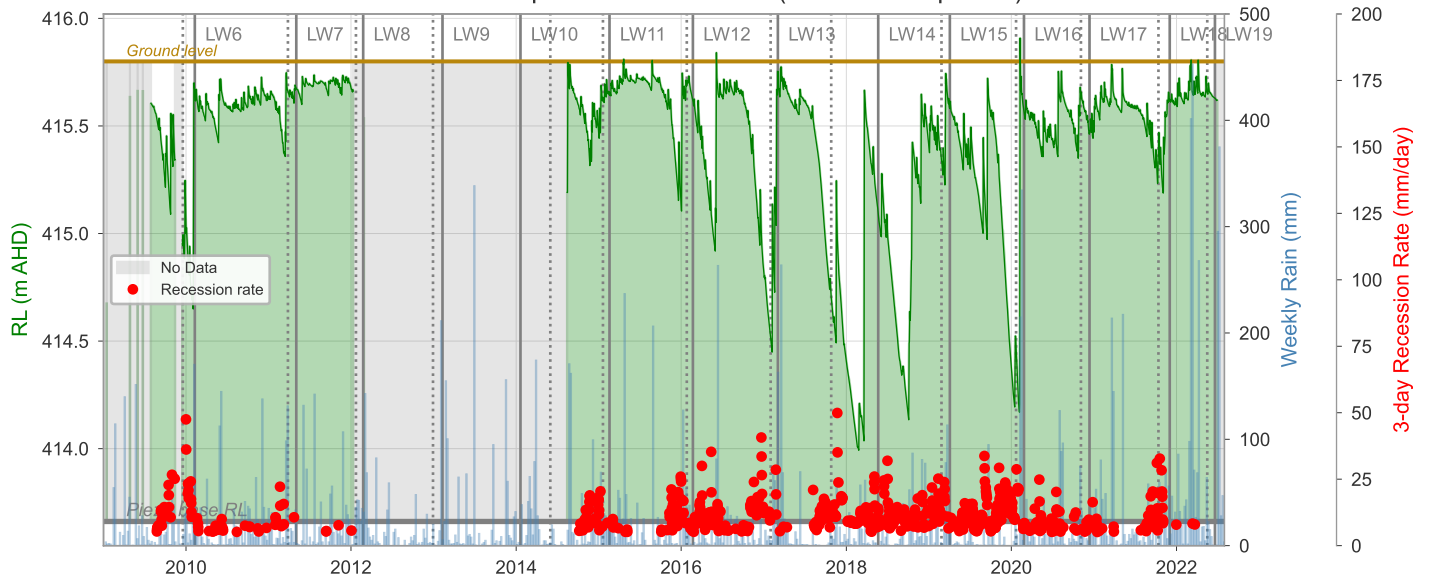
Dendrobium Swamp 23: Piezometer 01 (Within swamp EEC)



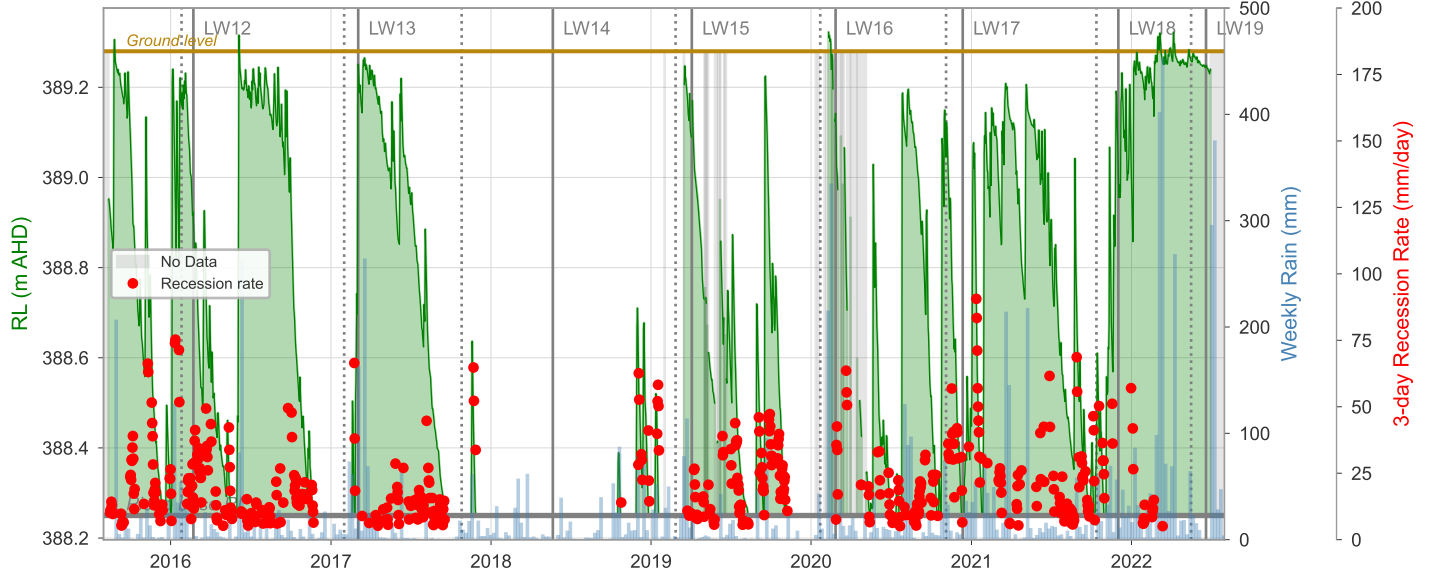
Dendrobium Swamp 23: Piezometer 02 (Within swamp EEC)



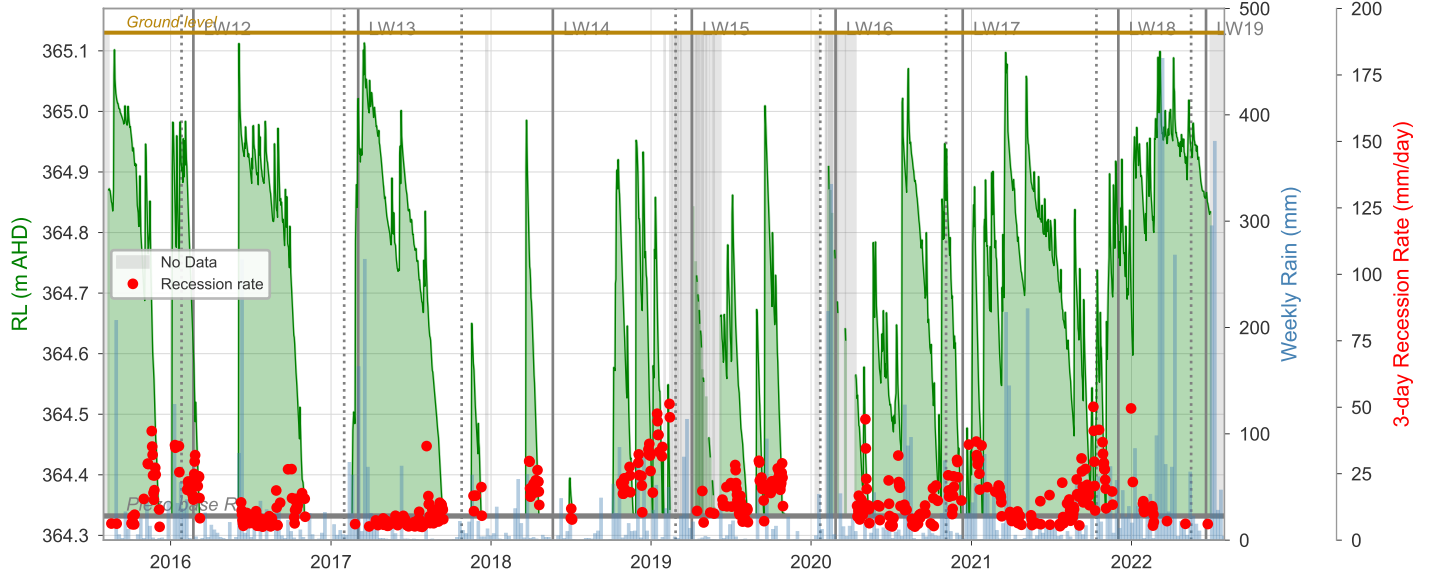
Dendrobium Swamp 25: Piezometer 01 (Within swamp EEC)



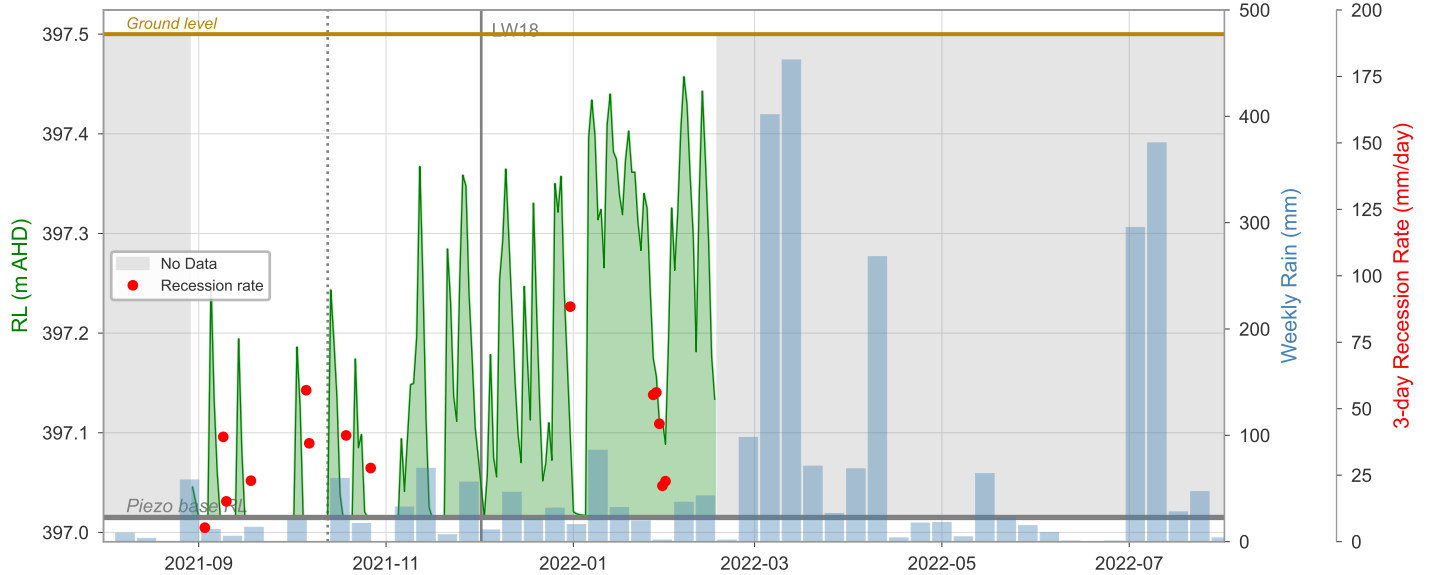
Dendrobium Swamp 33: Piezometer 01 (Within swamp EEC)



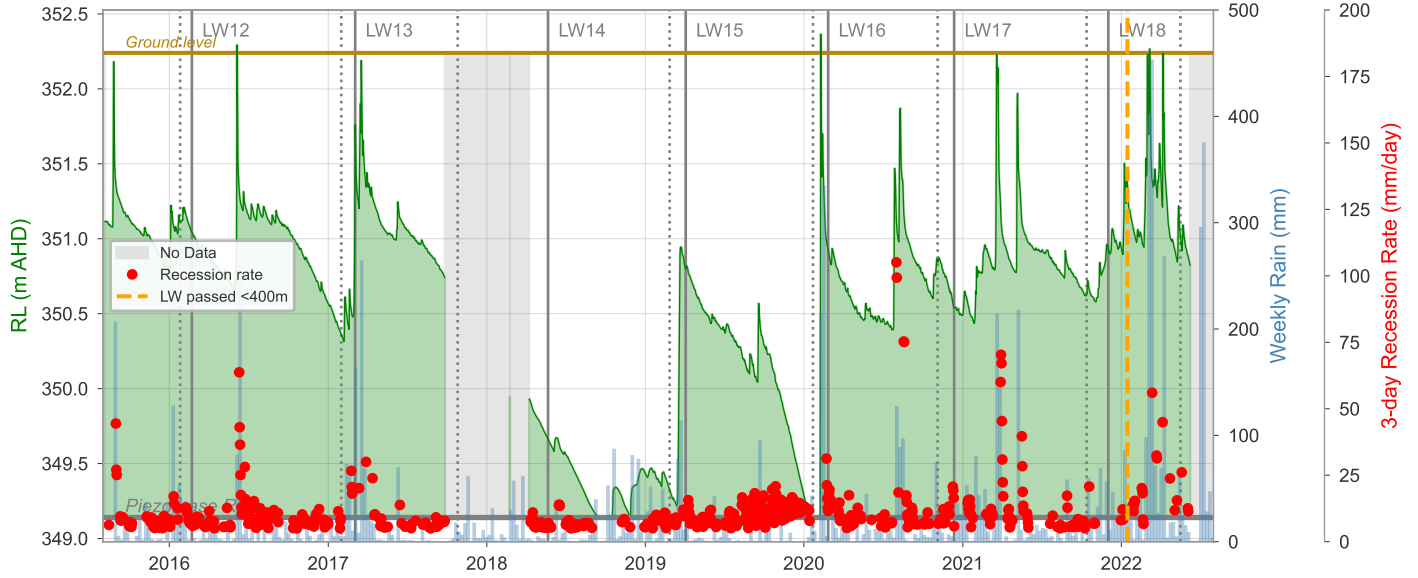
Dendrobium Swamp 33: Piezometer 03 (Within swamp EEC)



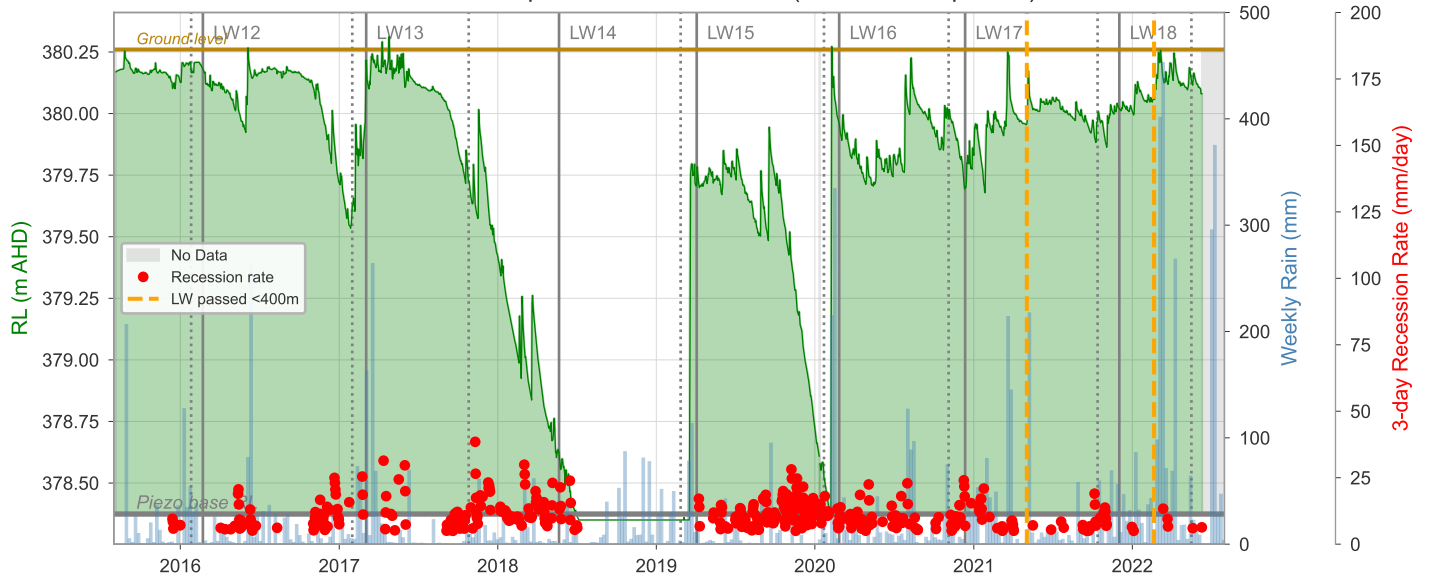
Dendrobium Swamp 34: Piezometer 01 (Within swamp EEC)



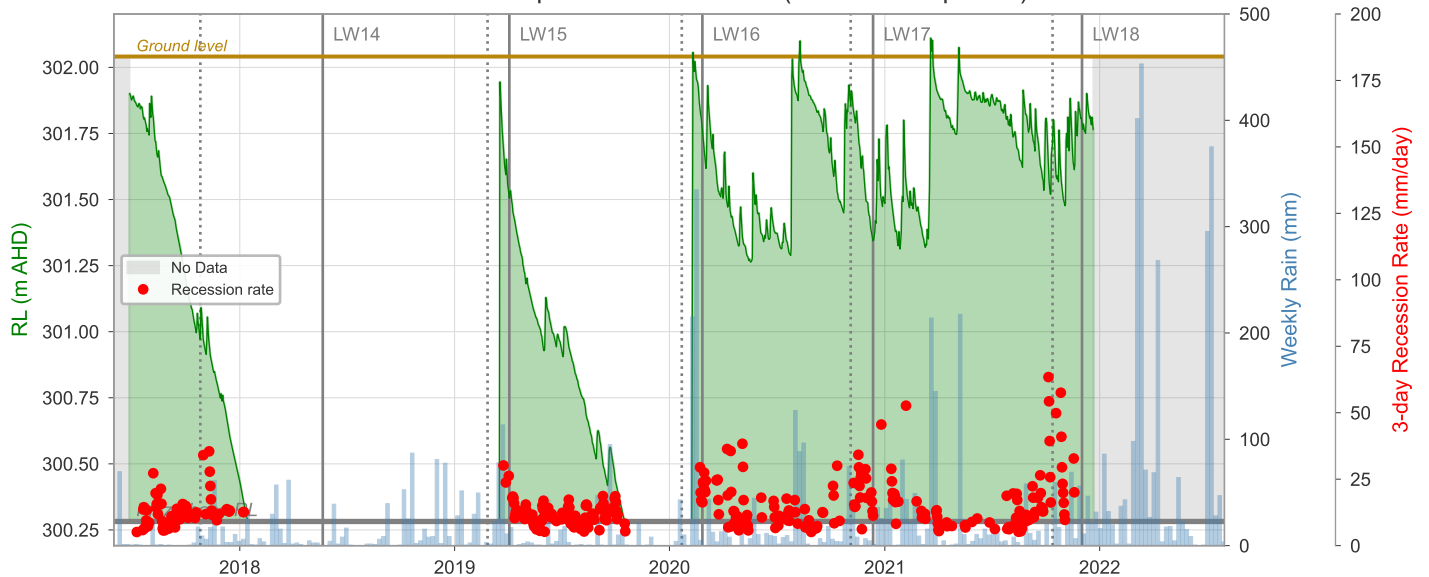
Dendrobium Swamp 35B: Piezometer 01 (Within swamp EEC)



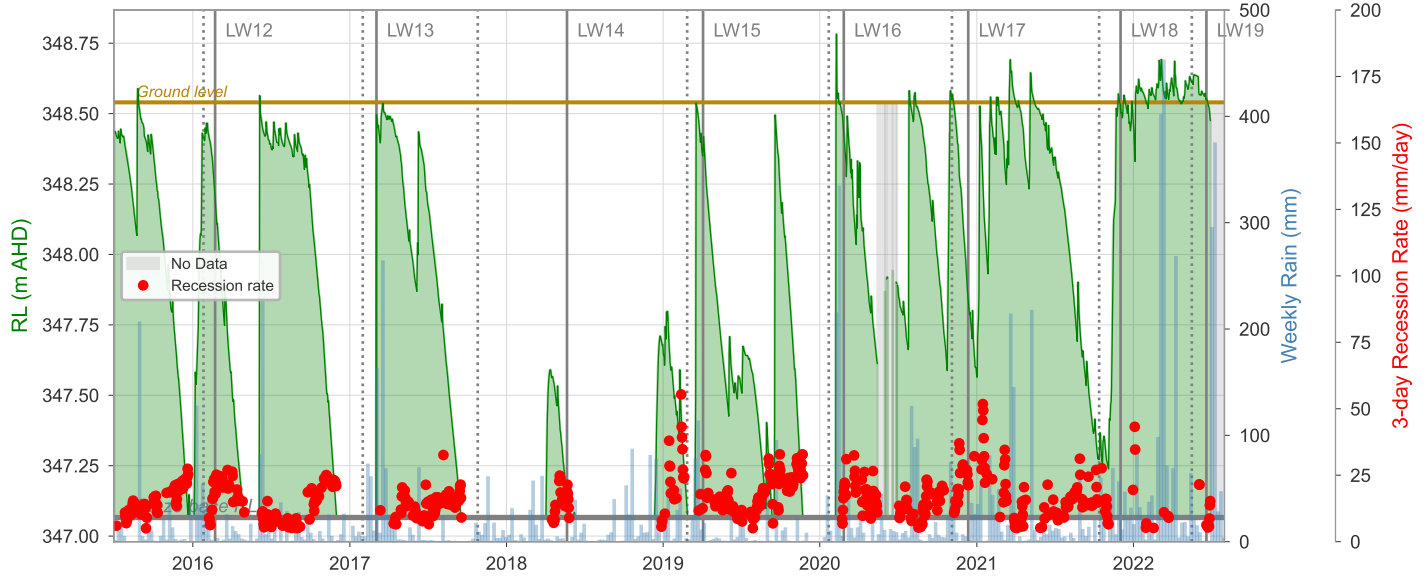
Dendrobium Swamp 35A: Piezometer 01 (Within swamp EEC)



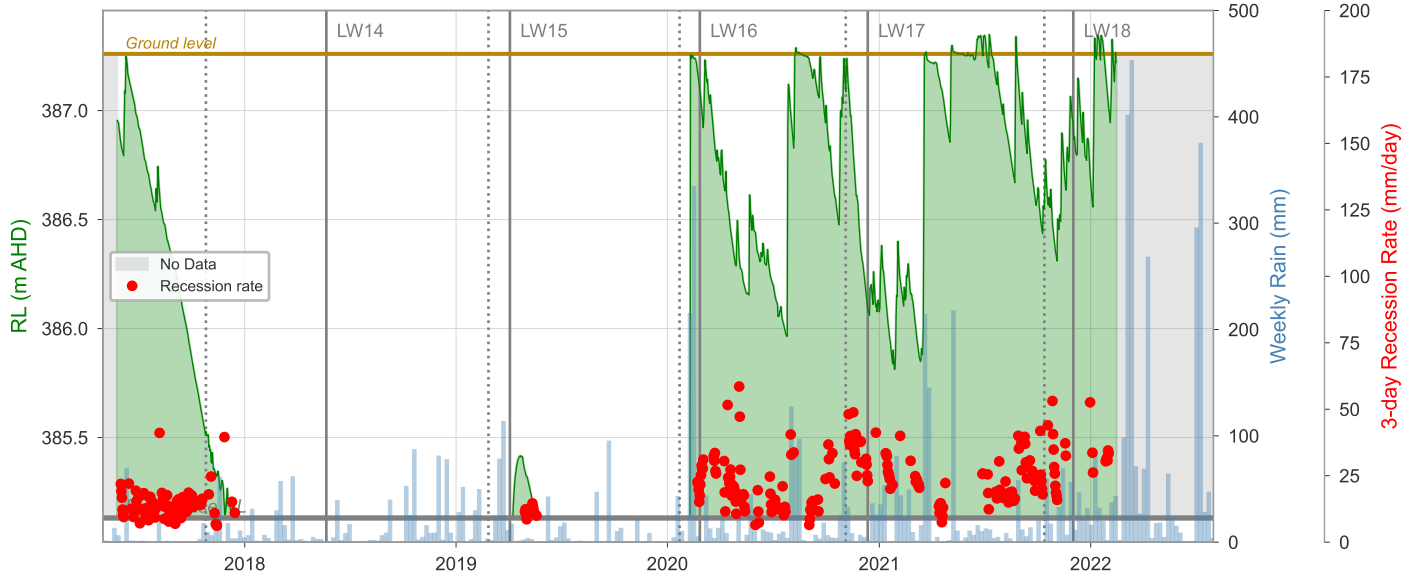
Dendrobium Swamp 83: Piezometer 01 (Within swamp EEC)



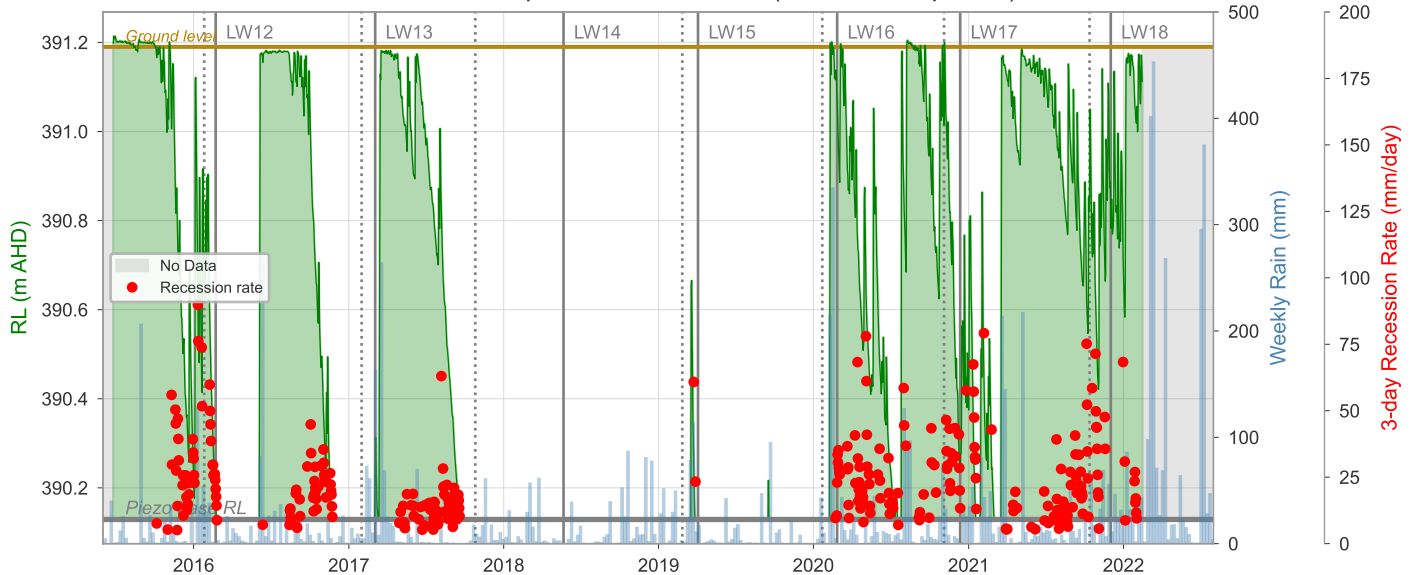
Dendrobium Swamp 84: Piezometer 01 (Within swamp EEC)



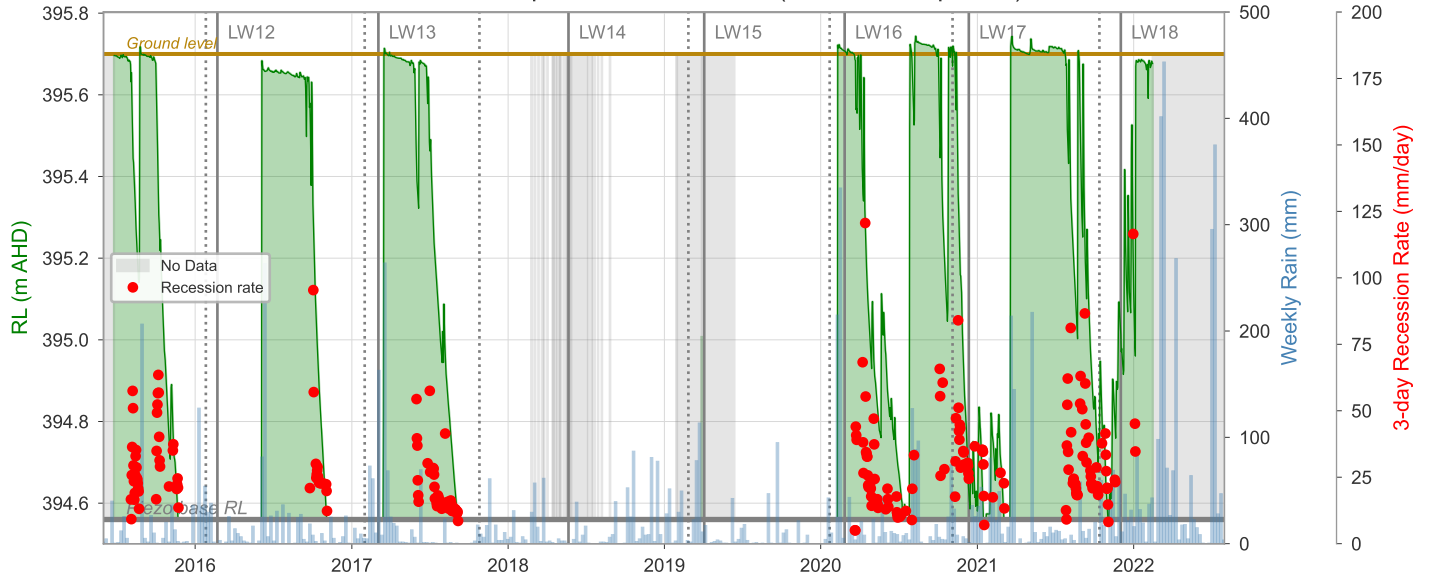
Dendrobium Swamp 85: Piezometer 03 (Within swamp EEC)



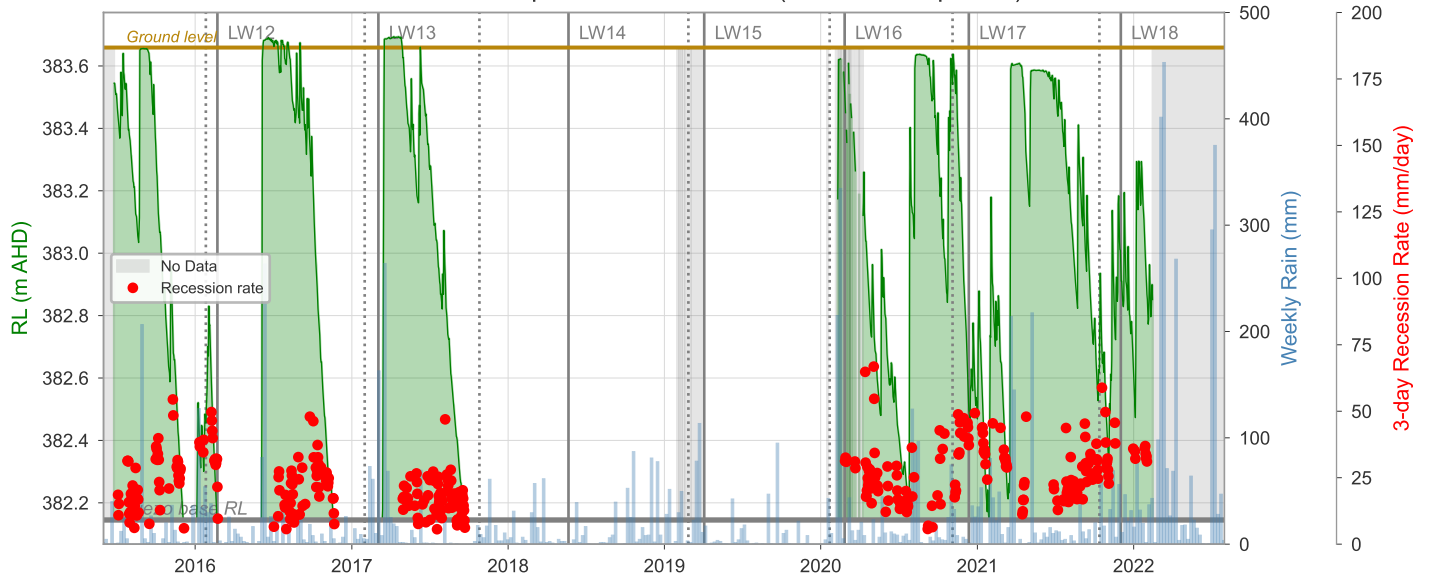
Dendrobium Swamp 85: Piezometer 02 (Within swamp EEC)



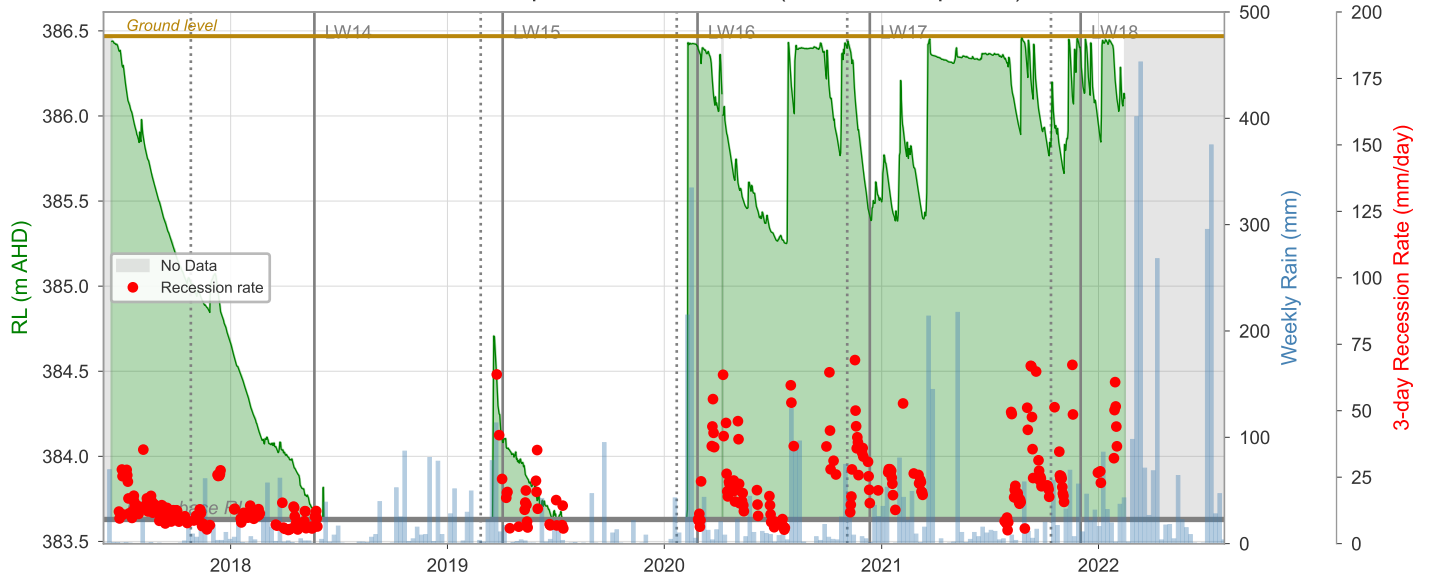
Dendrobium Swamp 85: Piezometer 01 (Within swamp EEC)



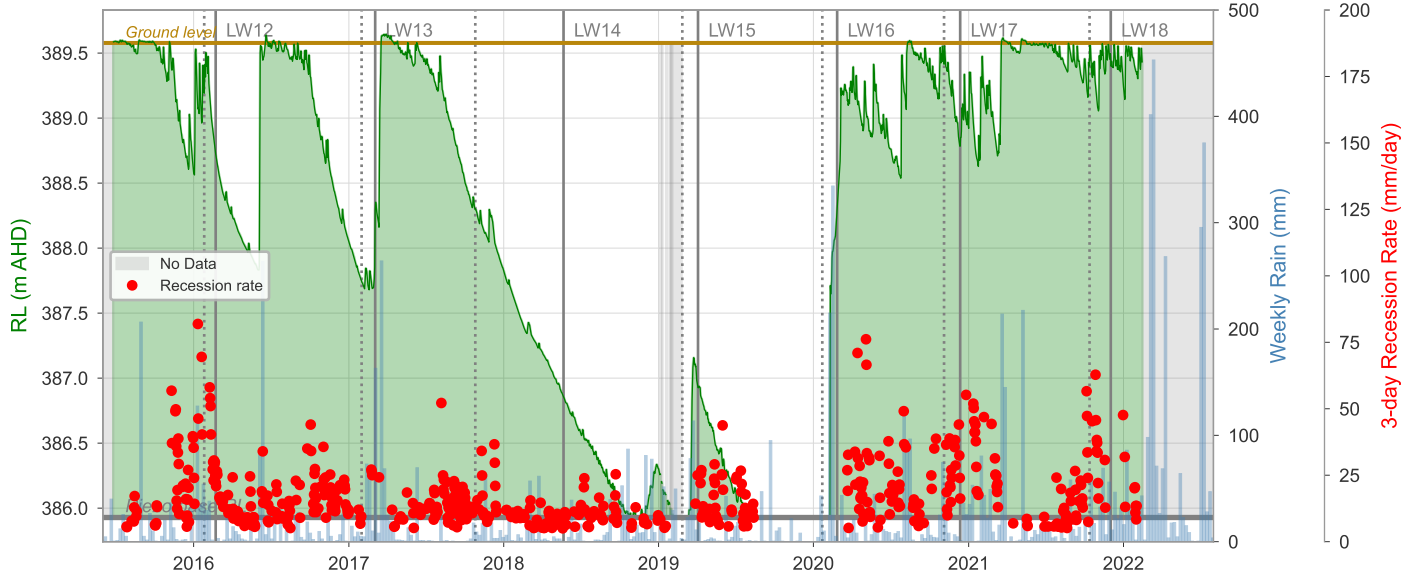
Dendrobium Swamp 86: Piezometer 01 (Within swamp EEC)



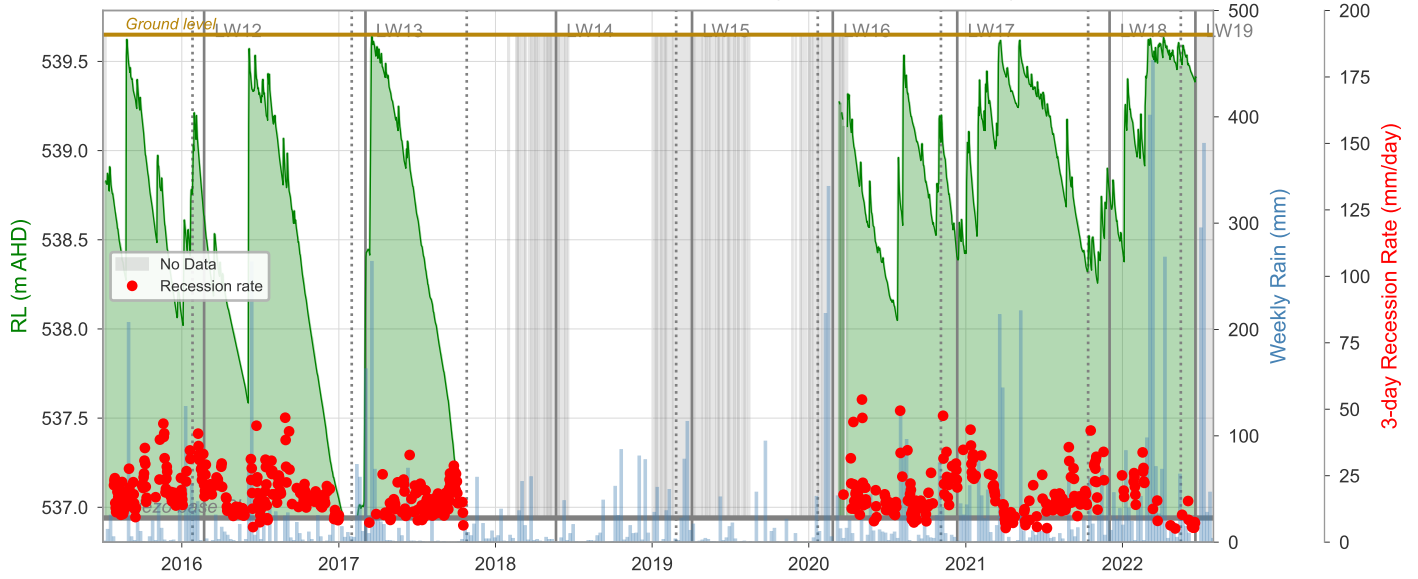
Dendrobium Swamp 86: Piezometer 03 (Within swamp EEC)



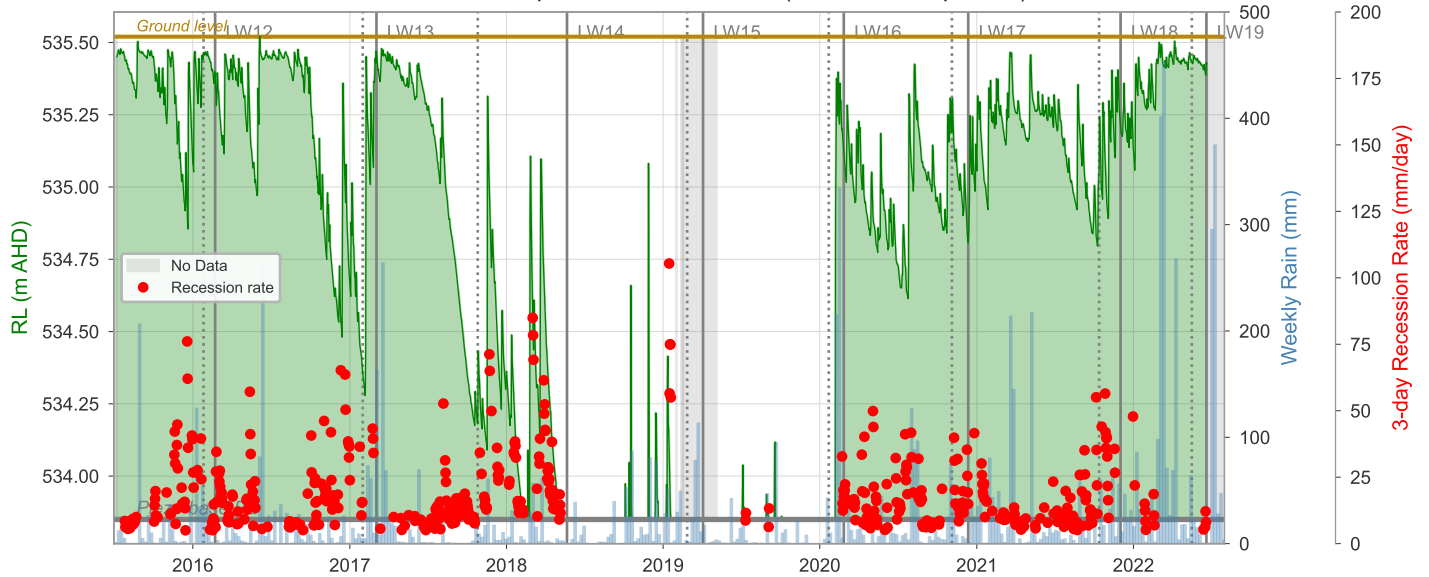
Dendrobium Swamp 86: Piezometer 02 (Within swamp EEC)



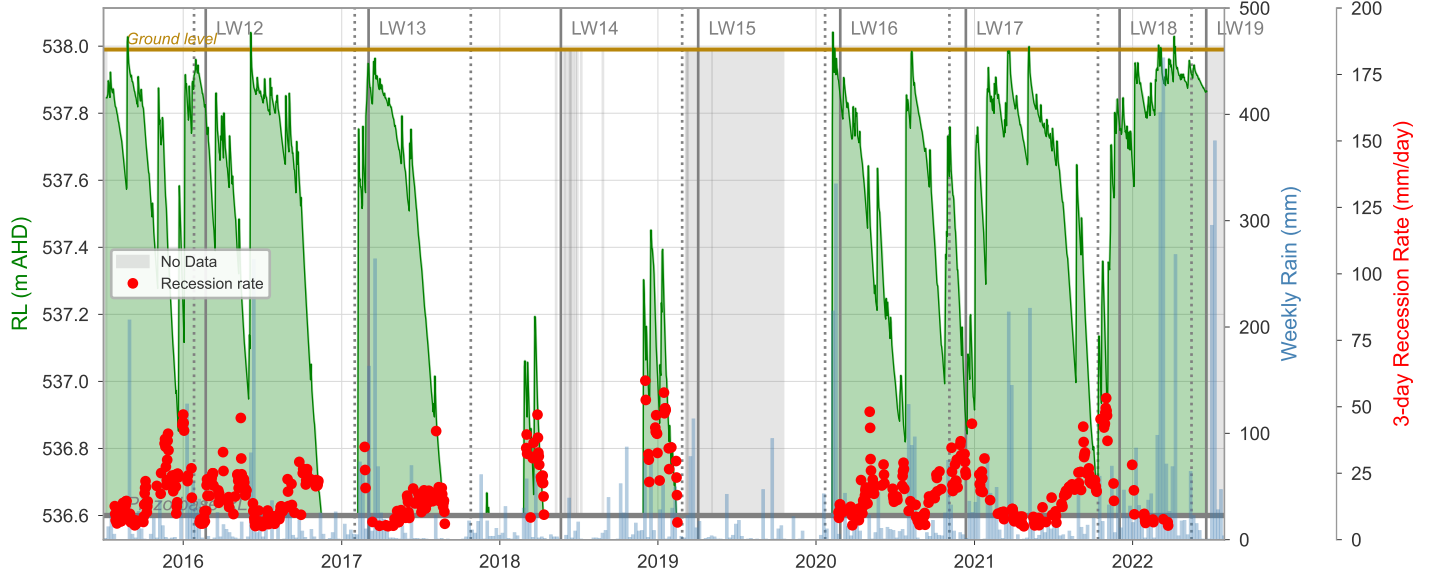
Dendrobium Swamp 87: Piezometer 01 (Within swamp EEC)



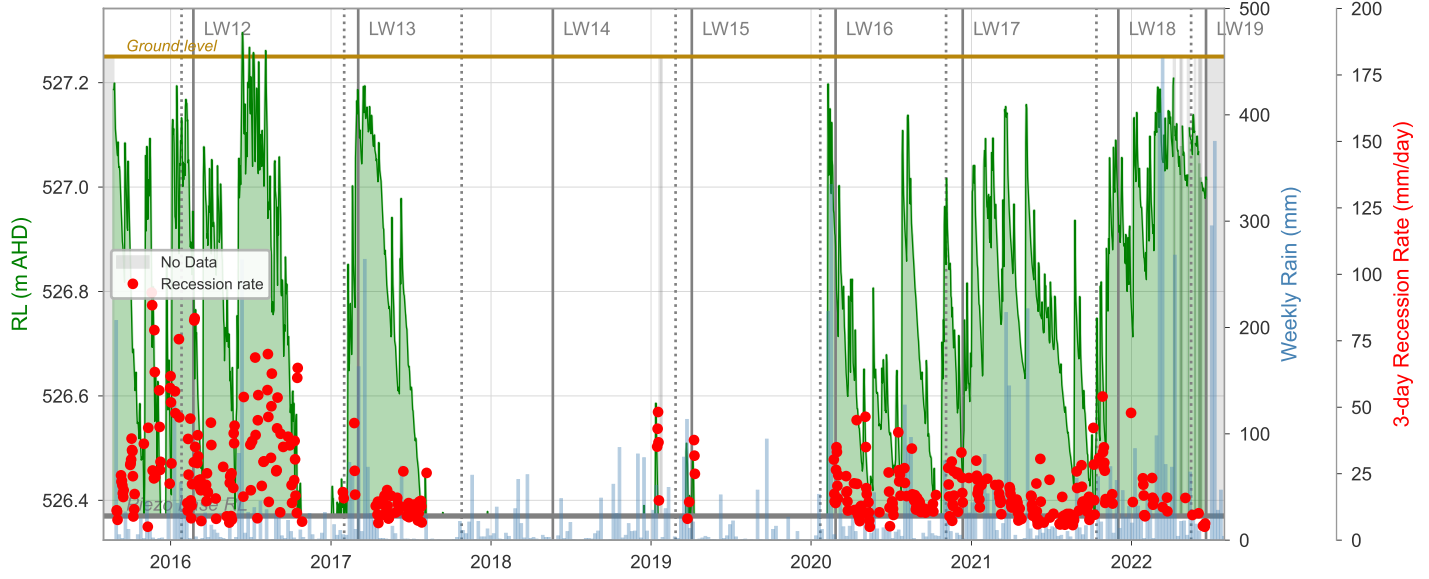
Dendrobium Swamp 87: Piezometer 02 (Within swamp EEC)



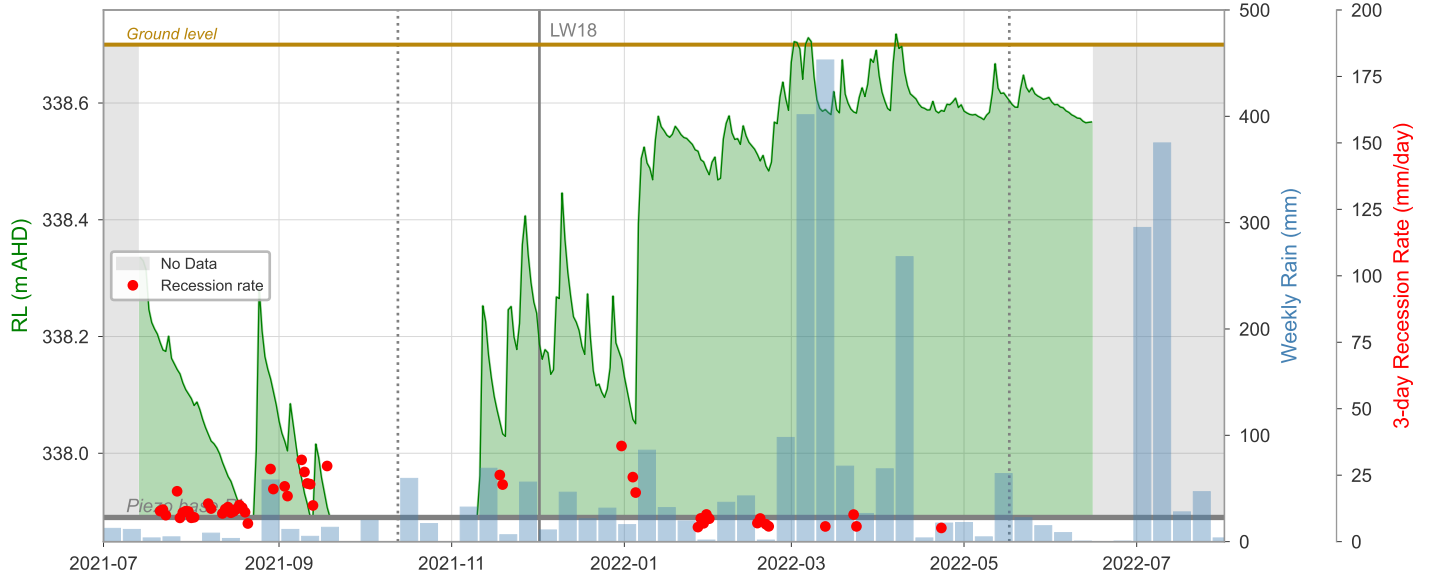
Dendrobium Swamp 88: Piezometer 01 (Within swamp EEC)



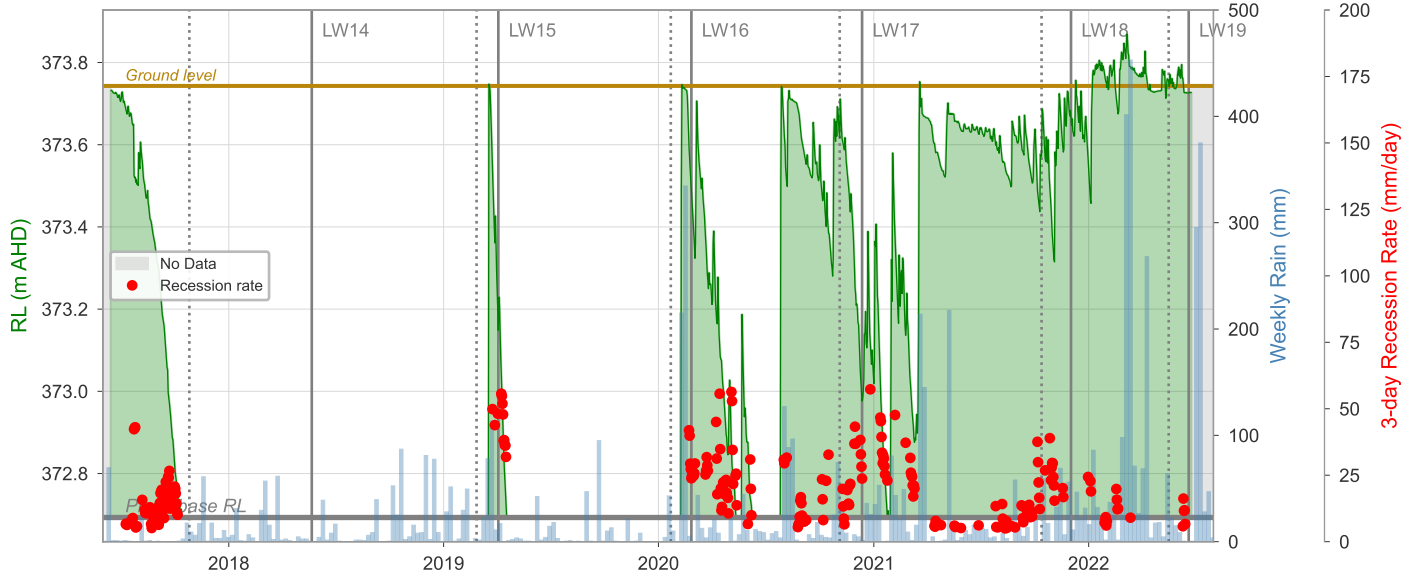
Dendrobium Swamp 88: Piezometer 02 (Within swamp EEC)



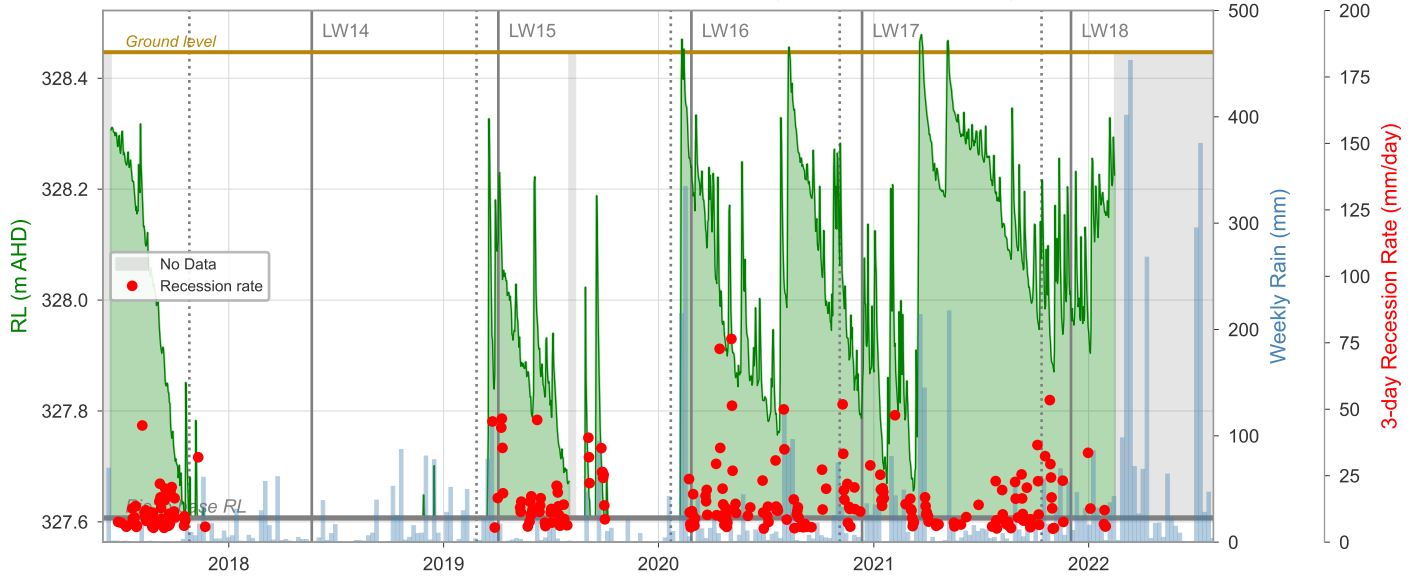
Dendrobium Swamp 95: Piezometer 01 (Within swamp EEC)



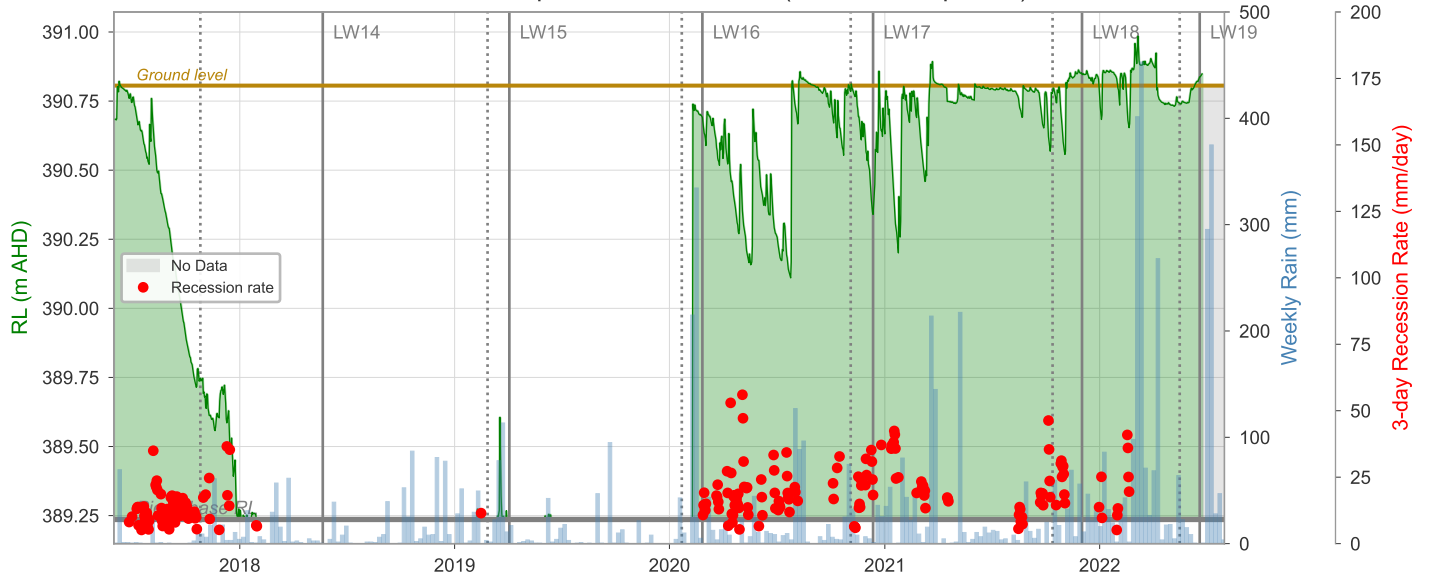
Dendrobium Swamp 97: Piezometer 01 (Within swamp EEC)



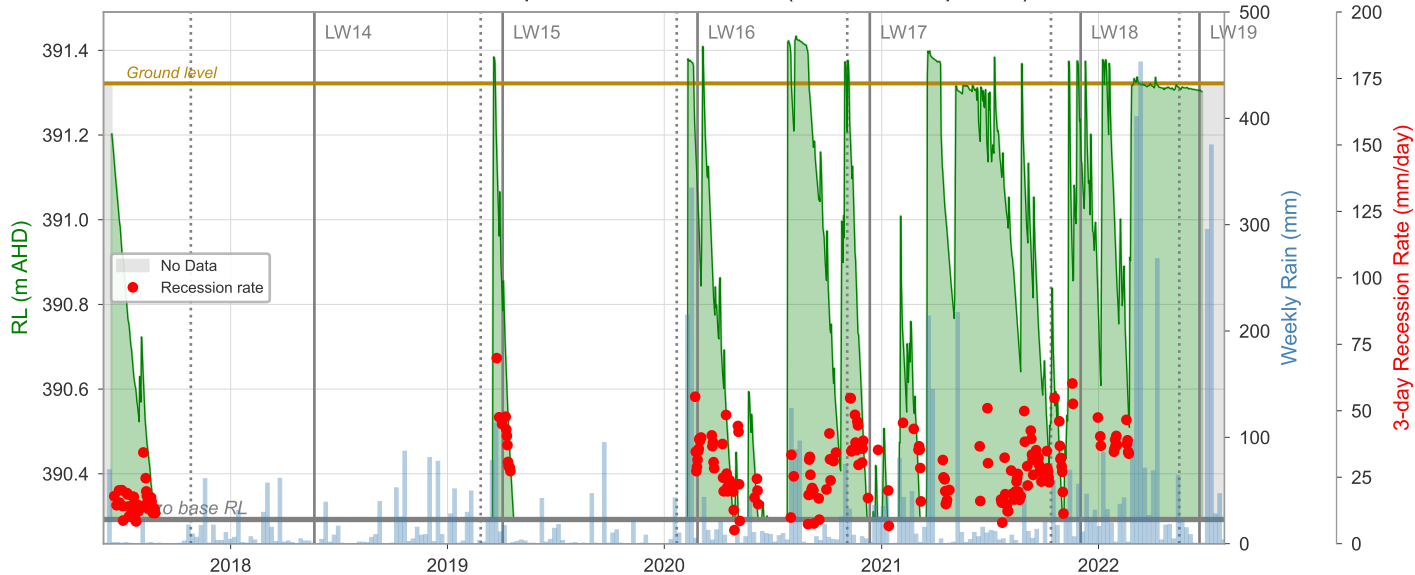
Dendrobium Swamp 98: Piezometer 01 (Within swamp EEC)



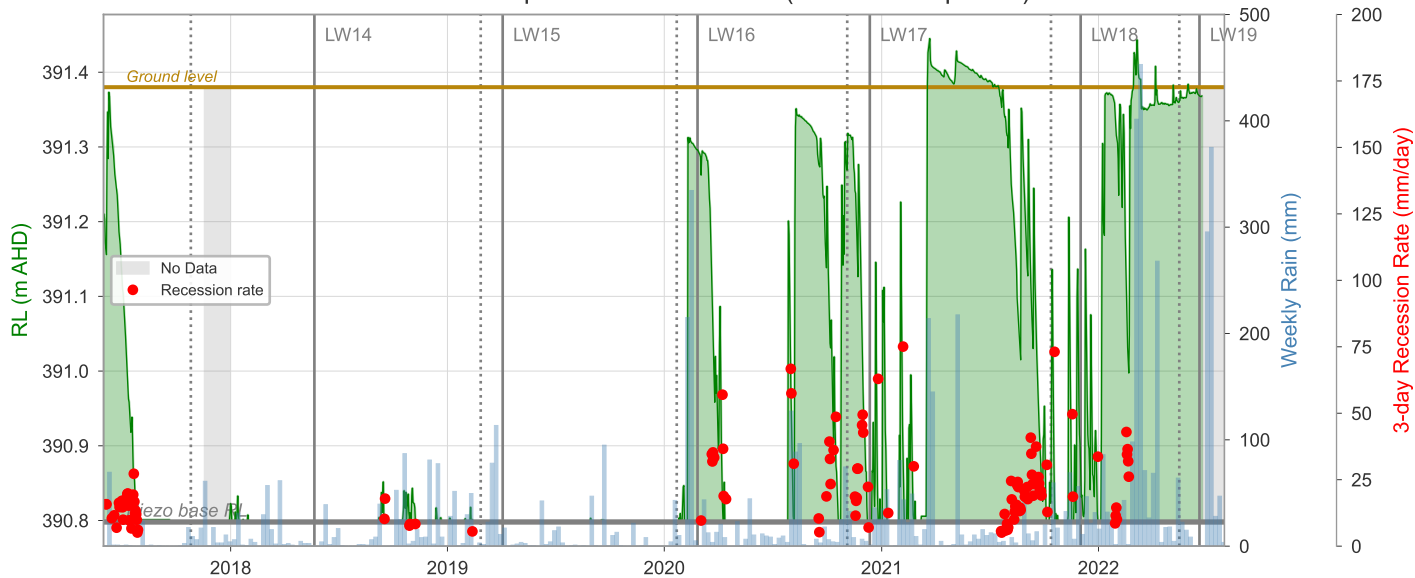
Dendrobium Swamp 99: Piezometer 01 (Within swamp EEC)



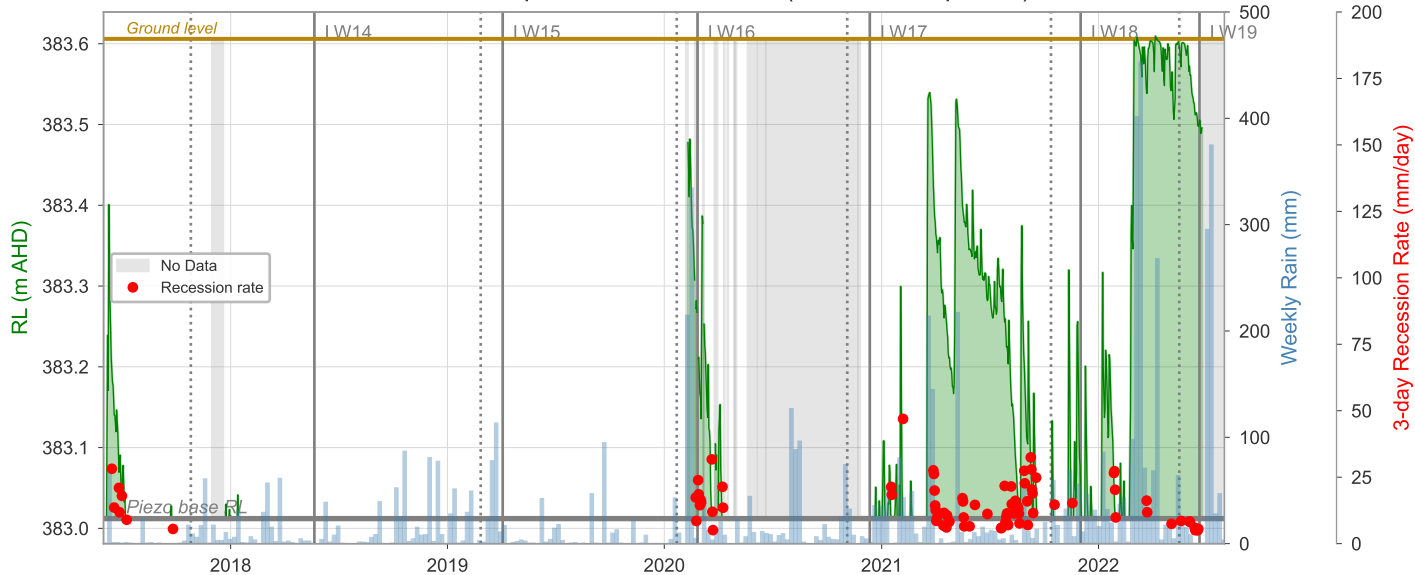
Dendrobium Swamp 100: Piezometer 01 (Within swamp EEC)



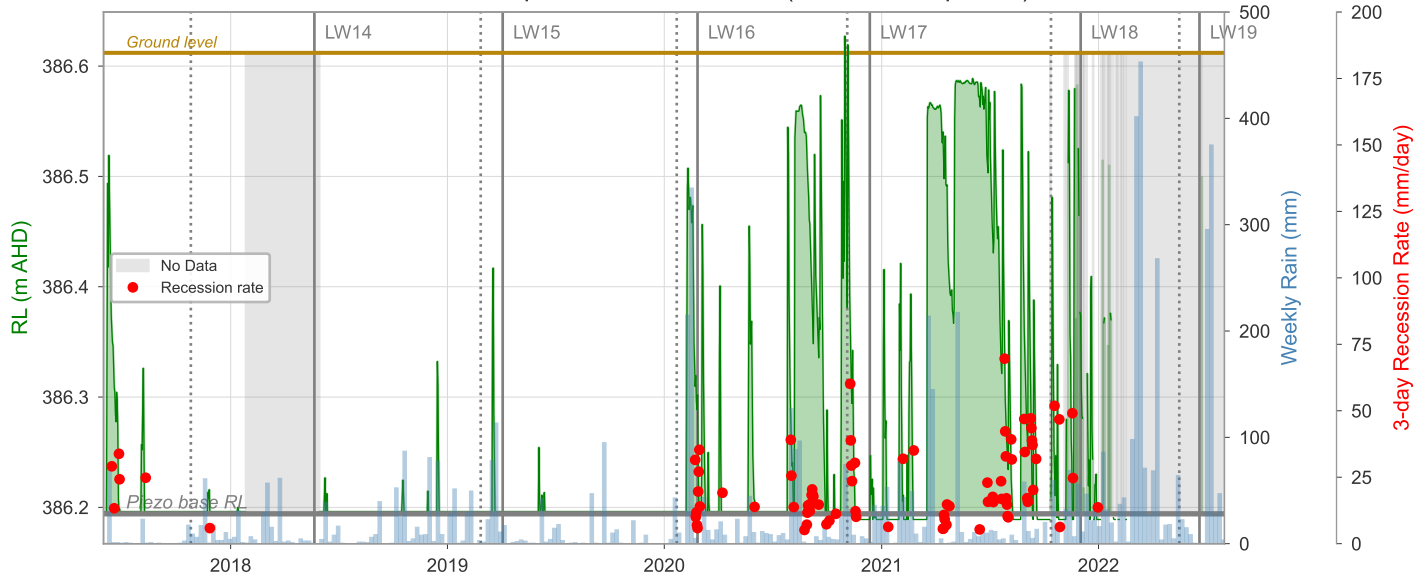
Dendrobium Swamp 101: Piezometer 01 (Within swamp EEC)



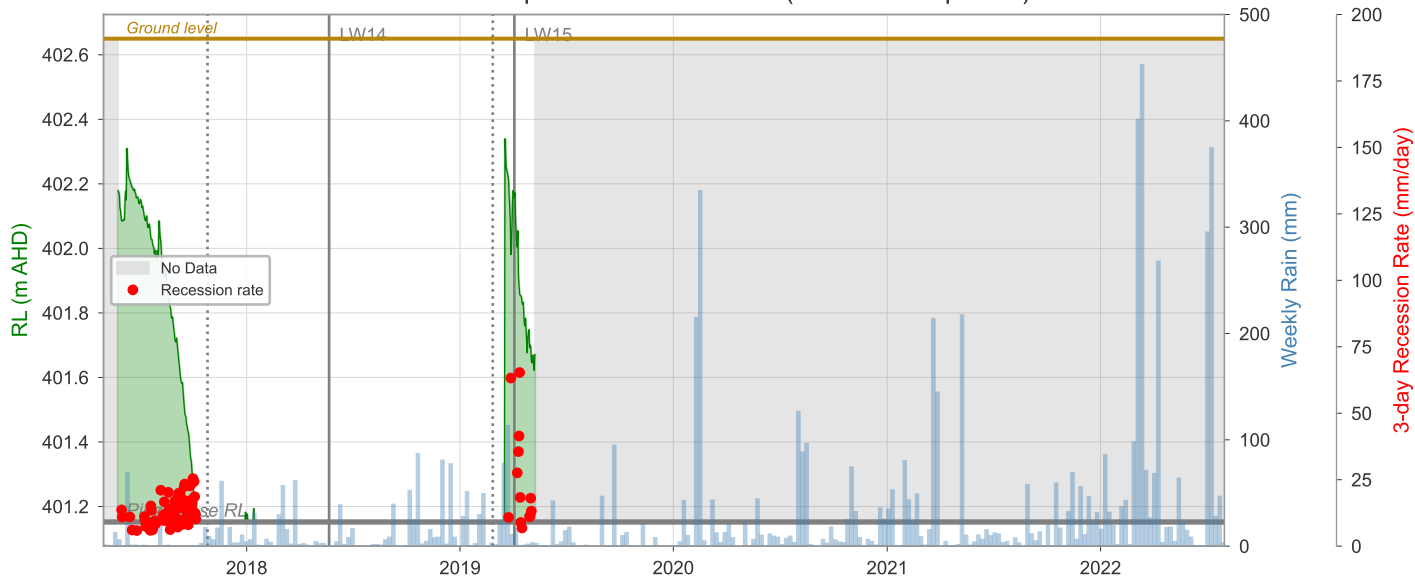
Dendrobium Swamp 103: Piezometer 01 (Within swamp EEC)



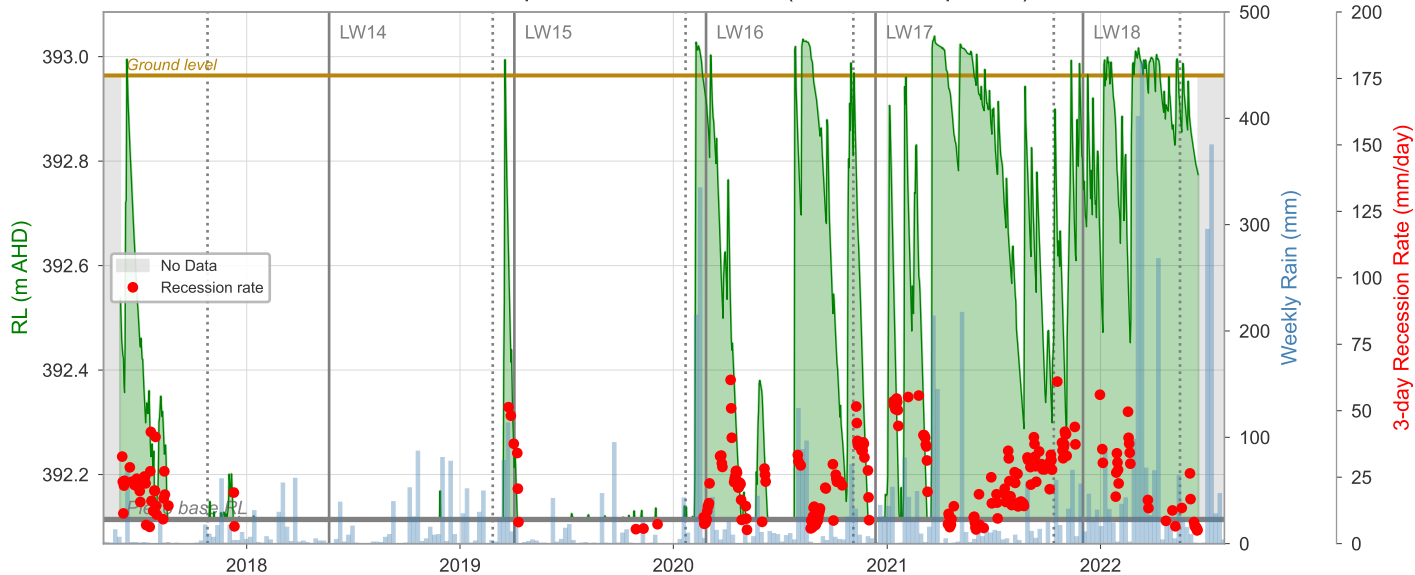
Dendrobium Swamp 105: Piezometer 01 (Within swamp EEC)



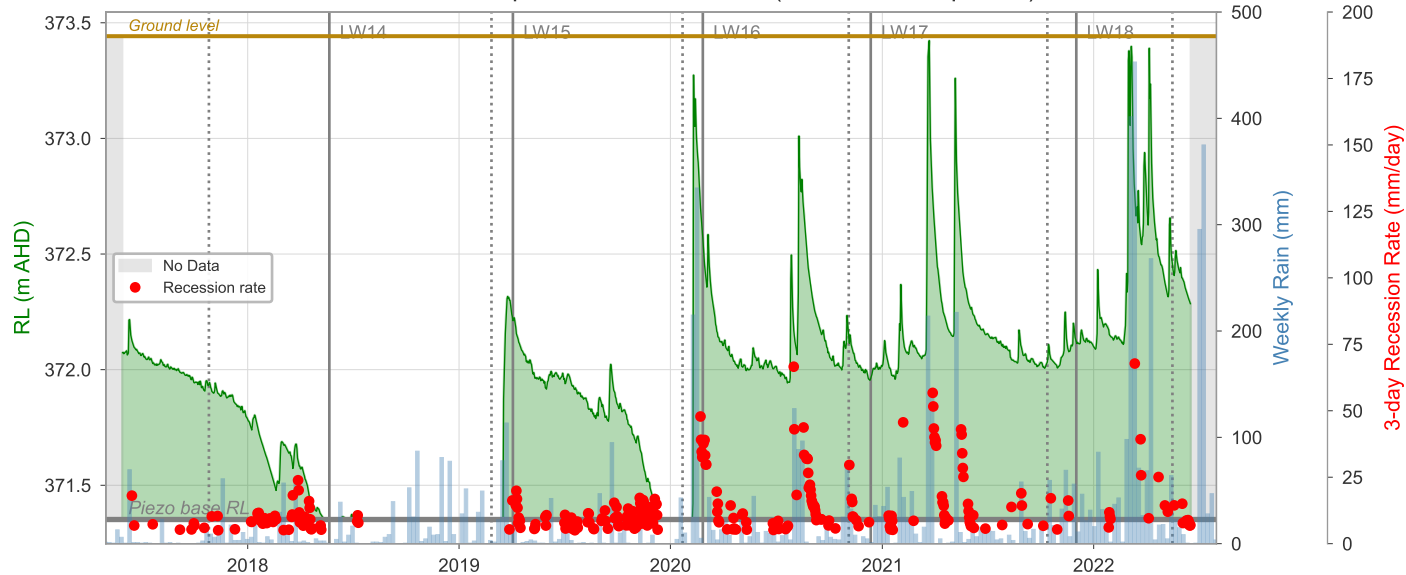
Dendrobium Swamp 106: Piezometer 01 (Within swamp EEC)



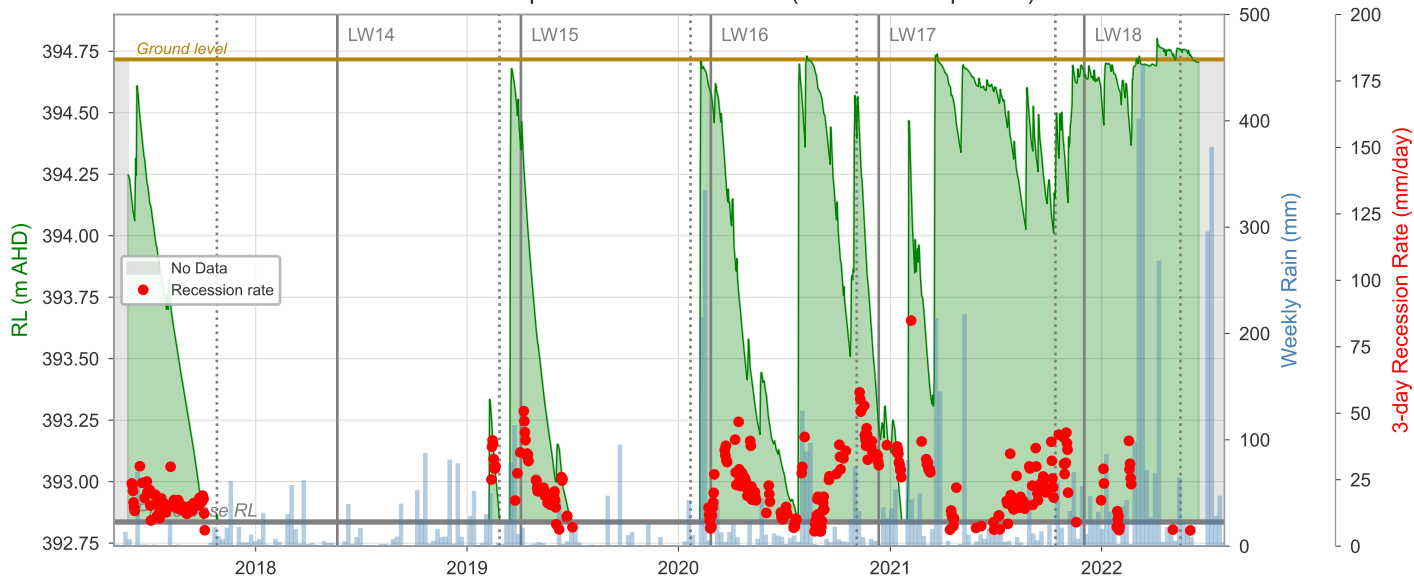
Dendrobium Swamp 107: Piezometer 01 (Within swamp EEC)



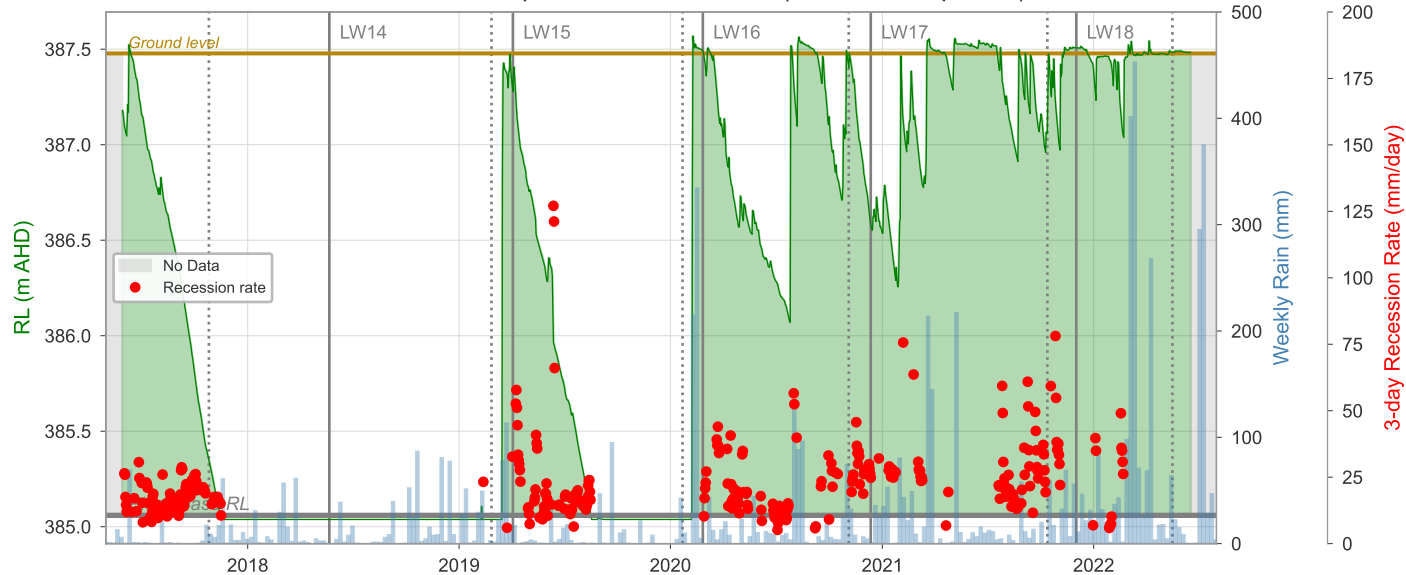
Dendrobium Swamp 108: Piezometer 01 (Outside swamp EEC)



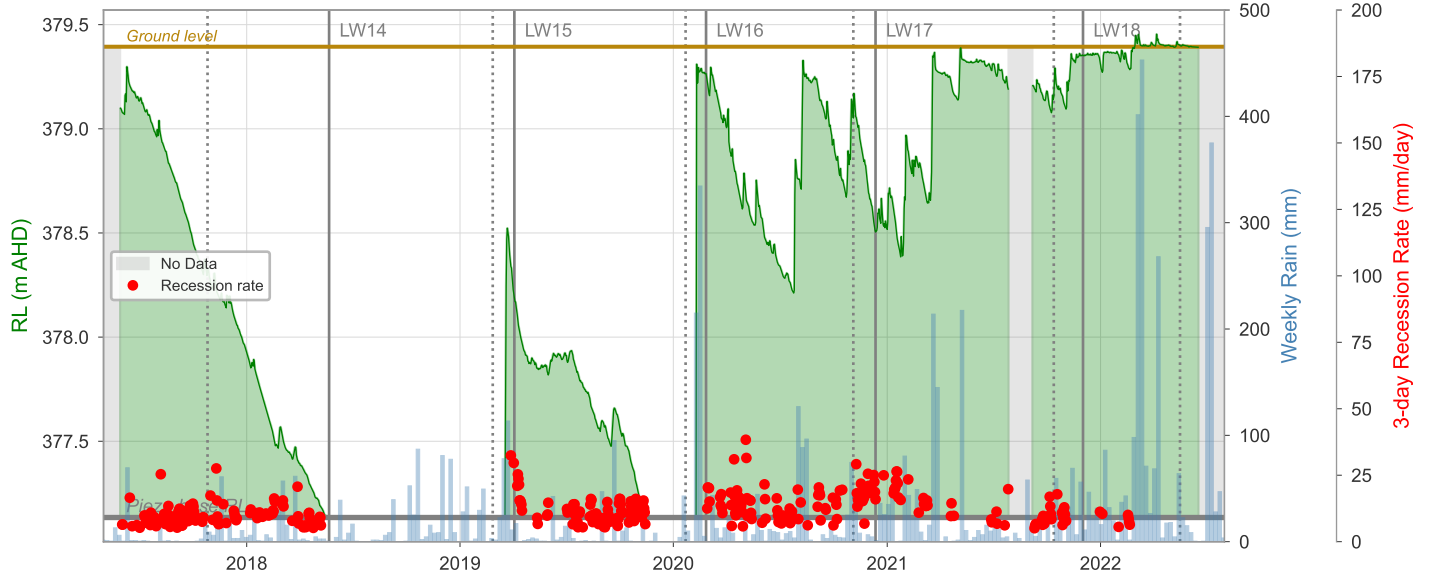
Dendrobium Swamp 109: Piezometer 01 (Within swamp EEC)



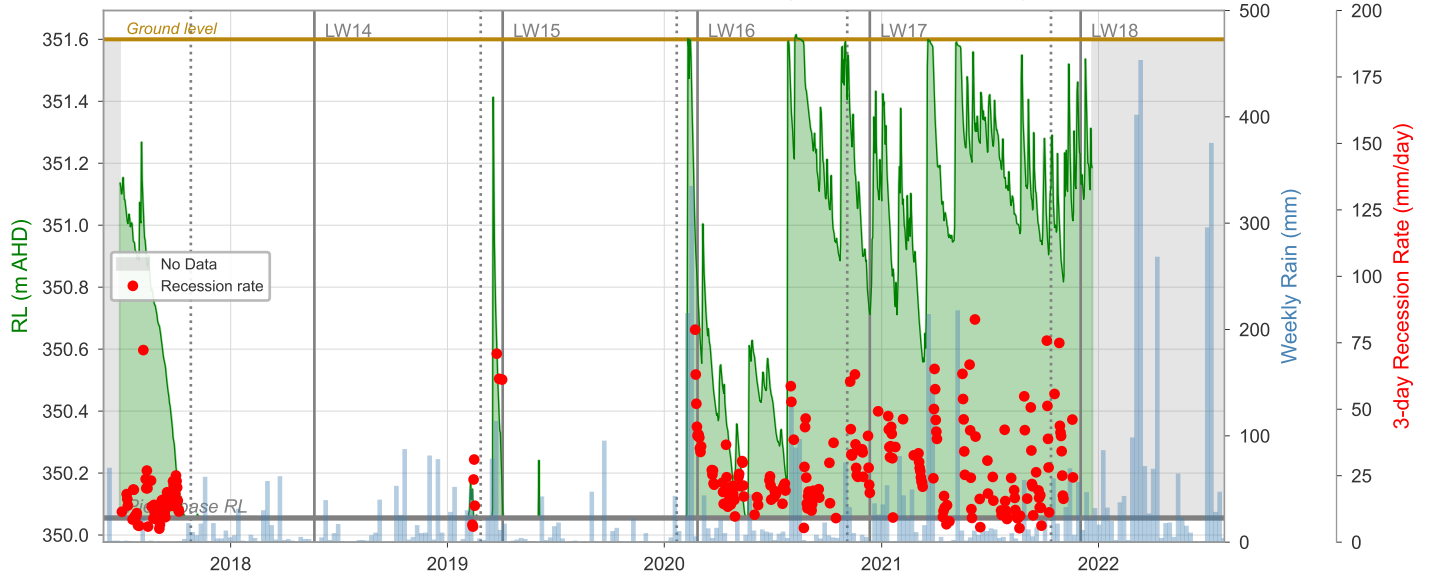
Dendrobium Swamp 110: Piezometer 01 (Within swamp EEC)



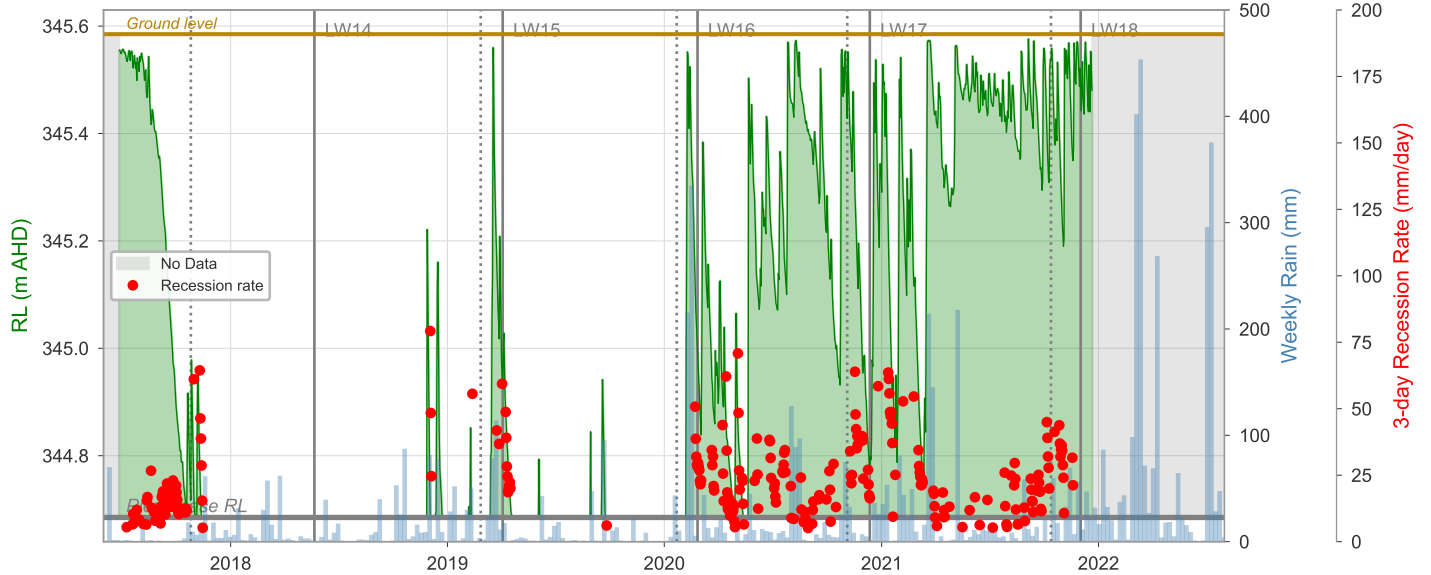
Dendrobium Swamp 111: Piezometer 01 (Within swamp EEC)



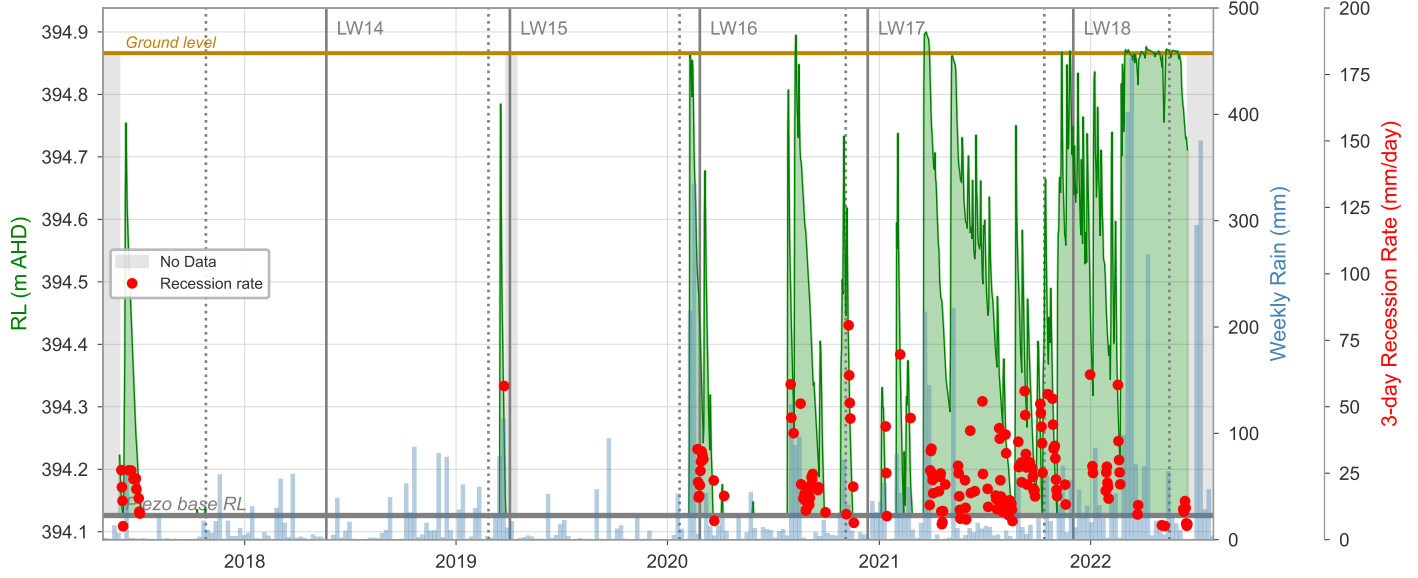
Dendrobium Swamp 112: Piezometer 01 (Within swamp EEC)



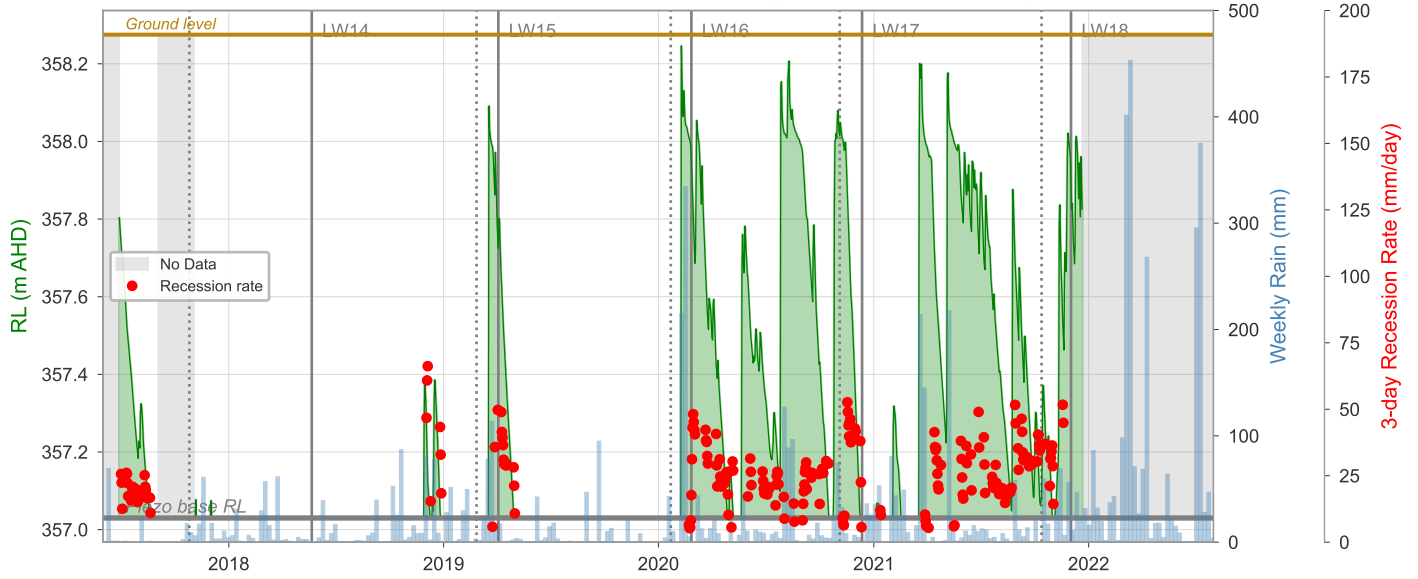
Dendrobium Swamp 113: Piezometer 01 (Within swamp EEC)



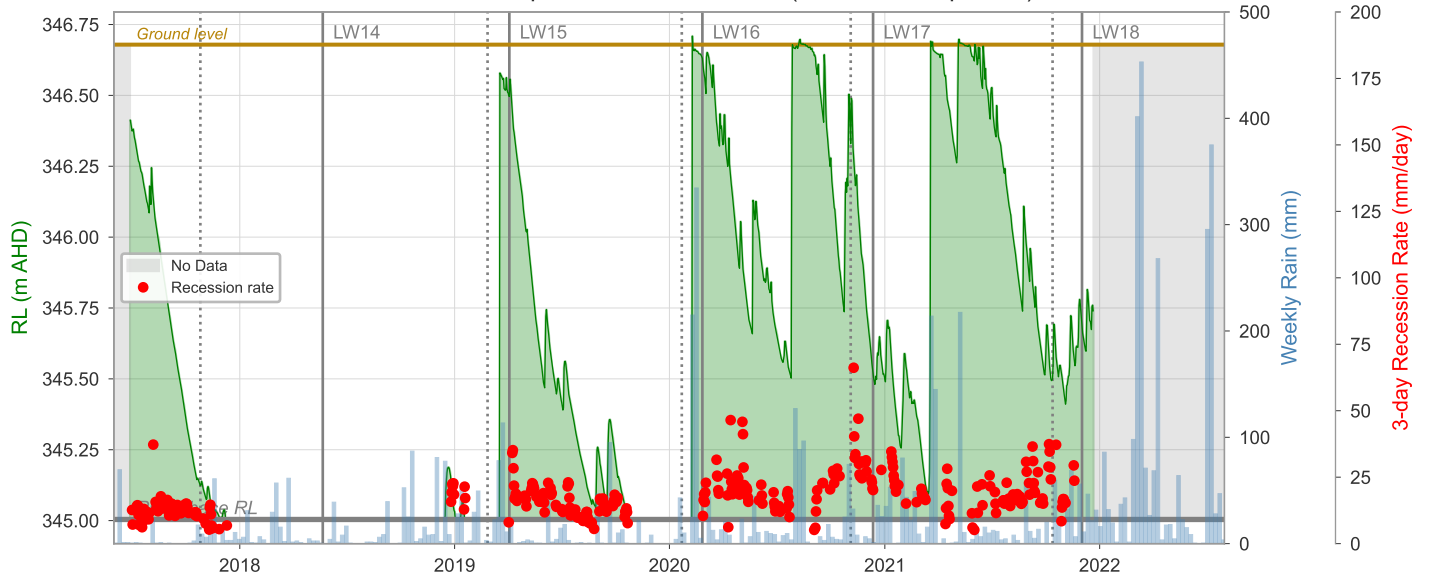
Dendrobium Swamp 114: Piezometer 01 (Within swamp EEC)



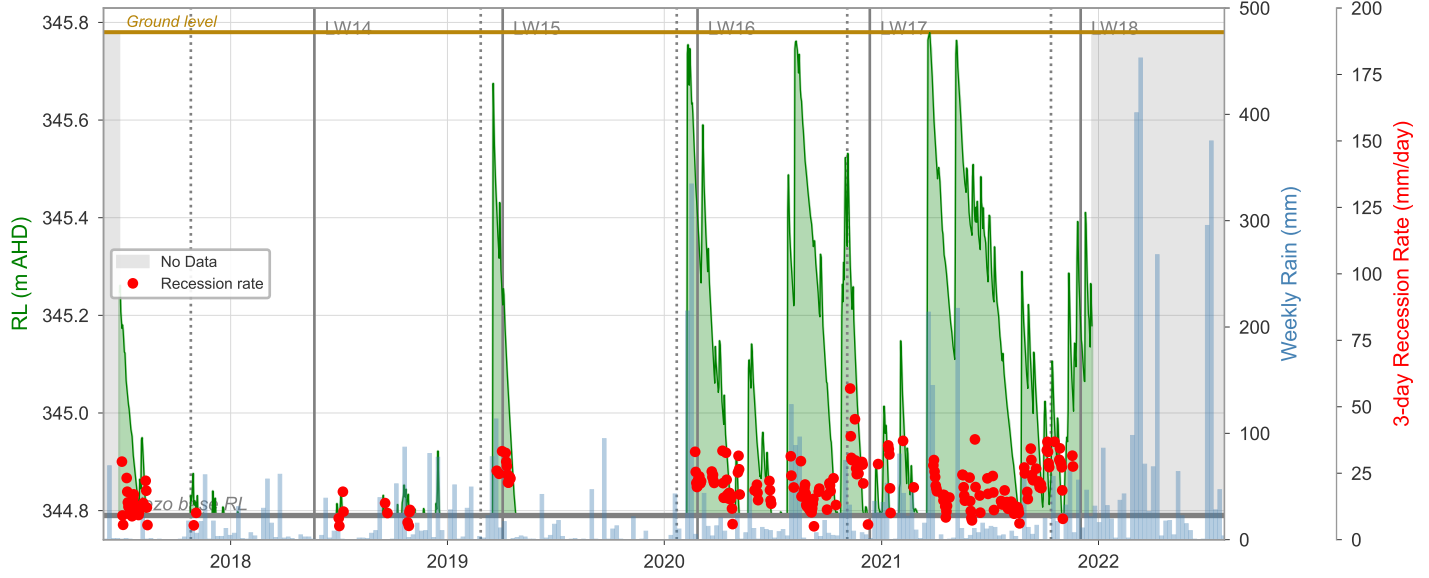
Dendrobium Swamp 115: Piezometer 01 (Outside swamp EEC)



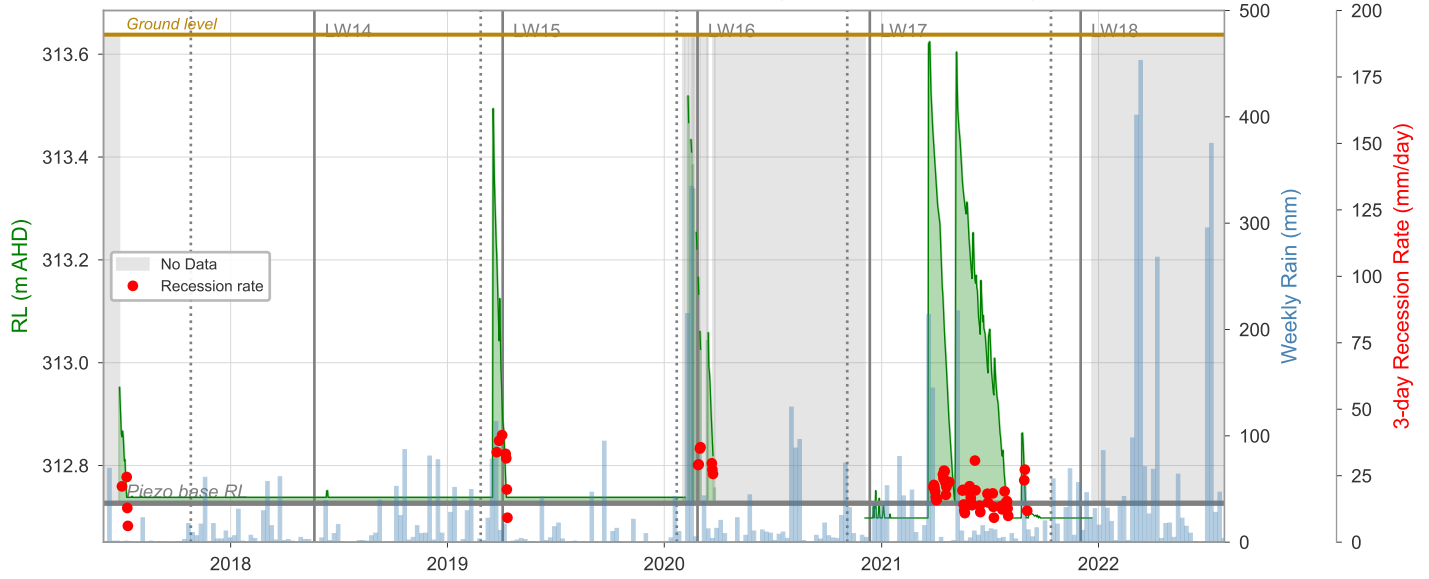
Dendrobium Swamp 116: Piezometer 01 (Within swamp EEC)



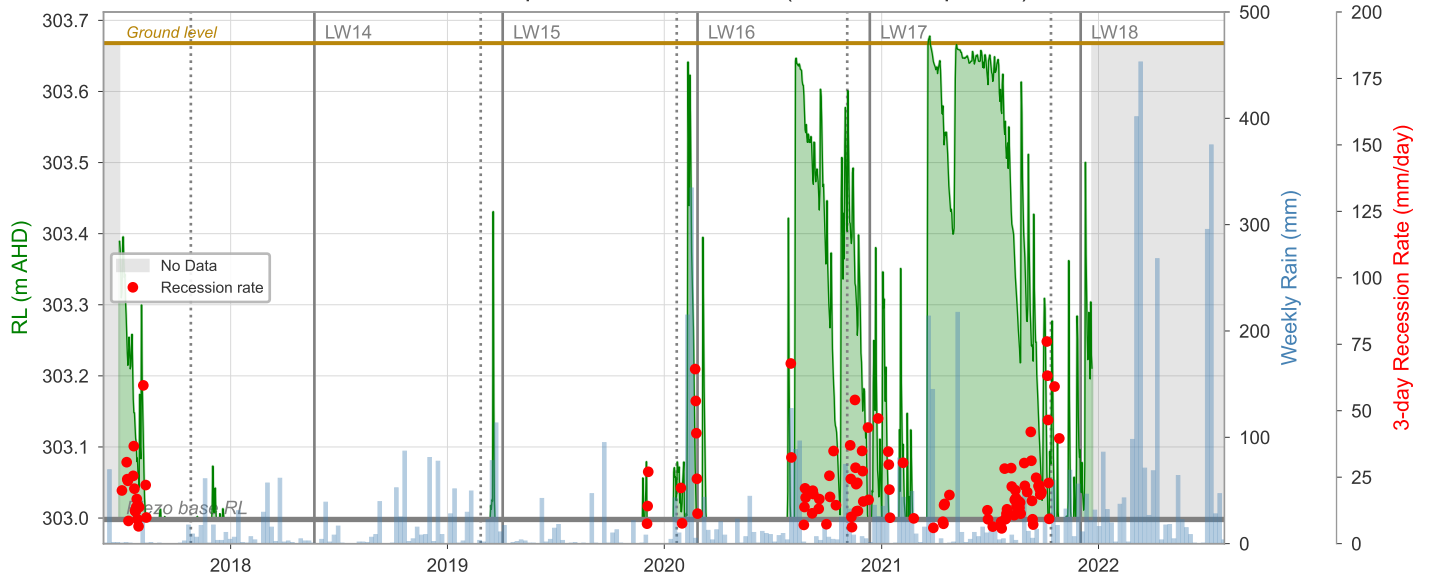
Dendrobium Swamp 117: Piezometer 01 (Within swamp EEC)



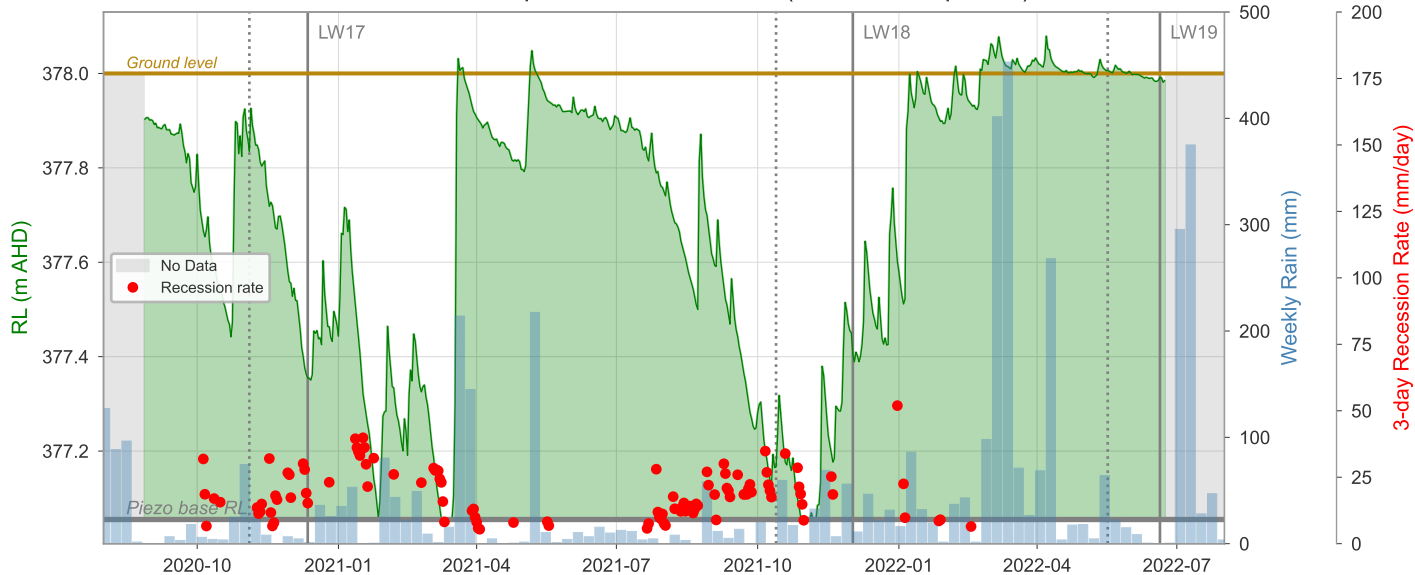
Dendrobium Swamp 118: Piezometer 01 (Outside swamp EEC)



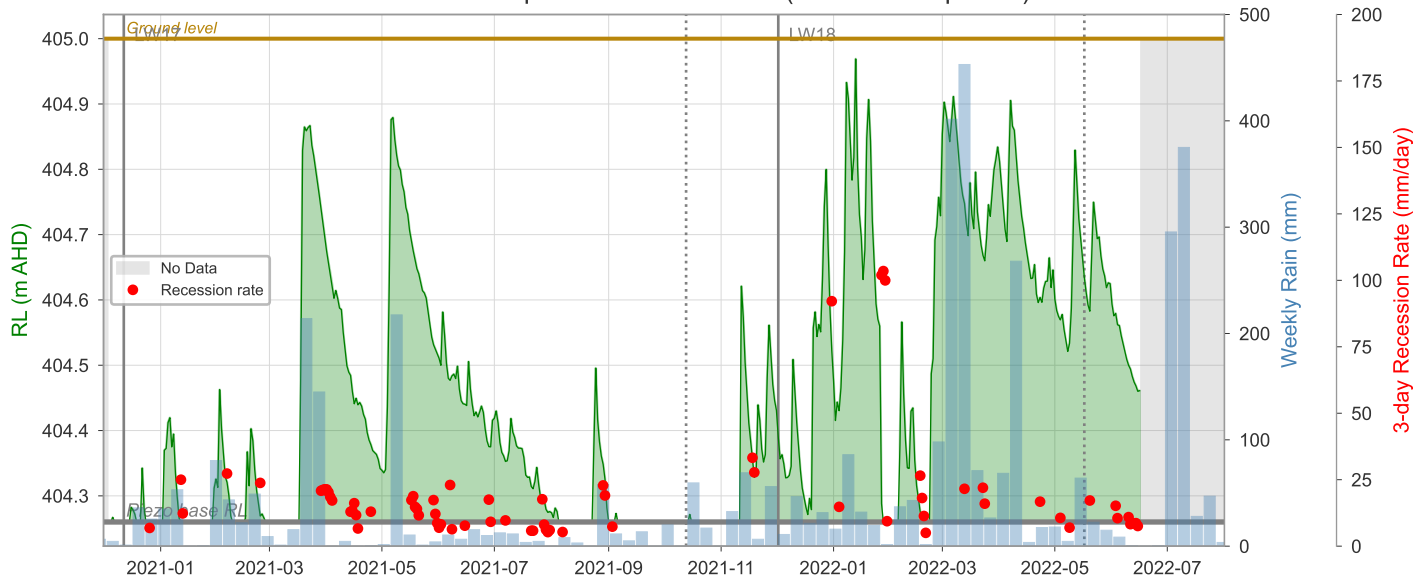
Dendrobium Swamp 119: Piezometer 01 (Within swamp EEC)



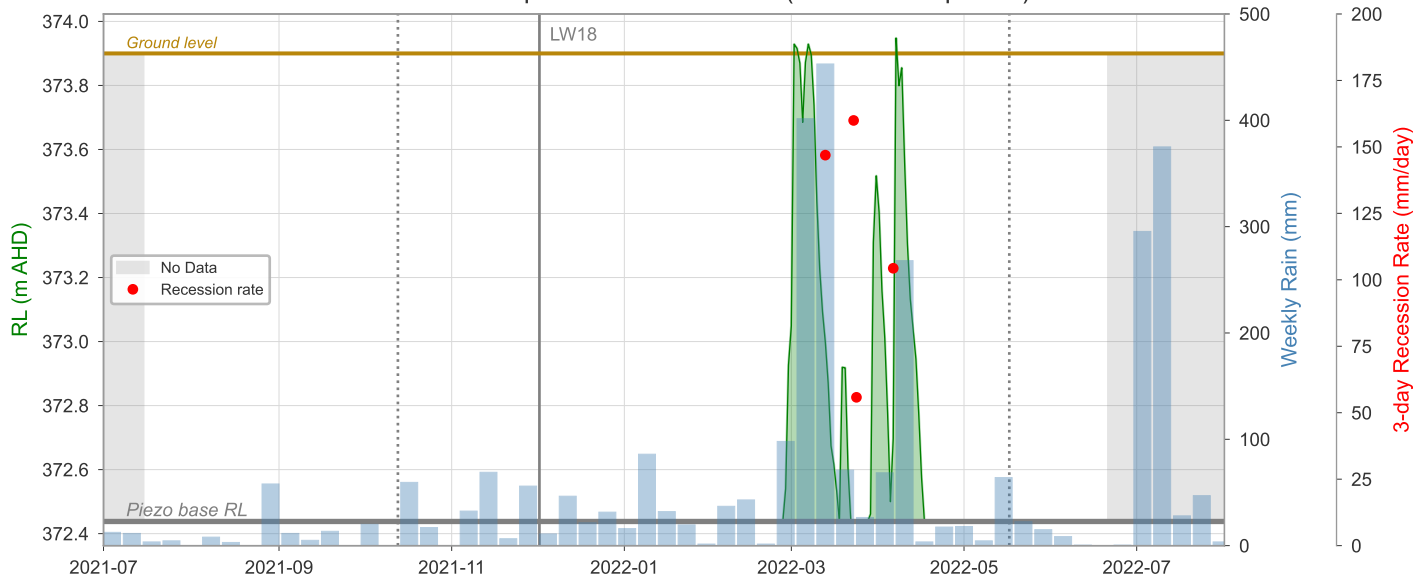
Dendrobium Swamp 144: Piezometer 01 (Within swamp EEC)



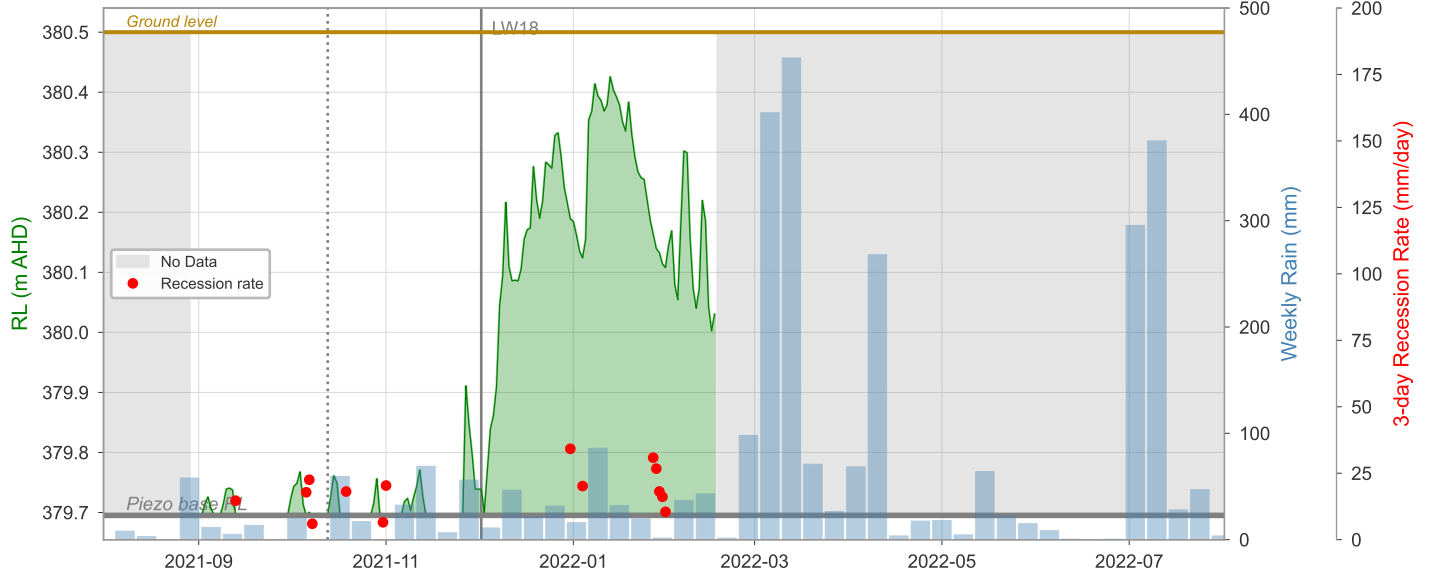
Dendrobium Swamp 145: Piezometer 01 (Within swamp EEC)



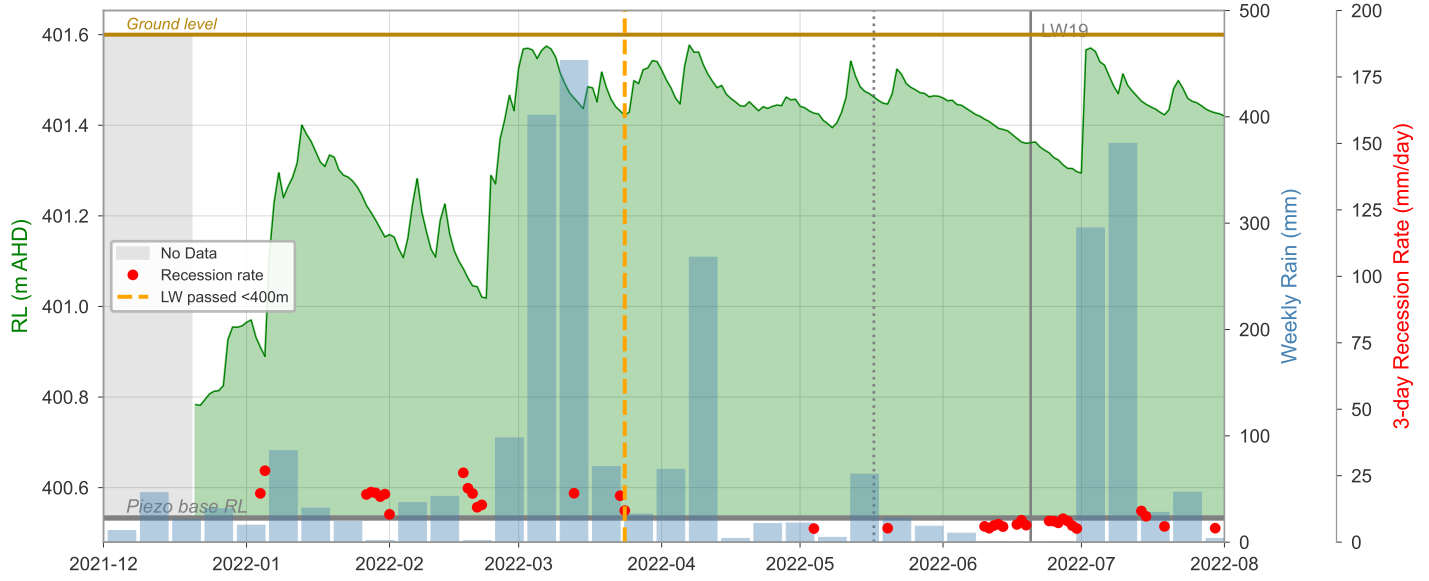
Dendrobium Swamp 146: Piezometer 01 (Within swamp EEC)



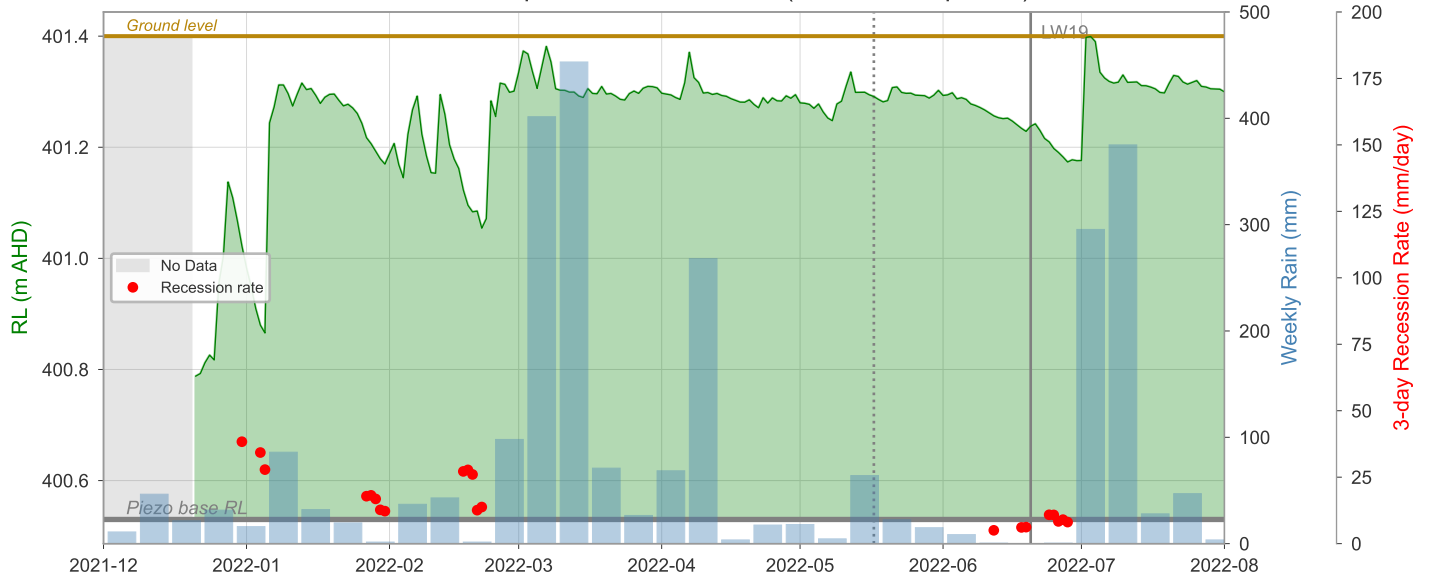
Dendrobium Swamp 148: Piezometer 01 (Within swamp EEC)



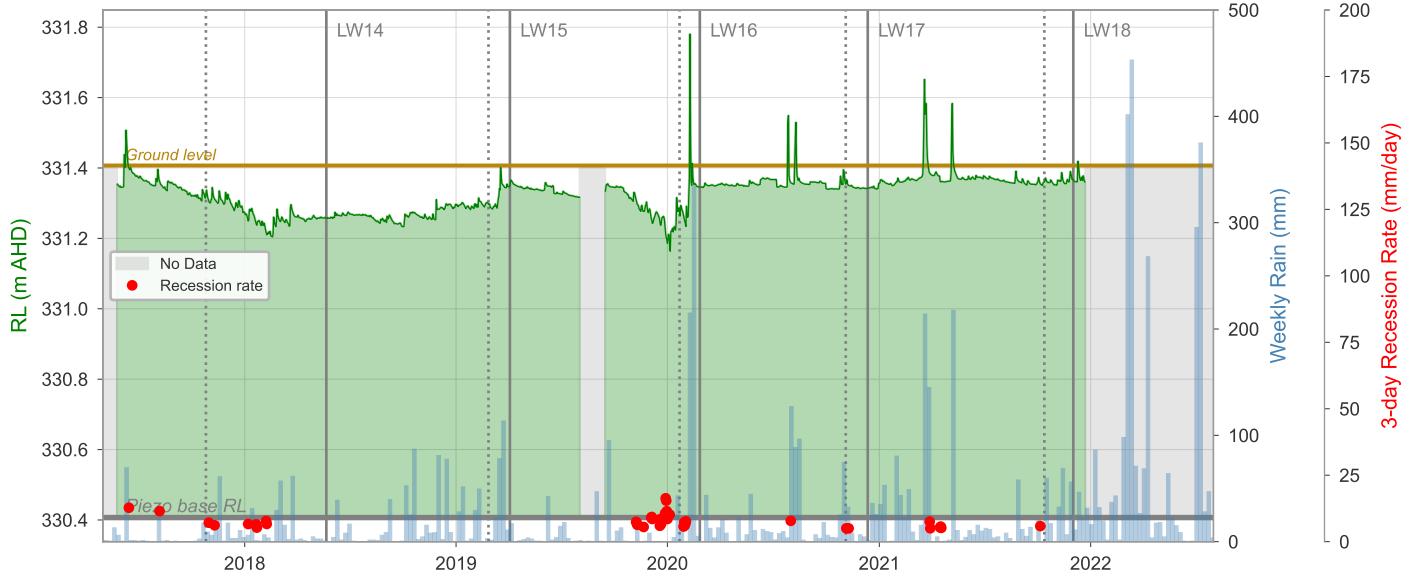
Dendrobium Swamp 150: Piezometer 01 (Within swamp EEC)



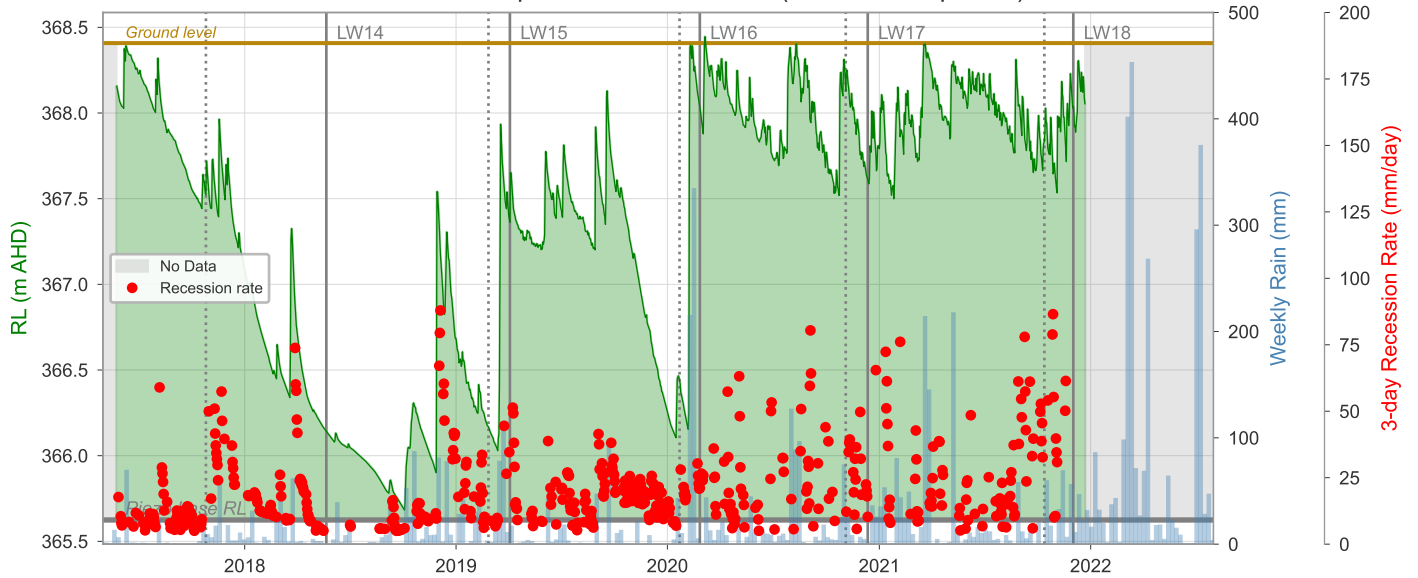
Dendrobium Swamp 151: Piezometer 01 (Within swamp EEC)



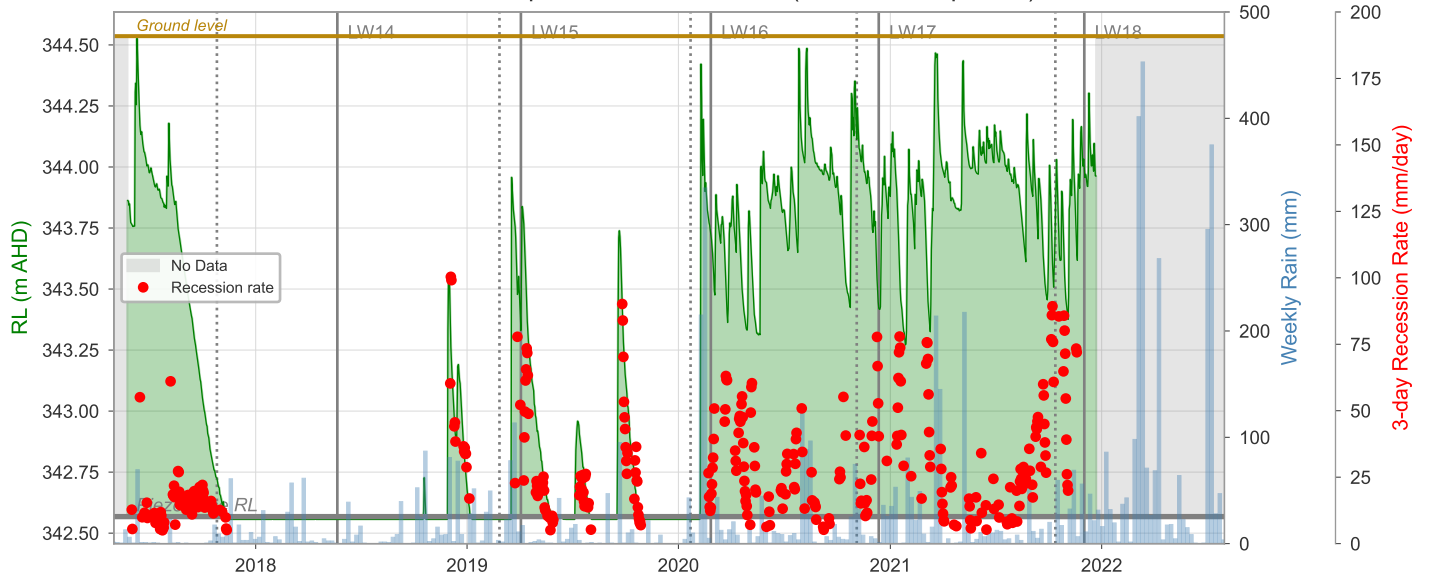
Dendrobium Swamp A07: Piezometer 03 (Within swamp EEC)



Dendrobium Swamp A07: Piezometer 01 (Within swamp EEC)

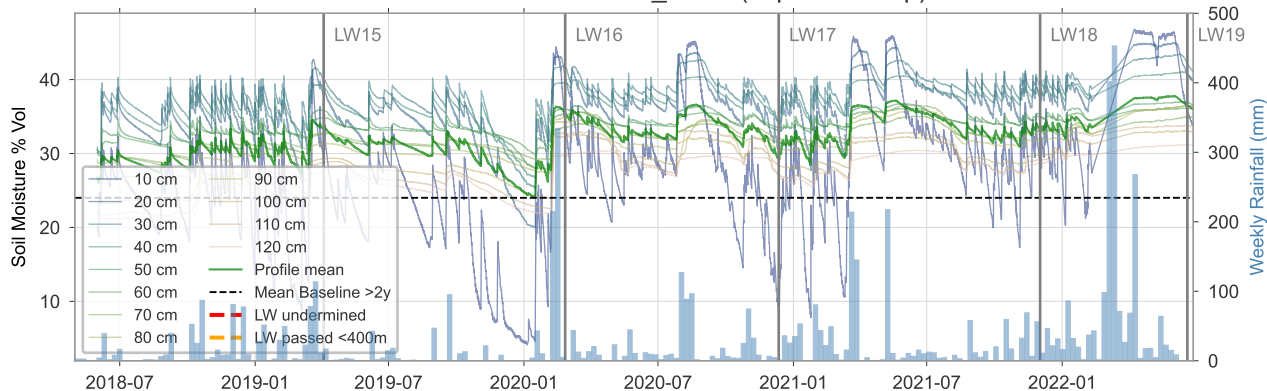


Dendrobium Swamp A07: Piezometer 02 (Within swamp EEC)

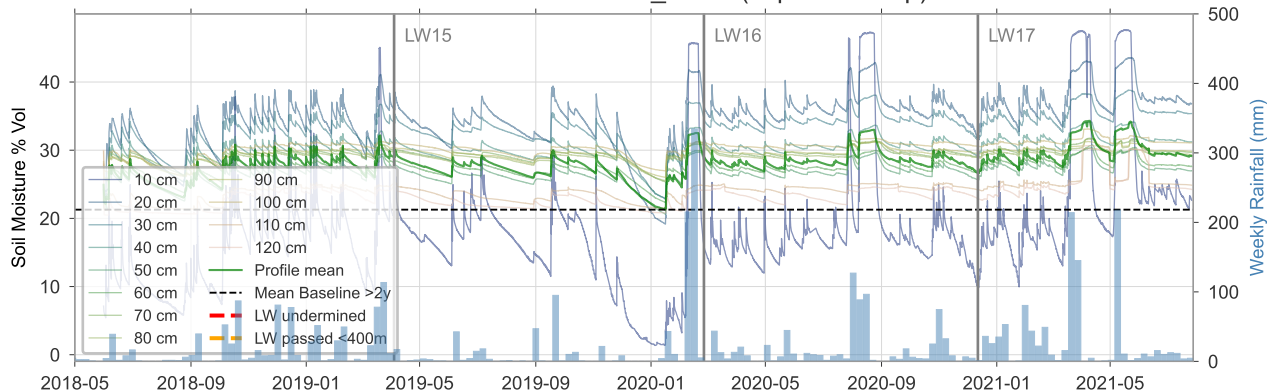


Appendix E: Soil moisture hydrographs

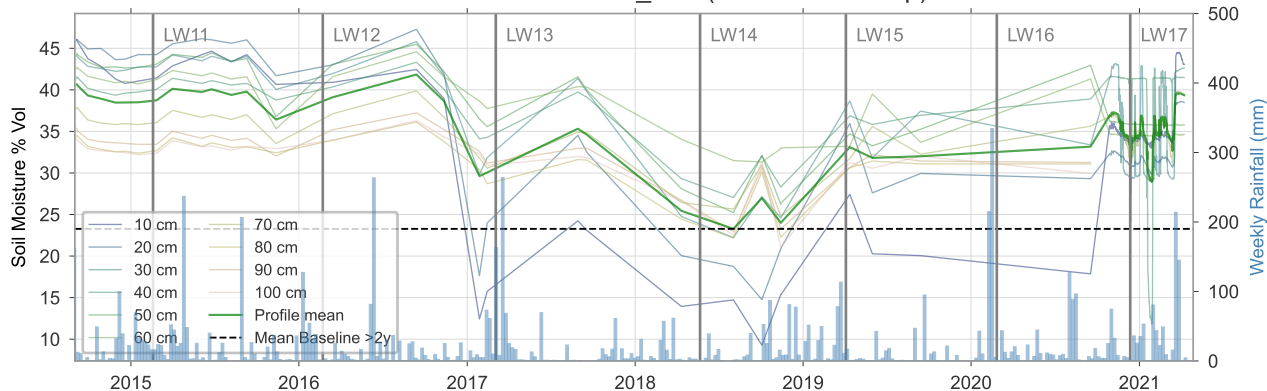
Soil moisture sensor S01b_S02iii (impact swamp)



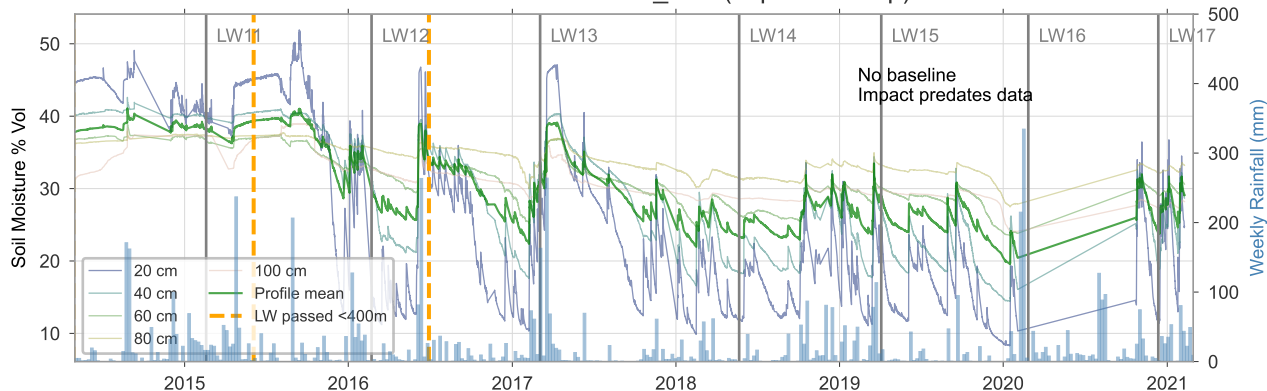
Soil moisture sensor S01b_S02iv (impact swamp)



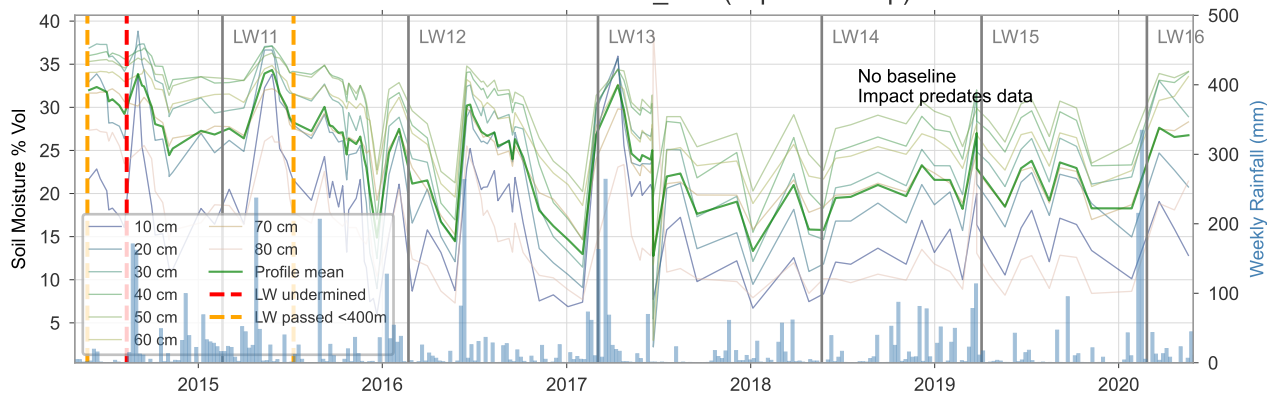
Soil moisture sensor S02_S01 (reference swamp)



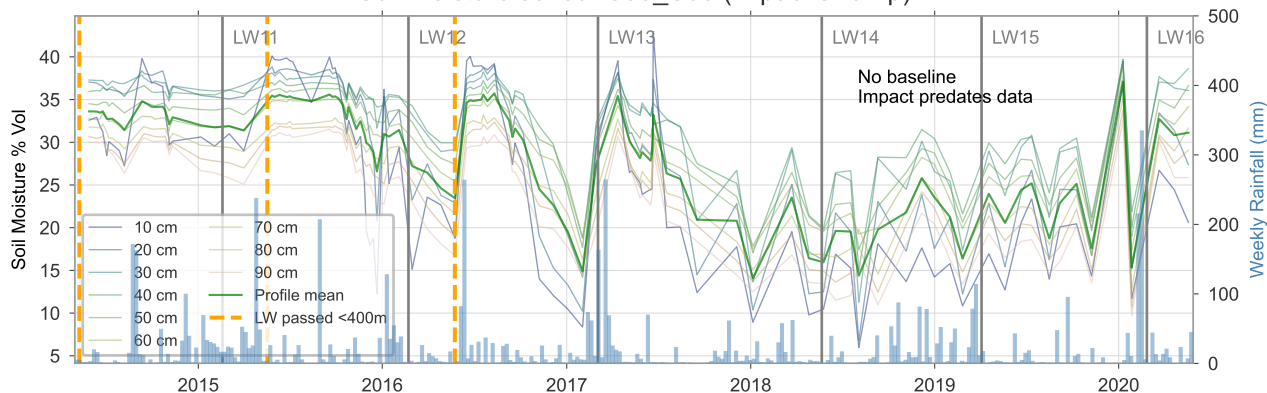
Soil moisture sensor S05_S01 (impact swamp)



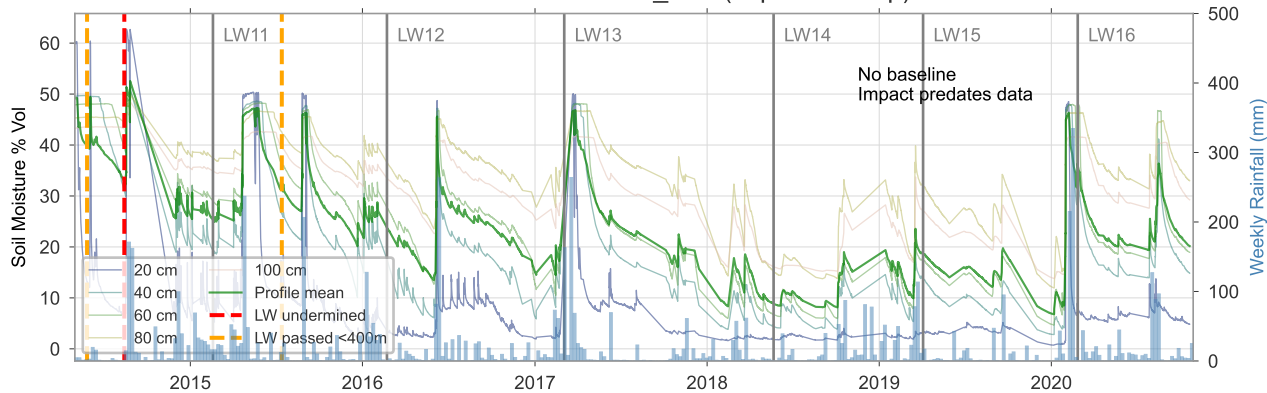
Soil moisture sensor S05_S02 (impact swamp)



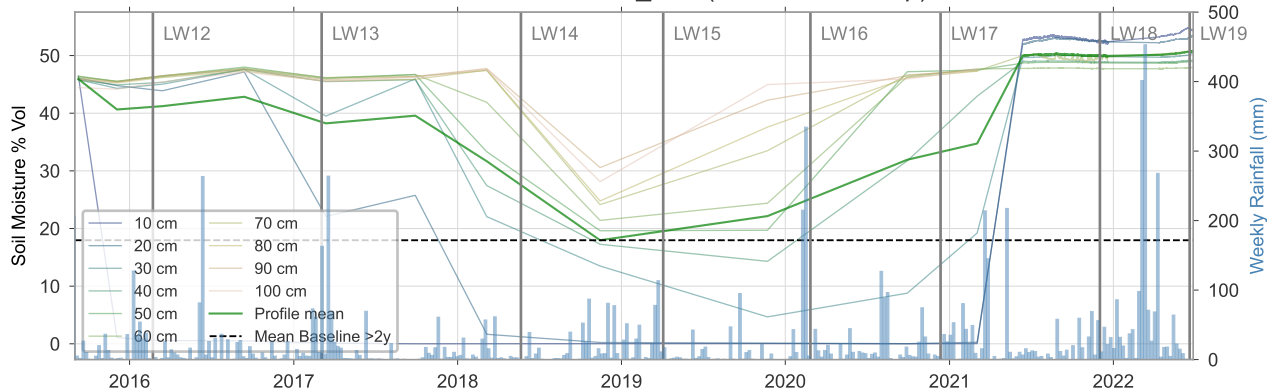
Soil moisture sensor S05_S05 (impact swamp)



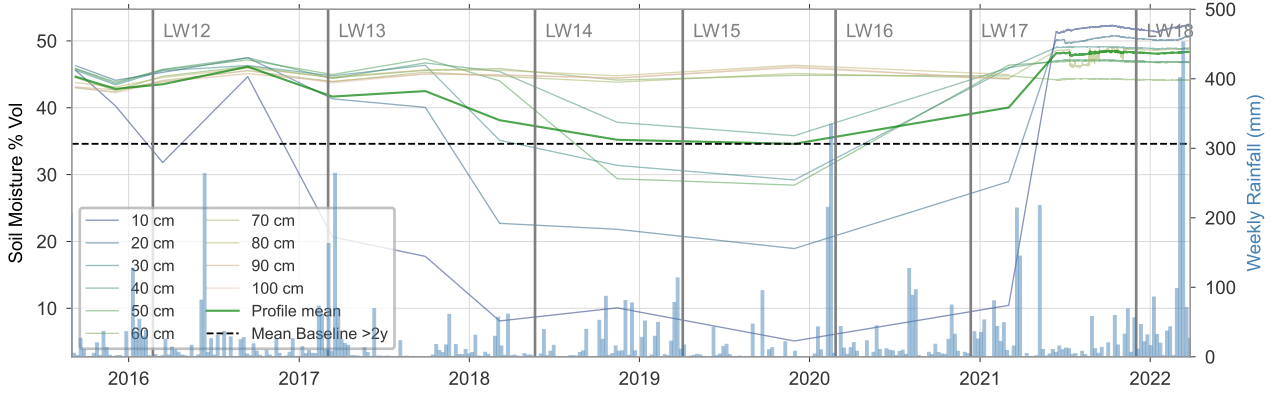
Soil moisture sensor S05_S08 (impact swamp)



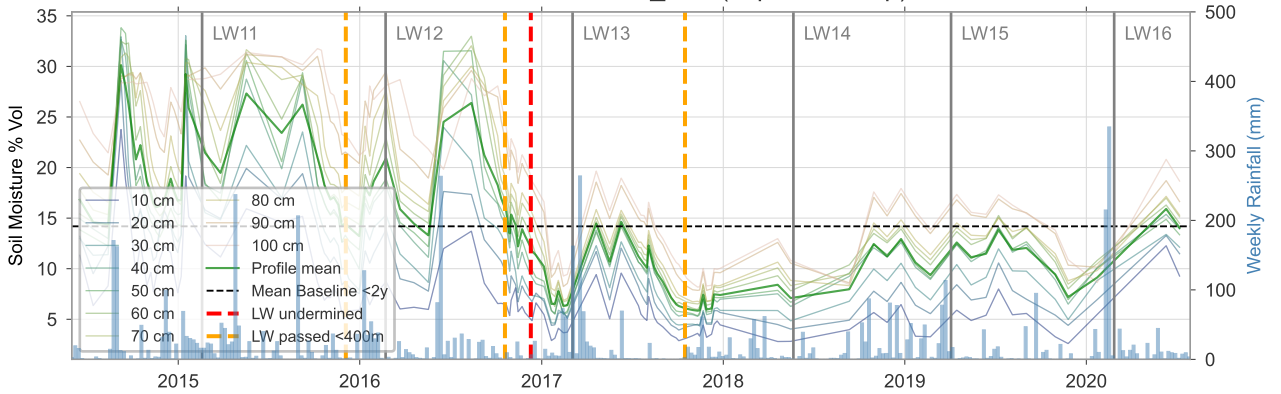
Soil moisture sensor S07_S05 (reference swamp)



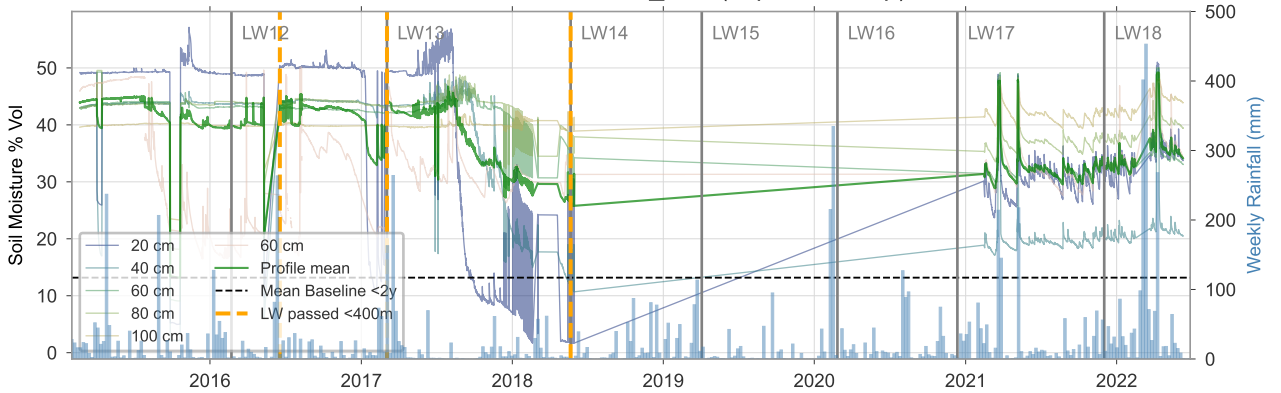
Soil moisture sensor S07_S06 (reference swamp)



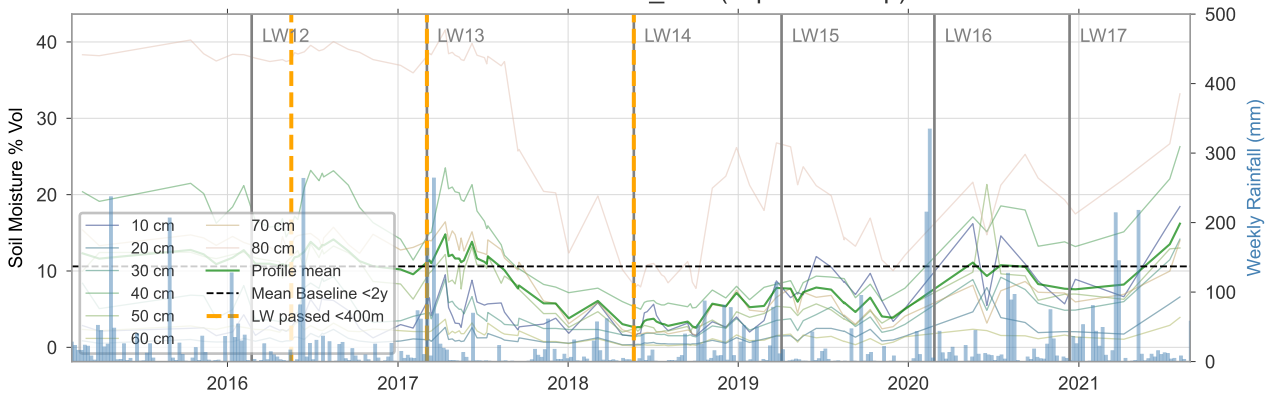
Soil moisture sensor S08_S05 (impact swamp)



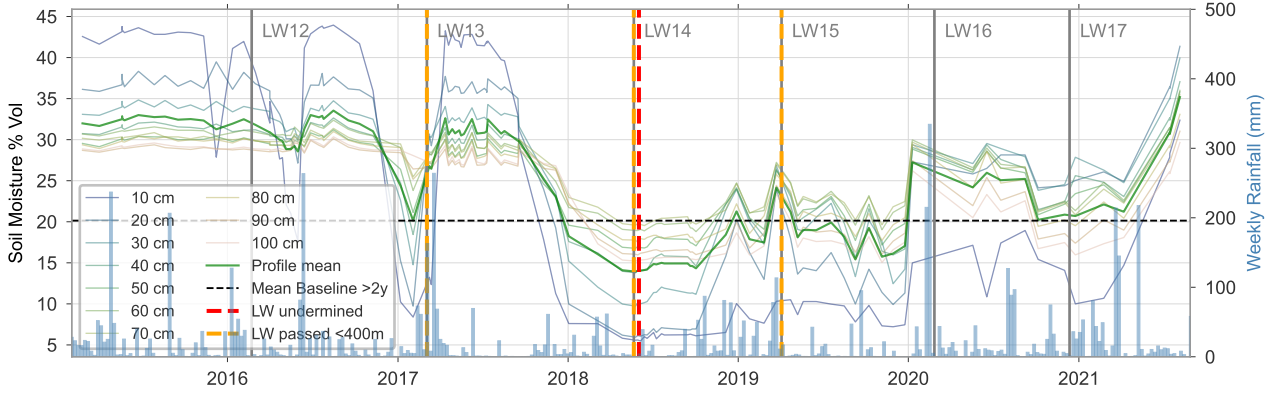
Soil moisture sensor S11_S01 (impact swamp)



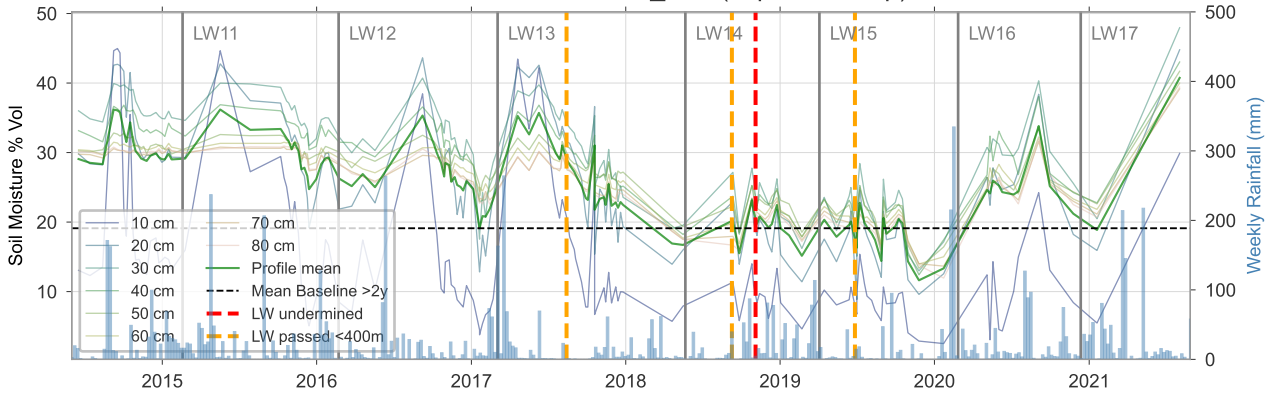
Soil moisture sensor S11_S02 (impact swamp)



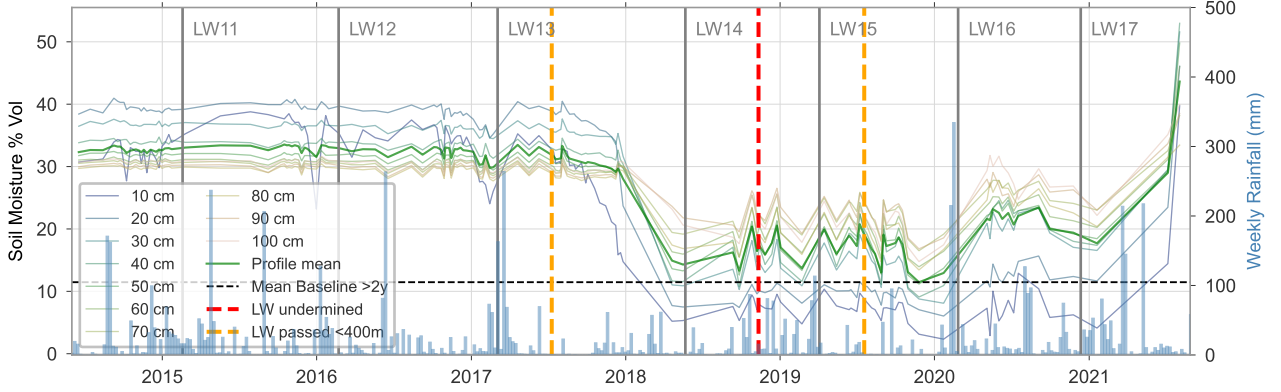
Soil moisture sensor S11_S05 (impact swamp)



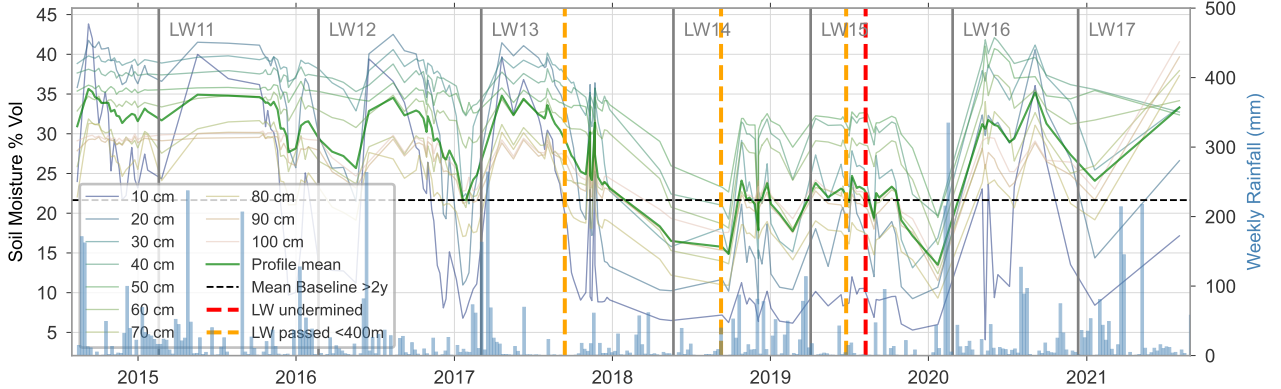
Soil moisture sensor S13_S01 (impact swamp)



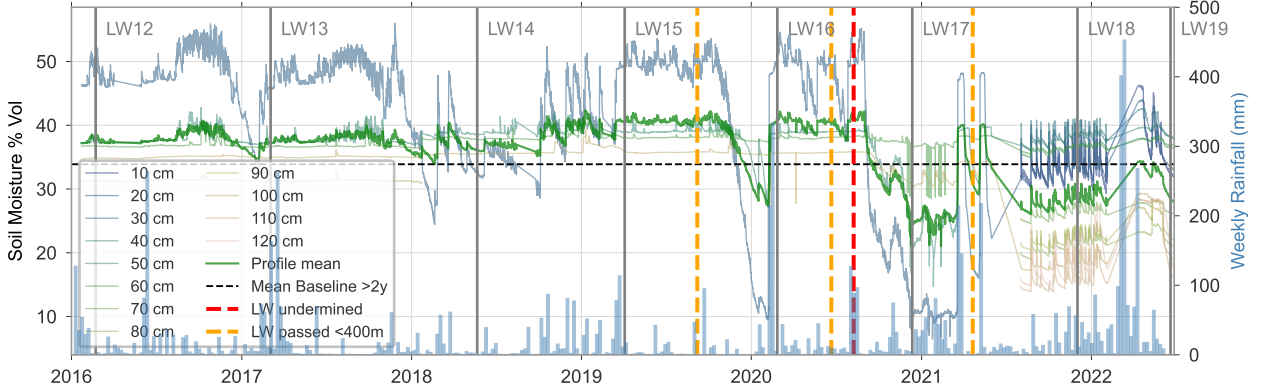
Soil moisture sensor S13_S02 (impact swamp)



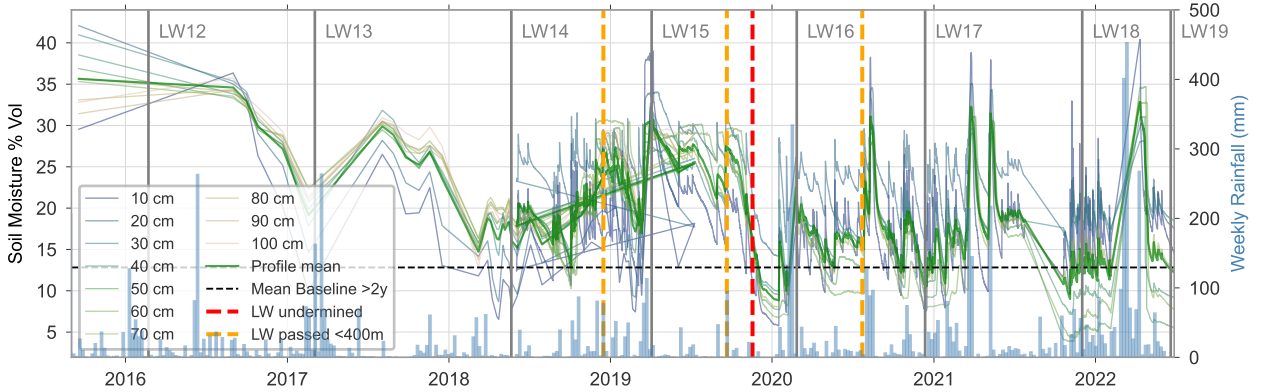
Soil moisture sensor S13_S03 (impact swamp)



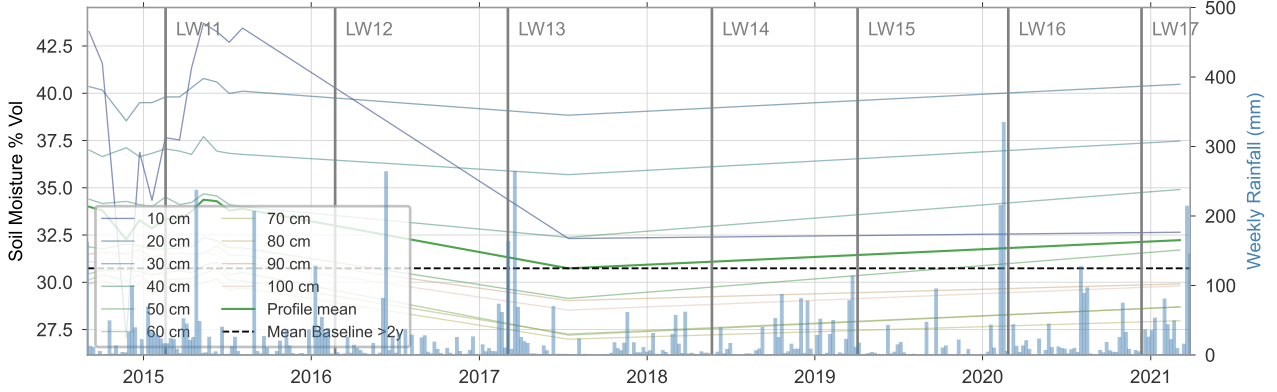
Soil moisture sensor S14_S01 (impact swamp)



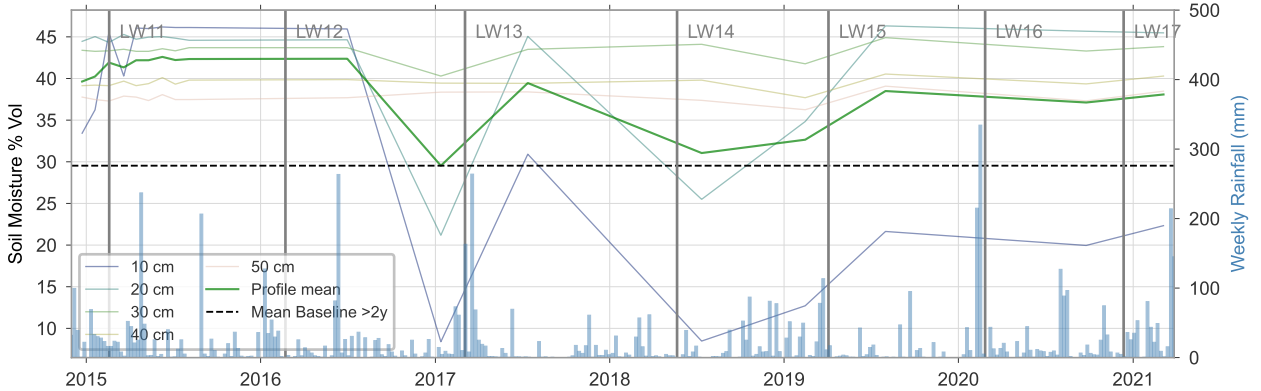
Soil moisture sensor S14_S02 (impact swamp)



Soil moisture sensor S15a_Piezo (impact swamp)



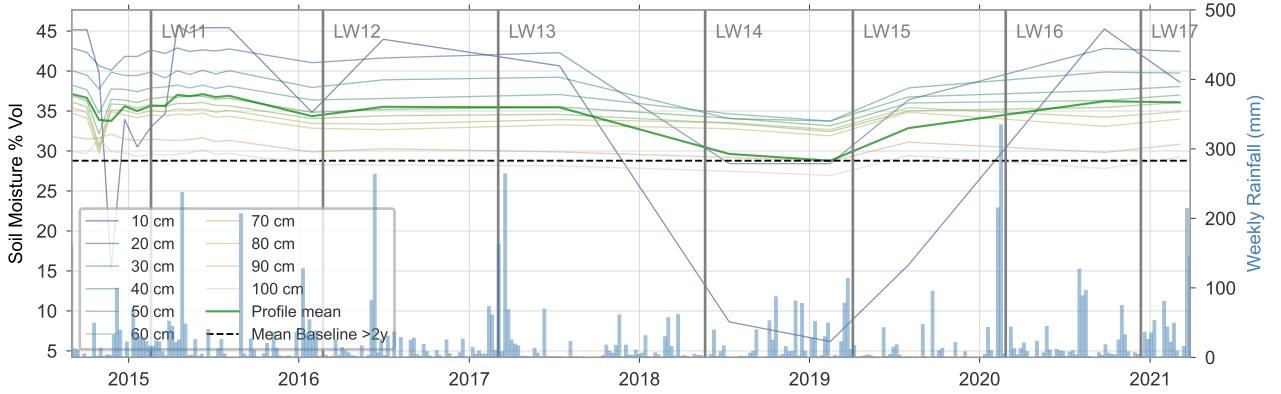
Soil moisture sensor S15a_S01 (impact swamp)



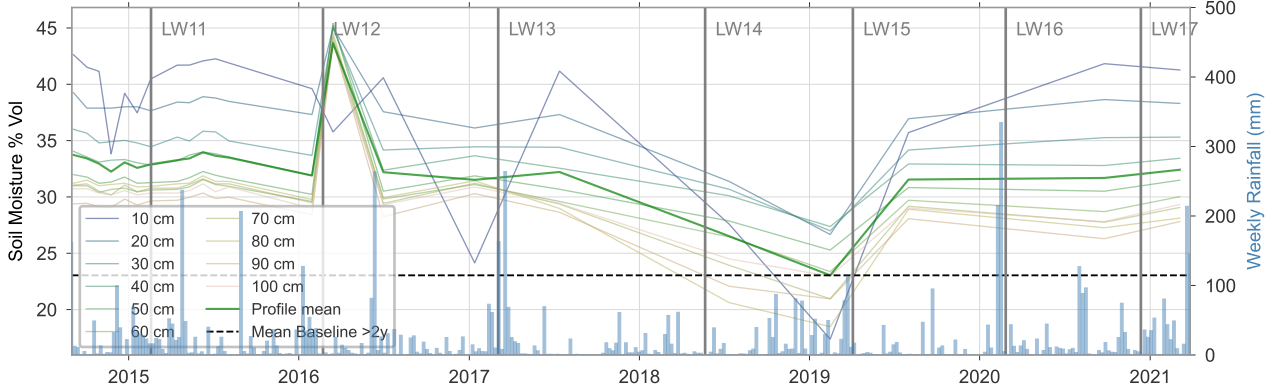
Soil moisture sensor S15a_S03 (impact swamp)



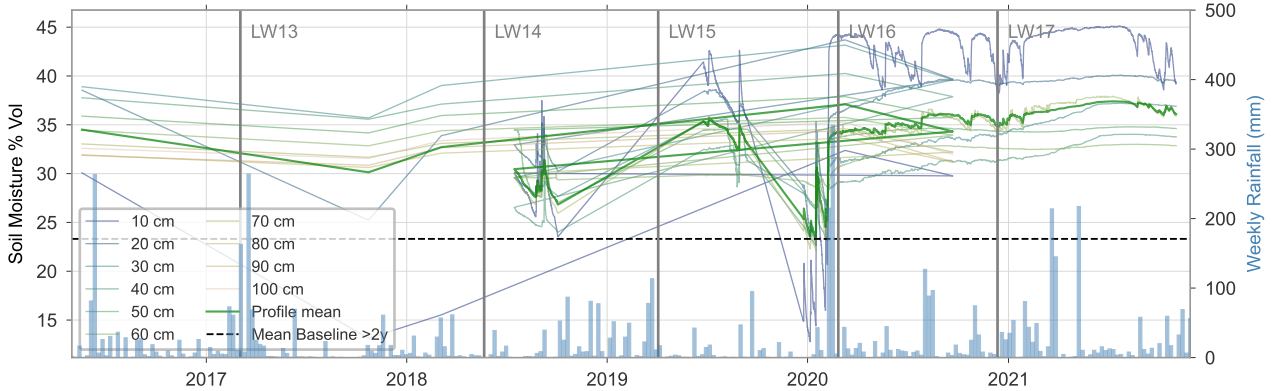
Soil moisture sensor S15A_S04_SEDGEINFLOW (impact swamp)



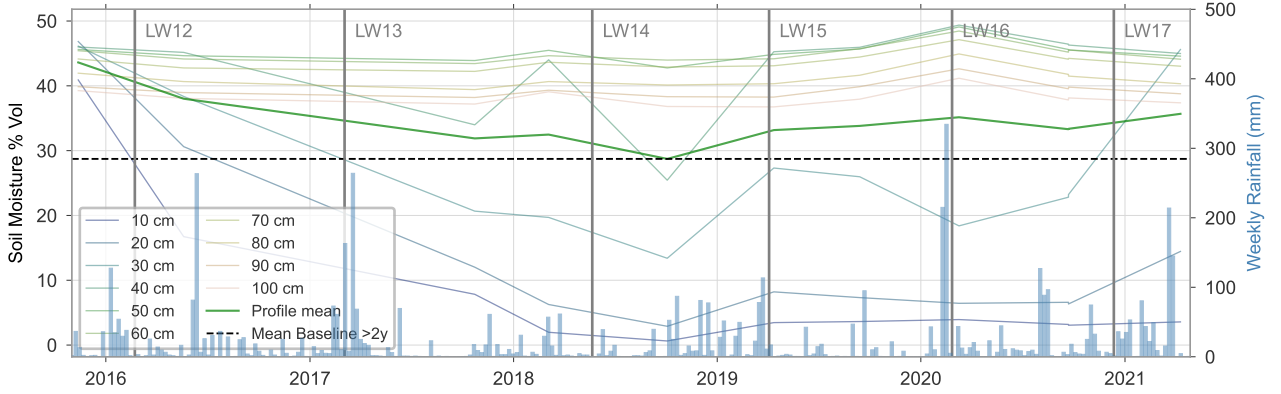
Soil moisture sensor S15a_S06 (impact swamp)



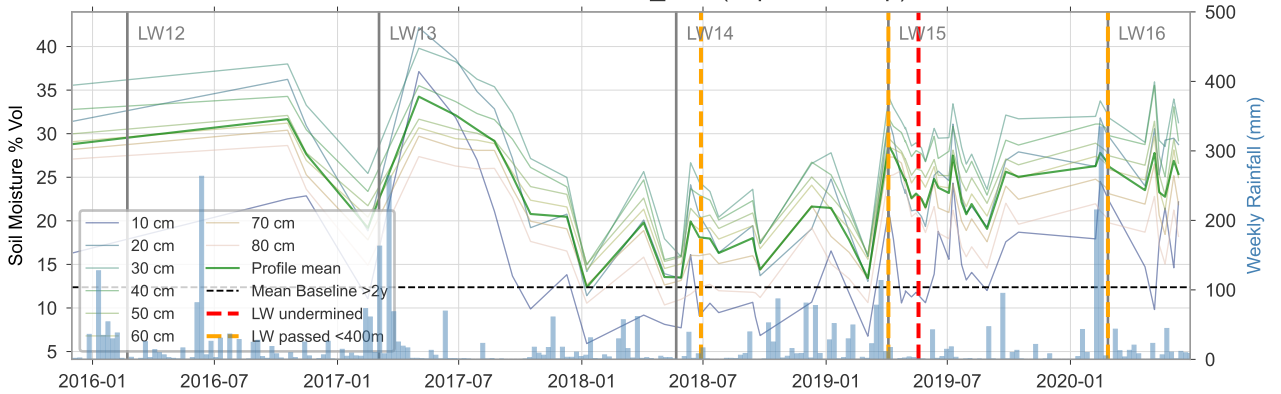
Soil moisture sensor S22_S01 (reference swamp)



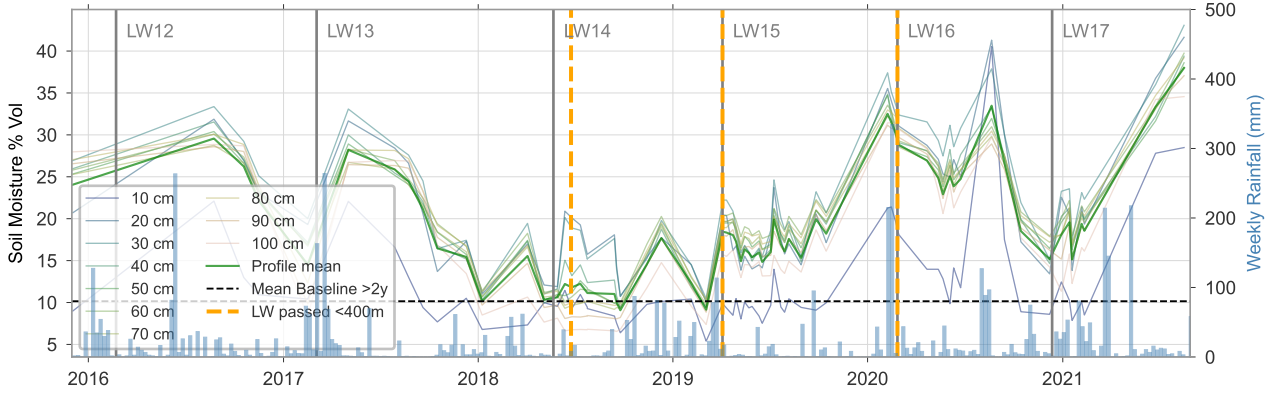
Soil moisture sensor S22_S02 (reference swamp)



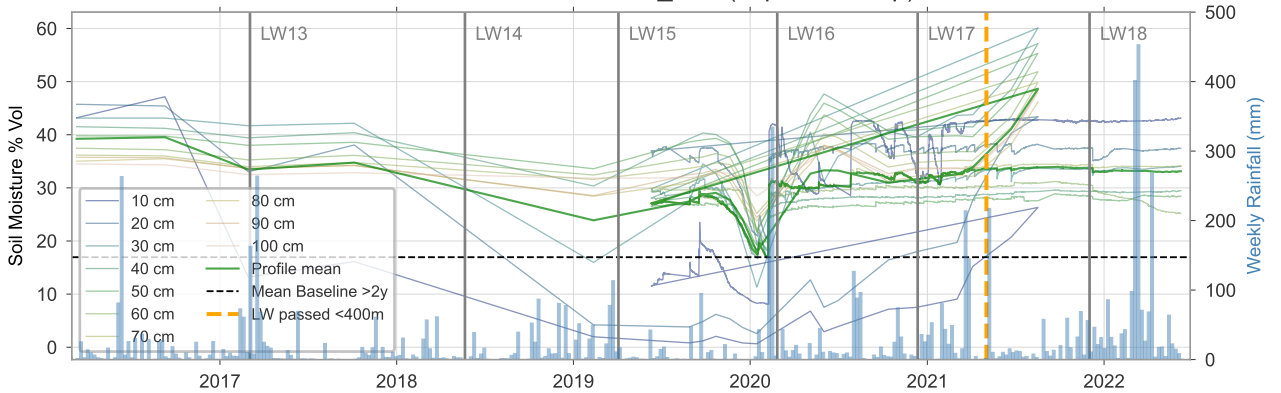
Soil moisture sensor S23_S01 (impact swamp)



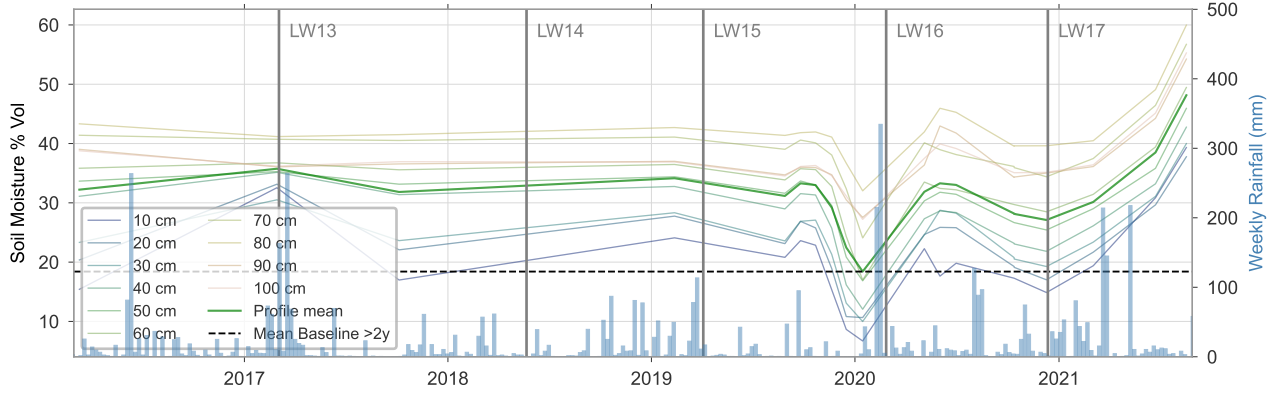
Soil moisture sensor S23_S02 (impact swamp)



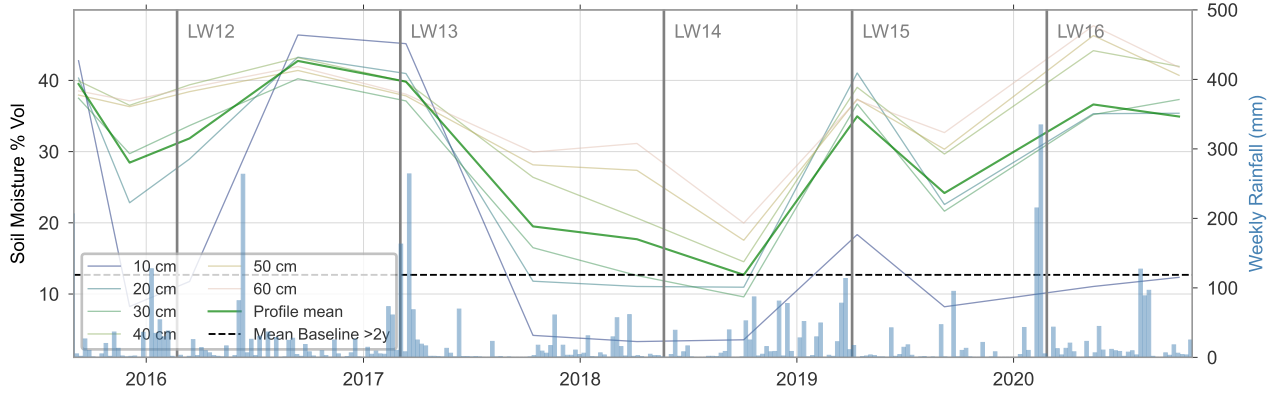
Soil moisture sensor S35a_S01 (impact swamp)



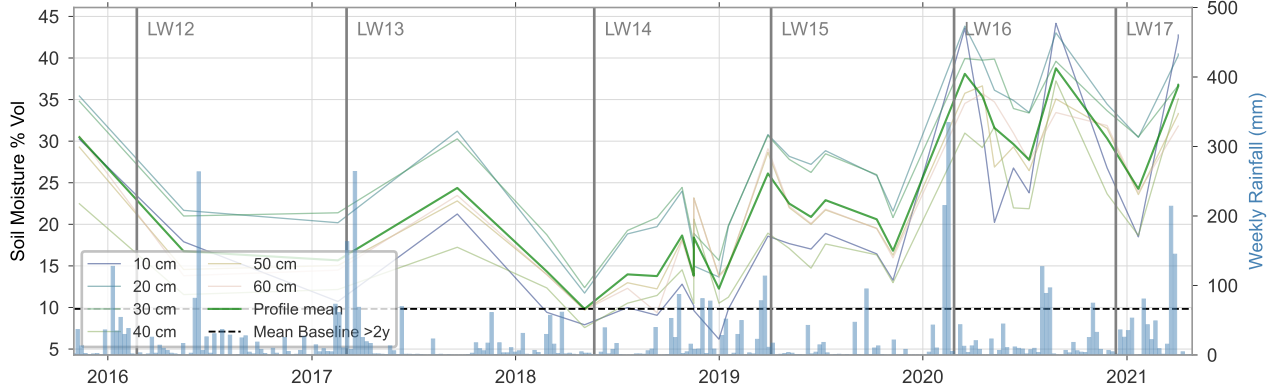
Soil moisture sensor S35b_S01 (impact swamp)



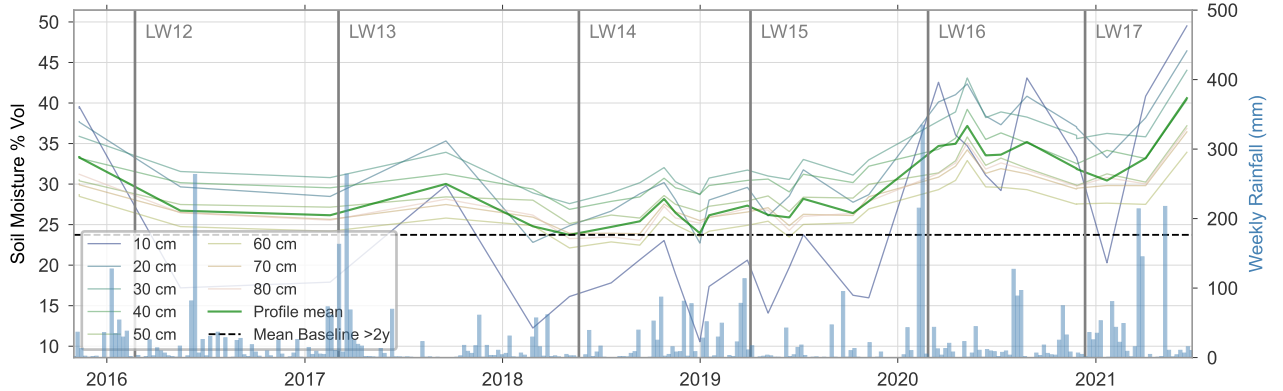
Soil moisture sensor S84_S01 (reference swamp)



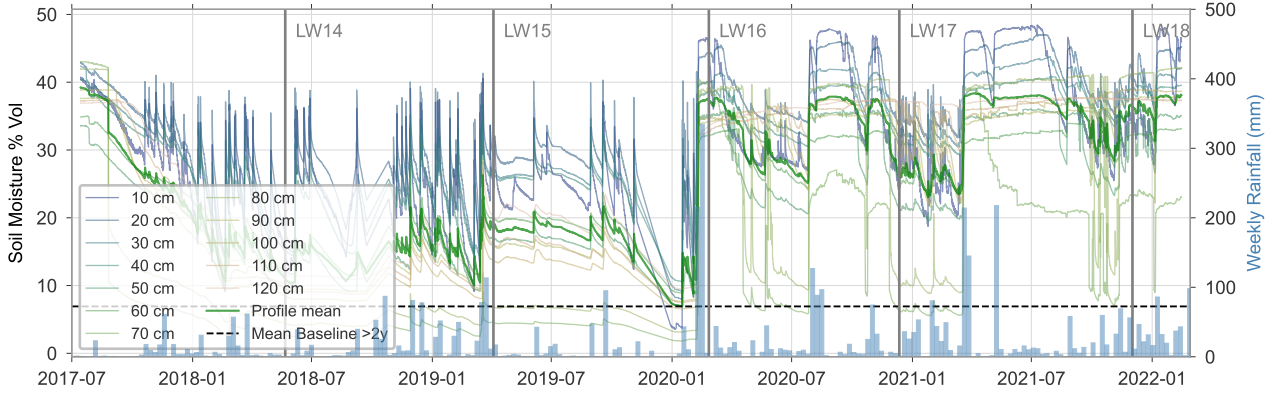
Soil moisture sensor S85_S01 (reference swamp)



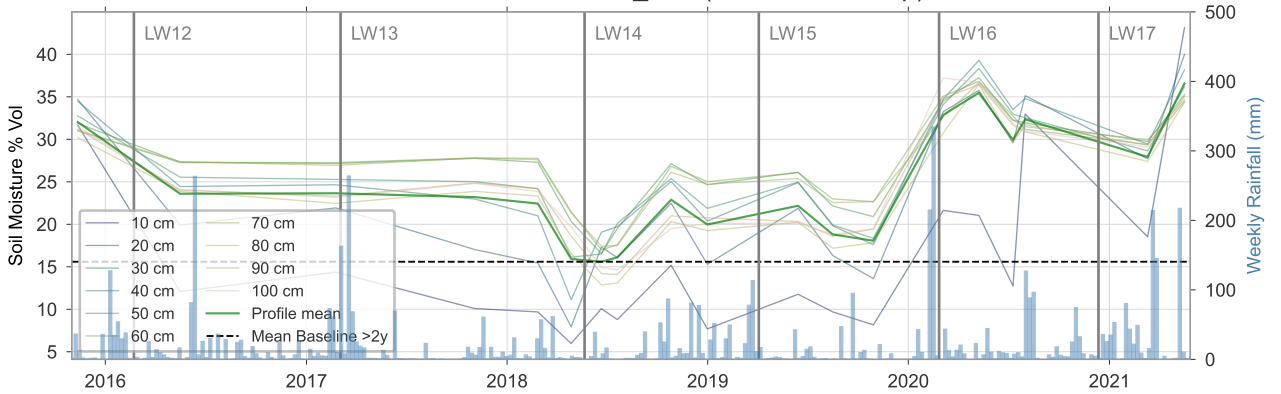
Soil moisture sensor S85_S02 (reference swamp)



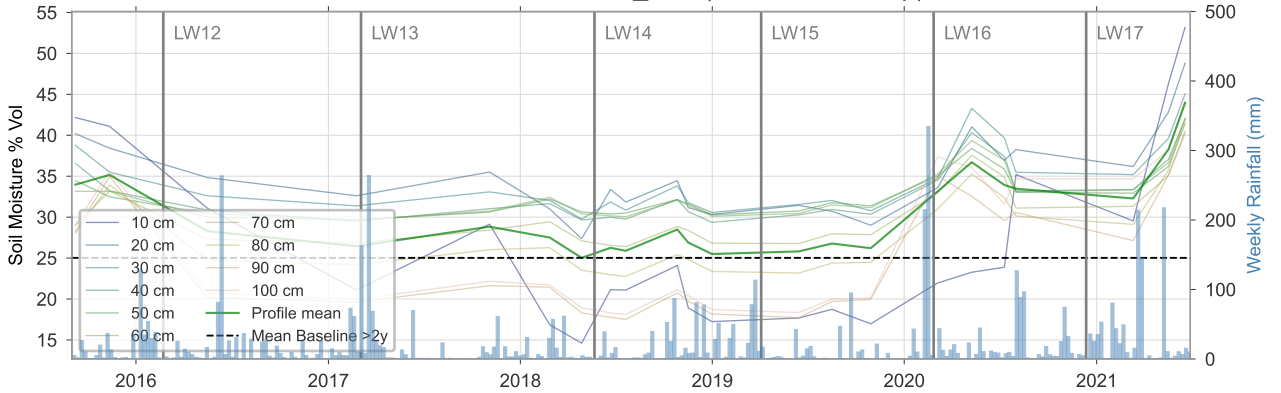
Soil moisture sensor S85_S03 (reference swamp)



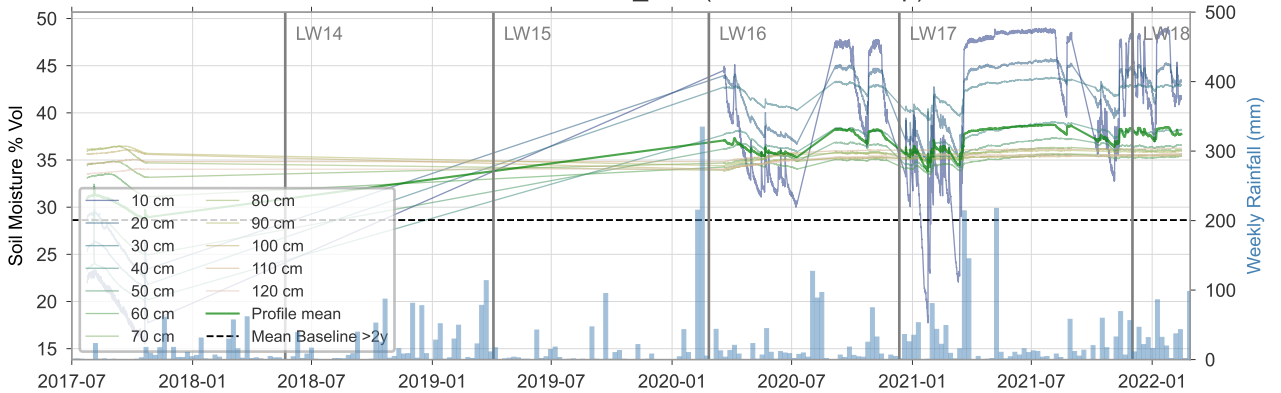
Soil moisture sensor S86_S01 (reference swamp)



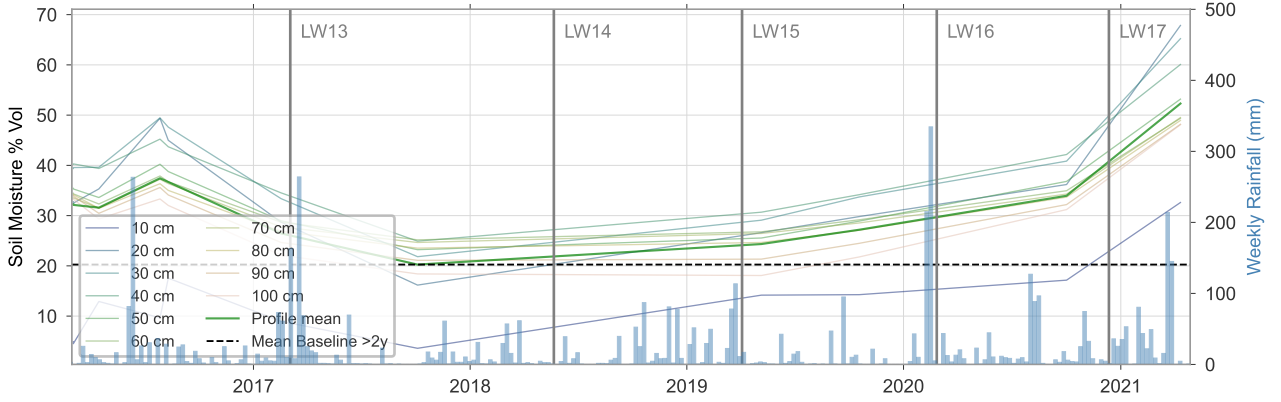
Soil moisture sensor S86_S02 (reference swamp)



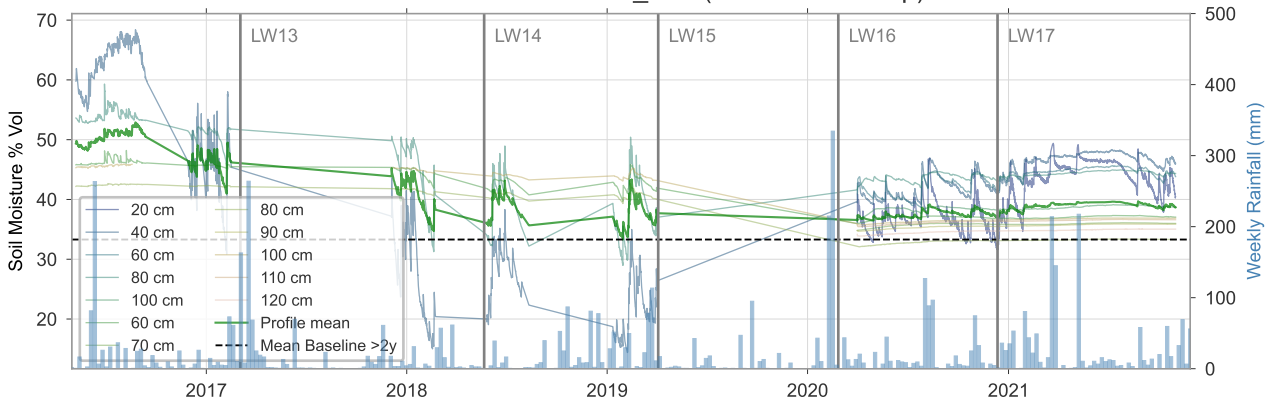
Soil moisture sensor S86_S03 (reference swamp)



Soil moisture sensor S87_S01 (reference swamp)



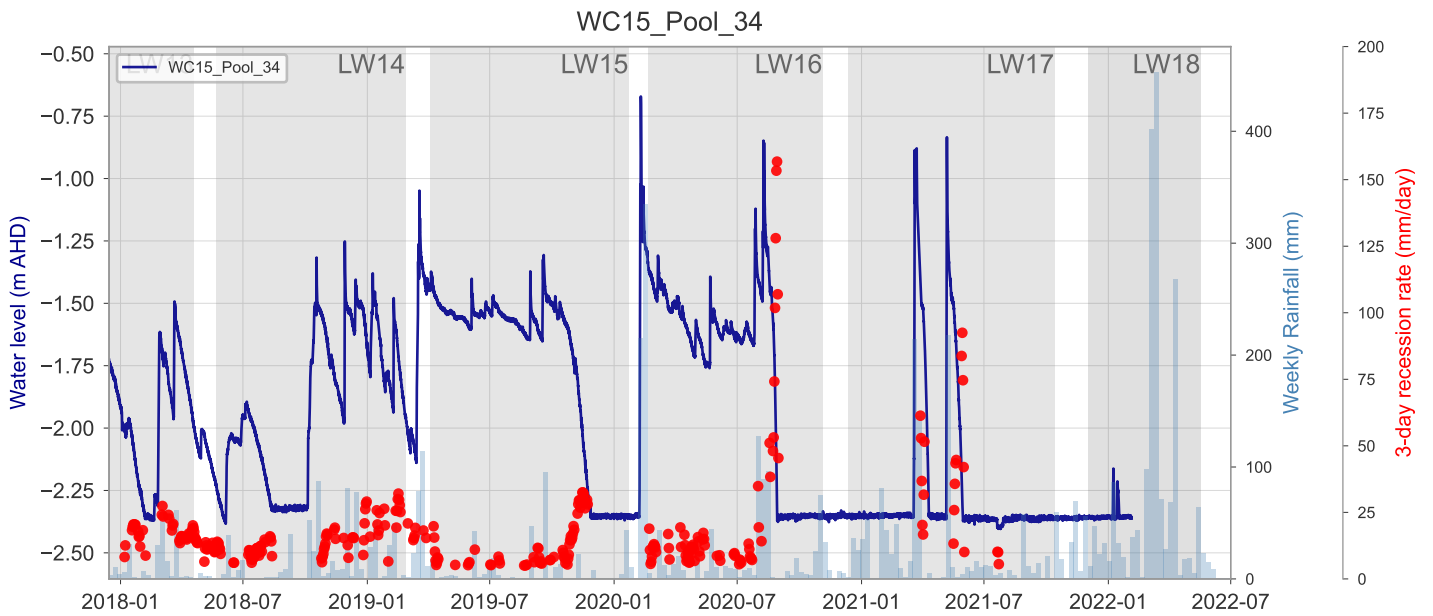
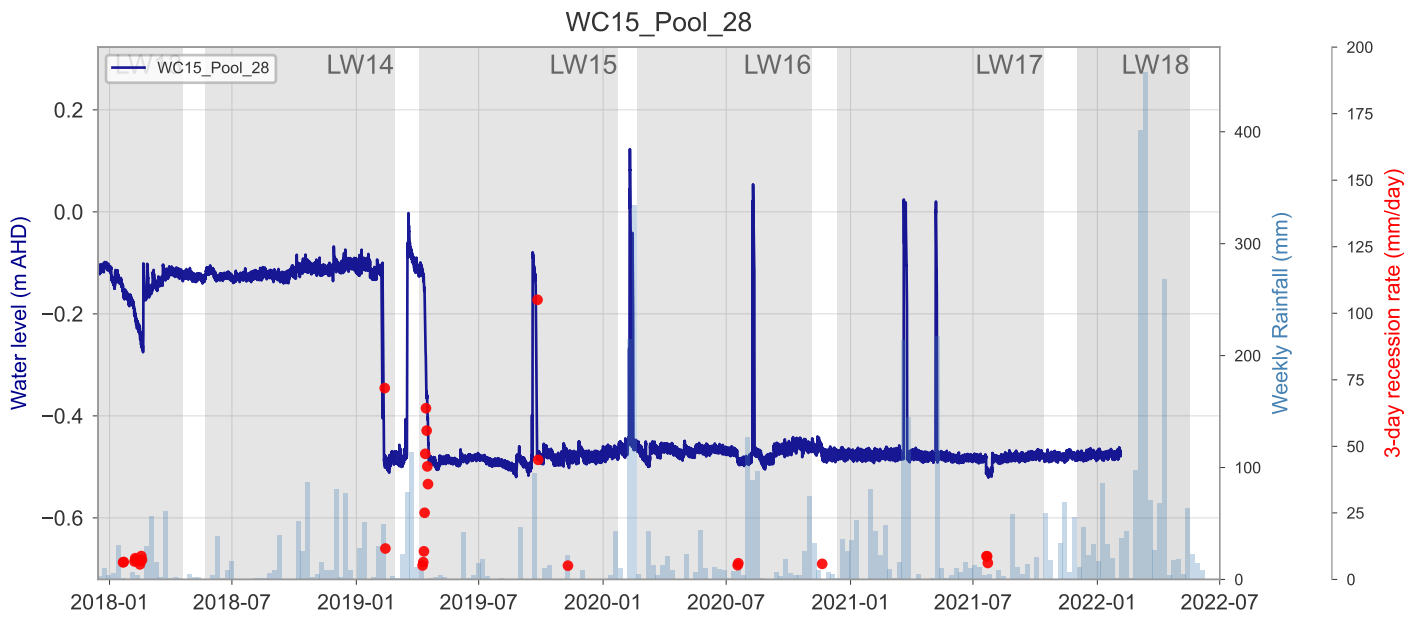
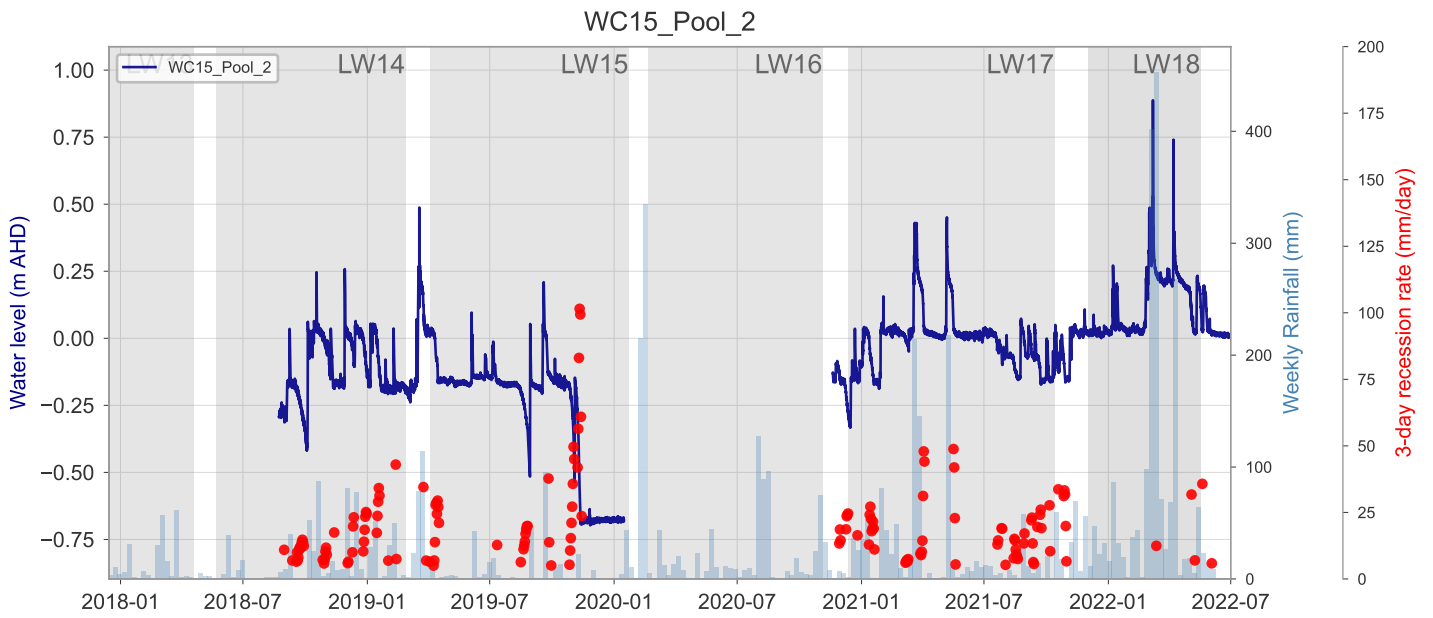
Soil moisture sensor S87_S02 (reference swamp)

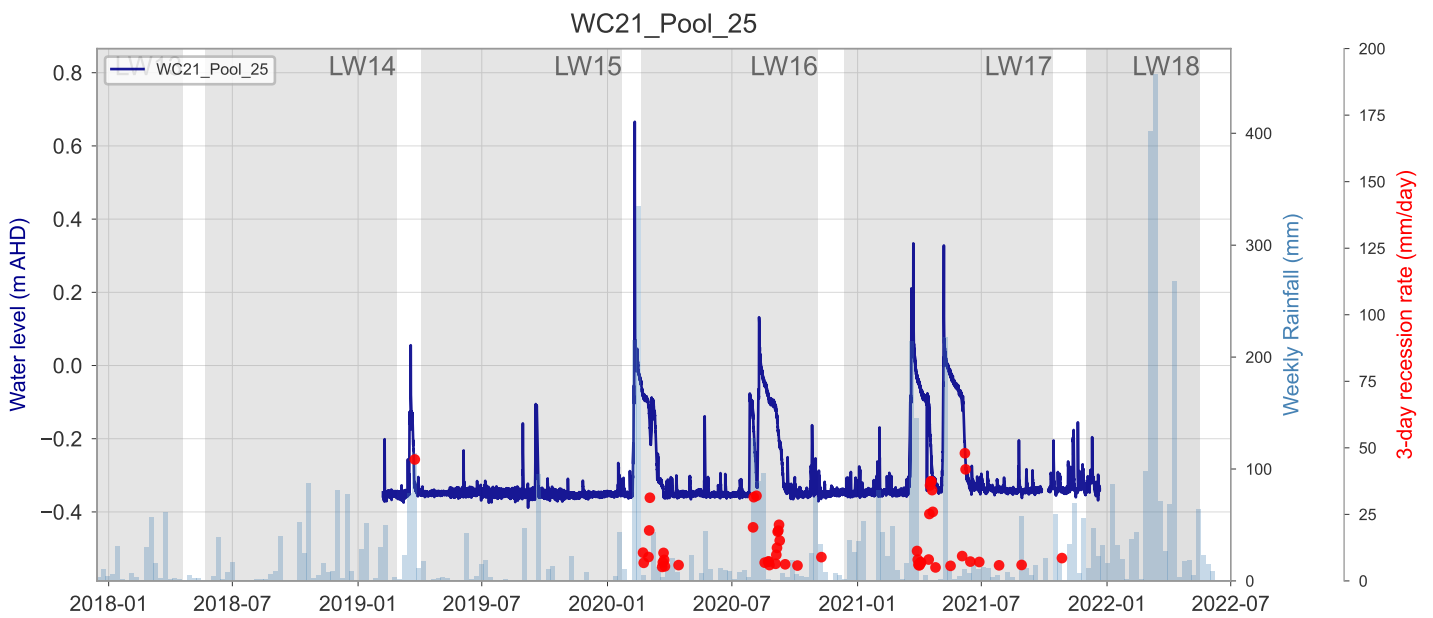
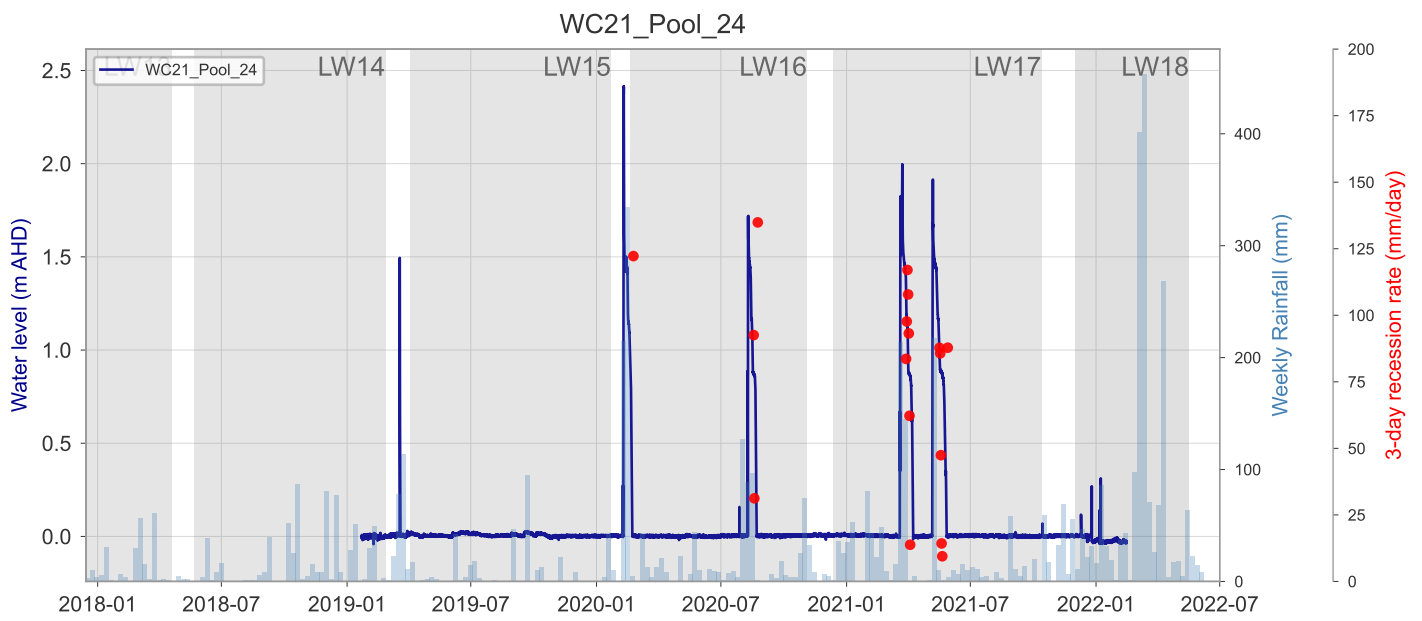
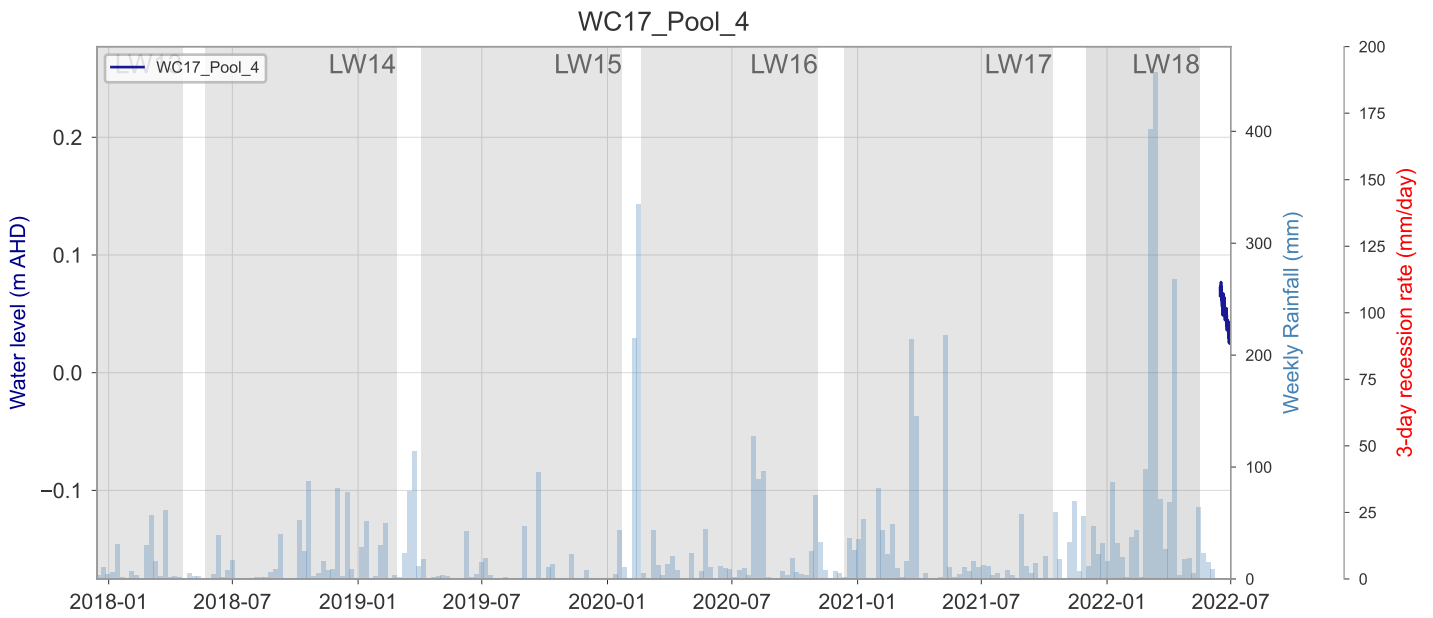


Soil moisture sensor S88_S01 (reference swamp)

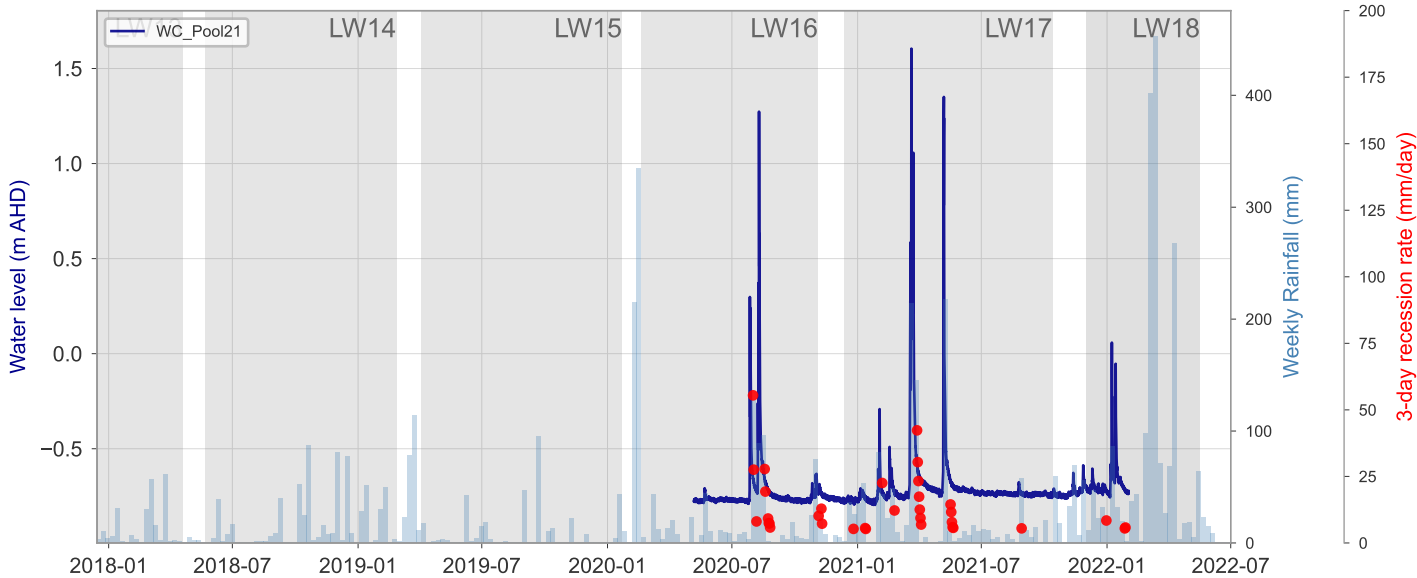


Appendix F: Stream pool level hydrographs

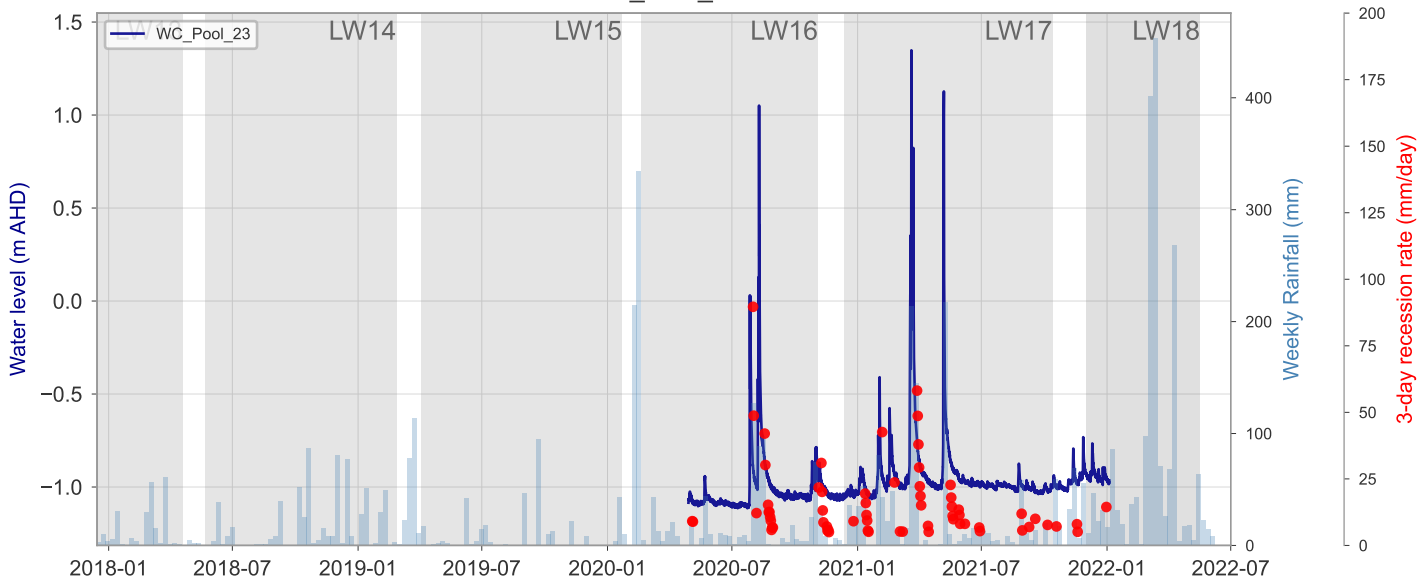




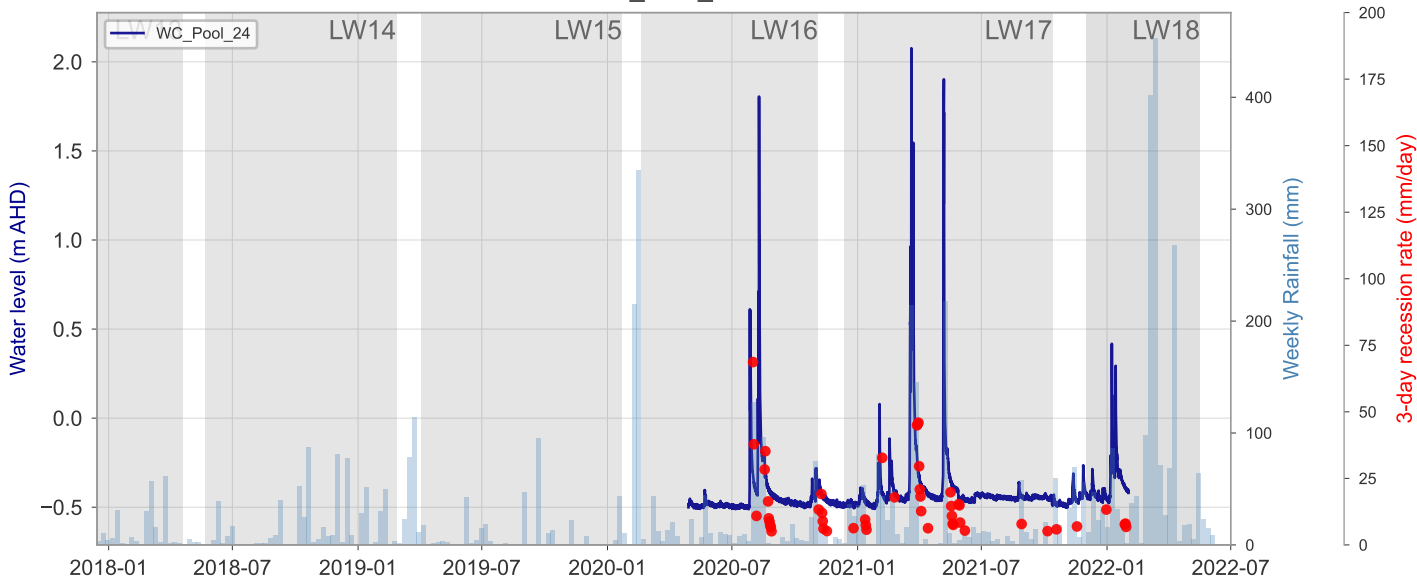
WC_Pool21



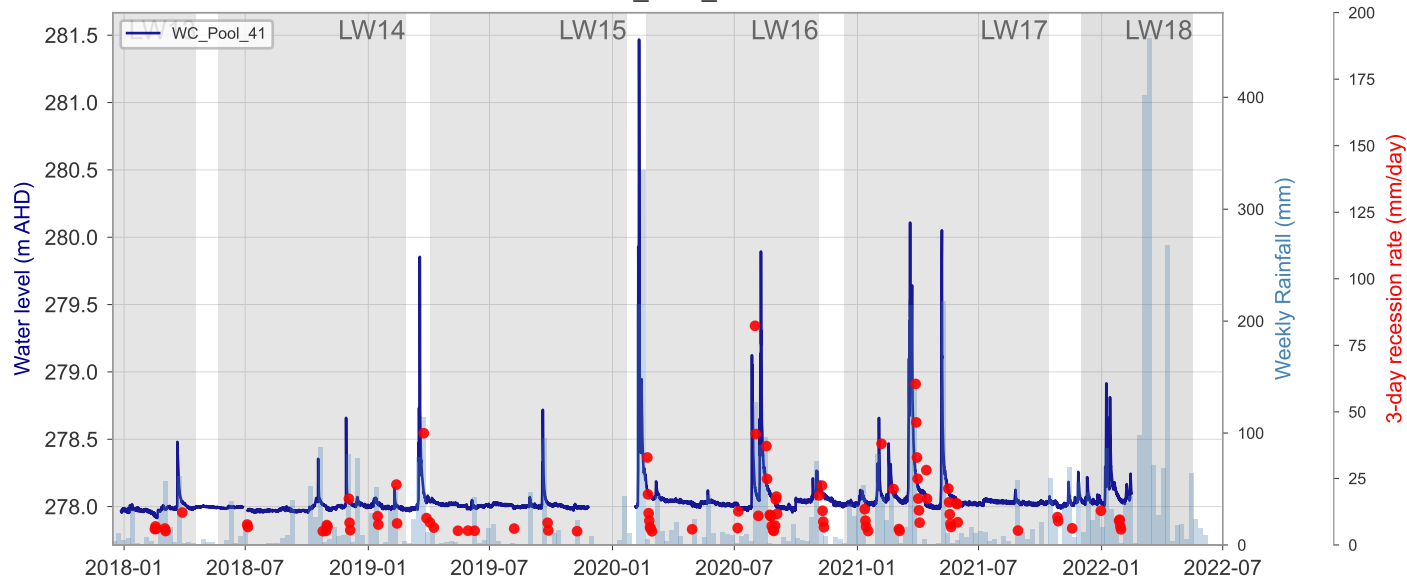
WC_Pool_23



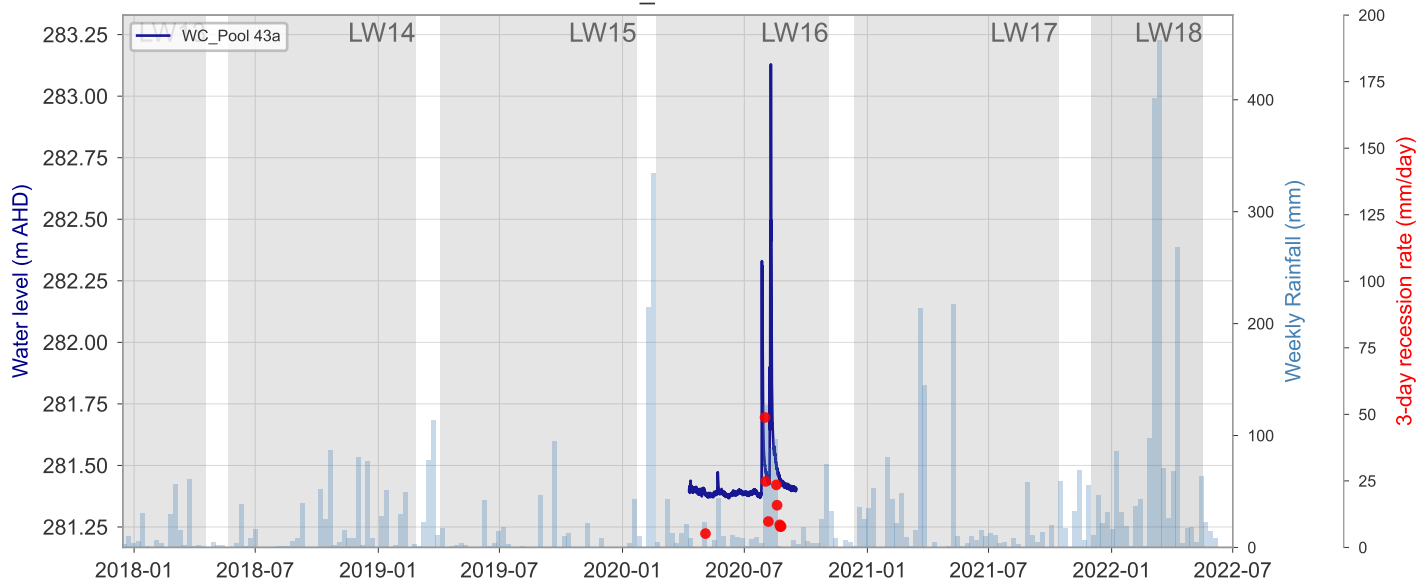
WC_Pool_24



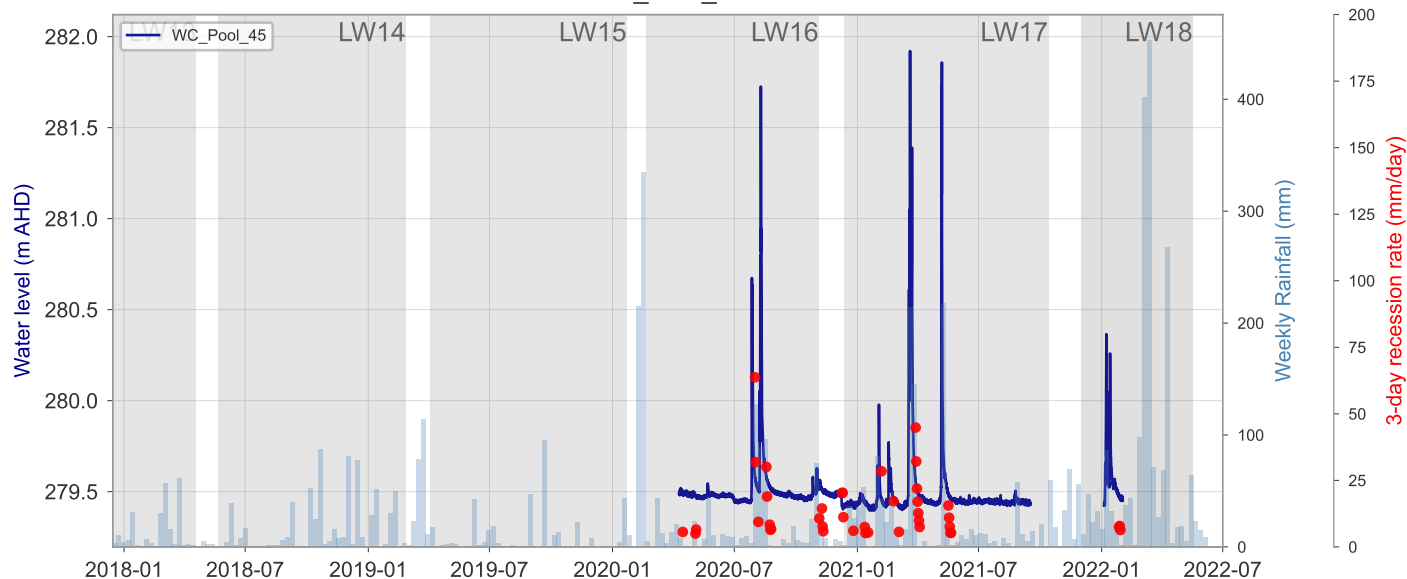
WC_Pool_41



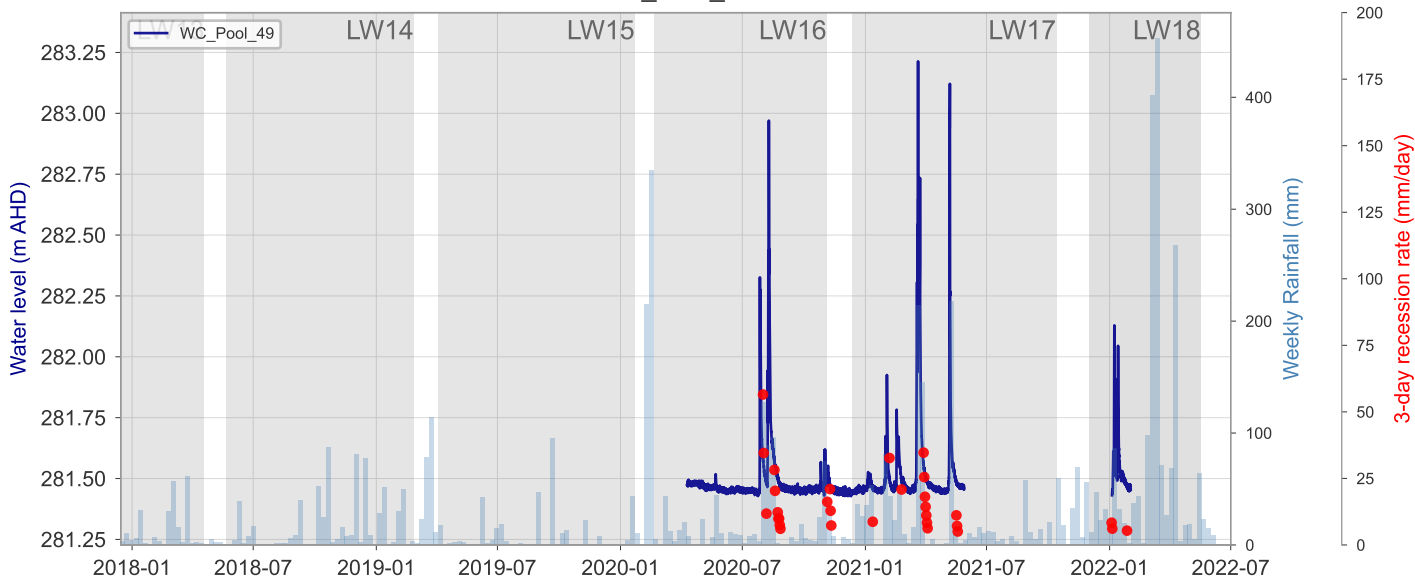
WC_Pool_43a



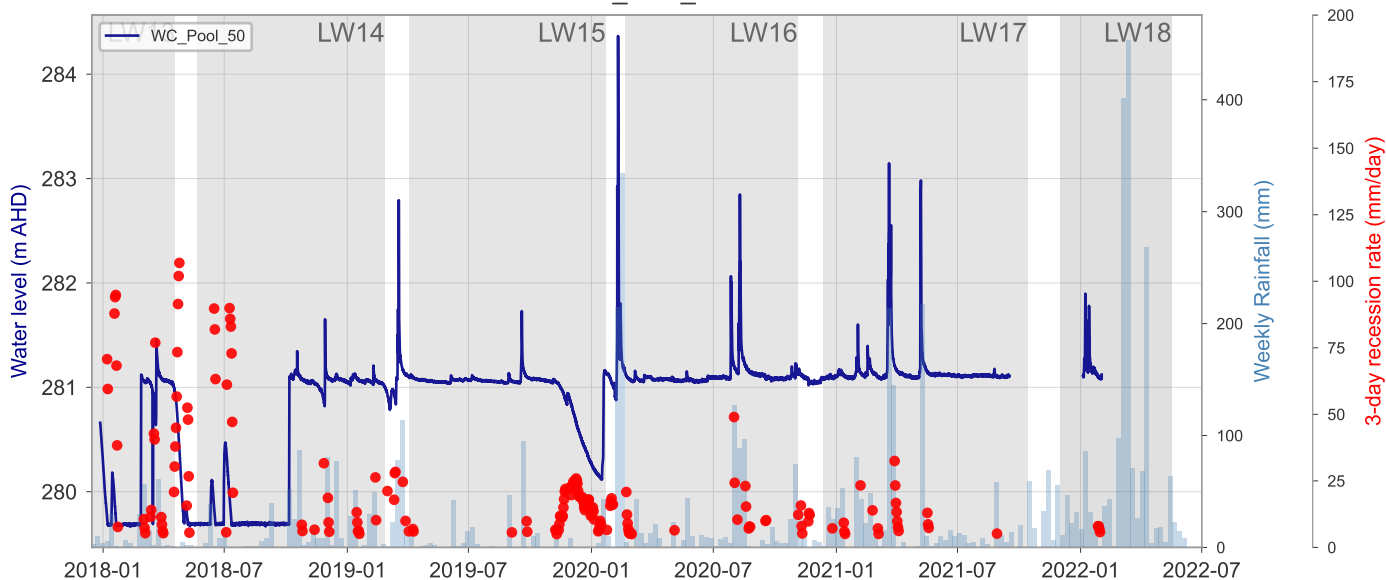
WC_Pool_45



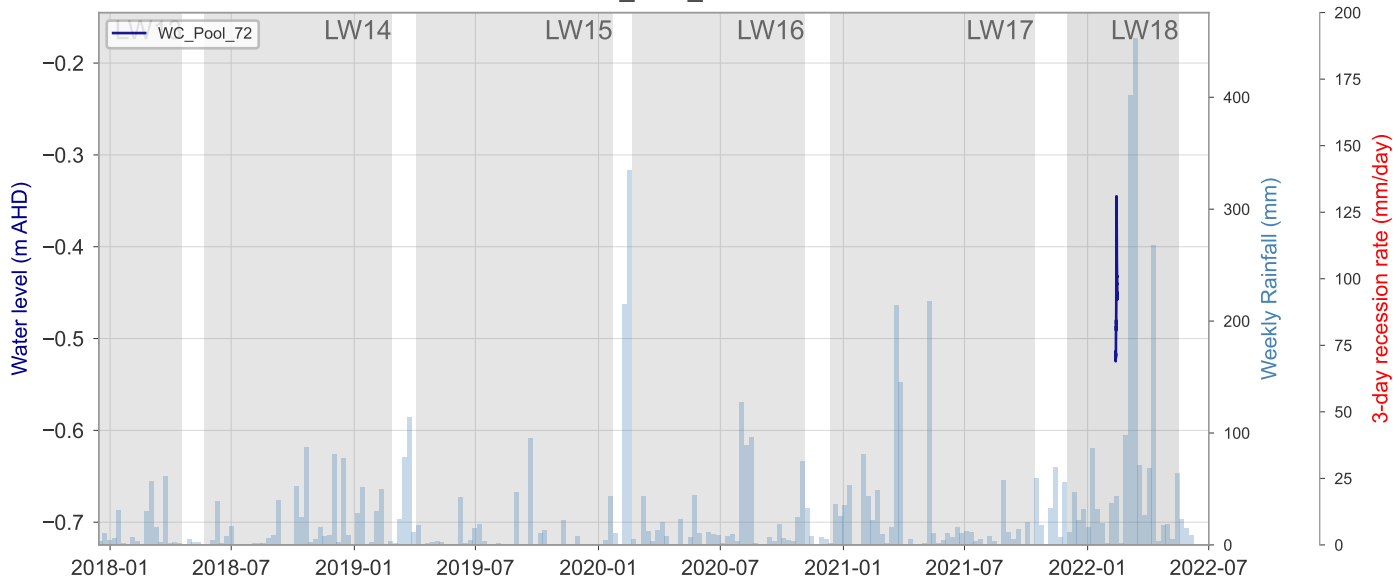
WC_Pool_49



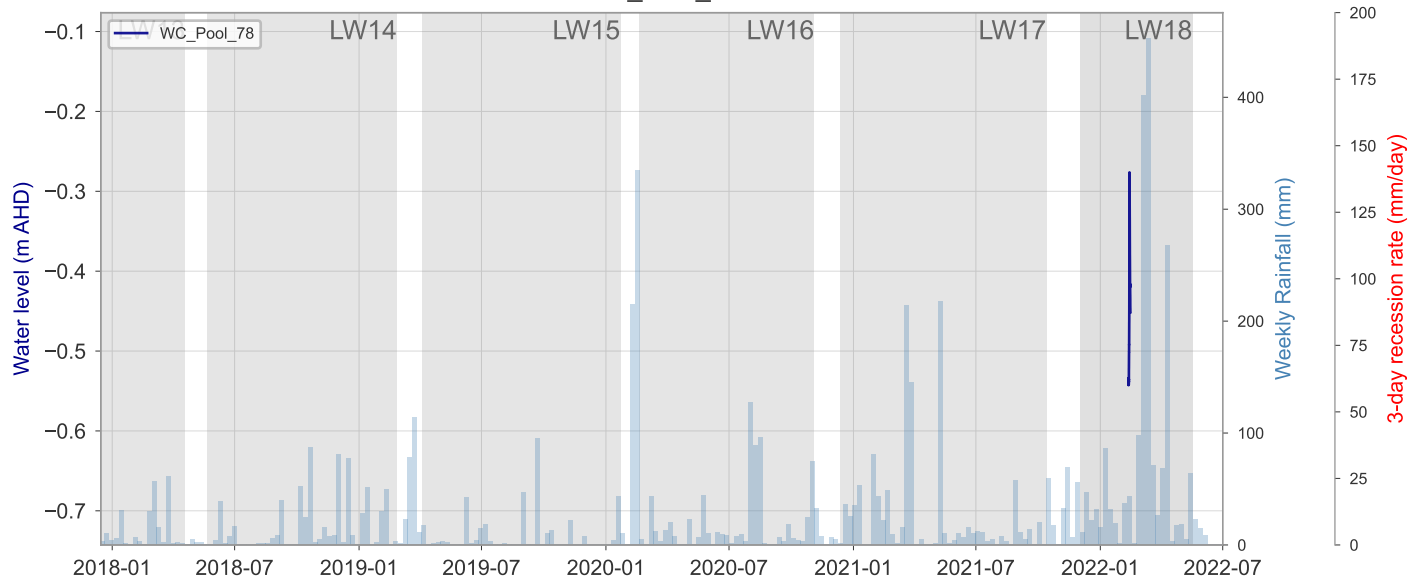
WC_Pool_50



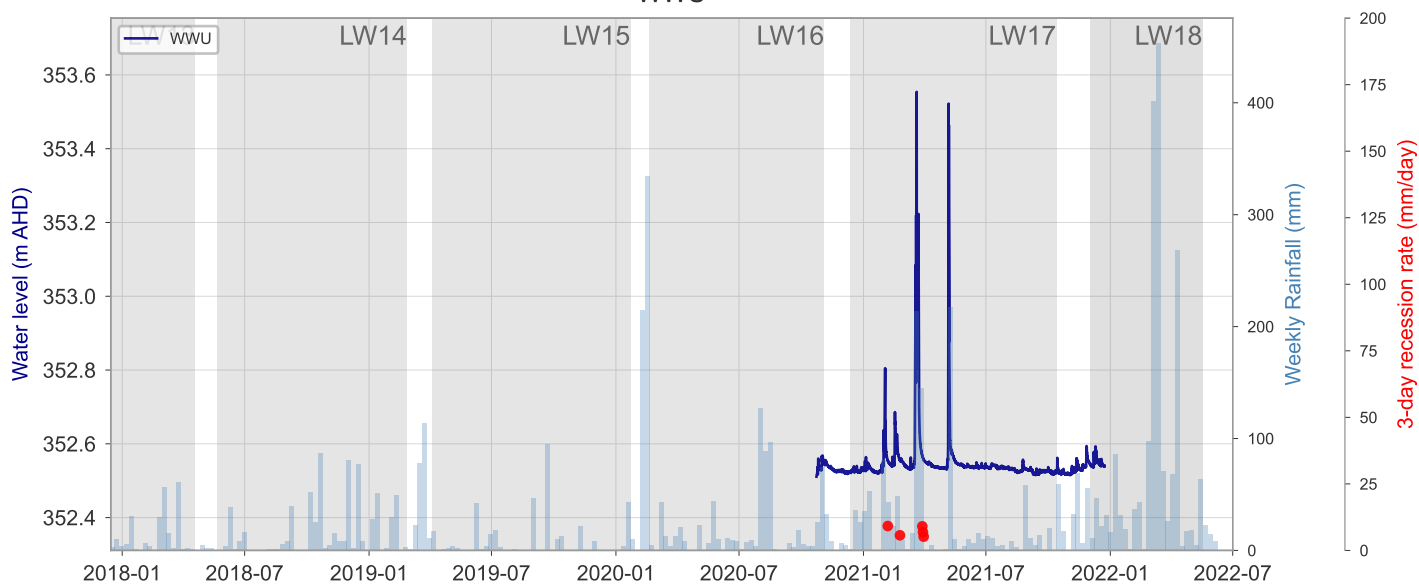
WC_Pool_72



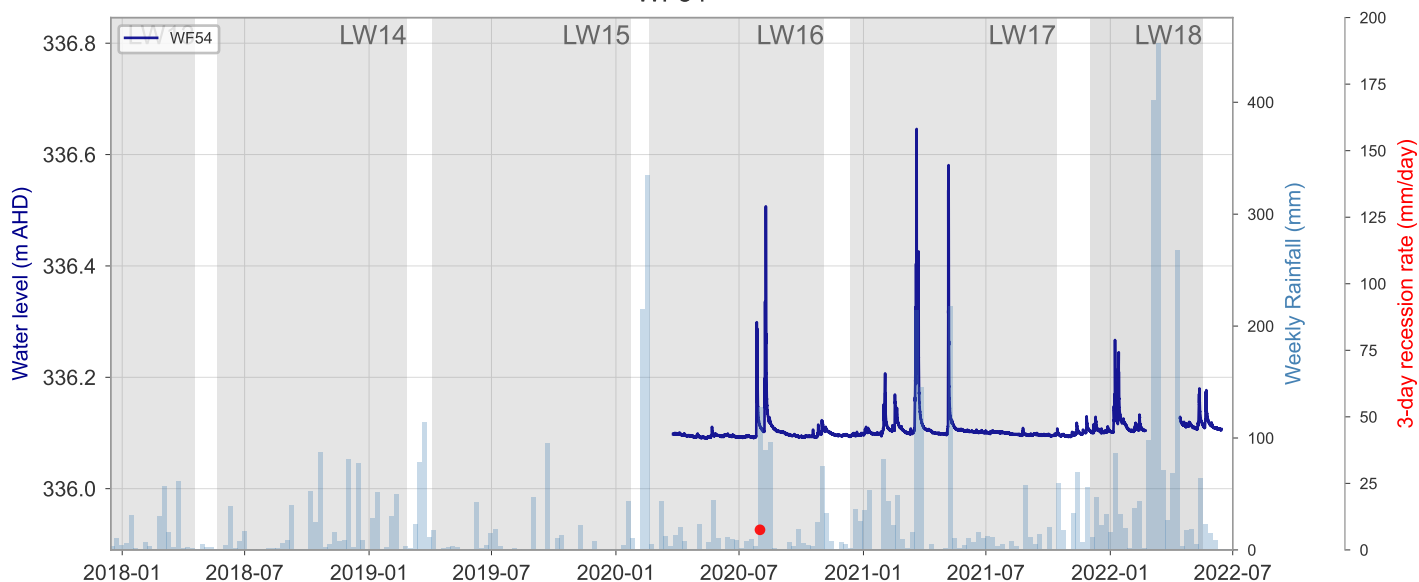
WC_Pool_78



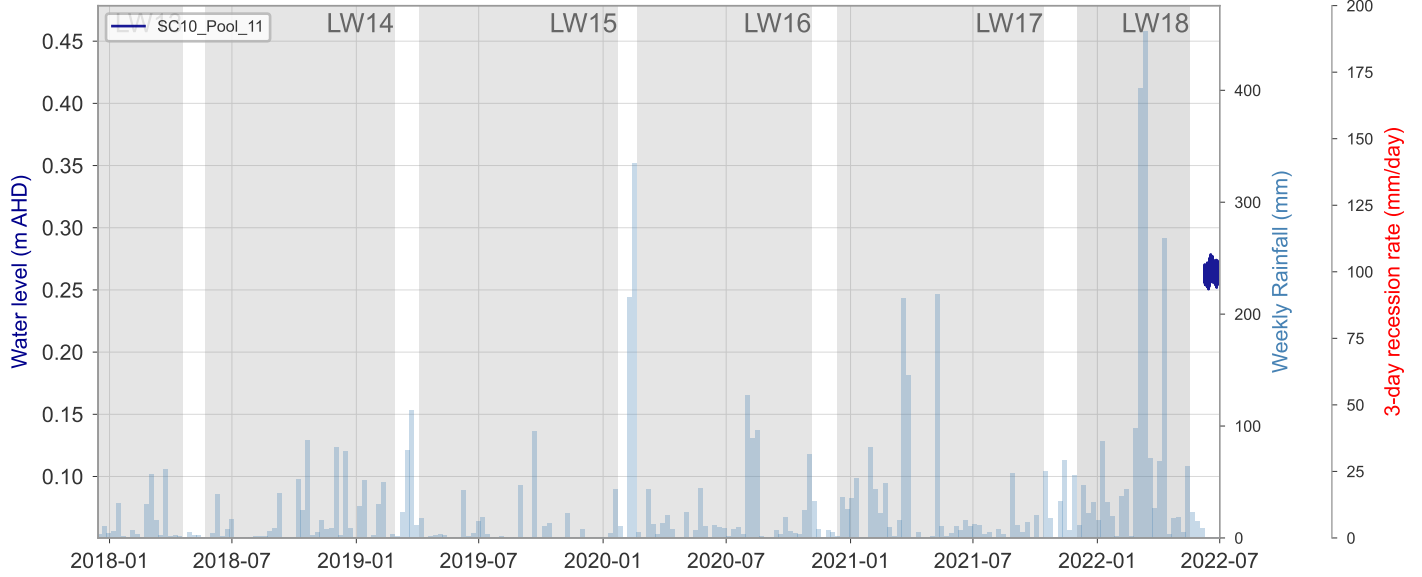
WWU



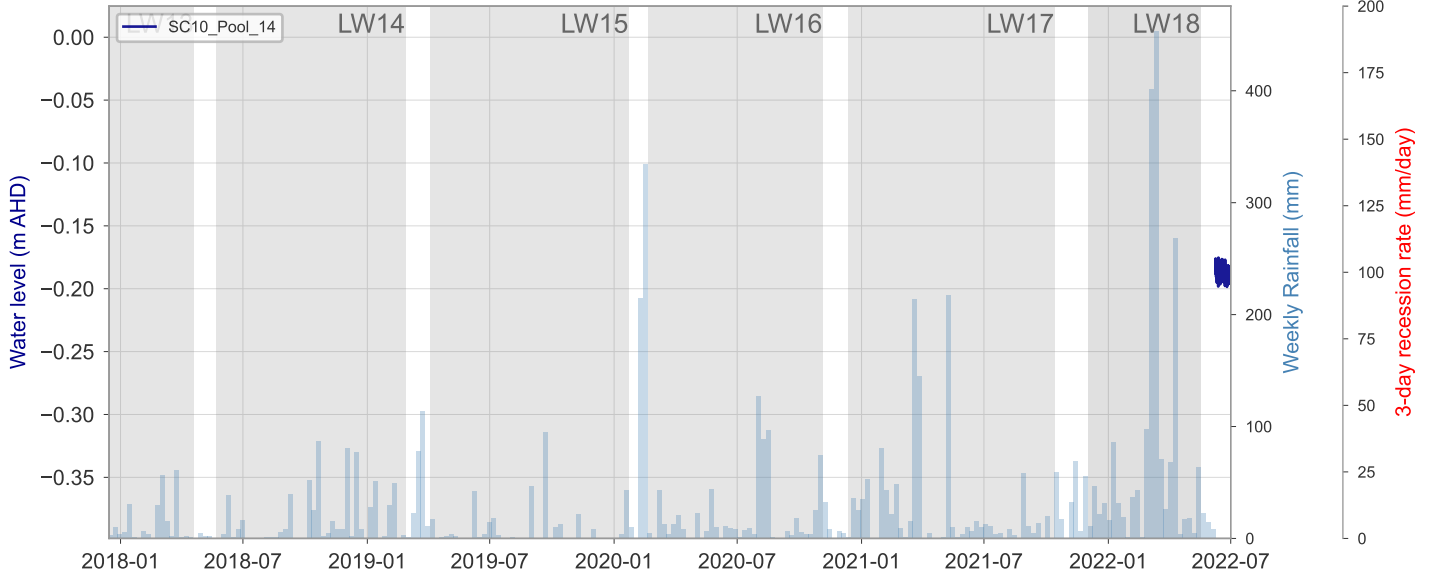
WF54



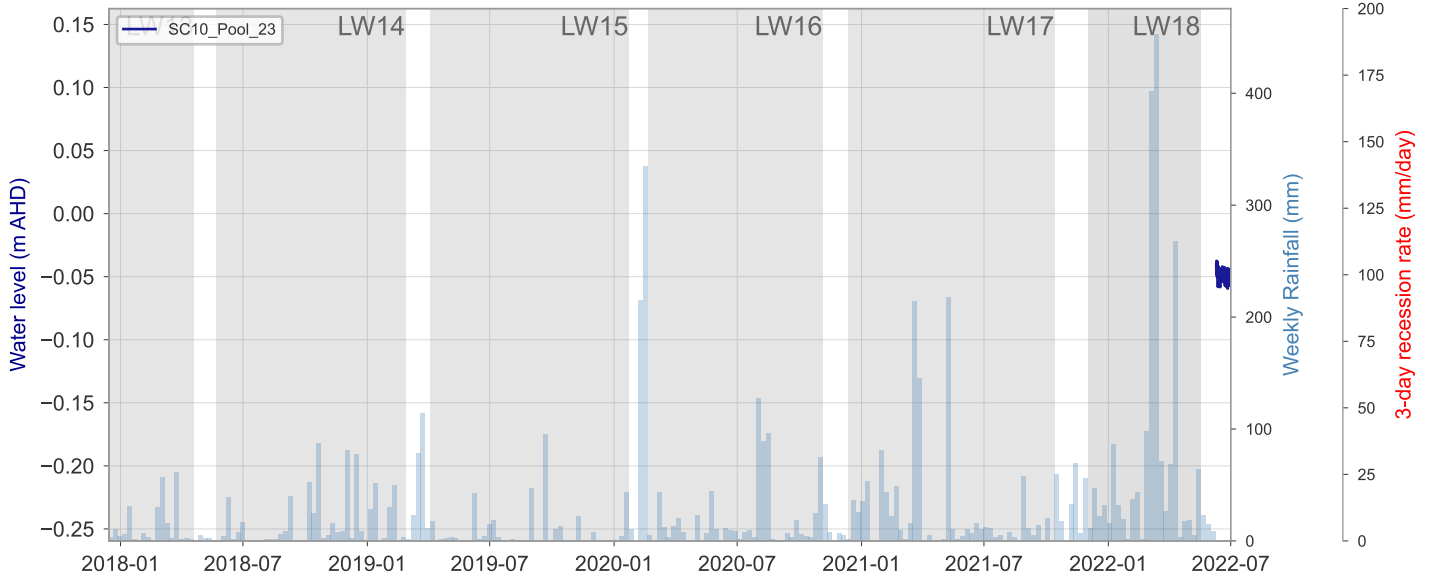
SC10_Pool_11



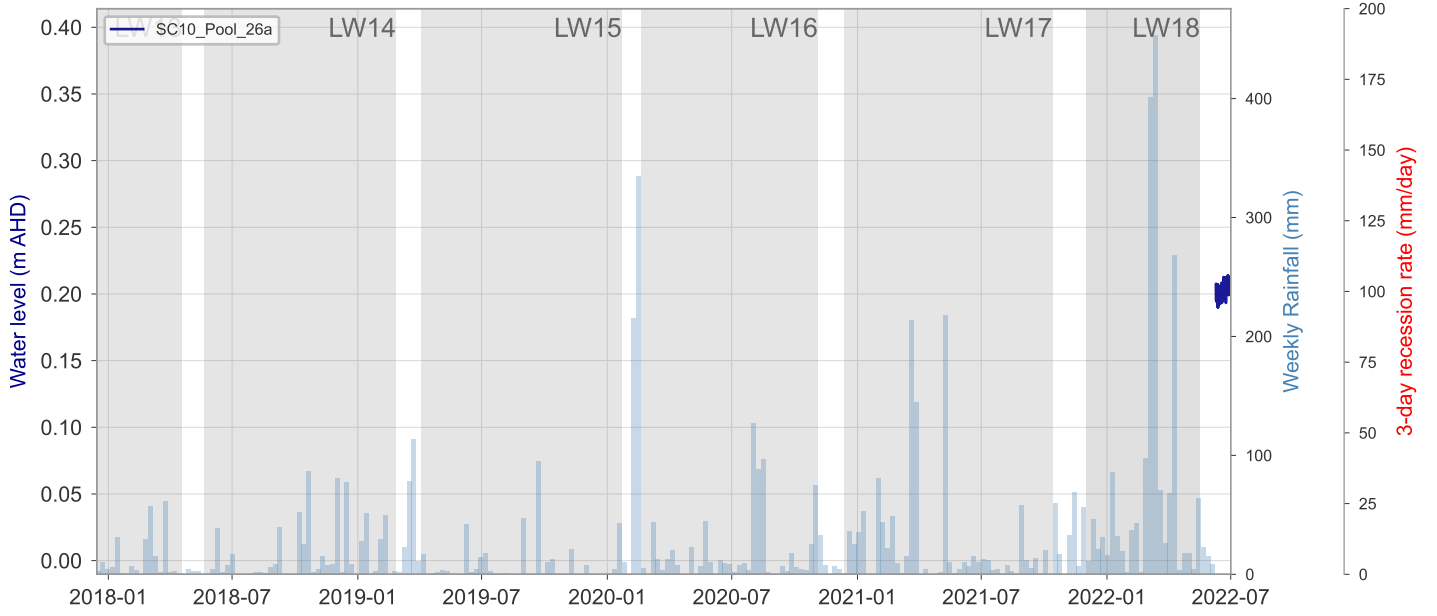
SC10_Pool_14



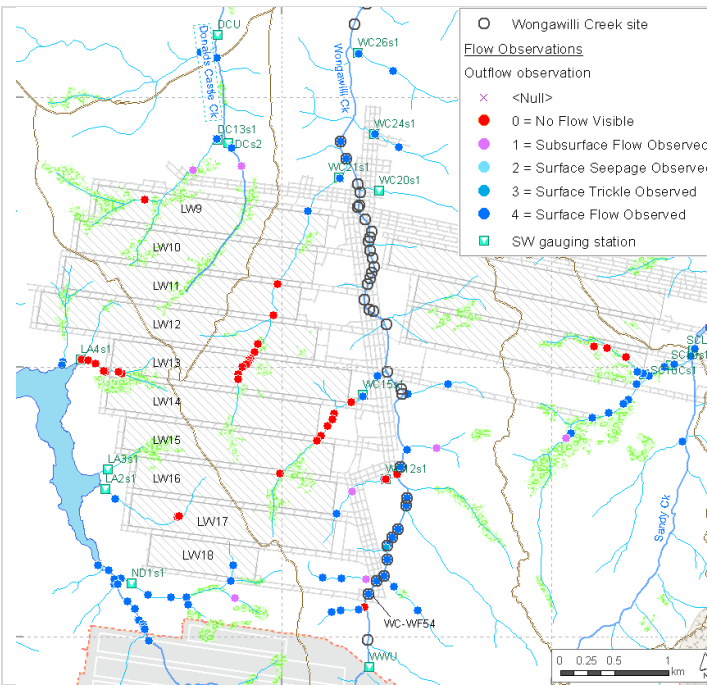
SC10_Pool_23



SC10_Pool_26a

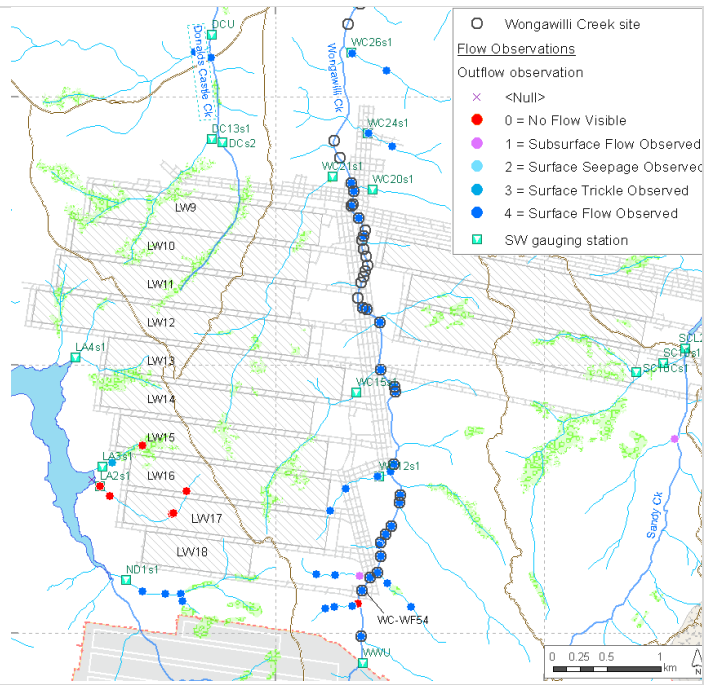


Appendix G: Watercourse flow observations



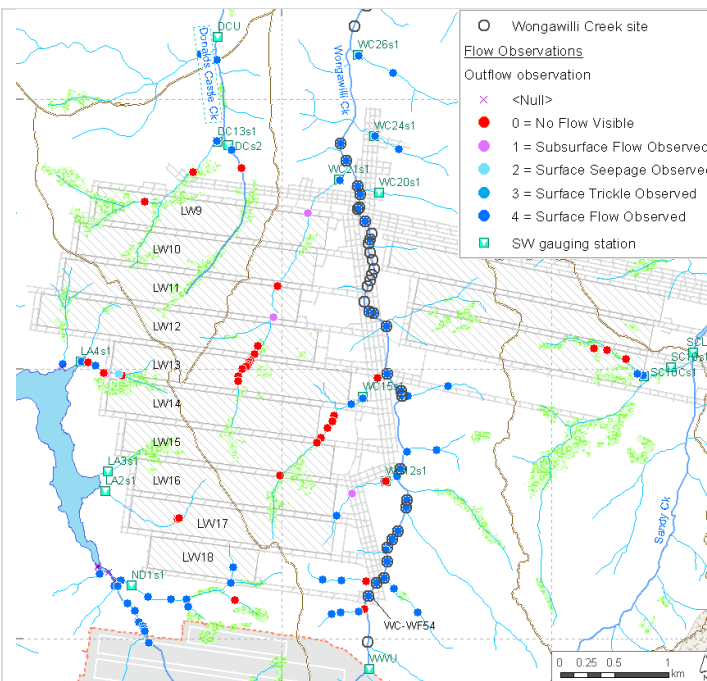
Flow observations: Dec-2021

Assessment D not triggered



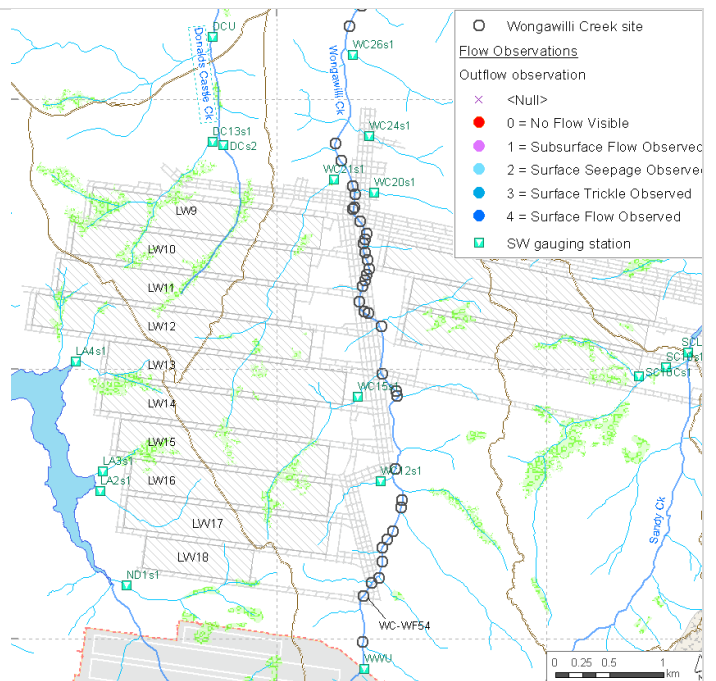
Flow observations: Jan-2022

Assessment D not triggered*



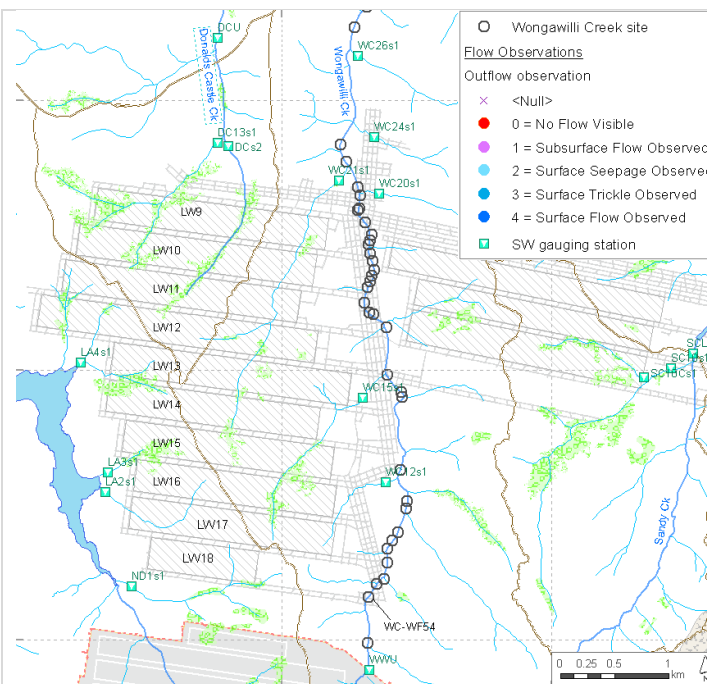
Flow observations: Feb-2022

Assessment D not triggered*



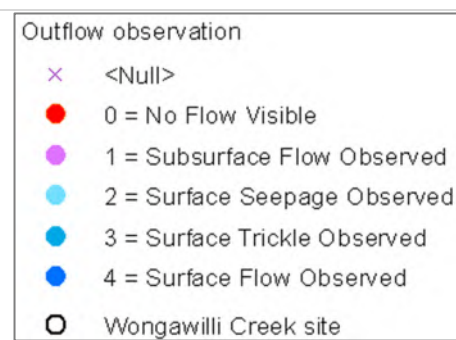
Flow observations: Mar-2022

Catchment closed *

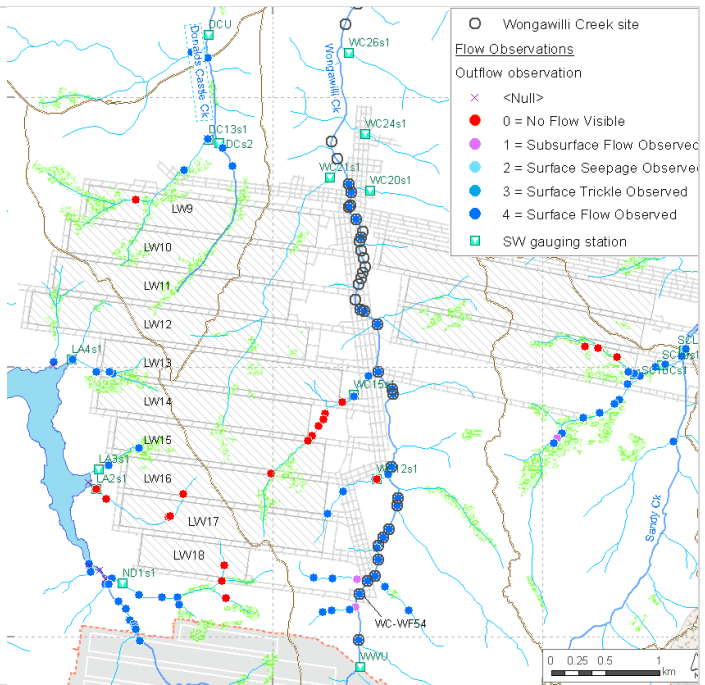
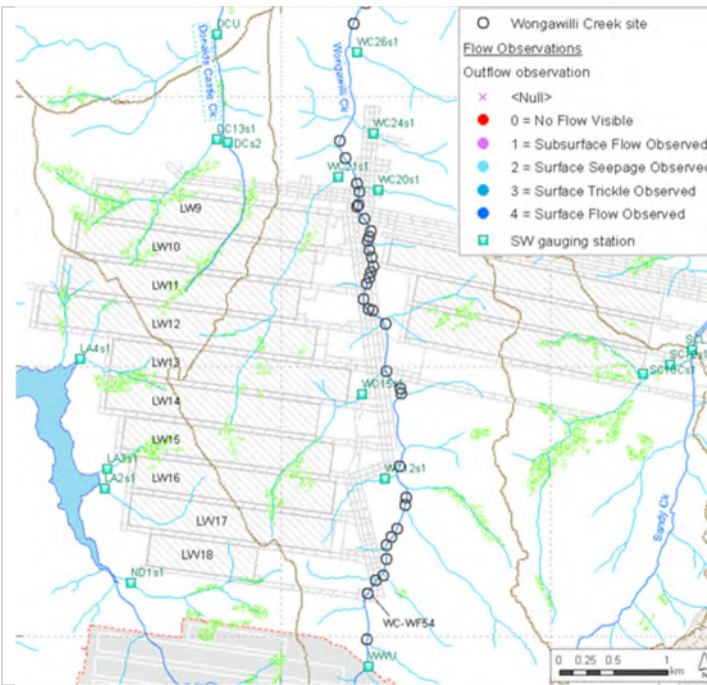


Flow observations: Apr-2022

Catchment closed *



* No or limited observations along Wongawilli Ck because of catchment closure due to high rainfall.



Flow observations: May-2022

Catchment closed *

Flow observations: Jun-2022

Assessment D not triggered

Outflow observation

- × <Null>
- 0 = No Flow Visible
- 1 = Subsurface Flow Observed
- 2 = Surface Seepage Observed
- 3 = Surface Trickle Observed
- 4 = Surface Flow Observed
- Wongawilli Creek site

* No or limited observations along Wongawilli Ck because of catchment closure due to high rainfall.

E:\DENDROBIUM\Reports\HGE014\TARP_DISW Flow observations during LW18.docx
 E:\DENDROBIUM\GIS\Maps\Deliverable\EoP18\SWobservations_EOP18.mxd



Appendix H: Rainfall-runoff modelling

H1. AWBM comparison: DCU – Donalds Castle Creek

This catchment incorporates the headwater sub-catchments DC13 and DCS2, and was mined under at the commencement of Longwall 9, and again by Longwalls 10-12, and marginally by Longwall 13. Longwalls 14-18 are beyond it (to the south). About 60% of the DCU catchment is not mined under.

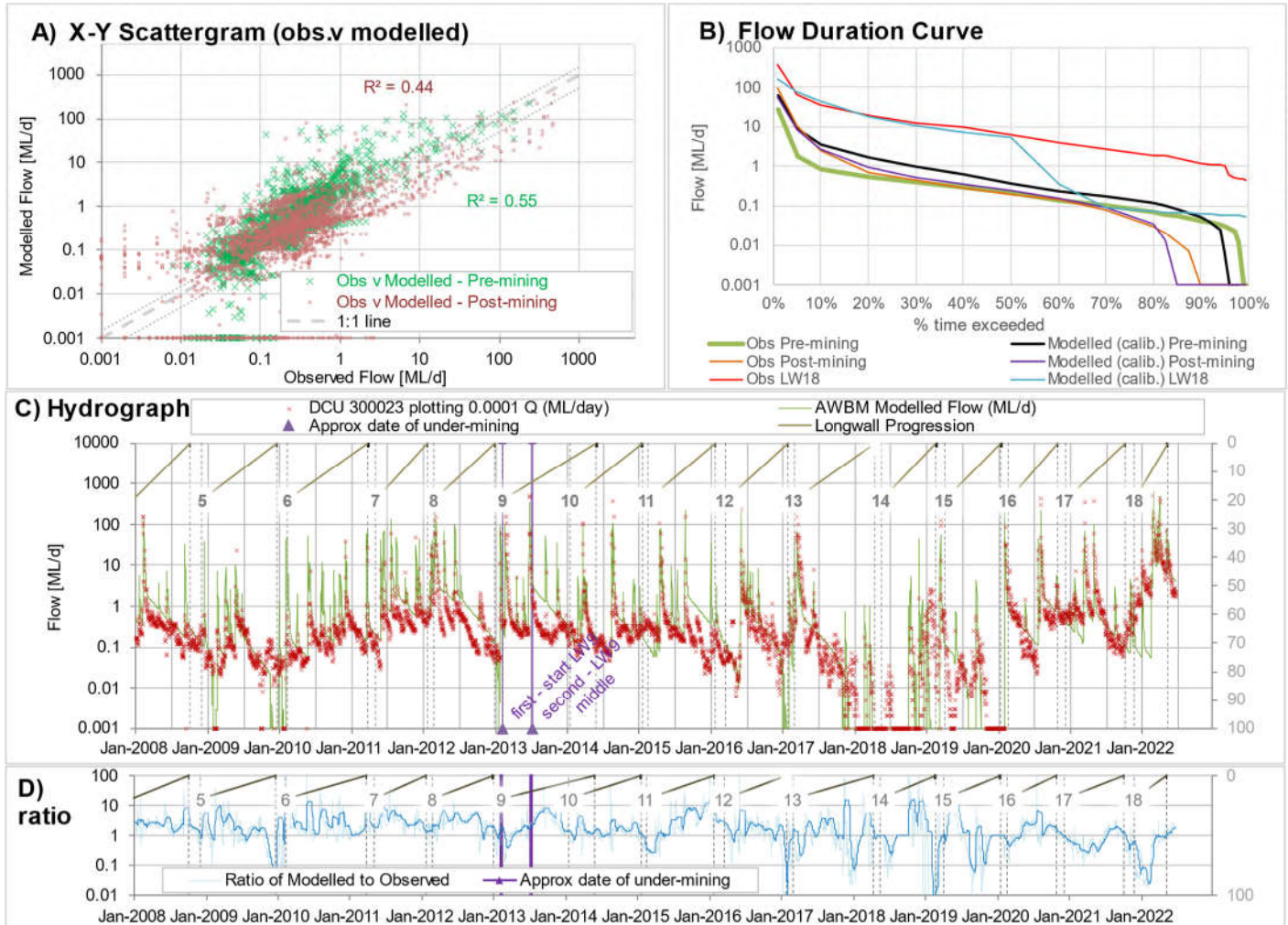


Figure H1 Comparison of observed flow against AWBM simulated flow: DCU

- A** This shows that during the pre-mining period the model is a moderate fit to observed data (this site continues to be difficult to calibrate a rainfall-runoff model for). This fit is essentially the same in the post-mining period. Simulation of the very lowest flows remains the main weakness.
- B** Confirms the reasonable match between modelled and observed flows for the pre-mining period (black vs green) and illustrates that the model still predicts the range of flows reasonably well for the subsequent post-mining period (purple vs orange). The model underestimates lower flows during Longwall 18, but the observed higher flows are well simulated in 2022 (blue vs red).
- C** The hydrograph shows a reasonable match between observed flows up until early 2013 (the start of Longwall 9), including periods of zero flow, and the match is the same after that time. The model is considered to capture the 2017-19 drought and flow in the subsequent wetter 2020-22 period quite well. Generally, the flow recessions are matched to a reasonable degree, but there is scope for more improvement in this catchment (it remains the most difficult hydrograph to match. There is no discernible systematic change in behaviour).
- D** The pre-mining ratio of modelled to observed flows shows the ratio hovers at about 1 (i.e. a good match between observed and modelled). The post-mining ratio, including during Longwall 18, oscillates around 1, and is similar to the pre-mining behaviour (e.g. see 2009).

Catchment discharge after Longwall 18: For the complete post-mining period, the water balance [$Q_{sim} + ET_{sim}$] is $\leq -6\%$ of average P_{obs} (+7%) **former TARP – Not triggered**

Assessment: The flow duration curves suggest that there is no systematic reduction in flow, especially at low flows, during Longwall 18. This is in agreement with the agreed TARP assessment using Reference Sites.

H2. AWBM comparison: WC12 – Wongawilli Creek tributary

The end of Longwall 15 skirted the north-western edge of this sub-catchment and to within 250 m of the watercourse itself. Longwall 16 mined within 40 m of WC12, and Longwall 17 mined under this watercourse. No landscape impacts (cracking, iron-staining) have been reported by IMCEFT.

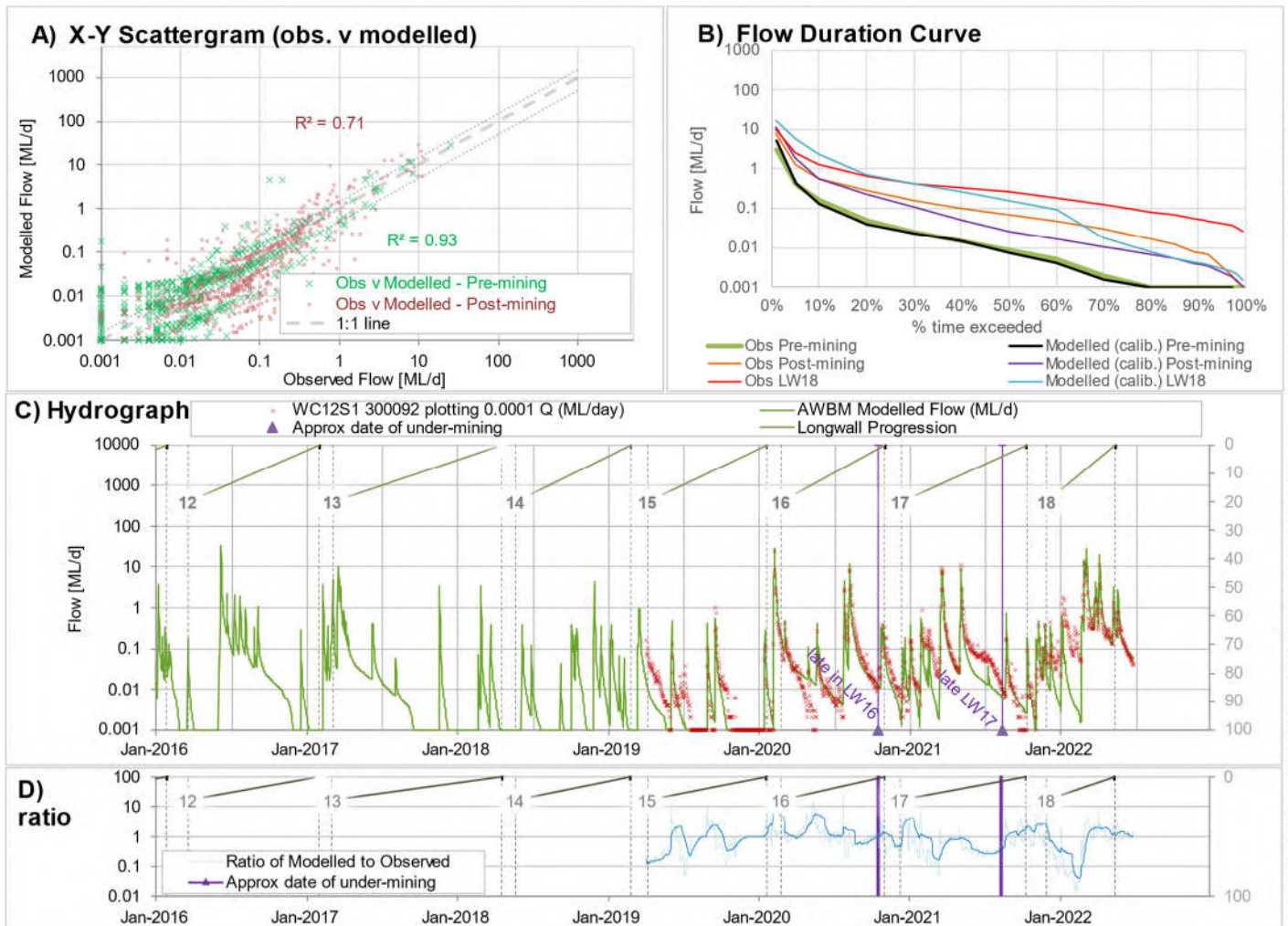


Figure H2 Comparison of observed flow against AWBM simulated flow: WC12

- A** This shows that during the short (560 day) pre-mining period the model is a reasonable fit to observed data, with the fit weaker in the shorter (605-day) post-mining period, but still moderate ($R^2 = 0.71$).
- B** Confirms the moderate match between modelled and observed flows for the pre-mining period (black vs green) and illustrates that the model underestimates flow for the subsequent post-mining period (purple vs orange) and the Longwall 18 period (red vs blue).
- C** The hydrograph shows a reasonable match between observed flows up until Oct-2020 (as Longwall 16 approaches WC12), including two periods of zero flow during the 2019 drought, and the match is the similar after that time. The model is considered to capture flow in the subsequent wetter 2020-22 period moderately well. Generally, the flow recessions are well matched, but some are over-estimated and some under-estimated in both the pre-mining and post-mining periods. There is no discernible systematic change in behaviour.
- D** The pre-mining ratio of modelled to observed flows shows the ratio oscillates at approximately 1 (i.e. a good match between observed and modelled). The post-mining ratio during Longwall 18 has oscillated around 1, and has behaved similarly to the pre-mining ratio.

Catchment discharge after Longwall 18:

For the complete post-mining period, the water balance [$Q_{sim} + ET_{sim}$] is < -6% of average P_{obs} (-7%)

former TARP – Level 1

Assessment:

This assessment suggests that mining effects on surface water flows are likely to occur in this sub-catchment, yet are relatively minor (Level 1).

H3. AWBM comparison: WWL – Wongawilli Creek (lower)

Wongawilli Creek lies between Areas 3A and 3B. The watercourse is not directly mined under by longwalls, but some tributaries (e.g. WC21, WC15, among others) have been mined under by Area 3A and 3B longwalls, including Longwall 17. Longwall 18 is outside the Wongawilli Creek catchment.

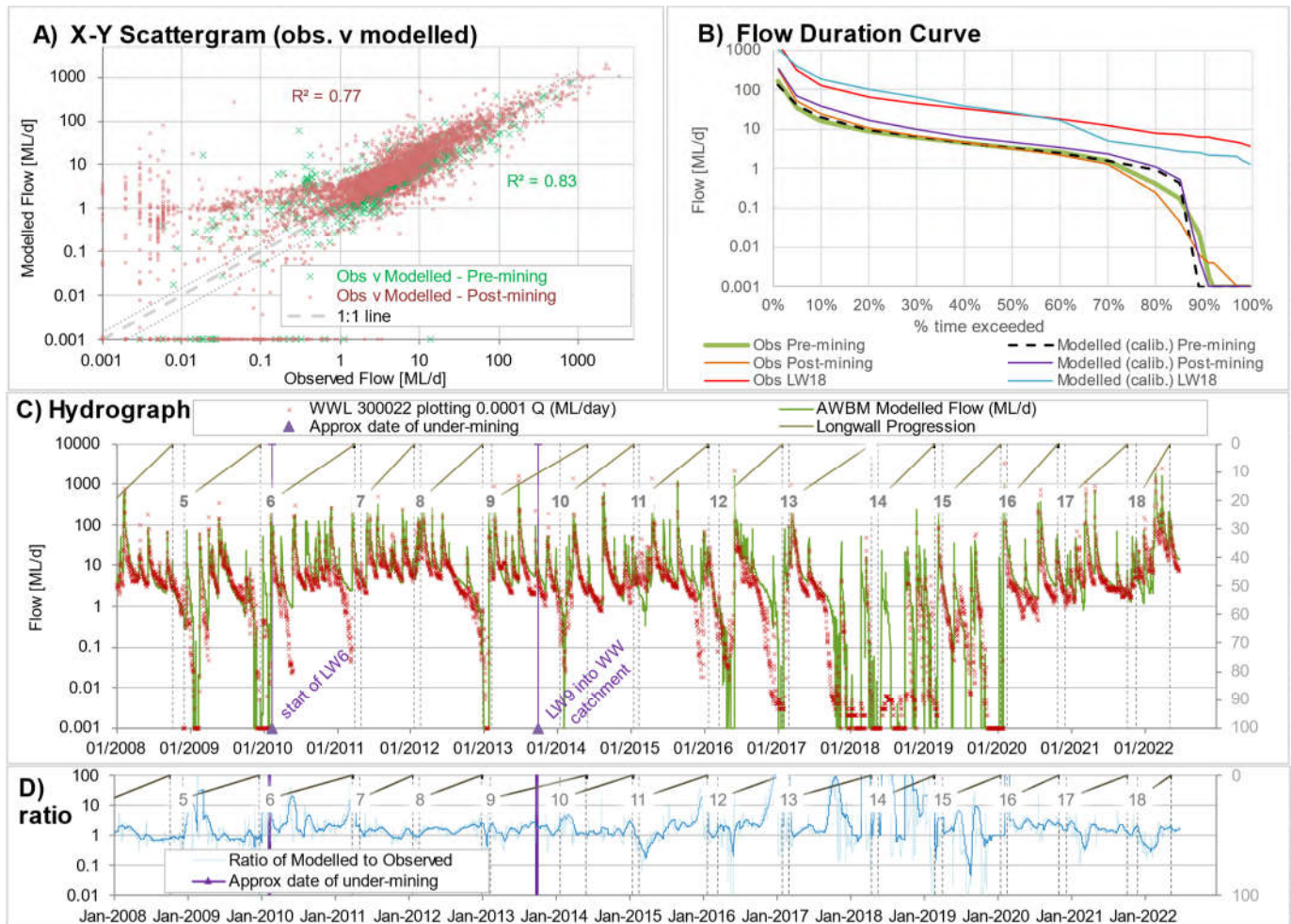


Figure H3 Comparison of observed flow against AWBM simulated flow: WWL

- A** This shows that during the pre-mining period the model is a good fit to observed data ($R^2 = 0.83$). This fit is slightly weaker, but still good ($R^2 = 0.77$), in the post-mining period.
- B** Confirms the reasonable match between modelled and observed flows for the pre-mining period (black vs green) and illustrates that the model still predicts the range of flows reasonably well for the subsequent post-mining period (purple vs orange) as well as reasonably well during the extremely wet Longwall 18 period (the model underestimates low flows during this period).
- C** The hydrograph shows a reasonable match between observed flows up until Feb-2010 (the start of Longwall 6), including two periods of zero flow, and the match is the same after that time. The model is considered to capture the 2017-19 drought and flow in the subsequent wetter 2020-22 period quite well. Generally, the flow recessions are well matched, but some are over-estimated and some under-estimated in both the pre-mining and post-mining periods. There is no discernible systematic change in behaviour.
- D** The pre-mining ratio of modelled to observed flows shows the ratio hovers at about 1 (i.e. a good match between observed and modelled). The post-mining ratio, including during Longwall 18, oscillates around 1, and is similar to the pre-mining behaviour, although during droughts (when flows were frequently close to 0), the ratio is more variable.

Catchment discharge after Longwall 18:

For the complete post-mining period, the water balance [$Q_{sim} + ET_{sim}$] is $\geq -6\%$ of average P_{obs} (-2.7%)

former TARP – Not triggered

Assessment:

The above analysis does not suggest any reduced sub-catchment flow / yield that can be discerned beyond natural variability or model/method accuracy. This is consistent with the agreed TARP assessment using Reference Sites.

H4. AWBM comparison: ND1S1 on Native Dog tributary ND1

ND1 is a tributary to Native Dog Creek, which flows into Lake Avon. Elouera Colliery longwalls are within or close to this sub-catchment, but were not directly beneath this watercourse or its tributaries. Dendrobium Longwall 18 mined under the upstream parts of the ND1 catchment.

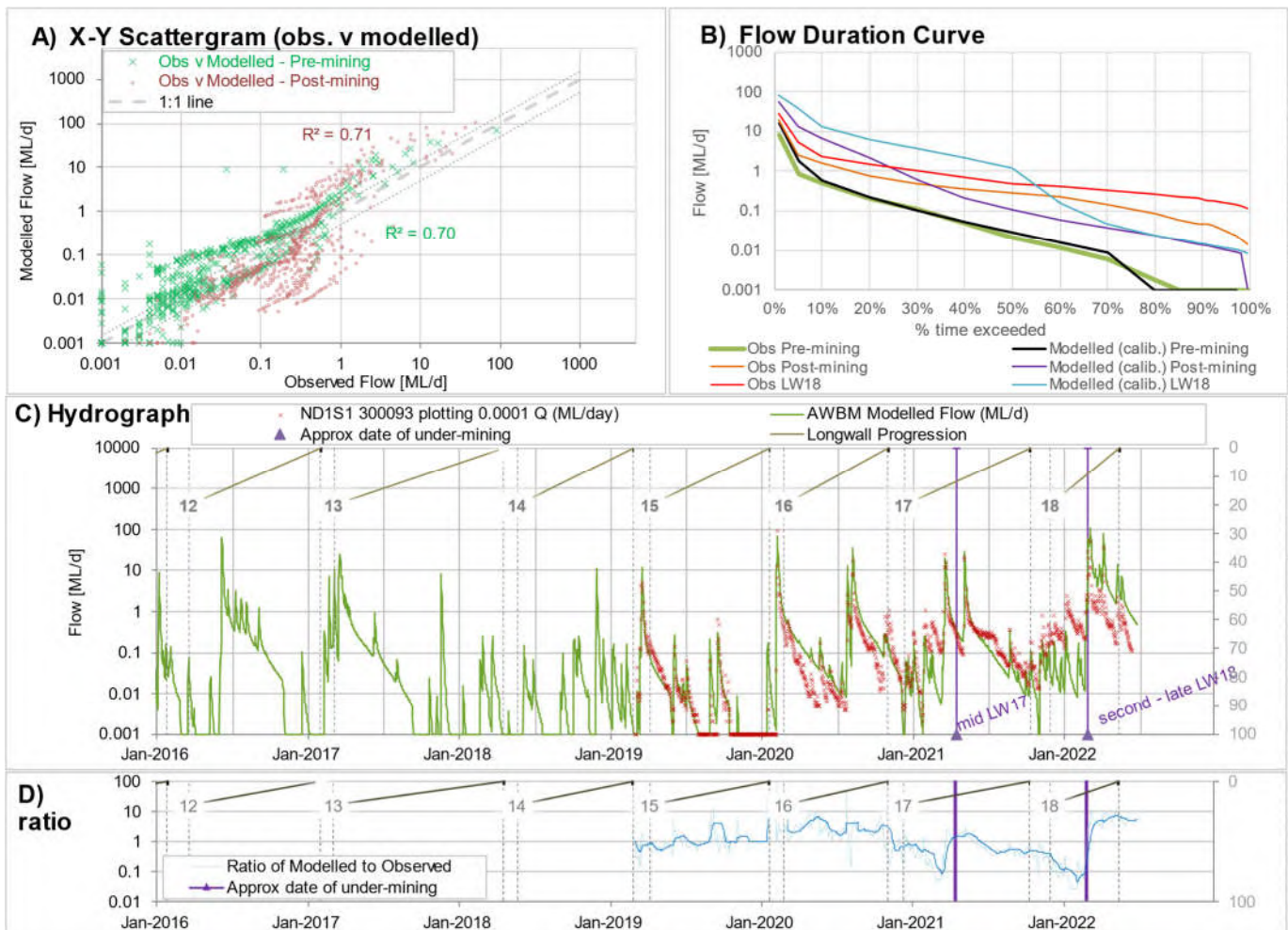


Figure H3 Comparison of observed flow against AWBM simulated flow: ND1

- A** This shows that during the pre-mining period the model is a good fit to observed data. The fit is similar ($R^2 = 0.71$) in the post-mining period.
- B** Confirms the reasonable match between modelled and observed flows for the pre-mining period (black vs green) and illustrates that the model still tends to under-estimate low flows for the subsequent post-mining period (purple vs orange) as well as during Longwall 18.
- C** The hydrograph shows a reasonable match between observed flows up until Jan-2022 (late in Longwall 17), including two periods of zero flow early in Longwall 17. The model then underestimates flows in the summer of 2021-22 (as at other sites). From Mar-2022, observed flows are consistently below modelled, and the timing of this suggests that Longwall 18 has affected (reduced) flow .
- D** The pre-mining ratio of modelled to observed flows shows the ratio hovers at about 1 (i.e. a good match between observed and modelled), but tends to >1 during 2020. Halfway through Longwall 18, the ratio has increased significantly, suggesting a mining effect.

Catchment discharge after Longwall 17:

For the complete post-mining period, the water balance [$Q_{sim} + ET_{sim}$] is $< -18\%$ of average P_{obs} (-30%)

former TARP – Level 3.

Assessment:

The above analysis suggests that halfway through Longwall 18, this site has been affected by mining, which is unsurprising given that Longwall 18 passes under the headwaters of this sub-catchment, including tributary ND1C. This is different to the agreed TARP assessment using Reference Sites.

H5. Parameters used for AWBM by modelled sub-catchment

AWBM was first developed by W. Boughton in the early 1990s (Boughton, 2004; Boughton and Chiew, 2003). The model takes average rainfall and potential evaporation across a catchment as inputs on a daily timestep. The user provides parameters to describe the relative area and soil moisture storage capacity of three stores covering the catchment (**Figure H5**).

Based on these inputs and parameters, surface runoff and baseflow are calculated and then released from the relevant storage using a linear decay (K_{surf} or K_{base}). These decayed flows are summed to estimate total catchment outflow on a daily basis.

Most of the parameters relate in part to the simulated connected groundwater system in the catchment. For this project, AWBM has been populated and run via a spreadsheet version of the AWBM model.

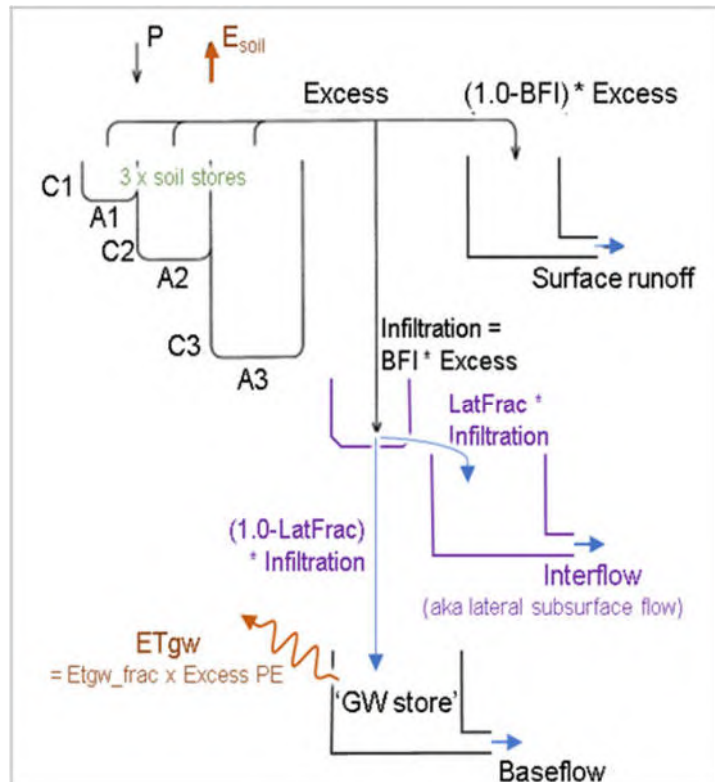


Figure H5. AWBM Rainfall-runoff model flow diagram

(modified from Boughton, 2004)

Table H1. AWBM parameters and inputs for selected Dendrobium catchment models

| SITE | A1 | A2 | A3 | Kbase | Klat | Ksurf | BFI | LatFrac | C1 | C2 | C3 | ETgw |
|--|--|-----------------|-----------------|----------|----------|----------|---|----------|-------|-------|------|----------|
| | area - fraction | area - fraction | area - fraction | fraction | fraction | fraction | fraction | fraction | mm | mm | mm | fraction |
| Donalds Castle Creek catchments | | | | | | | | | | | | |
| DCU | 0.08 | 0.20 | 0.72 | 0.99 | 0.7 | 0.30 | 0.60 | 0.7 | 0.04 | 0.25 | 0.40 | 0.006 |
| Wongawilli Creek catchments | | | | | | | | | | | | |
| WC12 | 0.1 | 0.55 | 0.35 | 0.982 | 0.85 | 0.35 | 0.32 | 0.7 | 0.015 | 0.175 | 0.35 | 0.01 |
| WWL | 0.3 | 0.35 | 0.35 | 0.992 | 0.8 | 0.20 | 0.60 | 0.7 | 0.015 | 0.15 | 0.25 | 0.04 |
| Lake Avon catchments | | | | | | | | | | | | |
| ND1 | 0.02 | 0.40 | 0.58 | 0.975 | 0.85 | 0.5 | 0.5 | 0.5 | 0.06 | 0.15 | 0.40 | 0.06 |
| SITE | DAILY RAINFALL INPUT | | | | | | EVAPORATION INPUTS | | | | | |
| Donalds Castle Creek catchments | | | | | | | | | | | | |
| DCU | Daily SILO Data Drill "DEN-South" to Oct-2007. Average of Dendrobium Centroid and A3B rainfall records used for Oct-2007-2021. | | | | | | Daily SILO "DEN-South" Pan Evaporation ('Evap'). Pan factor of 1. ET _{Gw} simulated from 0.6% of this sub-catchment. | | | | | |
| Wongawilli Creek catchments | | | | | | | | | | | | |
| WC12 | Average of SILO Data Drill "DEN-South" and WaterNSW Browns Road rainfall used. | | | | | | Daily SILO "DEN-South" Pan Evaporation ('Evap'). Pan factor of 1. ET _{Gw} simulated from 1% of this sub-catchment. | | | | | |
| WWL | Daily SILO Data Drill "DEN-South" to Oct-2007. Average of Dendrobium Centroid, A3B and SILO "DEN-South" rainfall records used for Oct-2007-2021. | | | | | | Daily SILO "DEN-South" Pan Evaporation ('Evap'). Pan factor of 1. ET _{Gw} simulated from 1% of this sub-catchment. | | | | | |
| Native Dog Creek catchments | | | | | | | | | | | | |
| ND1 | Average of Daily SILO Data Drill "DEN-South" and WaterNSW Browns Road to Oct-2007. Average of Dendrobium Centroid, A3B, Browns Rd and SILO "DEN-South" rainfall records used for Oct-2007-2021. | | | | | | Daily SILO "DEN-South" Pan Evaporation ('Evap'). Pan factor of 1. ET _{Gw} simulated from 6% of this sub-catchment. | | | | | |